

1 **Assessing groundwater salinity across Africa**

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12 In Africa groundwater is the principal source of drinking water (<https://washdata.org/>) and has significant
13 potential to improve food production as a source of irrigation water. Although volumes of stored
14 groundwater are high when compared to surface water, not all the available groundwater is suitable for
15 irrigation. One of the principal factors that limit the development of groundwater is salinity, but quantitative
16 evidence regarding the scale of salinization in Africa has been lacking. This paper presents an initial
17 analysis of the extent of groundwater salinization in Africa, bringing together groundwater salinity data
18 (n=12255) from across the continent. In our dataset *c.*80% of the samples have electrical conductivity values
19 less than 2000 $\mu\text{S}/\text{cm}$. Samples with high conductivity values of more than 2000 $\mu\text{S}/\text{cm}$ are restricted to
20 limited geological and climatic environments. The data reveals salinity does not affect the water security
21 status of most countries in Africa. However, salinity compounds water insecurity issues in arid low
22 groundwater recharge environments. The paper discusses the role of anthropogenic processes such as river
23 valley water resources development, extensive groundwater pumping, inter-basin groundwater transfer, and
24 irrigation in altering the salinity of groundwater bodies. The paper further elucidates the origin of
25 groundwater salinity by critically reviewing the natural and human-induced factors that control the
26 salinization of the various groundwater bodies across Africa. Existing case studies reveal several causes of

27 salinization, including i) human-induced salinization, ii) climate and hydrological change-induced paleo
28 groundwater salinity, iii) rock dissolution, and iv), saltwater encroachment.

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31 **1. Introduction**

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33 Driven by anthropogenic pressure and climate change, salinity has become a growing water security
34 challenge and the trend in salinity of fresh waters is increasing globally (Thorslund and van Vliet, 2020).

35 Water salinity will have an impact on the ability to meet water related Sustainable Development Goals
36 (SDGs 2, 6, 11, 12, etc.) in some regions (Flörke et al., 2019).

37

38 MacDonald et al. (2012) demonstrate that Africa's groundwater resources have the potential to transform
39 rural development if used sustainably. However, (Edmunds, 2012) points out that salinity may be one
40 important factor that restricts the use of groundwater for domestic use and food production, most notably
41 in semi-arid or arid areas. Groundwater quality is also increasingly incorporated into discussions of
42 sustainable groundwater use (Gleeson et al., 2020). High salinity water hampers plant growth, limiting its
43 use for crop irrigation. Irrigation in arid regions can lead to a buildup of soil salinity leaving areas unsuitable
44 for agriculture without significant flushing. Drinking water is also affected, and becomes significantly
45 unpalatable at total dissolved solids (salinity) levels greater than 1000 mg/L (WHO, 2011). High
46 groundwater salinity is one of the primary reasons for abandonment of newly drilled rural water supply
47 boreholes in arid regions (Kebede and Taye, 2021; Rivett et al., 2020; UNICEF, 2008; Pavelic et al., 2012).

48 There is growing evidence that high salinity drinking water is not only unpalatable but compromises
49 livelihoods and human health (Al Nahian et al., 2018; Naser et al., 2020; Dasgupta et al., 2016), for example,
50 salinity has critical health implications for blood pressure and kidney functions (Rosinger et al., 2021). For

51 all these reasons, it is important to know the extent of the salinity challenges facing Africa to help realize
52 the potential sustainable groundwater use has to produce transformative economic growth.

53

54 There is growing availability of global groundwater salinity data (Thorslund and van Vliet, 2020) and
55 scientific literature on deciphering the origins of groundwater salinity (Mirzavand et al., 2020), mapping
56 the extent of anthropogenic salinization of surface and groundwaters (Thorslund et al., 2021; Foster et al.,
57 2018), designing groundwater salinity management strategies (Khan et al., 2008; Fitch et al., 2016;
58 Lamontagne et al., 2005) and on assessing impacts of groundwater salinity (Russ et al., 2020). However,
59 because of the paucity of salinity data for Africa, the situation in Africa is poorly reflected in this global
60 debate. The lack of knowledge on the extent, origins, trends, and impacts of groundwater salinity hampers
61 the formulation of policy and management strategies.

