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Impact of Spatial Variability in Zooplankton Grazing Rates on Carbon Export Flux

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Key Points:

- Inverse modeling predicts strong spatial variability in global grazing dynamics for two zooplankton functional types
- Locally tuned zooplankton grazing dynamics improve the model's ability to reproduce satellite-derived phytoplankton biomass
- Locally tuned zooplankton grazing dynamics can decrease mean carbon flux by 17% and modify the routing of carbon export

Supporting Information:

Supporting Information may be found in the online version of this article.

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Abstract The biological carbon pump is a key controller of how much carbon is stored within the global ocean. This pathway is influenced by food web interactions between zooplankton and their prey. In global biogeochemical models, Holling Type functional responses are frequently used to represent grazing interactions. How these responses are parameterized greatly influences biomass and subsequent carbon export estimates. The half-saturation constant, or k value, is central to the Holling functional response. Empirical studies show k can vary over three orders of magnitude, however, this variation is poorly represented in global models. This study derives zooplankton grazing dynamics from remote sensing products of phytoplankton biomass, resulting in global distribution maps of the grazing parameter k . The impact of these spatially varying k values on model skill and carbon export flux estimates is then considered. This study finds large spatial variation in k values across the global ocean, with distinct distributions for micro- and mesozooplankton. High half-saturation constants, which drive slower grazing, are generally associated with areas of high productivity. Grazing rate parameterization is found to be critical in reproducing satellite-derived distributions of small phytoplankton biomass, highlighting the importance of top-down drivers for this size class. Spatially varying grazing dynamics decrease mean total carbon export by >17% compared to globally homogeneous dynamics, with increases in fecal pellet export and decreases in export from algal aggregates. This study highlights the importance of grazing dynamics to both community structure and carbon export, with implications for modeling marine carbon sequestration under future climate scenarios.

1. Introduction

The ocean plays a major role in mitigating the impact of climate change (Hoegh-Guldberg & Bruno, 2010). It is thought that over 20% of anthropogenic carbon dioxide emissions are stored within the global ocean (Friedlingstein et al., 2022). The biological carbon pump describes a suite of processes which can transport organic carbon from the surface ocean, to depths of over 1,000 m (Turner, 2015). This pathway is responsible for approximately 10% of the ocean's carbon inventory (DeVries, 2022). As carbon dioxide emissions are predicted to increase over the 21st Century, it is essential to fully understand the processes underlying carbon sequestration via the biological pump and predict how they will change in the future (Siegel et al., 2022).

Particulate organic carbon (POC) is exported out of the surface ocean as the fecal pellets of consumers or as aggregates of phytoplankton (Siegel et al., 2022). The rate that these forms of particulate carbon are exported via the biological carbon pump is directly influenced by zooplankton grazing. Grazing rates impact the biomass of both predator and prey (Rohr et al., 2024) and consequently the production of algal aggregates and fecal pellets. The production of aggregates and fecal pellets are influenced in opposite directions by grazing pressure. Slower grazing rates, for example, reduce zooplankton biomass and the number of fecal pellets produced by consumers, but increase the amount of prey biomass and thus algal aggregate formation. The proportion of total carbon export attributed to these two pathways varies globally. The contribution of algal aggregates increases in high phytoplankton biomass regions (e.g., the high latitudes and upwelling areas) where satiated or slower grazing zooplankton are not capable of consuming all the phytoplankton biomass (Siegel et al., 2014, 2022). In contrast, the subtropical gyres are characterized by lower phytoplankton biomass and a community dominated by microzooplankton grazing rapidly on small phytoplankton. The microzooplankton fecal pellets are readily consumed by the microbial loop in the euphotic zone due to their smaller size (Calbet & Landry, 2004), resulting in lower carbon export flux rates for this oceanic region.

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Biogeochemical (BGC) models can estimate particulate carbon using a combination of grazing rates and mortalities (e.g., Aumont et al. (2015)). In BGC models, grazing dynamics between predator and prey can be described by a food limited functional response (T. R. Anderson et al., 2010; Gentleman & Neuheimer, 2008; Vallina et al., 2014). This dictates how ingestion rates change with prey density (Gentleman & Neuheimer, 2008). The choice and parameterization of grazing functional responses can impact estimates of carbon export (T. R. Anderson et al., 2010). Holling Type II or Type III (Holling, 1959) grazing formulations are commonly used (Kearney et al., 2021; Rohr et al., 2022). These formulations require two parameters: the maximum grazing rate, g , and the half-saturation constant, k . The half-saturation constant represents the concentration of prey at which half the maximum grazing rate is reached (Gentleman & Neuheimer, 2008). Together these two parameters describe the shape and magnitude of the functional response. Ecologically, they represent the time taken to capture and consume prey (Rohr et al., 2022)—characteristics that vary with species physiology (Hansen et al., 1997; Hirst & Bunker, 2003). Although the functional response is described by both parameters, population dynamics are most sensitive to change in the k value (Rohr et al., 2022), which is the focus of this study.

Laboratory measurements of k values (Hansen et al., 1997; Hirst & Bunker, 2003) show a large range in k values across all species studied, spanning 0.96–6,000 mgC m³. Laboratory (Hansen et al., 1997), ecological (Barton et al., 2013) and modeling (Rohr et al., 2024) studies also point toward strong spatial variability in k values. However, even the most complex BGC models have fixed, globally homogenous, k values. These models can simulate spatial variability in grazing dynamics through the competition of multiple plankton functional types, but this likely does not capture the full physiological variability. Some models use mechanisms such as multi-prey responses (e.g., T. R. Anderson et al., 2010, 2015) and prey preferences (e.g., Aumont et al., 2015) to further emulate this variability, but these mechanisms are limited and there is little observational data to confirm what emergent grazing dynamics should be (Rohr et al., 2024). Furthermore, there is uncertainty around the impact of zooplankton grazing on carbon flux, which contributes to the large variability in global estimates of carbon export (Boyd, 2015; Rohr et al., 2024; Siegel et al., 2014).

We address these gaps by using an inverse modeling approach to estimate spatial variation in zooplankton k values. We find large variations with notable implications for carbon export. We first describe our approach, then our findings and then the implications of these for our overall understanding of zooplankton and carbon export.

2. Methodology

2.1. Overview

This study builds on the work of Siegel et al. (2014) which used satellite-derived estimates of Net Primary Productivity (NPP) and phytoplankton biomass to predict global grazing rates and subsequent estimates of carbon export. The work by Siegel et al. (2014) was extended by Archibald et al. (2019) to include diel vertical migrations (DVM) by zooplankton, allowing particulate organic matter to be exported via both passive sinking and the vertical movements of organisms. We modified the model by Archibald et al. (2019) to include explicit grazing and zooplankton biomass pools.

Here, we use a 0-D BGC box model to infer the optimal k parameters for both microzooplankton and meso-zooplankton, within each grid cell of a 1×1 degree global domain. We force this model with observed bottom-up controls (phytoplankton cell division rates) but allow it to prognostically compute NPP, phytoplankton biomass, zooplankton biomass and carbon export. Phytoplankton and zooplankton biomass pools are divided into two functional groups each (2P2Z). We then run a suite of simulations to determine what combination of k values is required to best match satellite derived phytoplankton biomass and thus infer the spatial distribution of grazing dynamics. Finally, to understand how more realistic zooplankton behavioral diversity influences marine carbon cycling, we compare global prognostic ecosystem biomass and carbon export from three model scenarios: a run using non-optimized, globally homogenous k values derived from literature (Baseline scenario); a run using optimized, globally homogenous k values (Global- k scenario), and a run using optimized, locally tuned k values (Local- k scenario). The model inputs (Section 2.2), the ecosystem sub-model (Section 2.3), the approach to determine optimized k values (Section 2.4), the carbon export sub-model (Section 2.5) and results analysis (Section 2.6) are discussed below.

2.2. Input Data

The model is forced by satellite-derived phytoplankton community mean growth rates, μ . The use of μ ensures coupling between NPP and grazing dynamics. Without this coupling, the top-down influence of grazing dynamics would be removed. This would make overgrazing of phytoplankton an impossibility, as there would always be NPP regardless of what the free-running biomass population is. μ was selected as an input over the explicit representation of nutrients to ensure observational forcing remained.

In the Carbon-based Productivity Model (CbPMv2) (Westberry et al., 2008), μ is computed from satellite derived chlorophyll-to-carbon ratios (Behrenfeld et al., 2005). NPP can then be derived using the relationship between μ and satellite-derived estimates of phytoplankton carbon biomass (P_{obs}), where

$$NPP = \mu P_{obs} \quad (1)$$

In CbPMv2, μ is computed for the bulk phytoplankton population; however, in this study we needed to differentiate the growth rates of two phytoplankton classes, to force our 2P2Z model. The distribution of particle backscatter can partition phytoplankton carbon biomass across size classes, but not their respective growth rates. To estimate the partitioning of growth rates into two size classes (μ_i) we assume a fixed allometric ratio, then determine the values required to produce bulk NPP from the two biomass pools. The following three equations are satisfied at each time step and location:

$$P_{obs} = PS_{obs} + PL_{obs} \quad (2)$$

$$NPP = \mu_{PS} PS_{obs} + \mu_{PL} PL_{obs} \quad (3)$$

$$\frac{\mu_{PL}}{\mu_{PS}} = \left(\frac{M_{PL}}{M_{PS}} \right)^{-0.25} \quad (4)$$

Both NPP and P_{obs} are derived from monthly climatologies presented in detail in Siegel et al. (2014). These climatologies are then interpolated to produce daily data. NPP values come from the Carbon-based Productivity Model (CbPMv2) (Westberry et al., 2008), which uses observations made by the Sea-viewing Wide-Field-of-view (SeaWiFS) satellite ocean color mission between 1997 and 2008 (McClain, 2009; Siegel et al., 2013, 2014). Phytoplankton biomass values (P_{obs}) are estimated using particulate backscattering coefficient data (Behrenfeld et al., 2005; Kostadinov et al., 2010; Siegel et al., 2013; Westberry et al., 2008). P_{obs} is partitioned into two size classes (PS and PL) using the slope of the particle size spectrum (Kostadinov et al., 2010; Siegel et al., 2014). M_i represents body size for the two size classes which has an allometric scaling constant of -0.25 applied in accordance with metabolic theory (e.g., West et al., 1997). This ensures the growth rate of small phytoplankton is always faster than microphytoplankton. For M_i , the same lower size limit implemented to partition P_{obs} is used, that is, 20 and 0.5 μm for PL and PS respectively (Kostadinov et al., 2010; Siegel et al., 2014). A maximum value for μ_i is set at 2 d^{-1} to correspond with the CbPMv2 data (Westberry et al., 2008). Observed minimum growth rates are approximately 0.1 d^{-1} , however the CbPMv2 model extrapolates this toward 0 (Westberry et al., 2008). In our study we use a minimum growth rate within the range of these two values (0.01 d^{-1}).

In the CbPMv2, all properties are assumed to be constant and distributed evenly within the mixed layer (Westberry et al., 2008). Within this study, phytoplankton biomass is assumed to be homogeneous across the mixed layer and negligible below the mixed layer depth as in Siegel et al. (2014). Integrated NPP is assumed constant across the euphotic zone as per Siegel et al. (2014). To enable the calculation of μ , depth integrated NPP is divided by the greater of euphotic zone depth (Z_{eu}) or mixed layer depth (Z_m), before dividing by volumetric phytoplankton biomass concentration. Depth data is interpolated from monthly climatologies also presented in detail in Siegel et al. (2014).

