

Higher-order interactions enhance the latitudinal tree diversity gradient

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The global decrease in species diversity from low to high latitudes is among the most robust biogeographic patterns^{1,2}. There is continuing debate on the contribution of conspecific negative density dependence (CNDD) to the latitudinal diversity gradient evident for trees^{3,4}. Theory suggests that CNDD based on pairwise interactions alone is not sufficient to explain the intricacies of diverse communities, because higher-order interactions (HOIs) may greatly modify these interactions^{5,6}. However, there has been a lack of empirical studies investigating how HOIs intertwine with pairwise interactions and how they may contribute to the latitudinal tree diversity gradient. Here we examined both pairwise interactions and HOIs across 32 large permanent forest plots, most in the northern hemisphere. We detected evidence of HOIs in 40% of the 1,543 species–plot combinations for tree growth, and 23% of the 1,340 such combinations for tree survival, with the strength of these interactions declining with latitude. HOIs were found to benefit rare species but disadvantage common species, suggesting a potential mechanism promoting species diversity. This stabilizing effect weakened towards higher latitudes, consistent with the latitudinal tree diversity gradient. Our findings reveal an important interplay between pairwise interactions and HOIs in promoting the latitudinal tree diversity gradient and help to clarify the contribution of CNDD to this biogeographic pattern.

The systematic decline in species diversity from the equator to the poles, known as the latitudinal diversity gradient, is among the most widely observed biogeographic patterns^{1,7}. The latitudinal decline in tree species diversity is particularly prominent and well documented^{2,8}. CNDD, whereby conspecific neighbours exert more negative effects on the performance of a focal tree than heterospecific neighbours⁹, is considered to be a primary mechanism that maintains local diversity¹⁰ and promotes the latitudinal tree diversity gradient^{11,12}. The lemma that CNDD regulates latitudinal diversity gradient posits that CNDD is stronger in tropical than in temperate forests⁴. However, empirical evidence for this proposition is mixed: some studies based

on field data across forest plots have reported significant declines in CNDD with increasing latitude¹² or decreasing species richness¹¹, whereas others did not support this trend^{3,13,14}. These inconsistent findings have led to a longstanding debate on whether there is a latitudinal pattern of CNDD and whether it contributes to latitudinal tree diversity gradient^{4,15–18}.

Although negative density dependence has been the main focus of interest in previous studies of the latitudinal tree diversity gradient, positive density dependence (facilitation) generated by mutualists (for instance, mycorrhizal fungi)¹⁹ could also play an important part in maintaining local species diversity^{20,21} and contribute to the

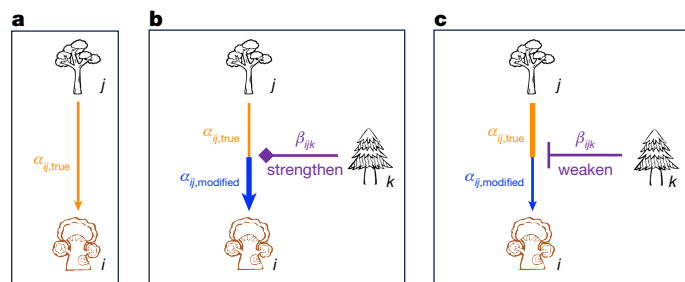


Fig. 1 | Effect of a neighbouring tree of species *j* on a focal tree of species *i* in the absence and presence of another neighbouring tree of species *k*. **a**, The pairwise effect of a neighbour of species *j* on the focal tree of species *i* in the absence of other neighbours is denoted $\alpha_{ij,true}$. **b,c**, In the presence of another neighbour of species *k*, the higher-order effect of *k* (initiator) on *i* (receiver) through *j* (transmitter), denoted β_{ijk} , could either strengthen (**b**) or weaken (**c**) $\alpha_{ij,true}$.

latitudinal diversity gradient^{22,23}. Moreover, existing studies have mostly been built on the assumption that organisms interact in a pairwise fashion^{5,24}, neglecting the potential effects of HOIs that emerge when pairwise interactions between two neighbouring trees are modified by other neighbours (Fig. 1). Specifically, the effect of a neighbouring tree of species *j* on a focal tree of species *i* in the absence of other trees, denoted $\alpha_{ij,true}$ (Fig. 1a), can be competitive ($\alpha_{ij,true} < 0$) or facilitative ($\alpha_{ij,true} > 0$), and conspecific ($j = i$) or heterospecific ($j \neq i$). This pairwise effect can be either strengthened (Fig. 1b) or weakened (Fig. 1c) in the presence of another neighbouring tree of species *k*. The higher-order effect of *k* (initiator) on species *i* (receiver) through *j* (transmitter) is denoted β_{ijk} . As a consequence, the effect of *j* on *i* in the presence of *k*, denoted $\alpha_{ij,modified}$ (implicitly incorporating β_{ijk}), can be stronger (Fig. 1b) or weaker (Fig. 1c) than $\alpha_{ij,true}$, depending on whether $\alpha_{ij,true}$ and β_{ijk} have the same or opposite signs. As an example, *Eucalyptus urophylla* (initiator) can inhibit the root growth of *Cryptocarya concinna* (transmitter) through allelopathic effects²⁵, thereby weakening the competitive effect of *C. concinna* on its neighbouring species (receiver).

There is accumulating evidence that HOIs regulate the survival and growth of trees^{26,27} and thus play a substantial part in the structuring of community assemblages^{28–30}. Consequently, estimating CNDD (when $\alpha_{ii} < \alpha_{ij} < 0$) without explicitly considering HOIs may evoke divergent latitudinal changes in CNDD and obscure the true contribution of CNDD to the latitudinal tree diversity gradient. To better understand the contributions of pairwise interactions and HOIs, we examined how their cumulative effects changed with species abundance and latitude. Pairwise interactions and HOIs may act as stabilizing forces promoting species diversity when their cumulative effects are negatively associated with species abundance (that is, rare species are less negatively or more positively affected). Such stabilizing effects may further reinforce the latitudinal tree diversity gradient if this cumulative effect–abundance relationship weakens with increasing latitude. It has remained unclear, however, whether HOIs are common in forests and how they contribute to the latitudinal tree diversity gradient.

In this study, we assembled census data from 32 large permanent forest plots spanning tropical to boreal forests (Fig. 2a and Supplementary Table 1) to address three key questions. (Q1) Are HOIs prevalent among trees across forest plots? (Q2) How do pairwise interactions ($\alpha_{ij,true}$ and $\alpha_{ij,modified}$) and HOIs (β_{ijk}) vary with latitude? (Q3) How do latitudinal changes in HOIs contribute to the latitudinal tree diversity gradient (Fig. 2b)? To answer these questions, we estimated both pairwise interactions and HOIs from demographic growth (for 1,543 tree species–plot combinations) and survival models (for 1,340 tree species–plot combinations). With these data, we built three types of growth

and survival model: (1) null models with no biotic interactions, (2) pair-only models including only pairwise interactions and (3) HOI-inclusive models including both pairwise interactions and HOIs. The pairwise interactions estimated from pair-only models implicitly incorporate the effects of HOIs and thus represent $\alpha_{ij,modified}$ (blue arrows in Fig. 1), whereas pairwise interactions estimated from HOI-inclusive models isolate HOIs and thus represent $\alpha_{ij,true}$ (orange arrows in Fig. 1), the true interaction between two trees of species *i* and *j* in the absence of other neighbours. We then compared Akaike information criterion (AIC) values of the three types of model (Q1), tested whether the estimated pairwise interactions ($\alpha_{ij,true}$ and $\alpha_{ij,modified}$) and HOIs (β_{ijk}) declined with latitude (Q2), and evaluated how the cumulative effects of pairwise interactions and HOIs on growth and survival changed with species abundance and latitude (Q3).

Prevalence of HOIs across forest plots

Our results demonstrate strong statistical support for the prevalence of HOIs across 32 large permanent forest plots, as evidenced by the AIC values of HOI-inclusive growth models being at least two units lower than those of the alternative models for 40% of the 1,543 species–plot combinations across all 32 forest plots (Fig. 3a). Moreover, HOI-inclusive growth models significantly improved predictions of tree growth rates compared with the alternative models (Fig. 3b). Although HOIs have been reported at local scales in forests^{26,27} and other systems (for instance, bacteria–paramecium–protozoan⁷, amphibians³¹, microcrustaceans³², annual plants³³), our study shows that HOIs are ubiquitous in global forests. The widespread HOIs in forests highlight the importance and necessity of further exploration of their latitudinal patterns and ecological consequences.

Latitudinal decline in biotic interactions

Competitive ($\alpha < 0, \beta < 0$) and facilitative ($\alpha > 0, \beta > 0$) interactions were almost equally frequent across all the species–plot combinations (Extended Data Fig. 1). The average strength of true intraspecific pairwise interactions ($\alpha_{ii,true}$), both competitive and facilitative, declined rapidly with latitude (orange dots and lines in Fig. 4a). However, the average strength of true interspecific pairwise interactions ($\alpha_{ih,true}$), both competitive and facilitative, remains relatively constant across latitudes (orange dots and lines in Fig. 4b). These results are consistent with some previous work indicating that higher tree diversity is associated with stronger CNDD^{11,12}. HOI coefficients (except β_{ihh}), both competitive and facilitative, also declined with latitude (Fig. 4c–f).