62

63 The objectives of this paper are to i) highlight the extent of groundwater salinization in Africa and analyze
64 the link between groundwater security (recharge and storage) and groundwater salinity, ii) assess the natural
65 and human-induced causes of groundwater salinization, and iii) highlight emerging groundwater salinity
66 management strategies that are being used or could be used in Africa.

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69 **2. Methods**

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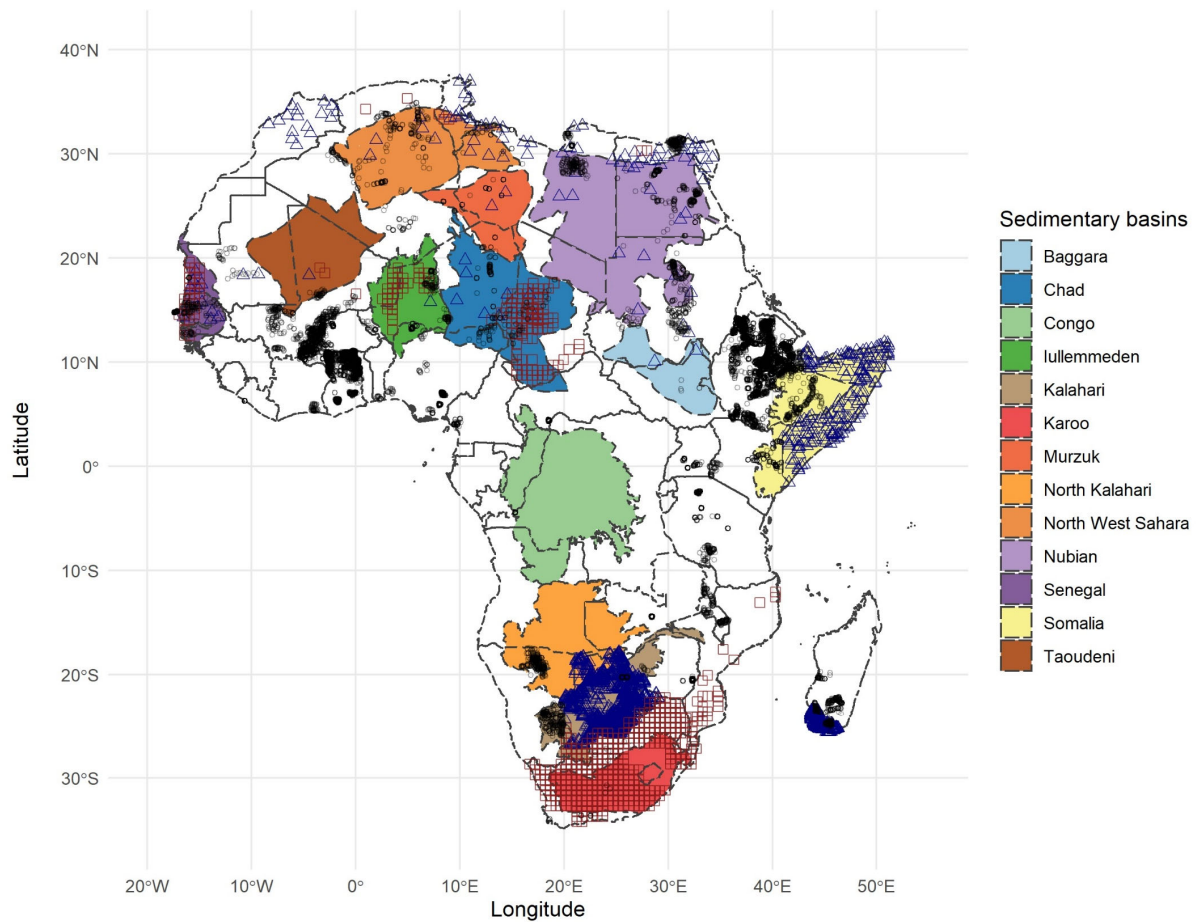
71 To develop a groundwater salinity map for Africa, we relied on secondary data sources including published
72 and unpublished reports and datasets. Since electrical conductivity (EC, $\mu\text{S}/\text{Cm}$) is easily measured, and
73 because it correlates strongly with total dissolved solids (TDS, mg/L), it is often used as a proxy for salinity
74 and is commonly found in water quality datasets even when anion or cation concentrations are not directly
75 measured. Thus, we used electrical conductivity as a proxy for groundwater salinity in this study. We