2.3. Ecosystem Sub-Model

A simple Phytoplankton-Zooplankton (2P2Z) model is constructed (Table 1). To run the ecosystem model, the global ocean is divided into a 1° latitude/longitude grid. The model is only run in grid cells with remote sensing products for a minimum of 10 out of 12 months. This limits the model to roughly between 50 and -50°N ,

Table 1
Model Variables, Parameters, and Forcing Fields

Parameters				
Symbol	Description	Value	Unit	Refs.
dt	Model time step	1	day	–
k_0	Microzooplankton half-saturation constant	–	mgC m^{-3}	–
k_1	Mesozooplankton half-saturation constant	–	mgC m^{-3}	–
b_{ZS}	Microzooplankton gross growth efficiency	0.3	–	Aumont et al. (2015)
b_{ZL}	Mesozooplankton gross growth efficiency	0.5	–	Archibald et al. (2019) and T. R. Anderson et al. (2010)
g_{ZS}	Maximum grazing rate of microzooplankton	2	d^{-1}	Rohr et al. (2022)
g_{ZL}	Maximum grazing rate of mesozooplankton	2	d^{-1}	Rohr et al. (2022)
m_p	Phytoplankton mortality	0.1	d^{-1}	Archibald et al. (2019) and Siegel et al. (2014)
m_{ZS}	Natural mortality microzooplankton	0.05	d^{-1}	Aumont et al. (2015) and Walker et al. (2019)
m_{ZL}	Natural mortality mesozooplankton	0.005	d^{-1}	Aumont et al. (2015)
PS_0/PL_0	Phytoplankton mortality refuge	0.2	mgC m^{-3}	Aumont et al. (2015)
ZS_0/ZL_0	Zooplankton mortality refuge	1	mgC m^{-3}	Archibald et al. (2019) and Stock and Dunne (2010)
p_{ZL}	Quadratic mortality mesozooplankton	0.02	$\text{m}^3\text{mgC}^{-1}\text{d}^{-1}$	Aumont et al. (2015) and T. R. Anderson et al. (2015)
$bact$	Bacterial remineralisation rate	0.025	d^{-1}	Aumont et al. (2015)
agg_i	Phytoplankton aggregation term	0.01/0.03	$\text{m}^3\text{mgC}^{-1}\text{d}^{-1}$	Stock et al. (2020) and Bisson et al. (2020)
p_{dvm}	Proportion of mesozooplankton that migrate	0.5	–	Archibald et al. (2019)
m_{fec}	Fraction of grazing going into fecal flux	0.3	–	Archibald et al. (2019) and Siegel et al. (2014)
f_{met}	Fraction of absorbed carbon metabolized	0.5	–	Archibald et al. (2019) and Steinberg and Landry (2016)
Forcing fields				
Symbol	Description	Unit	Refs.	
Z_{ml}	Mixed layer depth	m	Archibald et al. (2019) and Siegel et al. (2014)	
Z_{eu}	Depth of euphotic layer	m	Archibald et al. (2019) and Siegel et al. (2014)	
M	Phytoplankton growth rate	d^{-1}	Behrenfeld et al. (2005), Westberry et al. (2008), and Kostadinov et al. (2010)	
Prognostic variables				
Symbol	Description	Unit		
NPP	Net Primary Productivity	$\text{mgC m}^{-2}\text{d}^{-1}$		
PS	Small phytoplankton biomass	mgC m^{-3}		
PL	Microphytoplankton biomass	mgC m^{-3}		
ZS	Microzooplankton biomass	mgC m^{-3}		
ZL	Mesozooplankton biomass	mgC m^{-3}		
G_{ZS}	Microzooplankton grazing rate on Small phytoplankton	$\text{mgC m}^{-3}\text{d}^{-1}$		
$G_{ZL,PL}$	Mesozooplankton grazing rate on Microphytoplankton	$\text{mgC m}^{-3}\text{d}^{-1}$		
$G_{ZL,ZS}$	Mesozooplankton grazing rate on Microzooplankton	$\text{mgC m}^{-3}\text{d}^{-1}$		
G_{ZL}	Mesozooplankton combined grazing rate	$\text{mgC m}^{-3}\text{d}^{-1}$		
F_{eu}	Total POC flux out of the euphotic zone	$\text{mgC m}^{-2}\text{d}^{-1}$		
F_{alg}	Flux of algal aggregates out of the euphotic zone	$\text{mgC m}^{-2}\text{d}^{-1}$		
F_{fec}	Flux of fecal pellets out of the euphotic zone	$\text{mgC m}^{-2}\text{d}^{-1}$		
J_{dvm}	DVM-mediated export flux	$\text{mgC m}^{-2}\text{d}^{-1}$		
J_{met}	Respired DIC produced in twilight zone by migrating ZL	$\text{mgC m}^{-2}\text{d}^{-1}$		
J_{fec}	Fecal pellets produced in twilight zone by migrating ZL	$\text{mgC m}^{-2}\text{d}^{-1}$		
ER	Export Ratio	–		

Table 1
Continued

Prognostic variables		
Symbol	Description	Unit
<i>DER</i>	DVM Export Ratio	–
<i>DRR</i>	Respiration Ratio	–
<i>RD</i>	Weighted depth of respiration	–
f_{fec}	Fraction of fecal pellets in the euphotic zone	–
p_{met}	Fraction of metabolism in the twilight zone	–

covering approximately $2.93 \times 10^8 \text{ km}^2$ of the global ocean or just over 80% of its total surface area. This avoids estimation bias in polar regions due to seasonal ice and cloud cover. The model is run with a daily time step and spun up until quasi-equilibrium is reached. Results are taken from the last year of the model run.

The rate of change, per day, in biomass within the mixed layer (Z_{ml}) for each size class is given by

$$\frac{dPS}{dt} = \mu_{PS}PS - G_{ZS} - agg_{PS}PS^2 - m_P(PS - PS_0) - \frac{PS}{Z_{ml}} \frac{dz_{ml}}{dt} H\left(\frac{dz_{ml}}{dt}\right) \quad (5)$$

$$\frac{dPL}{dt} = \mu_{PL}PL - G_{ZL,PL} - agg_{PL}PL^2 - m_P(PL - PL_0) - \frac{PL}{Z_{ml}} \frac{dz_{ml}}{dt} H\left(\frac{dz_{ml}}{dt}\right) \quad (6)$$

$$\frac{dZS}{dt} = b_{ZS}G_{ZS} - m_{ZS}(ZS - ZS_0) - G_{ZL,ZS} \quad (7)$$

$$\frac{dZL}{dt} = b_{ZL}(G_{ZL,PL} + G_{ZL,ZS}) - m_{ZL}(ZL - ZL_0) - p_{ZL}ZL^2 \quad (8)$$

where *PS*, *PL*, *ZS*, and *ZL* represent biomass of small phytoplankton (0.5–20 μm, pico- and nanophytoplankton), microphytoplankton (20–200 μm), microzooplankton (20–200 μm) and mesozooplankton (>200 μm) respectively (Calbet & Calbet, 2008; Moriarty & O'Brien, 2013; Sieburth et al., 1978). The model does not resolve vertical or horizontal movement. Therefore, biomass represents the mean concentration within the mixed layer, with the assumption of even distribution. Natural mortality (m_i) terms have a lower threshold applied of 0.2 for phytoplankton (Aumont et al., 2015) and 1.0 for zooplankton (Archibald et al., 2019) for model stability. Algal aggregates are represented as quadratic mortality terms (agg_i) of plankton biomass (Aumont et al., 2015). This enables changes in biomass to be reflected in algal export. The influence of shear on aggregate formation (Aumont et al., 2015) is not represented due to the lack of vertical movement and other physical dynamics within the model. The last term in Equations 5 and 6 describes the dilution of biomass as the depth of the mixed layer increases (Archibald et al., 2019; Siegel et al., 2014). $H = 1$ if the change in mixed layer depth is less than or equal to zero, or $H = 0$ otherwise (Archibald et al., 2019; Evans & Parslow, 1985; Siegel et al., 2014). The sub-model is closed by a quadratic mortality term (p_{ZL}) for mesozooplankton, which represents grazing by higher trophic levels.

Zooplankton growth is the product of gross growth efficiency (b_i) and grazing (T. R. Anderson et al., 2015). Grazing rates (G_i) are based on Holling Type III (Holling, 1959) functional responses, where

$$G_{ZS} = \frac{g_{ZS} PS^2}{k_0^2 + PS^2} ZS \quad (9)$$

$$G_{ZL,PL} = \frac{g_{ZL} PL^2}{k_1^2 + PL^2} ZL \quad (10)$$

$$G_{ZL,ZS} = \frac{g_{ZL} ZS^2}{k_1^2 + ZS^2} ZL \quad (11)$$

k_0 and k_1 are half-saturation constants and g_i are maximum grazing rates. There is no prey preference (Aumont et al., 2015) for mesozooplankton grazing and no multiple prey feeding response (T. R. Anderson et al., 2010, 2015) as this fundamentally changes the relationship between k and the prey distribution (T. R. Anderson et al., 2015; Gentleman et al., 2003; Rohr et al., 2022). The grazing terms allow for two-way coupling between zooplankton and their prey, so that grazing rates are influenced by both predator and prey biomass. A Type III functional response is chosen due to increased stability, its suitability to coarse resolution global models and improved reproduction of seasonal population dynamics compared to Type II (Rohr et al., 2022, 2024).

2.4. Optimization of Grazing Dynamics

This study aims to assess the impact of locally tuned grazing dynamics on outputs from a coupled ecosystem-carbon export model. To do this three scenarios are considered (Section 2.1). The same ecosystem-carbon export model and bottom-up forcing is used for all three scenarios and all parameters except k are kept constant. For the Baseline and Global- k scenarios, the same pair of k values is used for every grid cell location (i.e., they are globally homogeneous). For the Baseline scenario, the median k values from 40 models reviewed by Rohr et al. (2022) are used. These are 40 and 80 mgC m⁻³ for microzooplankton and mesozooplankton, respectively. For the Global- k scenario, a single pair of globally optimized values are used. For the Local- k scenario, k values are locally tuned, at every grid cell location. The optimization process for both the globally optimized pair, and locally tuned values is detailed below.

For each model grid point, the optimum half-saturation constant for both microzooplankton (k_0) and mesozooplankton (k_1) grazing is assessed using an inverse modeling approach. Multiple simulations of the model are run, each with a different set of k_0 and k_1 values. The output of the model is then compared to the climatological seasonal cycle of phytoplankton biomass (Behrenfeld et al., 2005; Siegel et al., 2013, 2014; Westberry et al., 2008). The k values that most closely reproduce these satellite-derived biomass values are then selected as the “optimum” k values. To find the optimum values, a cost function is used. Several different cost functions were analyzed which produced consistent results (Figures S1 and S2 in Supporting Information S1). The cost function presented here (Equation 12) is the sum of the normalized absolute average error (Stow et al., 2009) for both small phytoplankton and microphytoplankton, computed across the full seasonal cycle. This represents the degree of agreement in the size and alignment of the seasonal cycle between model and observations. Here, the term “observations” refers to the satellite-derived phytoplankton biomass. A value of zero indicates a perfect match and alignment with observations.

$$\text{Cost} = nAAE_{PS} + nAAE_{PL} \quad (12)$$

where,

$$nAAE_{P(i)} = \frac{AAE_{P(i)}}{\sigma_o} \quad (13)$$

$$AAE_{P(i)} = \sum \frac{|P(i)_{obs} - P(i)_{mod}|}{n} \quad (14)$$

σ_o is the standard deviation of the observed data, which represents observed temporal variance in the seasonal cycle; n is the total number of observations across the climatological year; i is the size class and $P(i)_{obs}$ and $P(i)_{mod}$ are observation and model values of phytoplankton respectively. The absolute average error is normalized by the standard deviation of the observed climatology, to enable comparisons of relative errors in high and low productivity regions. Due to the uncertainty in zooplankton observational estimates (Strömberg et al., 2009) a zooplankton term was not included in the cost function.

To maximize computational efficiency, two routines of k optimization are carried out. The first coarse resolution routine uses 15 log spaced values of k_0 and k_1 (mgC m⁻³): 16, 20, 26, 33, 43, 54, 70, 89, 114, 146, 187, 239, 306, 392 and 501. These are within the range of empirical and model estimates presented in Rohr et al. (2022). The model is run for each possible pair of half-saturation constants, at every grid cell (i.e., a total of 225 (15 × 15) runs at each grid cell). The pair of half-saturation constants that produces the lowest cost are then selected as the optimal values at that location. The result is a distribution of optimal k values across the global ocean.