Strong negative correlations were detected between $\alpha_{ii,true}$ and β_{ihh} ($r = -0.96$) and between $\alpha_{ih,true}$ and β_{ihh} ($r = -0.89$) (Extended Data Fig. 2). This meant that both intraspecific (α_{ii}) and interspecific (α_{ih}) pairwise interactions were more strongly weakened by heterospecific neighbours ($\beta_{ihh} : \alpha_{ii} \leftarrow h$ and $\beta_{ihh} : \alpha_{ih} \leftarrow h$) than by conspecific neighbours. Consequently, the latitudinal signal of intraspecific pairwise interactions modified by HOIs ($\alpha_{ii,modified}$) became much weaker (blue dots and lines in Fig. 4a) or even disappeared (Supplementary Text 2.3), because β_{ihh} weakened $\alpha_{ii,true}$ more strongly in low latitudes than in high latitudes (Fig. 4e). By contrast, interspecific pairwise interactions ($\alpha_{ih,true}$) were weakened uniformly by relatively constant β_{ihh} across latitudes (Fig. 4f), resulting in relatively constant modified interspecific pairwise interactions ($\alpha_{ih,modified}$) (blue dots and lines in Fig. 4b). Our findings suggest that HOIs, by modulating pairwise interactions across latitudes, may reconcile the divergent observations of latitudinal changes in CNDD^{3,11,12}.

HOIs enhance latitudinal tree diversity gradient

Previous studies have primarily focused on competitive pairwise interactions (that is, CNDD) and their contribution to the latitudinal tree

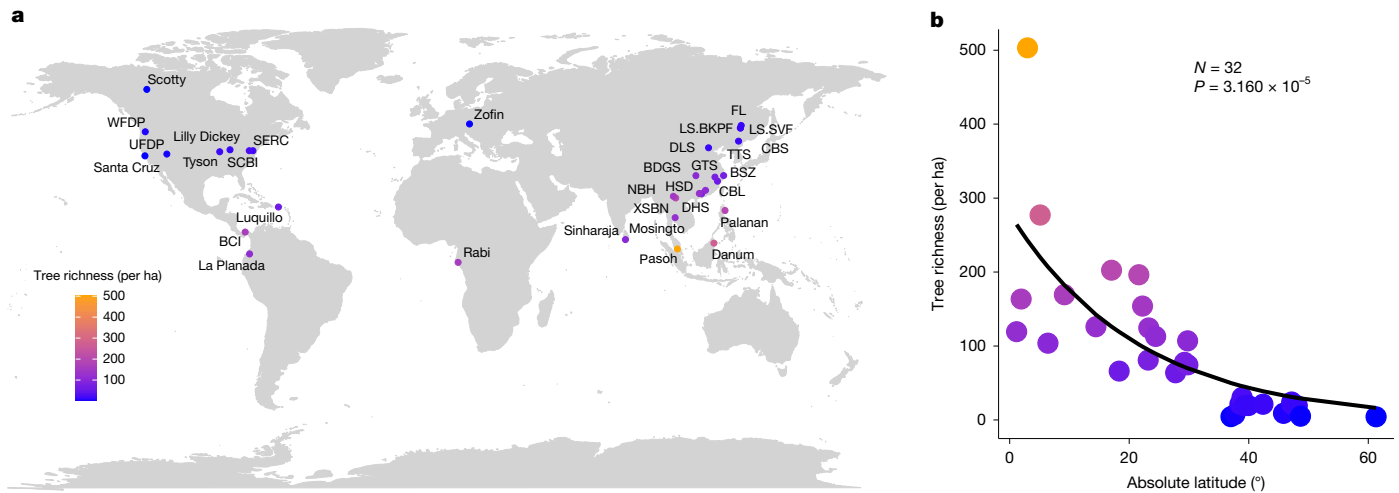


Fig. 2 | Latitudinal tree diversity gradient across 32 forest plots worldwide. **a**, Geographical distribution of the 32 forest plots. **b**, Latitudinal tree diversity gradient (number of tree species per hectare, distinguished by point colour). The latitudinal trend in tree diversity was fitted using an exponential

regression model; the solid line shows the fitted relationship. The significance of the regression coefficient was assessed using a two-sided *t*-test ($t = -4.892$, d.f. = 30, $P = 3.160 \times 10^{-5}$, $N = 32$ plots).

diversity gradient^{3,4,12}. By contrast, our results show that facilitative pairwise interactions and HOIs are also common and strong (Fig. 4 and Extended Data Fig. 1), highlighting the necessity of considering their overall contributions (including both competitive and facilitative effects) to the latitudinal tree diversity gradient (Fig. 2b). Here we assessed the relative changes in growth rate caused by cumulative effects (both competitive and facilitative effects) of pairwise interactions and HOIs separately and examined how they changed with species abundance and latitude. The relative changes in growth rate caused by cumulative effects of pairwise interactions and HOIs both declined with species abundance, shifting from beneficial for rare species to detrimental for common species (Fig. 5). This pattern implies that both the cumulative effects of pairwise interactions and HOIs promote the growth rates of rare species but suppress those of common species, thereby potentially favouring species diversity at local scales. Moreover, the stabilizing effect of pairwise interactions

changed little across latitudes (the *P* value for the interaction between abundance and latitude for the relative changes caused by pairwise interactions was 0.736; Table 1), whereas the stabilizing effect of HOIs became weaker (marginally significant change) towards higher latitudes (the *P* value of the interaction between abundance and latitude for the relative change caused by HOIs was 0.093; Table 1). Taken together, these findings suggest that latitudinal changes in HOIs enhance the latitudinal tree diversity gradient, whereas latitudinal changes in pairwise interactions contribute little to this gradient.

Our study documents the prevalence of HOIs and their potential contribution to the latitudinal tree diversity gradient. For accurate interpretation of our results, we note some limitations of our analyses. First, the present study estimated pairwise interactions and HOIs separately for tree growth and survival (Extended Data Figs. 3–5 and Extended Data Table 1), thereby overlooking their covarying effects

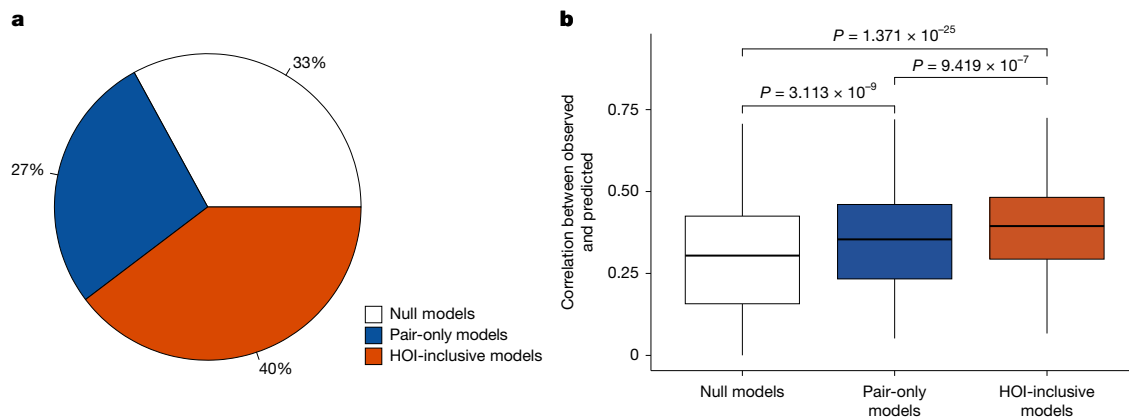


Fig. 3 | Evidence of HOIs from growth models across 32 forest plots. **a**, Percentages of 1,543 tree species–plot combinations across the 32 plots for which each of the three classes of growth models were best supported (AIC at least two units lower than that of simpler models). The white, blue and orange slices indicate the percentages of species supporting the null models (models without including biotic interactions), pair-only models (models including only pairwise interactions) and HOI-inclusive models (models including both pairwise interactions and HOIs), respectively. **b**, Correlations between

observed and predicted growth rates by null, pair-only and HOI-inclusive models for the 609 species supported by the HOI-inclusive models. Box plots show the median (centre line) and 25th and 75th percentiles (box limits), with whiskers extending to the most extreme data points within 1.5× the interquartile range. Differences among models were assessed using pairwise two-sided *t*-tests ($N = 609$ species); exact *P* values are shown in the figure. No adjustment was made for multiple comparisons.