76 compiled 12488 data points (Figure 1) from historic records held in published and unpublished reports (list
77 provided in annex i of Supplementary Materials) such as those held by the British Geological Survey (BGS),
78 published IAEA technical documents, UNESCO reports, and national databases. We also digitized 39
79 regional, national, and sub-national maps held by the BGS which contain data on groundwater salinity from
80 across Africa. The final dataset was used to construct groundwater salinity maps and to conduct statistical
81 analysis. We discarded data points lacking geographic coordinates, leaving a dataset with 12255 individual
82 electrical conductivity measurements. The dataset also includes recently published salinity data points from
83 South Africa (n=23, mean SEC from time-series data) and Egypt (n=251) (Thorslund and van Vliet, 2020).
84 The groundwater salinity mapping does not include an attempt to interpolate/extrapolate beyond these
85 points. The dataset and digitized maps cover most of Africa, but the humid central Africa region (Congo,
86 DRC, Gabon, and Angola) has a very low data density. The data density mirrors the relative proportions of
87 groundwater use and the geographic focus of past groundwater research. Disregarding the uninhabited
88 deserts in the Sahel and Sahara, the high data density corresponds to regions where groundwater is generally
89 the sole or the primary source of water for all types of use, while the low data density corresponds to humid
90 regions where alternative surface water sources are readily available.

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94

95 *Figure 1. Distribution of groundwater electrical conductivity sample data points (black circles),*
 96 *and data points (blue triangles) and contours (red squares) from digitised maps shown along with*
 97 *country boundaries and large sedimentary basin aquifers from (Bonsor et al., 2018). Points in*
 98 *areas with less density of data are lightly coloured whereas points in areas of high data density*
 99 *are darker.*

100

101 To produce the salinity map, all data was converted to specific electrical conductivity (EC) in $\mu\text{S}/\text{cm}$.
 102 Sample or map data was converted to weight per unit volume in milligrams per litre (mg/l) where necessary
 103 and then to electrical conductivity using a conversion factor of 0.7, a reasonable approximation for the
 104 range of salinity values within our dataset (Rusydi, 2018). Then the conductivity values were classified into

105 five categories that were loosely based on classifications by (Freeze and Cherry, 1979) and the FAO (1992:
106 <https://www.fao.org/3/t0234e/t0234e00.htm>): non-saline groundwater (< 500 $\mu\text{S}/\text{cm}$); slightly saline (500
107 – 1000 $\mu\text{S}/\text{cm}$); moderately saline (1000 – 2000 $\mu\text{S}/\text{cm}$); highly saline (2000 – 15,000 $\mu\text{S}/\text{cm}$); and very
108 highly saline groundwater or brine (> 15,000 $\mu\text{S}/\text{cm}$). Data digitised from maps as polygons were then
109 converted to gridded points at a density of 0.5° and all map and sample point data were plotted on a map of
110 Africa (Figure 1) and coloured by conductivity classification (Figure 2). Often data digitised from maps
111 was only provided as a concentration or conductivity range. Hence, map classes were included within the
112 final classification and presented on our African conductivity map (Figure 2). Due to the nature of the map
113 data these were not included in the analysis described below, where raw values were required.