To improve the resolution of our optimization, a second optimization routine is then carried out. At each location, the sampling input of k values is calculated by first taking the optimum k values from the coarse optimization run. Next, an upper and lower limit for input values is calculated by $\pm 10\%$ in each direction from the optimized values. Numbers are rounded to the nearest integer. The model is then run for every pairing of integers between these two limits at that location. The cost function is then reevaluated at each grid cell location and a new optimized pairing of k values selected. For example, if the coarse optimization identifies $k_0 = 20$ and $k_1 = 26$ as optimal, we then rerun the simulation for all integer values (and pairings) of $18 \leq k_0 < = 22$ and $23 \leq k_1 < = 29$.

For the Global- k scenario, the globally homogenous pair of k values are estimated by globally integrating the cost function for each k pairing. The pair that produce the lowest cost value from this global integration are selected for the Global- k scenario. The results from the coarse resolution optimization routine are used for this integration. For the Local- k scenario, the optimum pair of k values estimated for every grid cell (from the second optimization routine) are used.

2.5. Carbon Export Sub-Model

The Archibald et al. (2019) carbon export model is used to examine the impact of grazing parametrization in this study. The carbon export model consists of two modules: the euphotic and twilight zone modules. In this study, the twilight zone component has remained unchanged from its description in Archibald et al. (2019). However, a few changes have been made to the euphotic zone module to reflect the use of fully coupled grazing terms and the use of μ as an input instead of NPP. These changes are described below.

Carbon export out of the euphotic zone (Total Export Flux) is the sum of the passive sinking flux (F_{eu}) and DVM-mediated flux (J_{dvm}).

$$\text{Total Export Flux} = F_{eu} + J_{dvm} \quad (15)$$

$$F_{eu} = F_{alg} + F_{fec} \quad (16)$$

F_{eu} is the sum of microphytoplankton algal aggregates (F_{alg}) and fecal pellets produced by mesozooplankton grazing (F_{fec}). Sinking algal aggregates from the euphotic layer (F_{alg}) are estimated as

$$F_{alg} = Z_{eu} (1 - bact) (agg_{ZL} PL^2 + 0.5 m_P PL) \quad (17)$$

where the microphytoplankton aggregation term and 50 % of microphytoplankton linear mortality contributes to sinking aggregates (Aumont et al., 2015). As in Aumont et al. (2015), the remaining 50 % of the linear mortality term is classed as small POC, which is retained in the euphotic zone. $bact$ is the proportion lost to bacterial remineralization (Aumont et al., 2015). Small phytoplankton do not contribute to algal export in the model, as its smaller cell size means aggregates are assumed to contribute to the microbial web in the euphotic zone, rather than sinking export flux (Archibald et al., 2019; Calbet & Landry, 2004).

Euphotic zone sinking fecal pellets (F_{fec}) and fecal pellets produced in the twilight (J_{fec}) are estimated as

$$F_{fec} = (p_{dvm} f_{fec} + (1 - p_{dvm})) (m_{fec} G_{ZL}) Z_{eu} \quad (18)$$

$$J_{fec} = p_{dvm} (1 - f_{fec}) (m_{fec} G_{ZL}) Z_{eu} \quad (19)$$

where G_{ZL} is the combined grazing rate for mesozooplankton on both prey types ($G_{ZL,PL} + G_{ZL,ZS}$), p_{dvm} is the proportion of mesozooplankton that participate in DVM, m_{fec} is the fraction of grazed carbon expelled as fecal pellets and f_{fec} is the proportion of fecal pellets expelled in the euphotic zone. DVM is treated as a single event and POC is a single pool of carbon that decays exponentially (Archibald et al., 2019). All carbon export parameter values (Table 1) were kept consistent with those detailed in Archibald et al. (2019). In this study, microzooplankton do not vertically migrate and their fecal pellets do not contribute to export flux as their smaller pellet size means they are assumed to be consumed by the microbial loop in the euphotic zone (Archibald et al., 2019; Calbet & Landry, 2004).

Table 2
Locally Tuned Microzooplankton (k_0) and Mesozooplankton (k_1)
Half-Saturation Constants Estimated Using the Cost Function

Local- k	k_0	k_1
Median	18	27
Geometric mean	27	38
Biomass-weighted mean	31	49
Range	14–551	14–551
IQR	17(14–31)	38(14–52)
Global- k	33	392
Baseline	40	80

Note. Half-saturation constant values are in mgC m^{-3} . Global- k and Baseline scenario k values are included below for comparison (NB: average statistics cannot be provided for these two scenarios due to the same value being used for every grid cell in the model domain).

Finally, the production of respired dissolved inorganic carbon (DIC) in the twilight zone (J_{met}) is estimated as

$$J_{met} = p_{met} p_{dvm} f_{met} (G_{ZL} - m_{fec} G_{ZL}) Z_{eu} \quad (20)$$

where p_{met} is the fraction of total metabolism that occurs in the twilight zone and f_{met} is the fraction of absorbed carbon that is metabolized. The contribution of both J_{met} and J_{fec} to DVM-mediated flux (J_{dvm}) is then described by

$$J_{dvm} = J_{met} + J_{fec} \quad (21)$$

2.6. Analysis

To analyze the output data from the optimization of grazing dynamics and the ecosystem-carbon export model, several additional metrics are considered.

2.6.1. Grazing Dynamics

The biomass-weighted k value (BW- k) combines both k_0 and k_1 values. It considers the optimal k value for each zooplankton class and their relative abundance at every grid cell location (BW- $k = (k_0 ZS + k_1 ZL)/Z$). This reflects the emergent grazing dynamics of the entire zooplankton community. The half-saturation constant and maximum grazing rate can be related to the prey capture efficiency, ϵ . The prey capture efficiency is calculated by dividing the maximum grazing rate (See Table 2) by the half-saturation constant for each grid cell (Rohr et al., 2022) ($\epsilon = g/k^2$). To understand the relationship between k and both NPP and prey biomass, a linear regression is fitted to log-normalized data.

2.6.2. Carbon Export

The export ratio represents the proportion of NPP exported as carbon from the euphotic zone. DVM export ratio (DER) is DVM-mediated export as a fraction of total carbon exported from the euphotic zone. The DVM respiration ratio (DRR) is the amount of respiration carried out by migrating zooplankton as a fraction of the integrated respiration from the twilight zone (Z_{eu} —1,000 m). The weighted depth of respiration (RD) is the increase in depth of DIC production and oxygen utilization, as a result of zooplankton vertical migrations. Carbon export metrics are described in further detail in Archibald et al. (2019).

3. Results

3.1. Distribution of Locally Tuned Grazing Dynamics

Local-tuning of k values results in high variability in inferred grazing dynamics (Figure 1 and Figures S1–S4 in Supporting Information S1). k values span a range of 537 mgC m^{-3} (Table 2). High k values are generally associated with highly productive regions (Figure 1c). Lower values are generally associated with the less productive subtropical oligotrophic gyres, with the exception of the eastern South Pacific, where maximum k values (551 mgC m^{-3}) are estimated for microzooplankton. Maximum k values are also found in the high latitudes of the southern hemisphere.

Zooplankton functional groups are characterized by different grazing dynamics. Microzooplankton k values estimated from local optimization are, on average, lower than mesozooplankton (median k values are 18 and 27 mgC m^{-3} respectively), suggesting faster grazing for the smaller size class (Table 2 and Table S1 in Supporting Information S1). In addition, the globally optimized pair of k values (estimated for the Global- k scenario - see Section 2.4) are 33 mgC m^{-3} and 392 mgC m^{-3} for micro- and mesozooplankton respectively. Microzooplankton are generally characterized by more efficient grazing than mesozooplankton, with the exception of the oligotrophic gyres (Figure S5 in Supporting Information S1).

The distribution of grazing dynamics also differs between zooplankton size classes. This is particularly evident in equatorial upwelling regions, where microzooplankton and mesozooplankton communities are characterized by low and high k values respectively (Figures 1a and 1b). These differences result in divergent relationships between k values, NPP and prey biomass for the two size classes. High k_0 values are associated with low NPP and

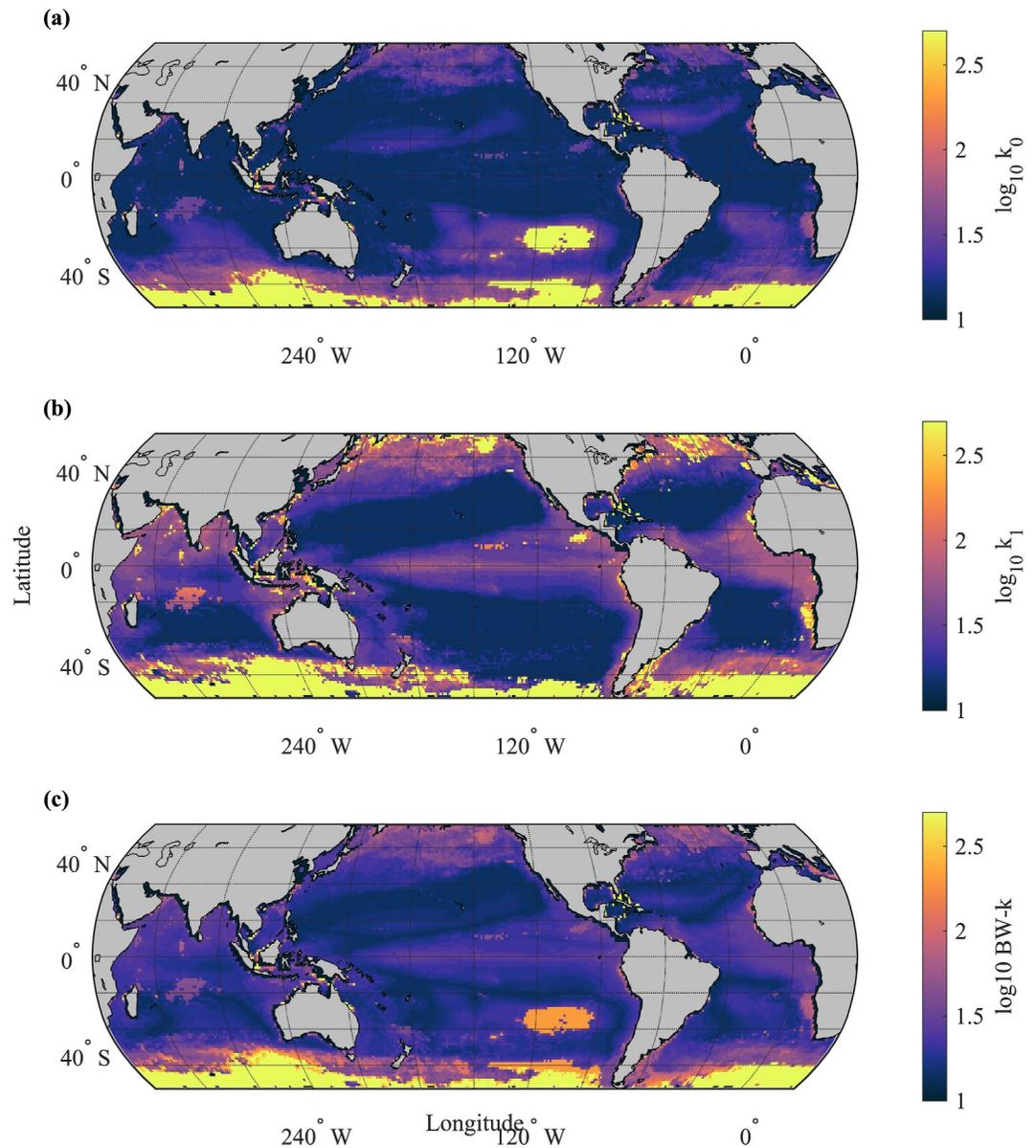


Figure 1. Locally tuned k values. (a) Microzooplankton half-saturation constants (k_0) estimated using the cost function (Equation 12). (b) Mesozooplankton half-saturation constants (k_1) estimated using the cost function. k values are in mgC m^{-3} . (c) Biomass-weighted k values (BW- k) which considers the optimal k value for each zooplankton size class and their relative abundance. BW- k reflects the overall grazing dynamics of the entire zooplankton community. The maximum and minimum values on the color bar represent the maximum/minimum k values sampled in the optimization.

prey biomass, whilst high k_1 values are associated with high NPP and prey biomass (Figure S6 in Supporting Information S1).