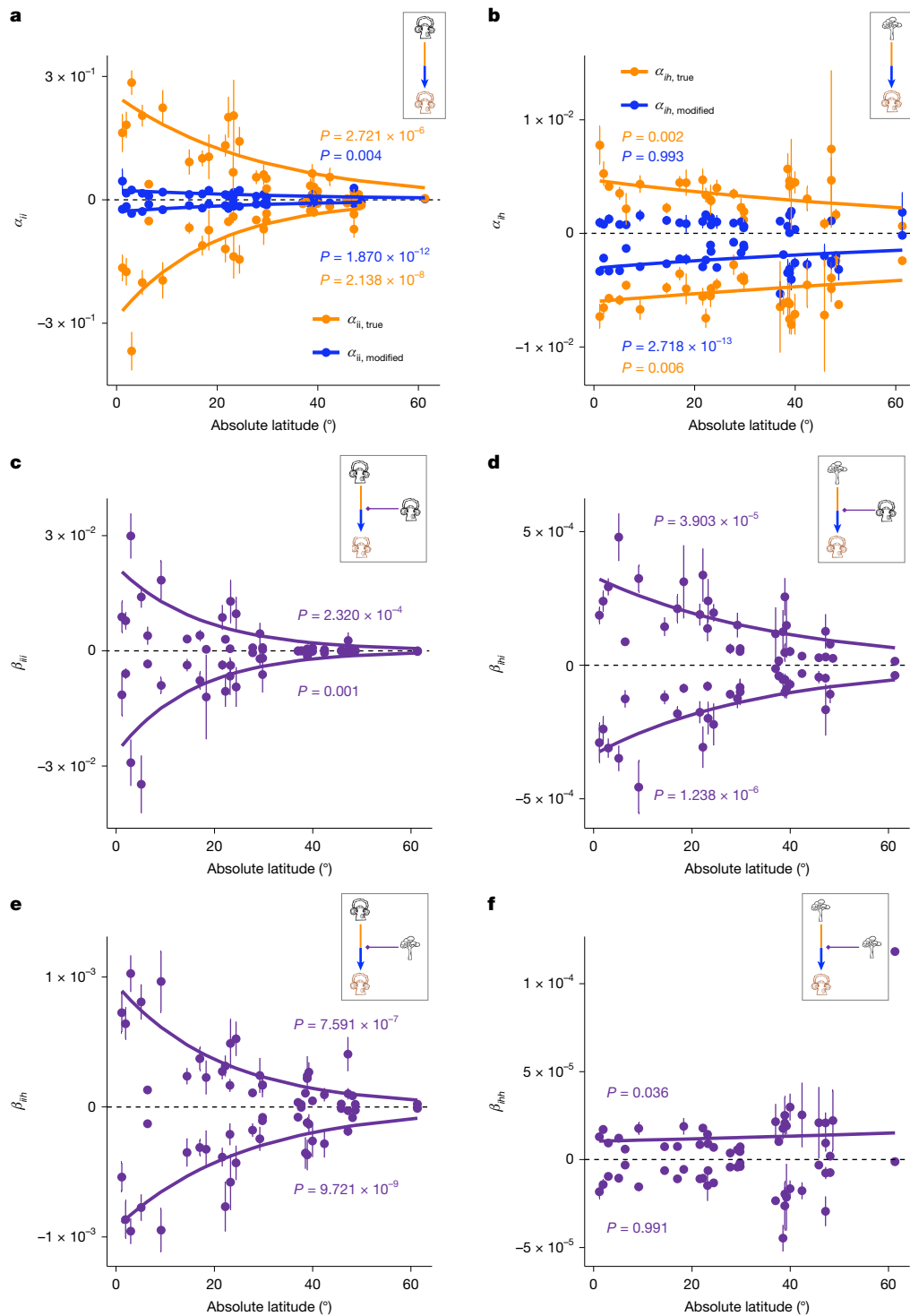


Fig. 4 | Latitudinal changes in pairwise and HOIs for growth models.

a, b, Latitudinal changes in intraspecific (α_{ii} , **a**) and interspecific (α_{ih} , **b**) pairwise interactions. The pairwise interactions estimated from pair-only models and HOI-inclusive models, denoted $\alpha_{modified}$ and α_{true} , are distinguished by blue and orange points (lines). **c, e**, Latitudinal changes in HOI coefficients in which intraspecific pairwise interactions are modified by conspecific neighbours (β_{ij} ; $\alpha_{ii} < i$, **c**) and heterospecific neighbours (β_{ih} ; $\alpha_{ii} < h$, **e**), respectively. **d, f**, Latitudinal changes in HOI coefficients in which interspecific pairwise interactions are modified by conspecific neighbours (β_{hi} ; $\alpha_{ih} < i$, **d**) and

heterospecific neighbours (β_{hh} ; $\alpha_{ih} < h$, **f**), respectively. The insets in each panel describe the six types of interaction (as in Fig. 1). Species-level pairwise interactions and HOIs were related to absolute latitude separately for competitive ($\alpha < 0, \beta < 0$) and facilitative ($\alpha > 0, \beta > 0$) interactions using exponential models. The significance of the regression coefficients was assessed using two-sided *t*-tests; exact *P* values are shown in the figure. Regression lines are shown only when the interaction strength changed significantly with latitude ($P < 0.05$). For clarity, plot-level mean \pm s.e.m. values are displayed rather than species-level estimates.

on population dynamics. Future studies should integrate multiple demographic processes (such as survival, growth and recruitment) to estimate overall species interactions influencing population

growth^{34,35}. Second, pairwise interactions and HOIs estimated at the neighbourhood scale should be upscaled to the community scale before their contribution to species diversity can be rigorously

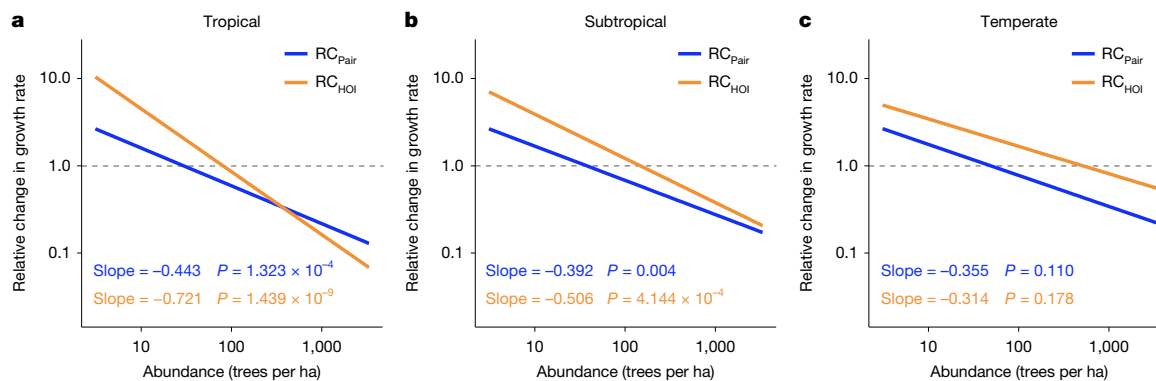


Fig. 5 | Predicted relationships between relative change in growth rate and species abundance across three latitudinal geographic zones.

a–c, Relative changes in growth rate caused by cumulative effects of pairwise interactions (RC_{Pair}) and HOIs (RC_{HOI}) are distinguished by blue and orange lines, respectively. Predictions are shown for three geographic zones³: tropical (0–23.5°), subtropical (23.5–35.0°), and temperate

(35.0–66.5°). Predictions were generated from the linear models summarized in Table 1, using the middle latitude of each zone (11.75° for the tropical zone, 29.25° for the subtropical zone and 45° for the temperate zone). Solid lines show model predictions, and shaded areas represent 95% confidence intervals.

assessed^{36,37}. To stimulate further investigation, we conducted a preliminary upscaling analysis and explored the potential contribution of HOIs to species coexistence and the latitudinal diversity gradient based on structural stability and found that forest plots with higher species richness were less susceptible to loss of species (Supplementary Text 3). Third, we assumed that pairwise interactions would be modified by HOIs in a linear form, consistent with the findings of a broad range of previous theoretical^{28,30,38} and empirical studies^{33,39}. However, a recent study explored alternative functional forms and found that nonlinear formulation (using exponential and sigmoid functions) more accurately predicted population trends of annual plants than the linear form⁴⁰; this warrants further investigation in forests.

How the latitudinal diversity gradient is maintained is a long-standing question in ecology and biogeography. We provide empirical evidence that HOIs are widespread in global forests. HOIs decline with latitude and weaken both the intraspecific (CNDD) and interspecific pairwise interactions, and thus enhance the latitudinal tree diversity gradient. Our findings help to resolve the debate on the role of pairwise CNDD in maintaining latitudinal diversity gradient and contribute to a shift from a paradigm of classical community ecology theory based solely on pairwise interactions towards a framework that incorporates facilitations and HOIs.

Table 1 | Summary of how relative changes in growth rate vary with species abundance and absolute latitude

Response	Effect	Estimate	s.e.	t	P value
RC_{Pair}	Intercept	1.494	0.507	2.945	0.003
	Abundance	-0.461	0.160	-2.888	0.004
	Latitude	-0.003	0.026	-0.095	0.924
	Abundance:latitude	0.002	0.007	0.336	0.736
RC_{HOI}	Intercept	3.586	0.531	6.752	2.07×10^{-11}
	Abundance	-0.864	0.167	-5.175	2.57×10^{-7}
	Latitude	-0.036	0.028	-1.309	0.19
	Abundance:latitude	0.012	0.007	1.680	0.093

RC_{Pair} and RC_{HOI} represent relative changes in growth rate caused by cumulative effects of pairwise interactions and HOIs, respectively. The relative change in growth rate was fitted by linear models using log-transformed species abundance, absolute latitude and their interaction ($N=1,543$ species).

Online content

Any methods, additional references, Nature Portfolio reporting summaries, source data, extended data, supplementary information, acknowledgements, peer review information; details of author contributions and competing interests; and statements of data and code availability are available at <https://doi.org/10.1038/s41586-026-10434-6>.

- Mittelbach, G. G. & McGill, B. J. *Community Ecology* (Oxford Univ. Press, 2019).
- Keil, P. & Chase, J. M. Global patterns and drivers of tree diversity integrated across a continuum of spatial grains. *Nat. Ecol. Evol.* **3**, 390–399 (2019).
- Hülsmann, L. et al. Latitudinal patterns in stabilizing density dependence of forest communities. *Nature* **627**, 564–571 (2024).
- Hülsmann, L., Chisholm, R. A. & Hartig, F. Is variation in conspecific negative density dependence driving tree diversity patterns at large scales? *Trends Ecol. Evol.* **36**, 151–163 (2021).
- Levine, J. M., Bascompte, J., Adler, P. B. & Allesina, S. Beyond pairwise mechanisms of species coexistence in complex communities. *Nature* **546**, 56–64 (2017).
- Gibbs, T. L. et al. When can higher-order interactions produce stable coexistence? *Ecol. Lett.* **27**, e14458 (2024).
- Hairton, N. G. et al. The relationship between species diversity and stability: an experimental approach with protozoa and bacteria. *Ecology* **49**, 1091–1101 (1968).
- Chu, C. et al. Direct and indirect effects of climate on richness drive the latitudinal diversity gradient in forest trees. *Ecol. Lett.* **22**, 245–255 (2019).
- Comita, L. S., Muller-Landau, H. C., Aguilar, S. & Hubbell, S. P. Asymmetric density dependence shapes species abundances in a tropical tree community. *Science* **329**, 330–332 (2010).
- Yenni, G., Adler, P. B. & Ernest, S. K. M. Strong self-limitation promotes the persistence of rare species. *Ecology* **93**, 456–461 (2012).
- Johnson, D. J., Beaulieu, W. T., Bever, J. D. & Clay, K. Conspecific negative density dependence and forest diversity. *Science* **336**, 904–907 (2012).
- LaManna, J. A. et al. Plant diversity increases with the strength of negative density dependence at the global scale. *Science* **356**, 1389–1392 (2017).
- Hille Ris Lambers, J., Clark, J. S. & Beckage, B. Density-dependent mortality and the latitudinal gradient in species diversity. *Nature* **417**, 732–735 (2002).
- Song, X., Lim, J. Y., Yang, J. & Luskin, M. S. When do Janzen–Connell effects matter? A phylogenetic meta-analysis of conspecific negative distance and density dependence experiments. *Ecol. Lett.* **24**, 608–620 (2020).
- Detto, M., Visser, M. D., Wright, S. J. & Pacala, S. W. Bias in the detection of negative density dependence in plant communities. *Ecol. Lett.* **22**, 1923–1939 (2019).
- LaManna, J. A., Mangan, S. A. & Myers, J. A. Conspecific negative density dependence and why its study should not be abandoned. *Ecosphere* **12**, e03322 (2021).
- Chisholm, R. A. & Fung, T. Comment on “Plant diversity increases with the strength of negative density dependence at the global scale”. *Science* **360**, eaar4685 (2018).
- Hülsmann, L. & Hartig, F. Comment on “Plant diversity increases with the strength of negative density dependence at the global scale”. *Science* **360**, eaar2435 (2018).
- Chen, L. et al. Differential soil fungus accumulation and density dependence of trees in a subtropical forest. *Science* **366**, 124–128 (2019).
- Aubrey, T. G. Positive density dependence acting on mortality can help maintain species-rich communities. *eLife* **9**, e57788 (2020).
- LaManna, J. A. et al. Consequences of local conspecific density effects for plant diversity and community dynamics. *Ecol. Lett.* **27**, e14506 (2024).
- Delavaux, C. S., Crowther, T. W., Bever, J. D., Weigelt, P. & Gora, E. M. Mutualisms weaken the latitudinal diversity gradient among oceanic islands. *Nature* **627**, 335–339 (2024).

23. Zhong, Y. et al. Arbuscular mycorrhizal trees influence the latitudinal beta-diversity gradient of tree communities in forests worldwide. *Nat. Commun.* **12**, 3137 (2021).
24. Bimler, M. D. & Mayfield, M. M. Ecology: lifting the curtain on higher-order interactions. *Curr. Biol.* **33**, R77–R79 (2023).
25. Chu, C. et al. Allelopathic effects of *Eucalyptus* on native and introduced tree species. *For. Ecol. Manag.* **323**, 79–84 (2014).
26. Lai, H., Chong, K. Y., Yee, A. T. K., Mayfield, M. M. & Stouffer, D. B. Non-additive biotic interactions improve predictions of tropical tree growth and impact community size structure. *Ecology* **103**, e03588 (2022).
27. Li, Y. et al. Beyond direct neighbourhood effects: higher-order interactions improve modelling and predicting tree survival and growth. *Nat. Sci. Rev.* **8**, nwa244 (2021).
28. Bairey, E., Kelsic, E. D. & Kishony, R. High-order species interactions shape ecosystem diversity. *Nat. Commun.* **7**, 1–7 (2016).
29. Grilli, J., Barabás, G., Michalska-Smith, M. J. & Allesina, S. Higher-order interactions stabilize dynamics in competitive network models. *Nature* **548**, 210–213 (2017).
30. Gibbs, T., Levin, S. A. & Levine, J. M. Coexistence in diverse communities with higher-order interactions. *Proc. Natl Acad. Sci. USA* **119**, e2205063119 (2022).
31. Wilbur, H. M. Competition, predation, and the structure of the *Ambystoma-Rana* sylvatica community. *Ecology* **53**, 3–21 (1972).
32. Neill, W. E. The community matrix and interdependence of the competition coefficients. *Am. Nat.* **108**, 399–408 (1974).
33. Mayfield, M. M. & Stouffer, D. B. Higher-order interactions capture unexplained complexity in diverse communities. *Nat. Ecol. Evol.* **1**, 62 (2017).
34. Zuidema, P. A., Jongejans, E., Chien, P. D., Doring, H. J. & Schieving, F. Integral projection models for trees: a new parameterization method and a validation of model output. *J. Ecol.* **98**, 345–355 (2010).
35. Merow, C. et al. Advancing population ecology with integral projection models: a practical guide. *Methods Ecol. Evol.* **5**, 99–110 (2014).
36. Wiegand, T. et al. Consequences of spatial patterns for coexistence in species-rich plant communities. *Nat. Ecol. Evol.* **5**, 965–973 (2021).
37. Wiegand, T. et al. Latitudinal scaling of aggregation with abundance and coexistence in forests. *Nature* **640**, 967–973 (2025).
38. Letten, A. D. & Stouffer, D. B. The mechanistic basis for higher-order interactions and non-additivity in competitive communities. *Ecol. Lett.* **22**, 423–436 (2019).
39. Buche, L., Bartomeus, I. & Godoy, O. Multitrophic higher-order interactions modulate species persistence. *Am. Nat.* **203**, 458–472 (2024).
40. Buche, L. et al. Neighbor density-dependent facilitation promotes coexistence and internal oscillation. *Ecol. Monogr.* **95**, e70040 (2025).

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Methods

Study sites and census data

Our study was based on multiple censuses of 32 large permanent forest dynamic plots, with an average plot size of 24.5 ha (range: 9–50 ha). Data were sourced from the Forest Global Earth Observatory (<http://www.forestgeo.si.edu>) and Chinese Forest Biodiversity Monitoring Network (<http://www.cfbiodiv.org>) (Fig. 2a and Supplementary Table 1). Plots span tropical to boreal terrestrial biomes with latitude ranging from 1.92° S to 61.30° N. All plots were established and censused several times following a standardized protocol⁴¹. In each census, all free-standing woody stems with a diameter at breast height (DBH) greater than 1 cm were tagged (unique ID), mapped (coordinates), identified (species identity), measured (DBH) and recorded (alive, dead or recruit). The census was repeated every 5 years to monitor forest dynamics (for instance, survival, growth and recruitment). Most plots were only censused twice. For a few plots with three or more censuses, we selected two consecutive censuses between 1998 and 2022 for analysis (Supplementary Table 1). Overall, we compiled data for more than 3 million trees of 5,000 species across the 32 study plots.

Growth and survival models with HOIs

We estimated species interactions from demographic growth and survival models using the compiled dynamic forest census data^{3,27}. For trees with more than one stem (that is, with multiple branches at height less than 1.3 m), we considered the survival and growth of the main stem. The growth of a focal tree f of a species i (Growth_{i_f}) was modelled as a function of its potential growth rate in the absence of neighbours (G_i), size (DBH_{i_f}), and neighbourhood pairwise (Pair_{i_f}) and higher-order effects (HOI_{i_f})^{27,42}:

$$\text{Growth}_{i_f} = G_i \times \text{DBH}_{i_f}^\gamma \times e^{\text{Pair}_{i_f}} \times e^{\text{HOI}_{i_f}}. \quad (1)$$

Similarly, the survival probability of a focal tree f of a species i (Survival_{i_f}) is modelled as²⁷:

$$\text{Survival}_{i_f} = \frac{1}{1 + e^{\lambda_i + \gamma_1 \times \text{DBH}_{i_f}^{-1} + \gamma_2 \times \text{DBH}_{i_f} + \gamma_3 \times \text{DBH}_{i_f}^2 + \text{Pair}_{i_f} + \text{HOI}_{i_f}}}, \quad (2)$$

where λ_i is the intrinsic survival probability in the absence of neighbours. The inverse of diameter ($\text{DBH}_{i_f}^{-1}$) is included to model rapid decline in mortality rate with increasing diameter, whereas the terms $\text{DBH}_{i_f}^2$ and DBH_{i_f} model the U-shaped senescence effect⁴³.