3.2. Impact of Locally Tuned Grazing Dynamics on Model Skill

Locally tuned k values improve model skill in comparison to globally homogenous k values (Figure 2). Mean cost (Equation 12) is reduced by 43% in the Local- k scenario, compared to the Global- k scenario. Therefore, the use of locally tuned k values improves the model's ability to reproduce satellite-derived phytoplankton biomass (Figure 3). In comparison, the use of optimized globally homogenous k values (Global- k) has a limited ability (−14%) in reducing model cost from the Baseline scenario. Reduction in cost values due to local-tuning is most evident in the tropics and subtropics, particularly productive upwelling regions (Figure 2). Despite improvements,

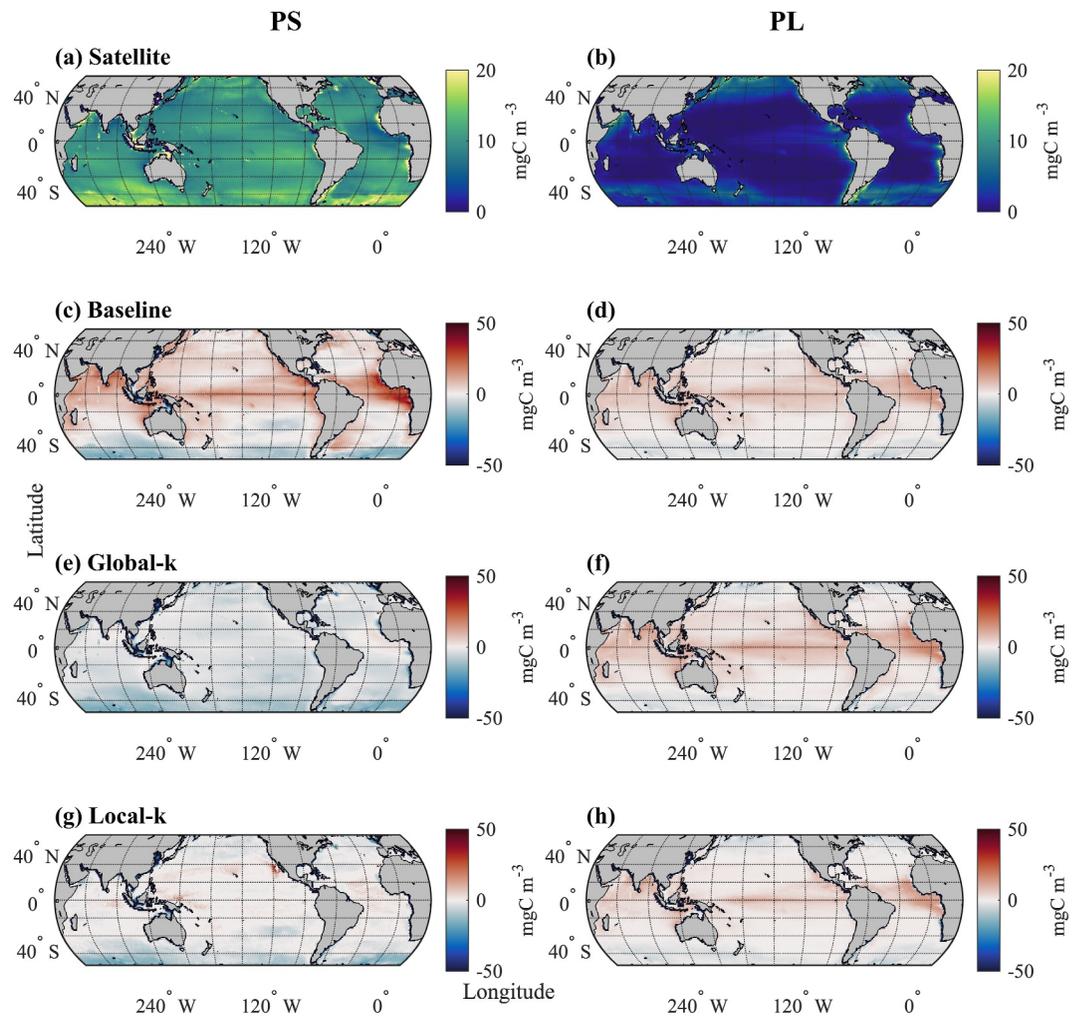


Figure 2. Cost values representing the difference between observations of phytoplankton biomass and the model phytoplankton biomass (Equation 12). (a) Cost values from the Baseline model scenario (non-optimized k values). (b) Percentage change in cost values in the Global- k scenario compared to the Baseline scenario. (c) Percentage change in cost values in the Local- k scenario compared to the Baseline.

microphytoplankton biomass estimates are the greatest source of error (78%) in the Local- k cost function (Figures S7 and S8; Table S3 in Supporting Information S1).

3.3. Ecosystem Impact

The reproduction of the remotely sensed small phytoplankton biomass distribution is greatly improved with the implementation of locally tuned k values (Figures 3 and 4a–4d). The use of optimized globally homogenous k values (Global- k) improves the reproduction of observed small phytoplankton but shows little of the regional variability found in observations (Figures 4a and 4c). This regional variability is only reproduced when locally tuned grazing dynamics are implemented (Figure 4d) highlighting the importance of top-down drivers for this size class. Without optimized grazing dynamics, a large overestimation of small phytoplankton biomass in equatorial upwelling occurs (Figures 3a and 4b). The Local- k model does a good job in estimating global small phytoplankton biomass with an annual mean (\pm S.D.) of $9.40 \pm 3 \text{ mgC m}^{-3}$, in comparison to $11.89 \pm 6 \text{ mgC m}^{-3}$ from satellite-derived estimates (Siegel et al., 2014) (Figure 4a). NPP is 52 and 43 Gt C yr^{-1} for the Local- k and Global- k runs respectively, compared to 87 Gt C yr^{-1} for the non-optimized baseline run (Figure S9 in Supporting Information S1).

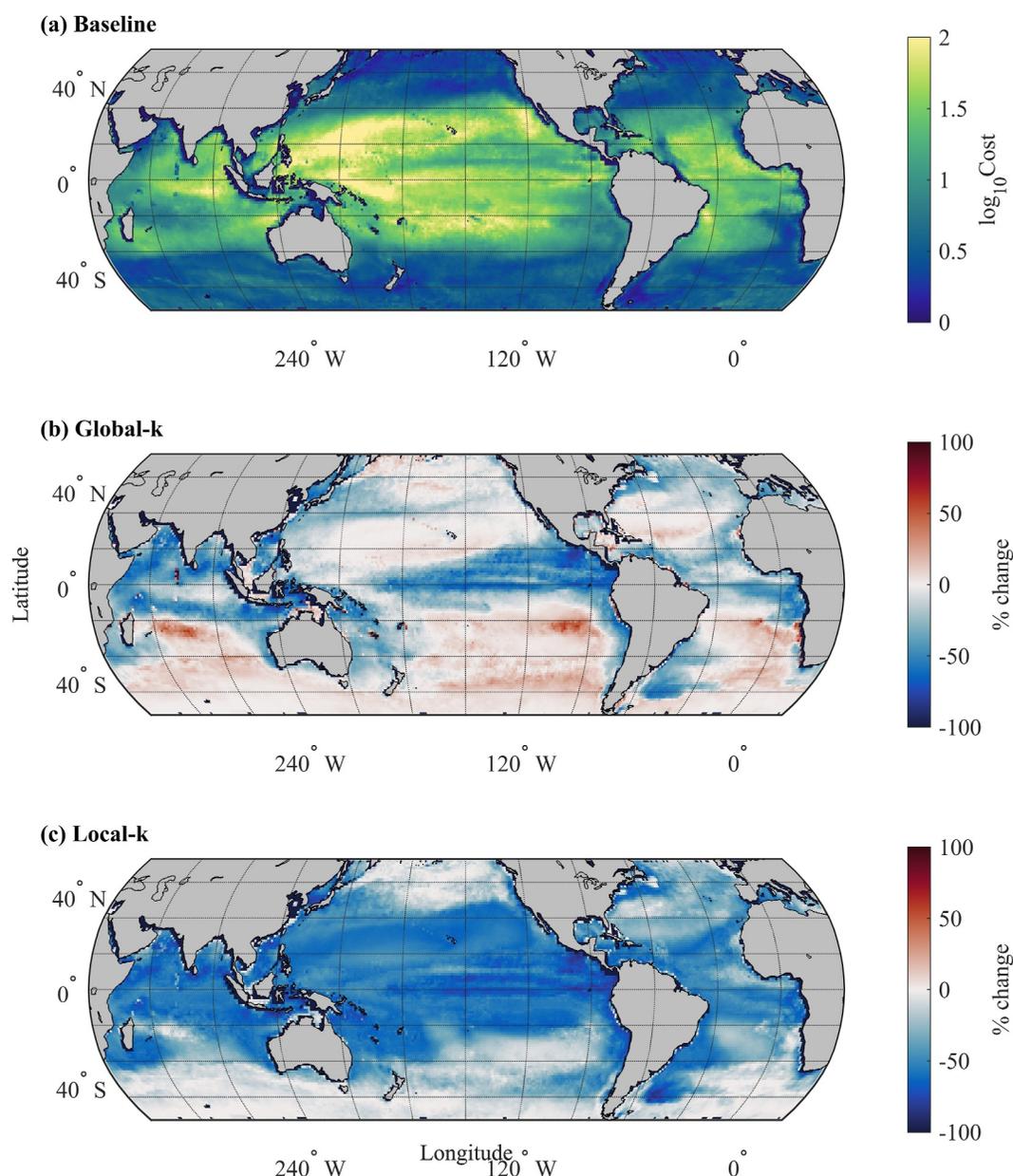


Figure 3. Absolute bias in modeled small phytoplankton (PS) and microphytoplankton (PL) biomass in comparison to satellite derived biomass ($P(i)_{mod} - P(i)_{obs}$). Three model scenarios are shown: Baseline (non-optimized k values), Global- k (globally optimized k value) and Local- k (locally tuned k values). Observational phytoplankton biomass values were calculated as per Siegel et al. (2014), using particulate backscattering coefficient data (Behrenfeld et al., 2005; Kostadinov et al., 2010; Siegel et al., 2013; Westberry et al., 2008).

Microphytoplankton observational distributions are reproduced well in all three model versions showing little difference when locally tuned grazing dynamics are applied (Figures 4e–4h), suggesting the distribution for this functional group is determined primarily by realistic bottom-up drivers. However, all three models overestimate biomass in equatorial upwelling areas (Figure 3). In the Local- k model, microphytoplankton has a global mean (\pm S.D.) of $3.05 \pm 3 \text{ mgC m}^{-3}$, compared to $2.62 \pm 5 \text{ mgC m}^{-3}$, from satellite-derived estimates (Siegel et al., 2014) (Figure 4e). In the subtropical oligotrophic gyres, microphytoplankton biomass estimates appear to closely emulate observations, despite higher cost values in this region. This is due to the low biomass in the region, which results in small changes producing large error values with normalization.