The pairwise effects of all neighbours on the focal tree i_f (Pair_{i_f}) can be decomposed into conspecific and heterospecific effects:

$$\text{Pair}_{i_f} = \alpha_{ii} \times n_{i,i_f} + \alpha_{ih} \times n_{h,i_f}, \quad (3)$$

where α_{ii} and α_{ih} are intraspecific and interspecific pairwise interaction coefficients, and n_{i,i_f} and n_{h,i_f} are conspecific and heterospecific pairwise neighbourhood crowding indices, respectively. Here we take a mean-field approximation⁴⁴ by replacing the species-specific interaction coefficients with a constant (average) heterospecific coefficient (α_{ih})^{36,37}. The higher-order effects of all neighbours and all neighbours' neighbours on focal tree i_f (HOI_{i_f}) can be decomposed into four components:

$$\text{HOI}_{i_f} = \beta_{iii} \times n_{ii,i_f} + \beta_{iih} \times n_{ih,i_f} + \beta_{ihh} \times n_{hi,i_f} + \beta_{hhh} \times n_{hh,i_f}, \quad (4)$$

where β_{iii} , β_{iih} , β_{ihh} and β_{hhh} are HOI coefficients associated respectively with intraspecific pairwise interactions modified by conspecific neighbours (β_{iii} : $\alpha_{ii} \leftarrow i$) or heterospecific neighbours (β_{iih} : $\alpha_{ii} \leftarrow h$), and interspecific pairwise interactions modified by conspecific neighbours (β_{ihh} : $\alpha_{ih} \leftarrow i$) or heterospecific neighbours (β_{hhh} : $\alpha_{ih} \leftarrow h$); and n_{ii,i_f} , n_{ih,i_f} , n_{hi,i_f} and n_{hh,i_f} are the corresponding higher-order neighbourhood

crowding indices. The conspecific and heterospecific pairwise neighbourhood crowding indices n_{i,i_f} and n_{h,i_f} add up the crowding contributions of all conspecific and heterospecific neighbours, respectively, around a focal tree within a given radius. The contribution of a neighbour is directly proportional to its size and inversely proportional to the distance to the focal individual (equations S3 and S4 in Supplementary Text 1). The contribution of a neighbour to the corresponding higher-order crowding indices n_{ii,i_f} and n_{hh,i_f} is further multiplied by its conspecific crowding index (equations S7 and S9 in Supplementary Text 1), and that to n_{ih,i_f} and n_{hh,i_f} is further multiplied by its heterospecific crowding index (equations S8 and S10 in Supplementary Text 1). The calculation of pairwise and higher-order neighbourhood crowding indices is summarized in Supplementary Text 1 according to our previously developed method²⁷.

Model fitting and evaluations

In each forest plot, we fitted demographic growth and survival models for each species with more than 100 trees (and also with at least 20 surviving and 20 dead trees for the survival model³) to ensure model performance and robustness. Overall, we fitted growth and survival models for 1,543 and 1,340 tree species–plot combinations, respectively. We log-transformed equation 1 to linearize the growth model and reduce model heteroscedasticity and residuals. We used logistic regression for tree survival (0 for death and 1 for alive). Overall, three classes of demographic models were constructed to determine the importance of tree size, pairwise interactions and HOIs: (1) null models without inclusion of biotic interactions, (2) pair-only models including pairwise interactions only, and (3) HOI-inclusive models including both pairwise interactions and HOIs. The pairwise interaction coefficients estimated from HOI-inclusive models were α_{true} (orange arrows in Fig. 1), because the effects of HOIs were isolated. By contrast, the pairwise coefficients estimated from pair-only models were α_{modified} (blue arrows in Fig. 1), because they implicitly incorporated effects of HOIs. Then, we compared the AIC values of the three models. The inclusion of HOIs was statistically supported when the AIC of the HOI-inclusive model was at least two units lower than those of the alternative models (Q1).

Latitudinal changes in biotic interactions

To explore the latitudinal trend in biotic interactions (Q2), we fitted exponential relationships between the estimated coefficients of pairwise interactions (α_{ii} and α_{ih}) and HOIs (β_{iii} , β_{iih} , β_{ihh} and β_{hhh}) and absolute latitude for competitive ($\alpha < 0$, $\beta < 0$) and facilitative ($\alpha > 0$, $\beta > 0$) interactions separately, given that they were nearly equally frequent (Extended Data Fig. 1). We further evaluated how HOIs (β_{ijk}) modified the true pairwise interactions ($\alpha_{ij,\text{true}}$) by examining their correlations.

Contribution of biotic interactions to latitudinal diversity gradient

To further assess how the latitudinal change in pairwise interactions and HOIs might contribute to the latitudinal tree diversity gradient (Q3), we first calculated the relative change in growth rate and survival probability caused by the neighbourhood cumulative effects of true pairwise interactions ($e^{\text{Pair}_{i_f}}$) and HOIs ($e^{\text{HOI}_{i_f}}$) separately for each focal tree (equations 1–4)^{27,33} and then averaged the relative change across trees for each species ($\text{RC}_{\text{pair}} = \sum_{f=1}^{N_i} e^{\text{Pair}_{i_f}} / N_i$, $\text{RC}_{\text{HOI}} = \sum_{f=1}^{N_i} e^{\text{HOI}_{i_f}} / N_i$). Finally, we explored how the relative changes (RC_{pair} and RC_{HOI}) varied with species abundance per hectare and absolute latitude using linear regression. The relative changes and species abundance were log-transformed to improve normality.

Results from the survival models (Extended Data Figs. 3–5 and Extended Data Table 1) were qualitatively consistent with those from the growth models. Therefore, we only reported the growth model results in the main text. Moreover, as demonstrated in Supplementary Text 2, the growth model results were robust to variations in parameter settings (for instance, for different neighbourhood

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radii), inclusion of spatial autocorrelation (with or without quadrats as random effects), uncertainty in estimation of interaction coefficients, and the distinction between small (DBH < 10 cm) and large (DBH ≥ 10 cm) trees. All analyses were conducted in R v.4.4.1 (ref. 45). The map in Fig. 2 was generated using R packages ggplot2 (v.4.0.0) and ggrepel (v.0.9.6).

Reporting summary

Further information on research design is available in the Nature Portfolio Reporting Summary linked to this article.

Data availability

The raw census data that support this study are available upon request and with permission of the principal investigators of the Forest Global Earth Observatory and Chinese Forest Biodiversity Monitoring Network networks (names and contact information of the principal investigators are provided in Supplementary Table 1). For some plots, the data are publicly available at <https://forestgeo.si.edu/explore-data>. The processed datasets supporting the findings of this study are publicly available at Figshare⁴⁶ (<https://doi.org/10.6084/m9.figshare.28426862>).

Code availability

The custom code used in this study is available at Figshare⁴⁶ (<https://doi.org/10.6084/m9.figshare.28426862>).

41. Condit, R. *Tropical Forest Census Plots: Methods and Results from Barro Colorado Island, Panama and a Comparison with Other Plots* (Springer, 1998).
42. Kunstler, G. et al. Plant functional traits have globally consistent effects on competition. *Nature* **529**, 204–207 (2016).

43. Monserud, R. A. & Sterba, H. Modeling individual tree mortality for Austrian forest species. *For. Ecol. Manag.* **113**, 109–123 (1999).
44. O'Dwyer, J. P. & Chisholm, R. A mean field model for competition: from neutral ecology to the Red Queen. *Ecol. Lett.* **17**, 961–969 (2014).
45. R Core Team. *R: A Language and Environment for Statistical Computing* (R Foundation for Statistical Computing, 2024).
46. Li, Y. et al. Data and code for: higher-order interactions enhance the latitudinal tree diversity gradient. *Figshare* <https://doi.org/10.6084/m9.figshare.28426862.v1> (2026).

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Author contributions Y. Li, C.C. and J.X. conceived the study in conversation with M.M.M. Y. Li and C.C. assembled the forest census data. Y.J. and Y. Li cleaned and formatted the data. J.X. and Y. Li analysed the data and generated figures and tables. Y. Li wrote the manuscript with input from C.C., J.X., S.J.W., M.M.M. and O.G. F.H. made substantial contributions during revisions. The authors (including A.A., K.J.A.-T., J.B., J.D.B., P.B., N.A.B., W.B., D.F.R.P.B., M.C., K.C., S.J.D., Q.D., S.E., A.F., E.S.F., G.S.G., Z.H., J.H., M.J., G.J., D.J.J., A.S.J., K.K., A.J.L., B.L., J.L., L.L., F.L., Y. Liu, Z.L., J.A.L., K.M., S.M.M., W.M., H.R.M., X.M., J.A.M., M.N., A.N., M.J.O., N.L.E.O., G.P., R.P.P., X.Q., H.R., G.R., L.J.V.R., P.S., G.S., Z.S., J.S., M.E.S., J.T., M.U., Xihua Wang, Xugao Wang, Y.W., T.L.Y., W.Y., M.Y., M.Z., Y.Z., J.Z., F.H. and C.C.) contributed forest census data and revised the manuscript.

Competing interests The authors declare no competing interests.

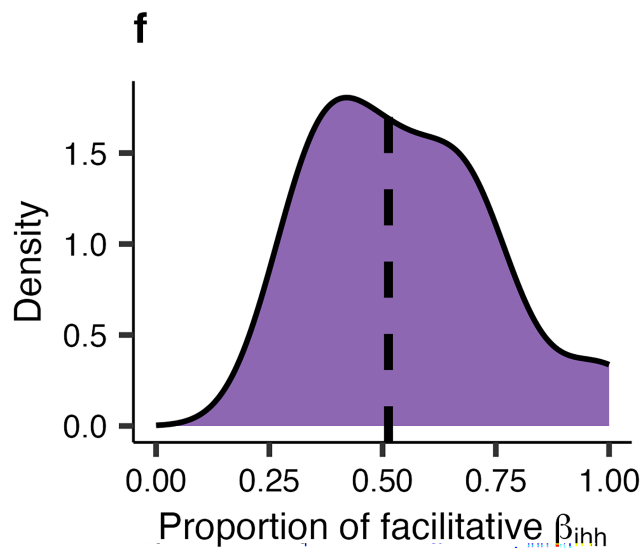
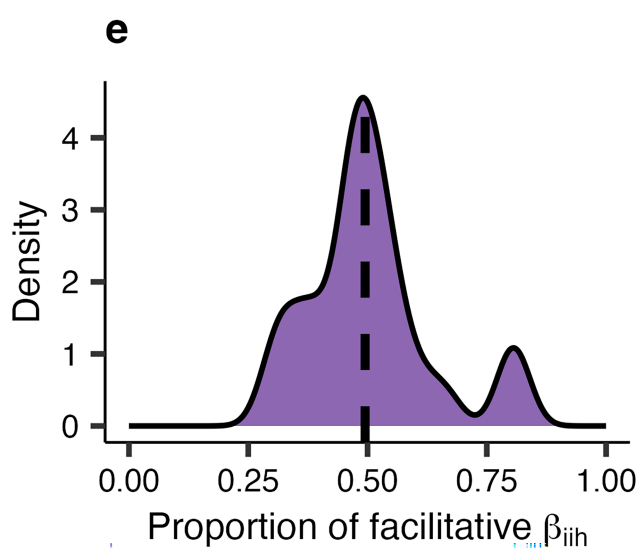
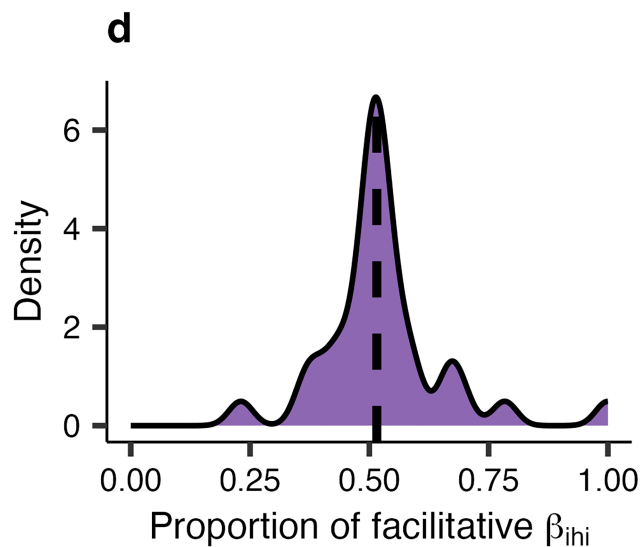
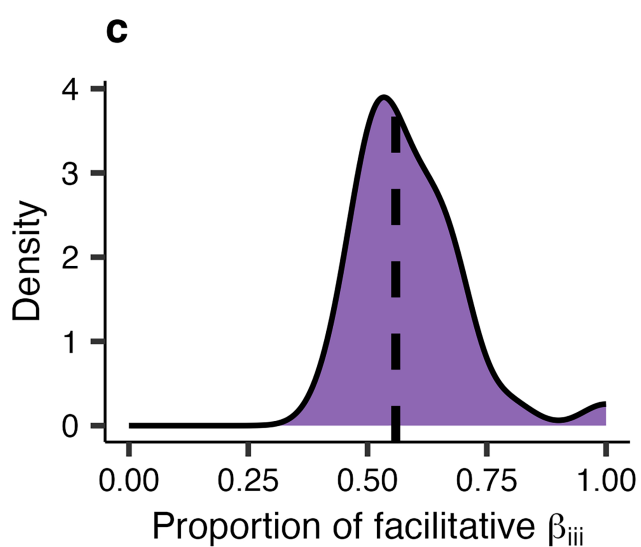
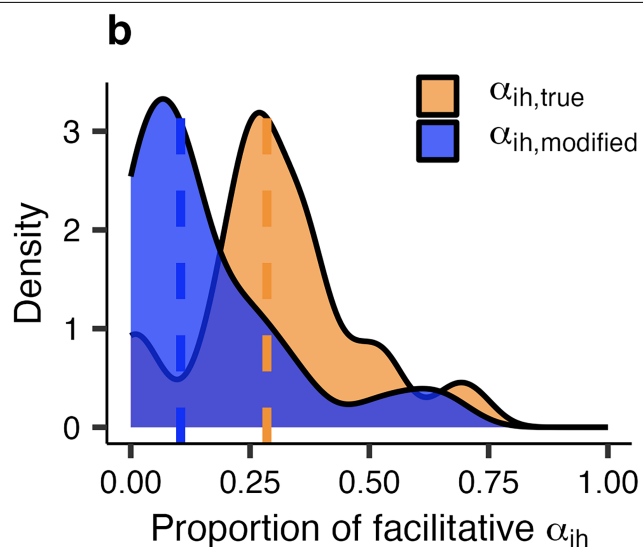
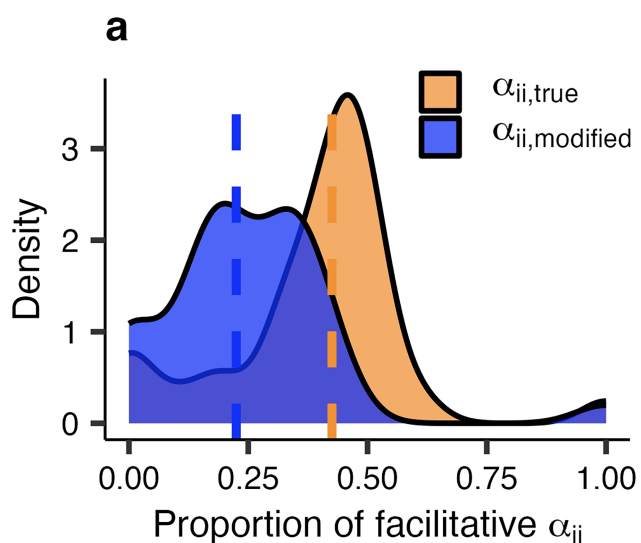
Additional information

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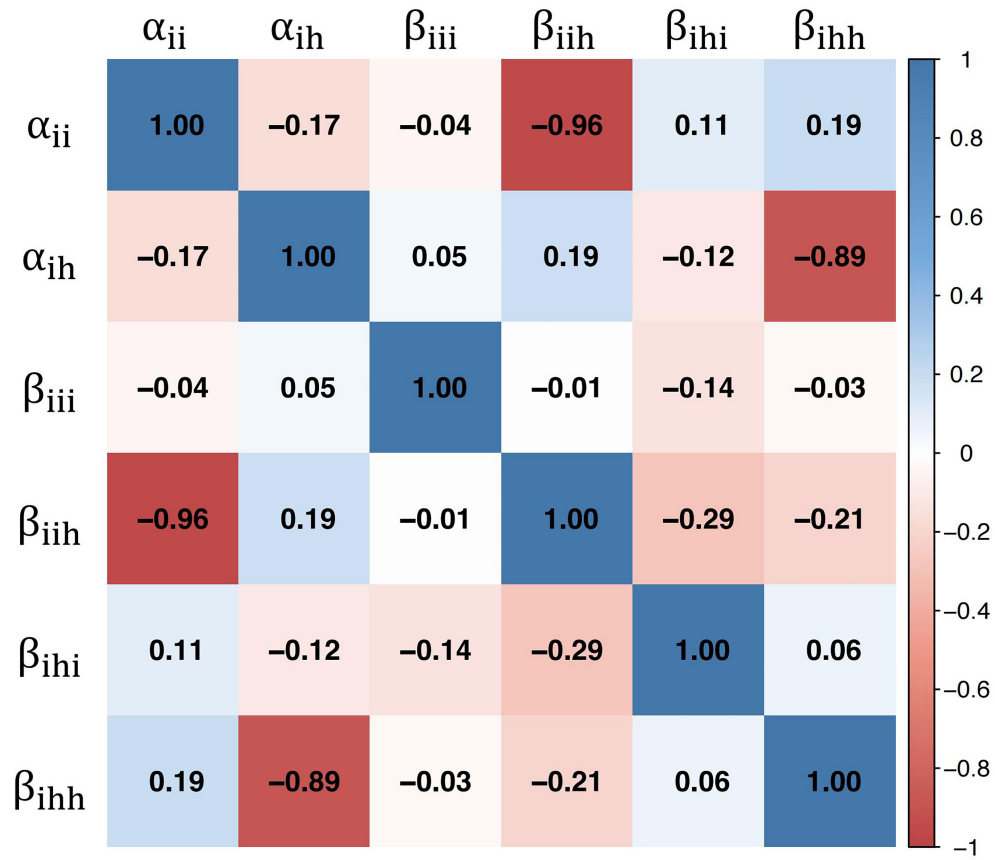


Extended Data Fig. 1 | See next page for caption.