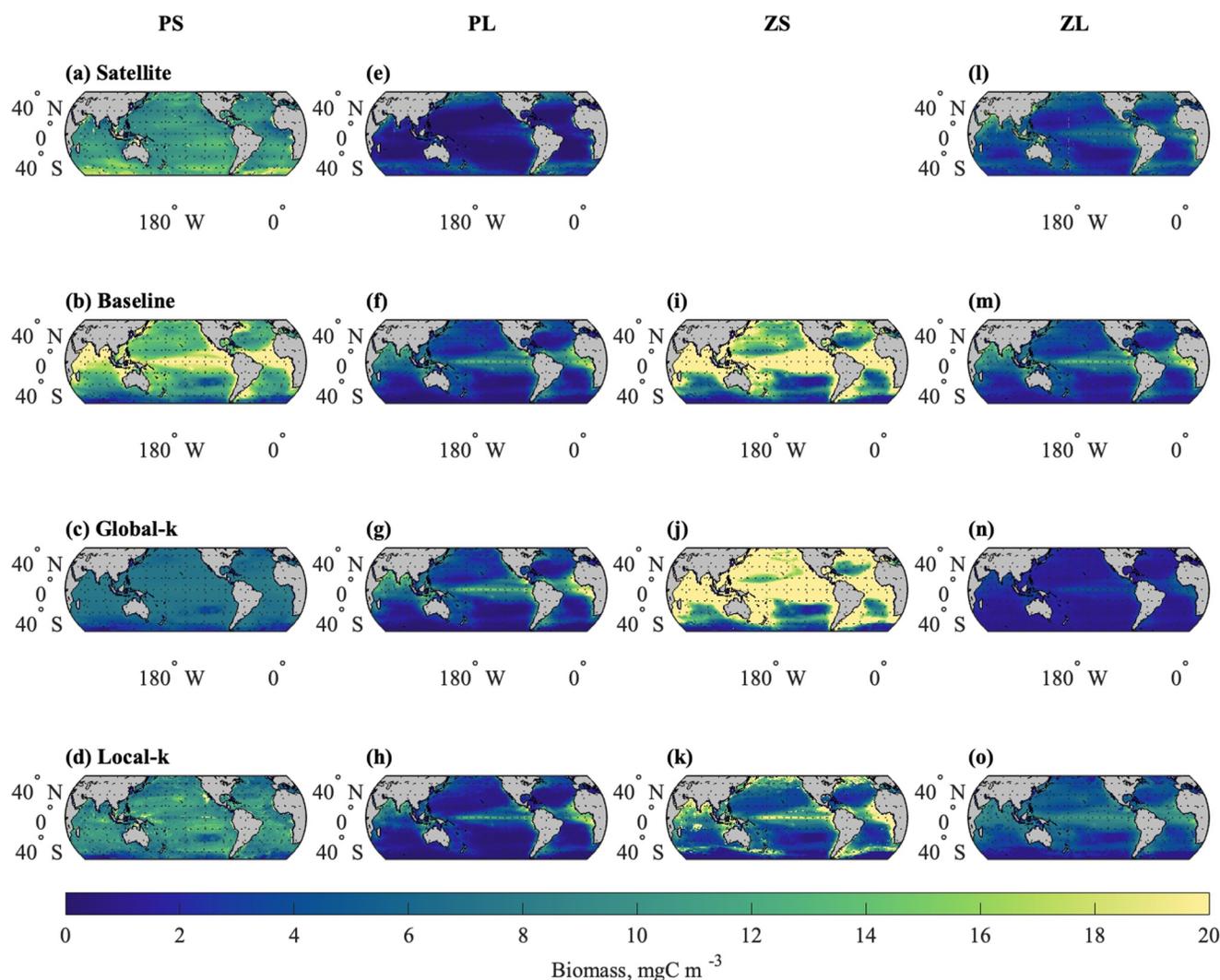


Figure 4. Small phytoplankton (PS), microphytoplankton (PL), microzooplankton (ZS) and mesozooplankton (ZL) biomass distributions estimated by the three model scenarios: Baseline, Global-*k* and Local-*k*. Satellite-derived biomass distributions are included for reference (not available for ZS). a & e are calculated as per Siegel et al. (2014), using particulate backscattering coefficient data (Behrenfeld et al., 2005; Kostadinov et al., 2010; Siegel et al., 2013; Westberry et al., 2008). Satellite-derived mesozooplankton biomass (l) is from Strömberg et al. (2009).

Local-tuning of *k* values improves zooplankton biomass estimates in comparison to observations (Figures 4i–4l and Figure S10 in Supporting Information S1). When only optimized homogenous global values are used, global mesozooplankton biomass is underestimated, with a mean (\pm S.D.) of $1.72 \pm 0.81 \text{ mgC m}^{-3}$ compared to $5.52 \pm 9 \text{ mgC m}^{-3}$ from Strömberg et al. (2009) (Figure 4i). This is as a result of the very high *k* value for mesozooplankton (392 mgC m^{-3}), resulting in very low grazing and therefore biomass. In contrast, mean (\pm S.D.) mesozooplankton biomass from the Local-*k* model is $5.07 \pm 3 \text{ mgC m}^{-3}$. All model versions underestimate mesozooplankton biomass in the higher latitudes of the northern hemisphere. Mean microzooplankton biomass estimates are greatly reduced with the implementation of locally tuned grazing dynamics, from $29.09 \pm 23 \text{ mgC m}^{-3}$ in the Global-*k* run to $8.85 \pm 11 \text{ mgC m}^{-3}$ in the Local-*k* run (Table 3).

3.4. Impact on Carbon Export

Local-tuning of *k* values decreases mean total carbon export by >17% (Table 3). The magnitude of change depends on the homogenous *k* values used for comparison (–35.64% in comparison to the Baseline and –17.07% in

Table 3
Comparison of Mean Global Carbon Export Estimates From the Three Model Scenarios

	Baseline		Global- <i>k</i>		Local- <i>k</i>	
	Mean	S.D.	Mean	S.D.	Mean	S.D.
Export flux (mgC m ⁻² d ⁻¹)	93.48	127.46	69.44	107.14	61.74	88.32
NPP (mgC m ⁻² d ⁻¹)	752.18	715.79	379.36	283.08	459.04	291.66
Export ratio	0.09	0.05	0.12	0.10	0.10	0.08
PS (mgC m ⁻³)	13.95	6.39	6.65	1.49	9.40	2.86
PL (mgC m ⁻³)	4.04	3.52	4.38	4.03	3.05	3.32
ZS (mgC m ⁻³)	16.69	11.59	29.09	29.91	8.85	11.21
ZL (mgC m ⁻³)	4.66	3.62	1.72	0.81	5.07	2.70
<i>G</i> _{ZS} (mgC m ⁻³ d ⁻¹)	6.33	8.08	3.00	2.63	3.73	3.38
<i>G</i> _{ZL} (mgC m ⁻³ d ⁻¹)	4.66	3.62	1.72	0.81	5.07	2.70
<i>G</i> _{ZL,PL} (mgC m ⁻³ d ⁻¹)	0.11	0.25	1.7 × 10 ⁻³	4.1 × 10 ⁻³	0.21	0.31
<i>G</i> _{ZL,ZS} (mgC m ⁻³ d ⁻¹)	1.12	1.91	0.06	0.12	0.88	0.84
<i>F</i> _{alg} (mgC m ⁻² d ⁻¹)	68.54	88.19	68.28	104.98	36.91	72.32
<i>F</i> _{fec} (mgC m ⁻² d ⁻¹)	20.70	33.14	0.97	1.84	20.52	18.76
<i>J</i> _{dvm} (mgC m ⁻² d ⁻¹)	4.24	6.53	0.20	0.36	4.31	3.65
DER	0.03	0.02	3.2 × 10 ⁻³	2.3 × 10 ⁻³	0.10	0.06
DRR	0.03	0.02	2.7 × 10 ⁻³	2.0 × 10 ⁻³	0.11	0.06

Note. NPP = Net Primary Productivity. *F*_{alg} = Euphotic export flux of algal aggregates. *F*_{fec} = Euphotic export flux of fecal pellets. *J*_{dvm} = DVM-mediated export flux. *G*_{ZL} = Mesozooplankton grazing rate on all prey types. *G*_{ZS} = Microzooplankton grazing rate on all prey types. *G*_{ZL,PL} = Mesozooplankton grazing on microphytoplankton. *G*_{ZL,ZS} = Mesozooplankton grazing on microzooplankton. PS = Small phytoplankton biomass. PL = Microphytoplankton biomass. ZS = Microzooplankton biomass. ZL = Mesozooplankton biomass. DER = DVM export ratio. DRR = DVM respiration ratio.

comparison to the Global-*k* scenario). Export values are generally high, with a total export flux of 7.19 Pg C yr⁻¹ for the Local-*k* scenario and 8.16 Pg C yr⁻¹ and 10.94 Pg C yr⁻¹ for the Global-*k* and Baseline scenarios respectively.

The routing of carbon is impacted by the implementation of locally tuned grazing dynamics (Figure 5). In the Local-*k* scenario, more carbon is exported as fecal pellets and less as algal aggregates, compared to the Global-*k* scenario (Table 3). When *k* values are locally tuned, carbon exported from pellets and aggregates are more similar in magnitude (annual mean of 36.91 and 20.52 mgC m⁻²d⁻¹ respectively) compared to model runs with homogenous *k* values. In contrast, in both the Baseline and Global-*k* model runs carbon export is dominated by algal aggregates. Local-tuning of *k* values also results in increased export (>21%) via vertically migrating zooplankton compared to the Global-*k* model run (Figure 5, Table 3).

The differences in Global-*k* and Local-*k* estimates from the Baseline are, in certain regions, on the scale of the values seen in the Baseline scenario. This is evident in the large standard deviations shown in Table 3. In the regions with the top 5% largest deviation in normalized export fluxes, between the Local-*k* and Baseline estimates, changes in annual average export exceed 76%. Productive upwelling regions, that are characterized by high carbon export rates, are the regions of greatest change from the local-tuning of *k* values (Figure 5). Changes in export estimates mirror changes in zooplankton grazing rates, for example, changes in fecal export from the euphotic zone (*F*_{fec}) and DVM-mediated export flux (*J*_{dvm}) are inversely proportional to those of mesozooplankton grazing.

Mesozooplankton grazing on microzooplankton is a greater contributor to carbon flux than grazing on microphytoplankton. In the Local-*k* model, 80% of mesozooplankton grazing constitutes grazing on microzooplankton (*G*_{ZL,ZS}), however this is decreased from 91% (Baseline-*k*) and 99% (Global-*k*) with homogenous *k* values.

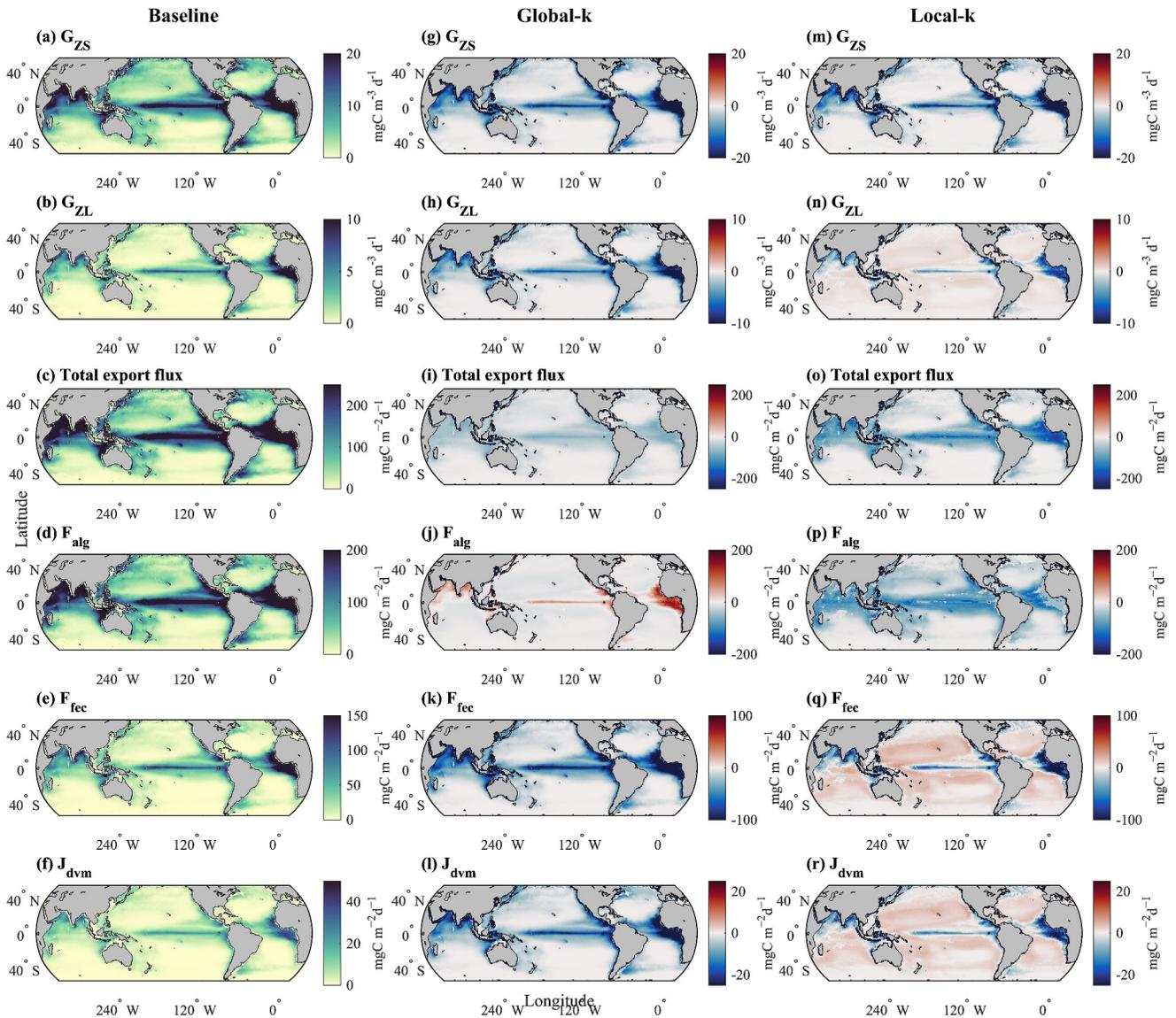


Figure 5. Changes in carbon export due to grazing parameterization. Three model runs are presented: Baseline, Global- k and Local- k . The outputs from the Baseline run are presented in plots (a–f). Plots (g–l) show the absolute change when changing the model input from the baseline run (non-optimized k values) to the Global- k run (globally optimized k values). Plots (m–r) show the absolute change when changing the model input from the baseline run to the Local- k run (locally tuned k values). F_{alg} = Euphotic export flux of algal aggregates. F_{fec} = Euphotic export flux of fecal pellets. J_{dvm} = DVM-mediated export flux. G_{ZL} = Mesozooplankton grazing rate on all prey types. G_{ZS} = Microzooplankton grazing rate on all prey types.

4. Discussion

One of the largest sources of uncertainty in the marine carbon cycle is zooplankton grazing (Rohr et al., 2023). In this study we used an inverse modeling approach to estimate spatial variation in zooplankton grazing dynamics and explored the subsequent impact of these dynamics on modeling marine ecosystems and carbon export. The focus of this study was the grazing parameter k , which is frequently used in global biogeochemical (BGC) models. We found that local-tuning of k results in high variability of inferred grazing dynamics. The local-tuning of k values improved the model's ability to reproduce satellite-derived phytoplankton biomass. Consequently, estimates of mean total carbon export decreased by > 17% compared to global tuning, with a greater proportion of export as fecal pellets and less as algal aggregates.

4.1. High Variability of Inferred Grazing Dynamics

Local optimization suggested high k values were generally associated with eutrophic ocean regions. This is consistent with a study in review (Rohr et al., 2024) which used an inverse modeling approach to infer high community k values for a single zooplankton group (combining k_0 and k_1) in equatorial upwelling regions and higher latitudes. In equatorial upwelling regions, communities are dominated by suspension-feeding copepods (Steinberg & Landry, 2016), whose slower grazing rates enable diatom blooms to form (Rohr et al., 2024). In the high latitudes, higher k values reduce grazing pressure on phytoplankton stocks. High latitude regions are characterized by fast growing phytoplankton species such as diatoms (Murphy et al., 2021), however, their growth is restricted by environmental conditions. Indeed, a review by Schmoker et al. (2013) found low phytoplankton growth rates in polar regions, compared to temperate locations. Higher k values in high latitude regions would prevent prey stocks from becoming fully depleted in the winter and allow spring/summer blooms to occur.

Generally low k values were estimated in the subtropical oligotrophic gyres, where communities are dominated by faster grazing microzooplankton, in particular pico- and nano-sized flagellates (Calbet & Calbet, 2008). Anomalous high k_0 values inferred in the hyper-oligotrophic South Pacific gyre (Ras et al., 2007) are in disagreement to the study by Rohr et al. (2024) and coincide with an underestimation of NPP, small phytoplankton biomass and near-zero growth rates. This suggests that in this region, small phytoplankton growth rates used to force the model may be too low. Growth rates are derived from observed NPP, which is divided by the greater of euphotic and mixed layer depth. The South Pacific gyre is characterized by a deep euphotic layer, which may have produced unrealistically low growth rates and anomalous k values for this size class. In addition, the lack of explicit representation of temperature within the model may affect growth and grazing rates in these extreme environments. The 1PIZ 3D model in Rohr et al. (2024) includes temperature which may negate these issues.

In this study, mesozooplankton showed on average, higher k values, and therefore slower prey capture times. This is consistent with ecological understanding (Barton et al., 2013). As the maximum grazing rate, g , was held constant across the model domain, variation in the k value represents variation in the prey capture rate, rather than consumption (Rohr et al., 2022). For both size classes, average half-saturation constants were in the lower quartile of empirical values (Hansen et al., 1997; Hirst & Bunker, 2003; Rohr et al., 2022). However, empirical estimates are from laboratory measurements of samples collected from a very narrow range of locations, with the majority from coastal regions in the northern hemisphere (e.g., fjords in Norway, coastal USA and UK, Japan) and none representing the open ocean (Hansen et al., 1997; Hirst & Bunker, 2003). These are also of individual species and are unlikely to be representative of the community mean values estimated here for each 1° grid cell (Rohr et al., 2022).

Locally tuned k values produced global distributions of grazing rates (G_i) that are consistent with other studies (Archibald et al., 2019; Siegel et al., 2014). Archibald et al. (2019) used satellite-derived estimates of NPP and phytoplankton biomass to predict global grazing rates. Mean G_{ZS} was $4.17 \text{ mgC m}^{-3} \text{ d}^{-1}$ and mean $G_{ZL,PL}$ was $0.98 \text{ mgC m}^{-3} \text{ d}^{-1}$ in Archibald et al. (2019). These represent grazing mortalities from satellite observations on phytoplankton, so include grazing losses by other groups not considered here (e.g., mesozooplankton grazing on small phytoplankton). However, the Local- k model improves the reproduction of these observationally derived grazing rates in comparison to globally homogenous k values (Table 3), with the potential to help address the uncertainty in global zooplankton grazing dynamics. In the Local- k scenario, on average 46% of NPP is lost to microzooplankton grazing on small phytoplankton and 2% of NPP is lost to mesozooplankton grazing on microphytoplankton. This supports findings by Bisson et al. (2020), where the average fraction of NPP consumed by mesozooplankton (<10%) was much lower than microzooplankton (>45%) across several model scenarios.

In this study, mesozooplankton were not assigned prey preferences, implicitly meaning both prey options were equally desirable. However, the microzooplankton biomass pool was larger than microphytoplankton. Consequently, the majority of the mesozooplankton grazing rate consisted of microzooplankton prey rather than phytoplankton. This is consistent with several field studies which have found a mesozooplankton preference for microzooplankton due to greater nutritional benefits (e.g., Campbell et al., 2009; Stoecker & Capuzzo, 1990; Vargas & González, 2004). However, this is not true for all species as filter feeders, for example, have been shown to have a preference for small-size phytoplankton, or no preference at all between phytoplankton and microzooplankton (Chen et al., 2017). These divergent strategies have potentially led to the large range of feeding

preferences presented in biogeochemical models (Rohr et al., 2022). Implementing different prey preferences may improve modeled grazing estimates and warrants further consideration.

4.2. Implications of Improved Model Skill for Future Studies

With potential to improve model skill, the local optimization of grazing dynamics could be advisable in future BGC modeling studies. This study shows that the competition generated by two zooplankton functional types isn't sufficient to emulate the global variability in grazing suggested by the local-tuning of k values. Other models currently emulate this dynamical variability using different methods, such as increased numbers of plankton functional groups (PFTs) (e.g., Dutkiewicz et al., 2021), prey switching (e.g., T. R. Anderson et al., 2010), or prey preferences (e.g., Aumont et al., 2015). In this study, four PFTs were used, in line with several modeling studies (e.g., Siegel et al., 2014), however, this groups together species with different functional traits, with different geographic distributions (Barton et al., 2013). Gelatinous salps, for example, graze preferentially on small phytoplankton, which leads to their prevalence in subtropical oligotrophic gyres (Barton et al., 2013). By explicitly representing more PFTs and their prey preferences, some of the impact of locally tuning k values may be reduced. A study by Le Quéré et al. (2016) showed that the explicit representation of krill in the Southern Ocean improved model skill in reproducing zooplankton dynamics. However, increasing the number of PFTs is computationally costly and unlikely to encompass the full extent of physiological and behavioral diversity found within the plankton community.

To enable the application of the grazing parameter k into larger BGC models, the relationship between k and environmental conditions needs to be further explored. Different species, ages, and sizes of zooplankton, even amongst the same functional group, graze with largely different k values (Rohr et al., 2022). Thus, the strong spatial variability in optimum k values implies strong variability in the composition of zooplankton communities within each functional group. Given the observational evidence of seasonal species succession (e.g., Colebrook, 1985), it is likely that a seasonally variant zooplankton community could require a seasonally variant k value, as well. This transition could potentially be incorporated into models without explicit competition, if k values could be shown to vary with environmental drivers. For instance, at the annual scale, inferred k values have been shown to vary with mean annual phytoplankton biomass (Rohr et al., 2024), but more work is required to determine the nature of this relationship with respect to other drivers (e.g., sea surface temperature) and across shorter, seasonal time scales (Text S1 in Supporting Information S1).

4.3. Reducing Uncertainty in Modeling Zooplankton

Large uncertainty exists in quantifying zooplankton biomass (Petrik et al., 2022), however the use of locally tuned k values improves the models ability to reproduce observational estimates. The observed mean (\pm S.D.) biomass of mesozooplankton is estimated as $5.9 \pm 10.6 \text{ mgC m}^{-3}$ by Moriarty and O'Brien (2013), which compares to $5.07 \pm 2.7 \text{ mgC m}^{-3}$ in the Local- k model scenario. In contrast, optimized globally homogenous k values worsen the model's ability to reproduce observational mesozooplankton biomass in comparison to the Baseline scenario (Table 3). This highlights a potential limitation of global optimization of grazing dynamics when only two zooplankton functional groups are used. This also suggests a possible reason why several BGC models underestimate mean global mesozooplankton biomass by a similar magnitude (Figure S10 in Supporting Information S1) (Aumont et al., 2015; Lovato et al., 2022). In addition, mean global microzooplankton estimates from the Global- k scenario are much greater than observations by Buitenhuis et al. (2013), where biomass is estimated to have a mean of $9.3 \pm 17.1 \text{ mgC m}^{-3}$ and a median of 3.1 mgC m^{-3} . These limitations occur despite improved model skill at reproducing the magnitude and distribution of satellite-derived phytoplankton. Local- k mean global biomass for microphytoplankton is within the range of estimates by Buitenhuis et al. (2013) (Table 3).

4.4. Implications for Predicting Carbon Export Under a Changing Climate

This study shows large variability in carbon export estimates driven by inferred grazing dynamics. There is a high degree of uncertainty in global export flux estimates, which vary between 5 and 12 PgC yr^{-1} (Siegel et al., 2022). Locally tuning grazing dynamics modifies carbon export estimates by >17% to coincide with this range (7.19 PgC yr^{-1}). The Global- k scenario estimates high carbon export, despite underestimating mesozooplankton biomass and therefore fecal export. This is due to a large proportion of algal aggregates. The substantial influence of this one model component on carbon export highlights one possible cause for uncertainty in carbon sequestration

estimates (Laufkötter et al., 2016). It is vital to reduce this uncertainty when modeling under different climate scenarios.

Algal and fecal estimates were closer to contributing equally to carbon export (as found within previous studies (Steinberg & Landry, 2016; Stock et al., 2014)) when grazing dynamics were locally optimized. The relationship between algal aggregates and fecal export highlights the balance between natural and grazing mortalities within BGC models, as the amount of biomass available for aggregates from mortality rates is impacted by grazing. The proportion of fecal versus algal export is determined by the zooplankton species present and their grazing dynamics (Steinberg & Landry, 2016). Local-tuning of grazing dynamics therefore has important consequences for climate models as the impact of climate change on plankton communities differs between species, trophic levels and geographic location (Cael et al., 2021), further influencing these two routes of carbon export.