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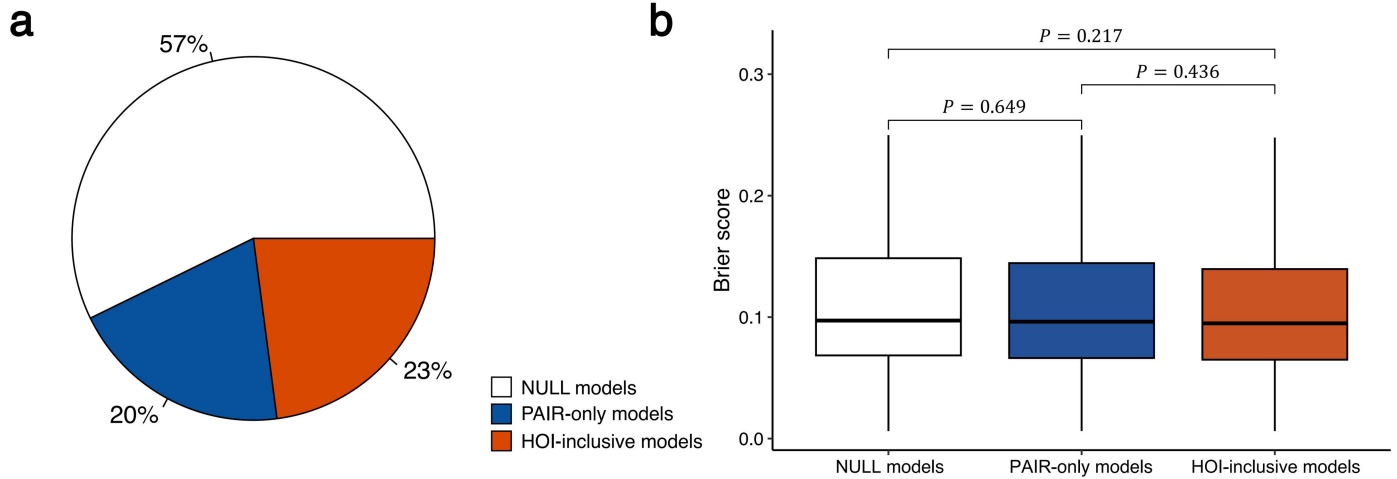
Extended Data Fig. 1 | Proportion of facilitative pairwise and higher-order interactions estimated from growth models. Panels (a) and (b) show the distribution of proportion of facilitative intraspecific (α_{ii}) and interspecific (α_{ih}) pairwise interactions, respectively. The pairwise interactions estimated from PAIR-only models and HOI-inclusive models, denoted as α_{modified} and α_{true} , are distinguished by blue and orange color. Panels (c) and (e) show the distribution of proportion of facilitative higher-order interaction coefficients

β_{iii} and β_{iir} , which represent the modifications of intraspecific pairwise interactions by conspecific neighbors ($\alpha_{ii} \leftarrow i$) and by heterospecific neighbors ($\alpha_{ii} \leftarrow h$), respectively. Panels (d) and (f) show the distribution of proportion of facilitative higher-order interaction coefficients β_{ihi} and β_{ihh} , corresponding to the modifications of interspecific pairwise interactions by conspecific neighbors ($\alpha_{ih} \leftarrow i$) and heterospecific neighbors ($\alpha_{ih} \leftarrow h$), respectively.



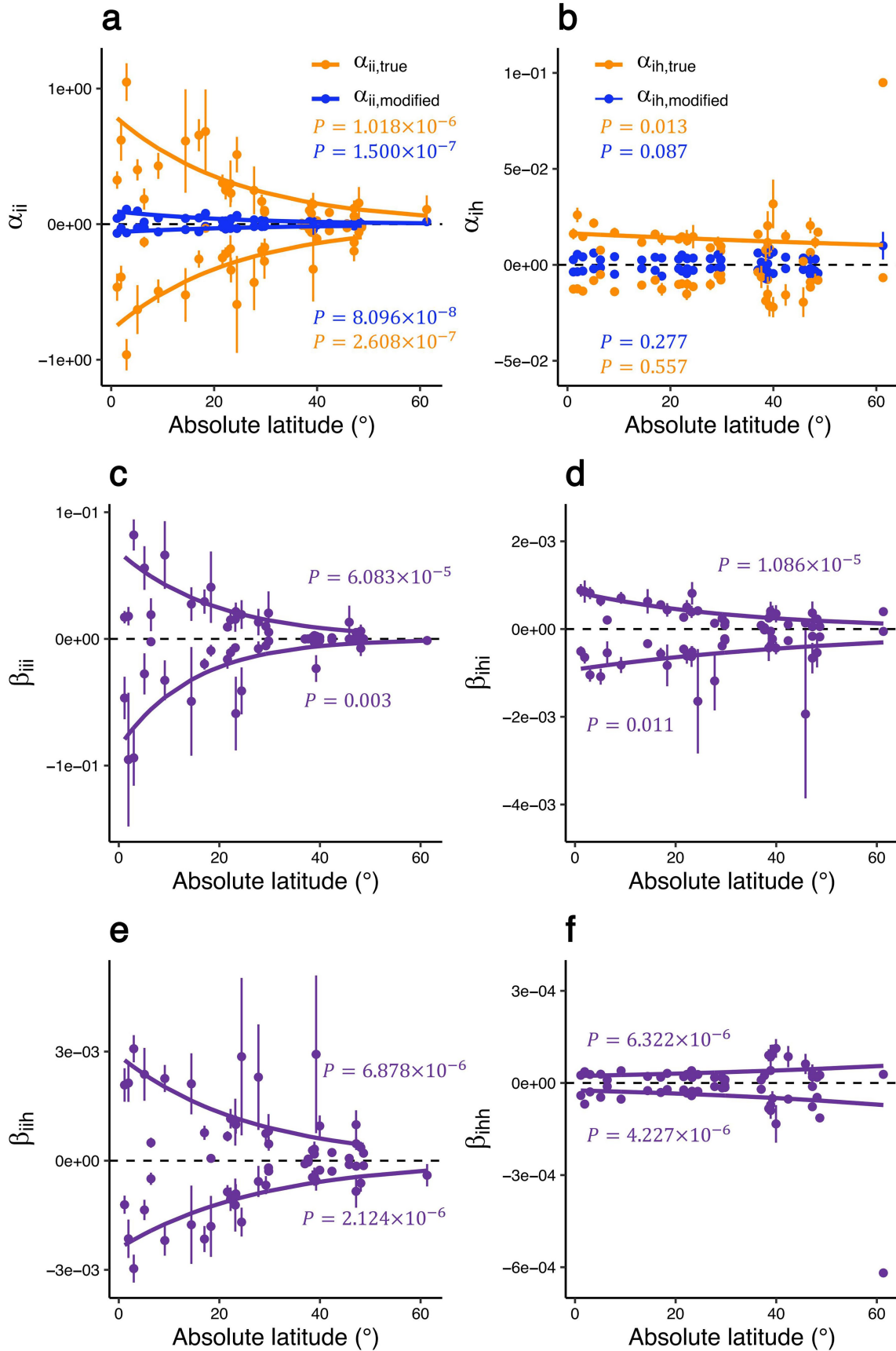
Extended Data Fig. 2 | Pearson correlation coefficients between different types of interactions estimated from growth models. Here, α_{ii} and α_{ih} denote intraspecific and interspecific pairwise interactions estimated from HOI-inclusive models. β_{iii} , β_{iih} , β_{ihi} and β_{ihh} are higher-order interaction coefficients referring to intraspecific pairwise interactions modified by

conspecific neighbors ($\alpha_{ii} \leftarrow i$), intraspecific pairwise interactions modified by heterospecific neighbors ($\alpha_{ii} \leftarrow h$), interspecific pairwise interactions modified by conspecific neighbors ($\alpha_{ih} \leftarrow i$), and interspecific pairwise interactions modified by heterospecific neighbors ($\alpha_{ih} \leftarrow h$), respectively.



Extended Data Fig. 3 | Evidence of higher-order interactions from survival models across 32 forest plots. Panel (a) shows the percentage of 1,340 tree species–plot combinations across the 32 plots for which each of the three classes of survival models was best supported (AIC at least 2 units lower than that of the two alternative models). The white, blue and orange regions indicate the percentage of species supporting the NULL models (models without including biotic interactions), PAIR-only models (models includes only pairwise interactions) and the HOI-inclusive models (models include both pairwise interactions and HOIs), respectively. Panel (b) is box-plot of the Brier score between observed survival outcome and predicted survival

probabilities by NULL, PAIR-only and HOI-inclusive models for the 306 species supported by the HOI-inclusive models. Brier score is calculated as $\frac{1}{n} \sum_{i=1}^n (y_i - p_i)^2$, where y_i is the observed survival outcome (0 for death and 1 for alive), p_i is the predicted survival probability for observation i , and n is the sample size for each species. A lower Brier score indicates more accurate predictions. Box plots show the median (center line), the 25th and 75th percentiles (box limits), and whiskers extending to the most extreme data points within $1.5 \times$ the interquartile range. Differences among models are assessed using pairwise two-sided t -test ($N = 306$ species); exact P values are shown in the figure. No adjustment is made for multiple comparisons.

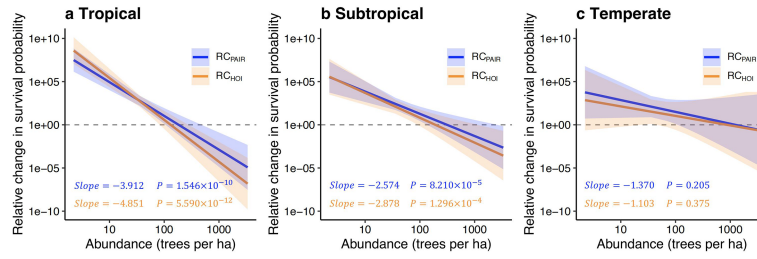


Extended Data Fig. 4 | See next page for caption.