The modeled distribution of carbon export corresponds to estimates by Stock et al. (2020), with coastal upwelling areas experiencing the greatest change from the local-tuning of k . These areas produce the highest export rates due to more efficient diatom-copepod food chains (Schmoker et al., 2013). However, the fast growth rates of diatoms makes them more susceptible to abrupt changes over the 21st Century in response to climate change (Cael et al., 2021). It is therefore vital to decrease uncertainty in grazing and export estimates in these areas. The subtropical oligotrophic gyres and the Southern Ocean are areas of lower carbon export due to the presence of smaller phytoplankton species which are lighter, sinking less carbon into the ocean interior (Calbet & Landry, 2004; Murphy et al., 2021; Schmoker et al., 2013). The highest flux estimates in the Southern Ocean occur closer to the Antarctic shelf edge, predominantly during summer months (Stock et al., 2020). However the polar extremes are out of the scope of this study due to the limitations of satellite observations in these areas (Siegel et al., 2014), which is a common issue with many plankton models (Cael et al., 2021; Dutkiewicz et al., 2021).

4.5. Limitations

There are several limitations of this modeling study. First, non- k parameters are held constant across the model domain, when in reality they are likely to vary in space and time. If these other parameters were tuned, the non-optimized (Baseline scenario) estimate of NPP may be reduced to coincide with the observed range (45–60 GtC yr⁻¹) (Le Quéré et al., 2016; Westberry et al., 2008) alongside the two optimized model scenarios. $G_{ZL,PL}$ and $G_{ZL,ZS}$ also use the same k value, which obscures whether changes in k_1 are biased to improve microphytoplankton estimates or small phytoplankton estimates (via microzooplankton) during the cost analysis. This may contribute to the overestimation of microphytoplankton in all model runs.

Second, grazing formulas are based on the Holling Type III functional response, however there is a lack of consensus within the modeling community about the most suitable functional response. T. R. Anderson et al. (2010) found that the use of different grazing formulations caused large variations in biomass, with diatoms most greatly affected. This resulted in carbon export predictions varying by as much as 25%. Modelers may also need to consider different response types for different zooplankton, under different prey conditions, especially filter feeders which have been observed to graze with a Type I response (Jeschke et al., 2004).

Third, in the model there are non-grazing losses that contribute to phytoplankton mortality, which could account for some of the variability seen between model scenarios. In the Local- k scenario, mean grazing losses for small phytoplankton (3.73 mgC m⁻³d⁻¹, Table 3) are greater than average non-grazing mortalities (0.92 mgC m⁻³d⁻¹ natural mortality and 1.14 mgC m⁻³d⁻¹ aggregate mortality). Mean microphytoplankton grazing losses (0.23 mgC m⁻³d⁻¹) are on average lower than natural and aggregation mortalities (0.28 and 0.68 mgC m⁻³d⁻¹ respectively). The detailed analysis of non-grazing losses was beyond the scope of this study. However, empirical studies have shown that grazing rates vary over three orders of magnitude (Hansen et al., 1997) whilst non-grazing losses are thought to show less natural variability (e.g., 0.13–0.96 d⁻¹ for viral-induced mortality (S. R. Anderson & Harvey, 2019)) and remain largely unconstrained (Roy et al., 2012). Thus, while non-grazing losses are explicitly simulated, they are much less likely to vary to the degree that zooplankton grazing parameters are, which are the focus of this study. Given the large non-grazing terms for microphytoplankton, further work into variable non-grazing loss parameters is warranted.

Fourthly, within the center of the oligotrophic gyres, microzooplankton were characterized by less efficient prey capture rates compared to mesozooplankton (Figure S5 in Supporting Information S1), however prey capture

efficiency should decline with size (Hansen et al., 1997; Rohr et al., 2022). This highlights a possible limitation of the model, potentially the functional groups used or model parameters. In this study, the same maximum grazing rate was used for both size classes, so prey capture efficiency is dominated by the half-saturation constant, or capture rate. In oligotrophic regions, smaller plankton dominate, so prey capture efficiencies for microzooplankton are more likely to be driven by consumption rather than capture rates, suggesting higher maximum grazing rates are needed to represent realistic capture efficiencies in these environments.

In addition, CbPMv2 was selected as the NPP forcing variable for consistency with Archibald et al. (2019), however other NPP models such as the Carbon, Absorption, and Fluorescence Euphotic-resolving (CAFE) (Silsbe et al., 2016) have been found to be more realistic, with consequences for carbon export (Bisson et al., 2018). Finally, the carbon export model does have several limitations which are discussed in detail in Archibald et al. (2019). In particular, the model is very sensitive to three parameters—*ffec*, *fmet* and *pdvm*, however these parameters remain unchanged to enable comparisons. The model also doesn't include small fecal pellets produced by microzooplankton in carbon export estimates as they are assumed to be retained in the euphotic zone. However, some studies suggest fecal export by this size class could contribute a significant portion of export flux, particularly in the subtropical oligotrophic gyres Bisson et al. (2020).

4.6. Future Considerations

The simplification of the coupled ecosystem-carbon export model means the results of this study should be considered as an example of an ecosystem model with and without spatially varying *k* values. Here, we have shown that highly spatially heterogeneous grazing dynamics are required to reproduce observed biomass when forced with observed bottom-up controls. This heterogeneity exceeds what is achievable from the explicit competition between two zooplankton functional types and has profound implications for the routing and magnitude of carbon export. Future models, particularly those concerned with ecosystem dynamics, high trophic levels and carbon export must reconcile with the possibility that even two zooplankton groups are insufficient to capture the true variability in top-down controls across the globe. More realistic representation of the global variability in zooplankton grazing dynamics may help shed light on the uncertainty in carbon export estimates under future climate scenarios.

Conflict of Interest

The authors declare no conflicts of interest relevant to this study.

Data Availability Statement

Climatologies used here are presented in detail in Siegel et al. (2014). CbPMv2 (Westberry et al., 2008) NPP, particulate backscatter (used to derive phytoplankton carbon biomass) and mixed layer depth data can be sourced from http://orca.science.oregonstate.edu/npp_products.php. World Ocean Atlas temperature and oxygen data used in the Archibald et al. (2019) carbon export model can be found at <https://coastwatch.pfeg.noaa.gov/erddap/index.html>.

References

- Anderson, S. R., & Harvey, E. L. (2019). Seasonal variability and drivers of microzooplankton grazing and phytoplankton growth in a subtropical estuary. *Frontiers in Marine Science*, 6, 174. <https://doi.org/10.3389/fmars.2019.00174>
- Anderson, T. R., Gentleman, W. C., & Sinha, B. (2010). Influence of grazing formulations on the emergent properties of a complex ecosystem model in a global ocean general circulation model. *Progress in Oceanography*, 57(1–4), 201–213. <https://doi.org/10.1016/j.pocean.2010.06.003>
- Anderson, T. R., Gentleman, W. C., & Yool, A. (2015). EMPOWER-1.0: An efficient model of planktonic ecosystems written in R. *Geoscientific Model Development*, 8(7), 2231–2262. <https://doi.org/10.5194/gmd-8-2231-2015>
- Archibald, K. M., Siegel, D. A., & Doney, S. C. (2019). Modeling the impact of zooplankton diel vertical migration on the carbon export flux of the biological pump. *Global Biogeochemical Cycles*, 33(2), 181–199. <https://doi.org/10.1029/2018GB005983>
- Aumont, O., Ethé, C., Tagliabue, A., Bopp, L., & Gehlen, M. (2015). PISCES-v2: An ocean biogeochemical model for carbon and ecosystem studies. *Geoscientific Model Development*, 8, 2465–2513. <https://doi.org/10.5194/gmd-8-2465-2015>
- Barton, A. D., Pershing, A. J., Litchman, E., Record, N. R., Edwards, K. F., Finkel, Z. V., et al. (2013). The biogeography of marine plankton traits. *Ecology Letters*, 16(4), 522–534. <https://doi.org/10.1111/ELE.12063>
- Behrenfeld, M. J., Boss, E., Siegel, D. A., Shea, D. M., Behrenfeld, M. J., Boss, E., et al. (2005). Carbon-based ocean productivity and phytoplankton physiology from space. *Global Biogeochemical Cycles*, 19, 1–14. <https://doi.org/10.1029/2004GB002299>
- Bisson, K. M., Siegel, D. A., & DeVries, T. (2020). Diagnosing mechanisms of ocean carbon export in a satellite-based food web model. *Frontiers in Marine Science*, 7, 518850. <https://doi.org/10.3389/FMARS.2020.00505/BIBTEX>