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Extended Data Fig. 4 | Latitudinal changes in pairwise and higher-order interactions for survival models. Panels (a) and (b) display the latitudinal changes in intraspecific (α_{ii}) and interspecific (α_{ij}) pairwise interactions, respectively. The pairwise interactions estimated from PAIR-only models and HOI-inclusive models, denoted as α_{modified} and α_{true} , are distinguished by blue and orange points (lines). Panels (c) and (e) show the latitudinal changes in higher-order interaction coefficients in which intraspecific pairwise interactions are modified by conspecific neighbors (β_{iij}) and by heterospecific neighbors (β_{ihh}), respectively. Panels (d) and (f) show the latitudinal changes in higher-order interaction coefficients in which interspecific pairwise

interactions are modified by conspecific neighbors (β_{iij}) and by heterospecific neighbors (β_{ihh}), respectively. The insets in each panel describe the six types of interaction (see Fig. 1). Species-level pairwise and higher-order interactions were related to absolute latitude separately for competitive ($\alpha < 0, \beta < 0$) and facilitative interactions ($\alpha > 0, \beta > 0$) using exponential models. Significance of the regression coefficients is assessed using two-sided *t*-test; exact *P* values are shown in the figure. Regression lines are shown only when the interaction strength changed significantly with latitude ($P < 0.05$). For clarity, we displayed plot-level mean values \pm SEM rather than species-level estimates.



Extended Data Fig. 5 | Predicted relationships between relative change in survival probability and species abundance across three latitudinal geographic zones. RC_{PAIR} and RC_{HOI} that represent relative changes in the odds of survival ($\frac{p}{1-p}$, p is survival probability) caused by cumulative effects of pairwise and higher-order interactions, are distinguished by blue and orange lines. Predictions are shown for three geographic zones: tropical zone

(0°–23.5°, a), subtropical zone (23.5°–35°, b), and temperate zone (35°–66.5°, c). Predictions are generated from the linear models summarized in Extended Data Table 1, using the middle latitude of each zone (i.e., 11.75° for tropical zone, 29.25° for subtropical zone and 45° for temperate zone). Solid lines show model predictions and shaded areas represent 95% confidence intervals.

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Extended Data Table 1 | Summary of relative changes in survival probability varying with species abundance and absolute latitude

Response	Effect	Estimate	Standard error	<i>t</i> value	<i>P</i> value
RC _{PAIR}	Intercept	24.136	2.977	8.106	1.17×10 ⁻¹⁵
	Abundance	-4.810	0.884	-5.438	6.39×10 ⁻⁸
	Latitude	-0.320	0.138	-2.322	0.020
	Abundance:Latitude	0.076	0.036	2.106	0.035

RC _{HOI}	Intercept	29.378	3.427	8.574	2.73×10 ⁻¹⁷
	Abundance	-6.176	1.018	-6.068	1.68×10 ⁻⁹
	Latitude	-0.487	0.158	-3.075	0.002
	Abundance:Latitude	0.113	0.042	2.699	0.007

RC_{PAIR} and RC_{HOI} represent relative changes in the odds of survival ($\frac{p}{1-p}$, *p* is survival probability) caused by cumulative effects of pairwise and higher-order interactions, respectively. The relative change in growth rate is fitted by linear models using log-transformed species abundance, absolute latitude and their interaction as predictors (*N*=1,340 species).

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|-----------------|---|
| Data collection | No software was used for data collection. |
| Data analysis | Data analyses were conducted in R version 4.4.1 using several packages, including DHARMA (v0.4.7), minpack.lm (v1.2-4) and ggplot2(v4.0.0). Additional packages used in the analyses are documented in the code. The custom code used in this study is available at Figshare: https://doi.org/10.6084/m9.figshare.28426862 . |

For manuscripts utilizing custom algorithms or software that are central to the research but not yet described in published literature, software must be made available to editors and reviewers. We strongly encourage code deposition in a community repository (e.g. GitHub). See the Nature Portfolio [guidelines for submitting code & software](#) for further information.

Data

Policy information about [availability of data](#)

All manuscripts must include a [data availability statement](#). This statement should provide the following information, where applicable:

- Accession codes, unique identifiers, or web links for publicly available datasets
- A description of any restrictions on data availability
- For clinical datasets or third party data, please ensure that the statement adheres to our [policy](#)

The raw census data that support this study are available upon request and with permission of the principal investigators of the ForestGEO and CForBio networks

(names and contact information of the PIs are provided in Supplementary Table 1). For some plots, the data are publicly available at <https://forestgeo.si.edu/explore-data>. The processed datasets supporting the findings of this study are publicly available at Figshare: <https://doi.org/10.6084/m9.figshare.28426862>.

Research involving human participants, their data, or biological material

Policy information about studies with [human participants or human data](#). See also policy information about [sex, gender \(identity/presentation\), and sexual orientation](#) and [race, ethnicity and racism](#).

Reporting on sex and gender	Not applicable.
Reporting on race, ethnicity, or other socially relevant groupings	Not applicable.
Population characteristics	Not applicable.
Recruitment	Not applicable.
Ethics oversight	Not applicable.

Note that full information on the approval of the study protocol must also be provided in the manuscript.

Field-specific reporting

Please select the one below that is the best fit for your research. If you are not sure, read the appropriate sections before making your selection.

Life sciences Behavioural & social sciences Ecological, evolutionary & environmental sciences

For a reference copy of the document with all sections, see nature.com/documents/nr-reporting-summary-flat.pdf

Ecological, evolutionary & environmental sciences study design

All studies must disclose on these points even when the disclosure is negative.

Study description	In this study, we assembled census data from 32 large permanent forest plots spanning tropical to boreal forests to address three key questions: (Q1) Are HOIs prevalent among trees across forest plots? (Q2) How do pairwise interactions and HOIs vary with latitude? (Q3) How do latitudinal changes in HOIs contribute to the latitudinal tree diversity gradient? To answer these questions, we estimated both pairwise interactions and HOIs from demographic growth (for 1,543 tree species–plot combinations) and survival models (for 1,340 tree species–plot combinations), respectively. With these data, we built three types of growth and survival models: (i) NULL models with no biotic interactions, (ii) PAIR-only models including only pairwise interactions, and (iii) HOI-inclusive models including both pairwise interactions and HOIs. We then compared AIC of the three types of models (Q1), tested whether the estimated pairwise interactions and HOIs declined with latitudes (Q2), and evaluated how the cumulative effects of pairwise interactions and HOIs on growth and survival changed with species abundance and latitude (Q3).
Research sample	The data used in this study were collected at 32 large permanent forest dynamic plots worldwide from the Forest Global Earth Observatory (ForestGEO, http://www.forestgeo.si.edu) and Chinese Forest Biodiversity Monitoring Network (CForBio, http://www.cfbiodiv.org).
Sampling strategy	Most plots had only been censused twice. For a few plots with three or more censuses, we selected two consecutive censuses between 1998 and 2022 for analysis. Overall, we compiled data for over three million trees of 5,000 species across 32 plots.
Data collection	All plots were established and censused multiple times following a standardized protocol. In each census, all free-standing woody stems with a diameter at breast height (DBH) larger than 1 cm were tagged (unique ID), mapped (coordinates), identified (species identity), measured (DBH), and recorded (alive, dead or recruit). The census was repeated every 5 years to monitor forest dynamics (e.g., survival, growth, and recruitment).
Timing and spatial scale	The 32 forest plots span tropical to boreal terrestrial biomes with latitude ranging from 1.92° S to 61.30° N. These plots vary in size between 9 and 50 ha (24.5 ha on average). At all plots included in this study, two or more censuses have been carried out with remeasurement intervals of approximately five years. Most plots had only been censused twice. For a few plots with three or more censuses, we selected two consecutive censuses between 1998 and 2022 for analysis.
Data exclusions	Observations of trees were excluded when information on coordinates, species, or status was missing. We fit demographic growth and survival models for each species with more than 100 trees (and additionally with at least 20 alive and dead observations for the survival model) to ensure model performance and robustness.
Reproducibility	This study is based on observational data, and no experiments are conducted.
Randomization	This study is based on observational data. Therefore, randomization into groups does not apply.

Blinding

Blinding is not relevant for this study because the data were not specifically collected to assess direct and higher-order interactions.

Did the study involve field work? Yes No

Reporting for specific materials, systems and methods

We require information from authors about some types of materials, experimental systems and methods used in many studies. Here, indicate whether each material, system or method listed is relevant to your study. If you are not sure if a list item applies to your research, read the appropriate section before selecting a response.

Materials & experimental systems

n/a	Involvement in the study
<input checked="" type="checkbox"/>	<input type="checkbox"/> Antibodies
<input checked="" type="checkbox"/>	<input type="checkbox"/> Eukaryotic cell lines
<input checked="" type="checkbox"/>	<input type="checkbox"/> Palaeontology and archaeology
<input checked="" type="checkbox"/>	<input type="checkbox"/> Animals and other organisms
<input checked="" type="checkbox"/>	<input type="checkbox"/> Clinical data
<input checked="" type="checkbox"/>	<input type="checkbox"/> Dual use research of concern
<input checked="" type="checkbox"/>	<input type="checkbox"/> Plants

Methods

n/a	Involvement in the study
<input checked="" type="checkbox"/>	<input type="checkbox"/> ChIP-seq
<input checked="" type="checkbox"/>	<input type="checkbox"/> Flow cytometry
<input checked="" type="checkbox"/>	<input type="checkbox"/> MRI-based neuroimaging

Plants

Seed stocks

Not applicable.

Novel plant genotypes

Not applicable.

Authentication

Not applicable.