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- Bisson, K. M., Siegel, D. A., DeVries, T., Cael, B. B., & Buesseler, K. O. (2018). How data set characteristics influence ocean carbon export models. *Global Biogeochemical Cycles*, *32*(9), 1312–1328. <https://doi.org/10.1029/2018GB005934>
- Boyd, P. W. (2015). Toward quantifying the response of the oceans' biological pump to climate change. *Frontiers in Marine Science*, *2*, 77. <https://doi.org/10.3389/FMARS.2015.00077/BIBTEX>
- Buitenhuis, E. T., Vogt, M., Moriarty, R., Bednaršek, N., Doney, S. C., Leblanc, K., et al. (2013). MAREDAT: Towards a world atlas of marine ecosystem data. *Earth System Science Data*, *5*(2), 227–239. <https://doi.org/10.5194/essd-5-227-2013>
- Cael, B. B., Dutkiewicz, S., & Henson, S. (2021). Abrupt shifts in 21st-century plankton communities. *Science Advances*, *7*(44), 8593–8622. <https://doi.org/10.1126/sciadv.abf8593>
- Calbet, A., & Calbet, A. (2008). The trophic roles of microzooplankton in marine systems. *ICES Journal of Marine Science*, *65*(3), 325–331. <https://doi.org/10.1093/ICESJMS/FSN013>
- Calbet, A., & Landry, M. R. (2004). Phytoplankton growth, microzooplankton grazing, and carbon cycling in marine systems. *Limnology & Oceanography*, *49*(1), 51–57. <https://doi.org/10.4319/lo.2004.49.1.0051>
- Campbell, R. G., Sherr, E. B., Ashjian, C. J., Plourde, S., Sherr, B. F., Hill, V., & Stockwell, D. A. (2009). Mesozooplankton prey preference and grazing impact in the western Arctic Ocean. *Deep Sea Research Part II: Topical Studies in Oceanography*, *56*(17), 1274–1289. <https://doi.org/10.1016/J.DSR2.2008.10.027>
- Chen, M., Liu, H., & Chen, B. (2017). Seasonal variability of mesozooplankton feeding rates on phytoplankton in subtropical coastal and estuarine waters. *Frontiers in Marine Science*, *4*, 186. <https://doi.org/10.3389/fmars.2017.00186>
- Colebrook, J. M. (1985). Continuous plankton records: Overwintering and annual fluctuations in the abundance of zooplankton. *Marine Biology*, *84*(3), 261–265. <https://doi.org/10.1007/BF00392495>
- DeVries, T. (2022). The ocean carbon cycle. *Annual Review of Environment and Resources*, *47*(1), 317–341. <https://doi.org/10.1146/ANNUREV-ENVIRON-120920-111307>
- Dutkiewicz, S., Boyd, P. W., & Riebesell, U. (2021). Exploring biogeochemical and ecological redundancy in phytoplankton communities in the global ocean. *Global Change Biology*, *27*(6), 1196–1213. <https://doi.org/10.1111/GCB.15493>
- Evans, G. T., & Parslow, J. S. (1985). A model of annual plankton cycles. *Biological Oceanography*, *3*, 327–347. <https://doi.org/10.1080/01965581.1985.10749478>
- Friedlingstein, P., Jones, M. W., O'Sullivan, M., Andrew, R. M., Bakker, D. C. E., Hauck, J., et al. (2022). Global carbon budget 2021. *Earth System Science Data*, *14*, 1917–2005. <https://doi.org/10.5194/ESSD-2021-386>
- Gentleman, W., Leising, A., Frost, B., Strom, S., & Murray, J. (2003). Functional responses for zooplankton feeding on multiple resources: A review of assumptions and biological dynamics. *Deep Sea Research Part II: Topical Studies in Oceanography*, *50*(22–26), 2847–2875. <https://doi.org/10.1016/J.DSR2.2003.07.001>
- Gentleman, W., & Neuheimer, A. B. (2008). Functional responses and ecosystem dynamics: How clearance rates explain the influence of satiation, food-limitation and acclimation. *Journal of Plankton Research*, *30*(11), 1215–1231. <https://doi.org/10.1093/plankt/fbn078>
- Hansen, P. J., Bjørnsen, P. K., & Hansen, B. W. (1997). Zooplankton grazing and growth: Scaling within the 2–2- μ m body size range. *Limnology & Oceanography*, *42*(4), 687–704. <https://doi.org/10.4319/LO.1997.42.4.0687>
- Hirst, A. G., & Bunker, A. J. (2003). Growth of marine planktonic copepods: Global rates and patterns in relation to chlorophyll a, temperature, and body weight. *Limnology & Oceanography*, *48*(5), 1988–2010. <https://doi.org/10.4319/lo.2003.48.5.1988>
- Hoegh-Guldberg, O., & Bruno, J. F. (2010). The impact of climate change on the world's marine ecosystems. *Science*, *328*(5985), 1523–1528. <https://doi.org/10.1126/science.1189930>
- Holling, C. S. (1959). The components of predation as revealed by a study of small-mammal predation of the European Pine Sawfly. *The Canadian Entomologist*, *91*(5), 293–320. <https://doi.org/10.4039/ENT91293-5>
- Jeschke, J. M., Kopp, M., & Tollrian, R. (2004). Consumer-food systems: Why type I functional responses are exclusive to filter feeders. *Biological Reviews*, *79*(2), 337–349. <https://doi.org/10.1017/S1464793103006286>
- Kearney, K. A., Bograd, S. J., Drenkard, E., Gomez, F. A., Haltuch, M., Hermann, A. J., et al. (2021). Using global-scale Earth system models for regional fisheries applications. *Frontiers in Marine Science*, *8*, 622206. <https://doi.org/10.3389/fmars.2021.622206>
- Kostadinov, T. S., Siegel, D. A., & Maritorena, S. (2010). Global variability of phytoplankton functional types from space: Assessment via the particle size distribution. *Biogeosciences*, *7*(10), 3239–3257. <https://doi.org/10.5194/BG-7-3239-2010>
- Laufkötter, C., Vogt, M., Gruber, N., Aumont, O., Bopp, L., Doney, S. C., et al. (2016). Projected decreases in future marine export production: The role of the carbon flux through the upper ocean ecosystem. *Biogeosciences*, *13*, 4023–4047. <https://doi.org/10.5194/BG-13-4023-2016>
- Le Quéré, C., Buitenhuis, E. T., Moriarty, R., Alvain, S., Aumont, O., Bopp, L., et al. (2016). Role of zooplankton dynamics for southern ocean phytoplankton biomass and global biogeochemical cycles. *Biogeosciences*, *13*(14), 4111–4133. <https://doi.org/10.5194/bg-13-4111-2016>
- Lovato, T., Peano, D., Butenschön, M., Matera, S., Iovino, D., Scoccimarro, E., et al. (2022). CMIP6 simulations with the CMCC Earth system model (CMCC-ESM2). *Journal of Advances in Modeling Earth Systems*, *14*(3). <https://doi.org/10.1029/2021MS002814>
- McClain, C. R. (2009). A decade of satellite ocean color observations. *Annual Review of Marine Science*, *1*, 19–42. <https://doi.org/10.1146/ANNUREV-MARINE.010908.163650>
- Moriarty, R., & O'Brien, T. D. (2013). Distribution of mesozooplankton biomass in the global ocean. *Earth System Science Data*, *5*(1), 45–55. <https://doi.org/10.5194/ESSD-5-45-2013>
- Murphy, E. J., Johnston, N. M., Hofmann, E. E., Phillips, R. A., Jackson, J. A., Constable, A. J., et al. (2021). Global connectivity of Southern Ocean ecosystems. *Frontiers in Ecology and Evolution*, *9*, 624451. <https://doi.org/10.3389/fevo.2021.624451>
- Petrik, C. M., Luo, J. Y., Heneghan, R. F., Everett, J. D., Harrison, C. S., & Richardson, A. J. (2022). Assessment and constraint of mesozooplankton in CMIP6 Earth system models. *Global Biogeochemical Cycles*, *36*(11), e2022GB007367. <https://doi.org/10.1029/2022GB007367>
- Ras, J., Claustre, H., & Uitz, J. (2007). Spatial variability of phytoplankton pigment distributions in the subtropical South Pacific Ocean: Comparison between in situ and predicted data. *Biogeosciences Discussions*, *4*, 3409–3451.
- Rohr, T., Anthony, J. R., Lenton, A., Chamberlain, M. A., & Shadwick, E. H. (2023). Zooplankton grazing is the largest source of uncertainty for marine carbon cycling in CMIP6 models. *Nature Communications Earth Environment*, *4*(1), 212. <https://doi.org/10.1038/s43247-023-00871-w>
- Rohr, T., Richardson, A., Lenton, A., Chamberlain, M., & Shadwick, E. (2024). The global distribution of grazing dynamics estimated from inverse modelling. *Geophysical Research Letters*, *51*(8), e2023GL107732. <https://doi.org/10.1029/2023GL107732>
- Rohr, T., Richardson, A. J., Lenton, A., & Shadwick, E. (2022). Recommendations for the formulation of grazing in marine biogeochemical and ecosystem models. *Progress in Oceanography*, *208*, 102878. <https://doi.org/10.1016/J.POCEAN.2022.102878>
- Roy, S., Broomhead, D. S., Platt, T., Sathyendranath, S., & Ciavatta, S. (2012). Sequential variations of phytoplankton growth and mortality in an NPZ model: A remote-sensing-based assessment. *Journal of Marine Systems*, *92*(1), 16–29. <https://doi.org/10.1016/J.JMARSYS.2011.10.001>

- Schmoker, C., Hernández-León, S., & Calbet, A. (2013). Microzooplankton grazing in the oceans: Impacts, data variability, knowledge gaps and future directions. *Journal of Plankton Research*, 35(4), 691–706. <https://doi.org/10.1093/PLANKT/FBT023>
- Sieburth, J. M. N., Smetacek, V., & Lenz, J. (1978). Pelagic ecosystem structure: Heterotrophic compartments of the plankton and their relationship to plankton size fractions. *Limnology & Oceanography*, 23(6), 1256–1263. <https://doi.org/10.4319/LO.1978.23.6.1256>
- Siegel, D. A., Behrenfeld, M. J., Maritorena, S., McClain, C. R., Antoine, D., Bailey, S. W., et al. (2013). Regional to global assessments of phytoplankton dynamics from the SeaWiFS mission. *Remote Sensing of Environment*, 135, 77–91. <https://doi.org/10.1016/J.RSE.2013.03.025>
- Siegel, D. A., Buesseler, K. O., Doney, S. C., Salliey, S. F., Behrenfeld, M. J., & Boyd, P. W. (2014). Global assessment of ocean carbon export by combining satellite observations and food-web models. *Global Biogeochemical Cycles*, 28(3), 181–196. <https://doi.org/10.1002/2013GB004743>
- Siegel, D. A., DeVries, T., Cetinić, I., & Bisson, K. M. (2022). Quantifying the ocean's biological pump and its carbon cycle impacts on global scales. *Annual Review of Marine Science*, 15(1), 329–356. <https://doi.org/10.1146/ANNUREV-MARINE-040722-115226>
- Silsbe, G. M., Behrenfeld, M. J., Halsey, K. H., Milligan, A. J., & Westberry, T. K. (2016). The CAFE model: A net production model for global ocean phytoplankton. *Global Biogeochemical Cycles*, 30(12), 1756–1777. <https://doi.org/10.1002/2016GB005521>
- Steinberg, D. K., & Landry, M. R. (2016). Zooplankton and the ocean carbon cycle. *Annual Review of Marine Science*, 9(1), 413–444. <https://doi.org/10.1146/annurev-marine-010814-015924>
- Stock, C. A., & Dunne, J. (2010). Controls on the ratio of mesozooplankton production to primary production in marine ecosystems. *Deep-Sea Research Part I Oceanographic Research Papers*, 57(1), 95–112. <https://doi.org/10.1016/j.dsr.2009.10.006>
- Stock, C. A., Dunne, J. P., Fan, S., Ginoux, P., John, J., Krasting, J. P., et al. (2020). Ocean biogeochemistry in GFDL's Earth system model 4.1 and its response to increasing atmospheric CO₂. *Journal of Advances in Modeling Earth Systems*, 12(10), e2019MS002043. <https://doi.org/10.1029/2019MS002043>
- Stock, C. A., Dunne, J. P., & John, J. G. (2014). Global-scale carbon and energy flows through the marine planktonic food web: An analysis with a coupled physical–biological model. *Progress in Oceanography*, 120, 1–28. <https://doi.org/10.1016/J.POCEAN.2013.07.001>
- Stoecker, D. K., & Capuzzo, J. M. (1990). Predation on protozoa: Its importance to zooplankton. *Journal of Plankton Research*, 12(5), 891–908. <https://doi.org/10.1093/PLANKT/12.5.891>
- Stow, C. A., Jolliff, J., McGillicuddy, D. J., Doney, S. C., Allen, J. I., Friedrichs, M. A., et al. (2009). Skill assessment for coupled biological/physical models of marine systems. *Journal of Marine Systems*, 76(1–2), 4–15. <https://doi.org/10.1016/J.JMARSYS.2008.03.011>
- Strömberg, K. H., Smyth, T. J., Allen, J. I., Pitois, S., & O'Brien, T. D. (2009). Estimation of global zooplankton biomass from satellite ocean colour. *Journal of Marine Systems*, 78(1), 18–27. <https://doi.org/10.1016/J.JMARSYS.2009.02.004>
- Turner, J. T. (2015). Zooplankton fecal pellets, marine snow, phytodetritus and the ocean's biological pump. *Progress in Oceanography*, 130, 205–248. <https://doi.org/10.1016/j.pocean.2014.08.005>
- Vallina, S. M., Ward, B. A., Dutkiewicz, S., & Follows, M. J. (2014). Maximal feeding with active prey-switching: A kill-the-winner functional response and its effect on global diversity and biogeography. *Progress in Oceanography*, 120, 93–109. <https://doi.org/10.1016/J.POCEAN.2013.08.001>
- Vargas, C. A., & González, H. E. (2004). Plankton community structure and carbon cycling in a coastal upwelling system. I. Bacteria, microprotozoans and phytoplankton in the diet of copepods and appendicularians. *Aquatic Microbial Ecology*, 34, 151–164. <https://doi.org/10.3354/AME034151>
- Walker, N. D., Susanto, H., Steinke, M., & Codling, E. A. (2019). Bottom-up and top-down control in a multitrophic system: The role of nutrient limitation and infochemical-mediated predation in a plankton food-web model. *Communication in Biomathematical Sciences*, 2, 65–84. <https://doi.org/10.5614/cbms.2019.2.2.1>
- West, G. B., Brown, J. H., & Enquist, B. J. (1997). A general model for the origin of allometric scaling laws in biology. *Science*, 276(5309), 122–126. <https://doi.org/10.1126/science.276.5309.122>
- Westberry, T., Behrenfeld, M. J., Siegel, D. A., & Boss, E. (2008). Carbon-based primary productivity modeling with vertically resolved photoacclimation. *Global Biogeochemical Cycles*, 22(2), 2024. <https://doi.org/10.1029/2007GB003078>

References From the Supporting Information

- Heneghan, R. F., Everett, J. D., Sykes, P., Batten, S. D., Edwards, M., Takahashi, K., et al. (2020). A functional size-spectrum model of the global marine ecosystem that resolves zooplankton composition. *Ecological Modelling*, 435, 109265. <https://doi.org/10.1016/J.ECOLMODEL.2020.109265>
- Yool, A., Palmiéri, J., Jones, C. G., de Mora, L., Kuhlbrodt, T., Popova, E. E., et al. (2021). Evaluating the physical and biogeochemical state of the global ocean component of UKESM1 in CMIP6 historical simulations. *Geoscientific Model Development*, 14(6), 3437–3472. <https://doi.org/10.5194/gmd-14-3437-2021>
- Yool, A., Popova, E. E., & Anderson, T. R. (2013). MEDUSA-2.0: An intermediate complexity biogeochemical model of the marine carbon cycle for climate change and ocean acidification studies. *Geoscientific Model Development*, 6(5), 1767–1811. <https://doi.org/10.5194/gmd-6-1767-2013>