114

115 To assess the compounded risk to water security caused by groundwater salinity, we explore the original
116 point sample data using maps, cross-plots, stacked bar-charts, box-plots (in the supplementary material),
117 and the Kruskal and Wilcoxon population tests (in the supplementary material). We also compare the original
118 sample point data for which values of electrical conductivity are available with: aridity; groundwater
119 recharge; and groundwater storage estimates (MacDonald et al., 2012)(MacDonald et al., 2021). As stated
120 previously, points converted from map data were not used in this analysis. It should be noted that the dataset
121 and the salinity maps are meant to provide a regional perspective on the state of groundwater salinity and
122 are not appropriate for local scale assessments. The dataset lacks 3D information due to the lack of depth
123 information within the original salinity datasets. The final dataset may be biased towards low salinity
124 groundwaters as high salinity groundwater is not the target for groundwater development through drilling
125 and water well salinity information is lacking. The other data bias is related to the regional disparity in data
126 density. In the hyper-arid, arid, and semi-arid environments, where groundwater is the predominant source
127 of drinking water, the data density is high.

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130 3. Results and Discussion

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132 3.1. The extent of aquifer salinization in Africa and water security

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134

135 *The extent of fresh and saline groundwater:*

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137 Figure 2 shows the EC distribution map constructed from the combined data points and the digitized maps
138 (see annex ii for other sets of maps, Supplementary Materials). Approximately 80% of the groundwater
139 samples within our dataset have EC values less than 2000 $\mu\text{S}/\text{cm}$ (Figure 3a). More than 50% of the
140 groundwater samples fall in the non-saline category ($\text{EC} < 500\mu\text{S}/\text{cm}$).The salinity dataset and map reveal
141 that the majority of the multi-layered sedimentary basin aquifers with large volumes of groundwater storage
142 (e.g. the Nubian Sandstone Aquifer System, The Chad Basin Aquifers, the Iullemeden-Taoudeni Aquifer
143 System, the Bagra basin, etc.) in Africa essentially hold non saline, slightly saline or moderately saline
144 groundwater resources, with salinity problems only encountered in specific regions, zones or depth levels
145 (Figure 3 a-d). The sedimentary basin with the highest frequency of saline groundwater ($\text{EC} > 2000\mu\text{S}/\text{cm}$)
146 is encountered in the North Western Saharan Aquifer underlying Tunisia, Algeria and Libya and in the
147 North Kalahari Sedimentary basin aquifer in Southern Africa (Figure 3b). By rock types, all aquifers contain
148 a substantial portion ($>60\%$) of low to moderate salinity groundwaters ($\text{EC} < 2000 \mu\text{S}/\text{cm}$) (Figure 3c). High
149 groundwater salinity values ($\text{EC} > 2000 \mu\text{S}/\text{cm}$) are observed in all rock types confirming the localized
150 nature of groundwater salinity in all rocks. Unconsolidated sediments and consolidated sediments with
151 inter-granular porosity host moderate salinity (median $\text{EC} > 1000 \mu\text{S}/\text{cm}$) groundwater bodies when
152 compared with basement, volcanics and consolidated fractured sedimentary aquifers which host non-saline
153 ($\text{EC} < 500 \mu\text{S}/\text{cm}$) or slightly saline groundwaters (EC values $< 1000 \mu\text{S}/\text{cm}$) (Figure 3c). The unconsolidated
154 sediments and the consolidated sediments with inter-granular porosity aquifers show the most frequent high

155 salinity values with $EC > 2000 \mu S/cm$. The basement rock areas, which account for 70% of drinking water
156 sources in Africa, hold essentially non-saline groundwater resources (median EC value $< 500 \mu S/cm$),
157 except for those in arid areas or in areas where wetland induced salinity may be locally encountered. The
158 volcanic aquifers which underlie most of the region in Eastern Africa (Ethiopia and Kenya) contain non-
159 saline and slightly saline groundwaters.

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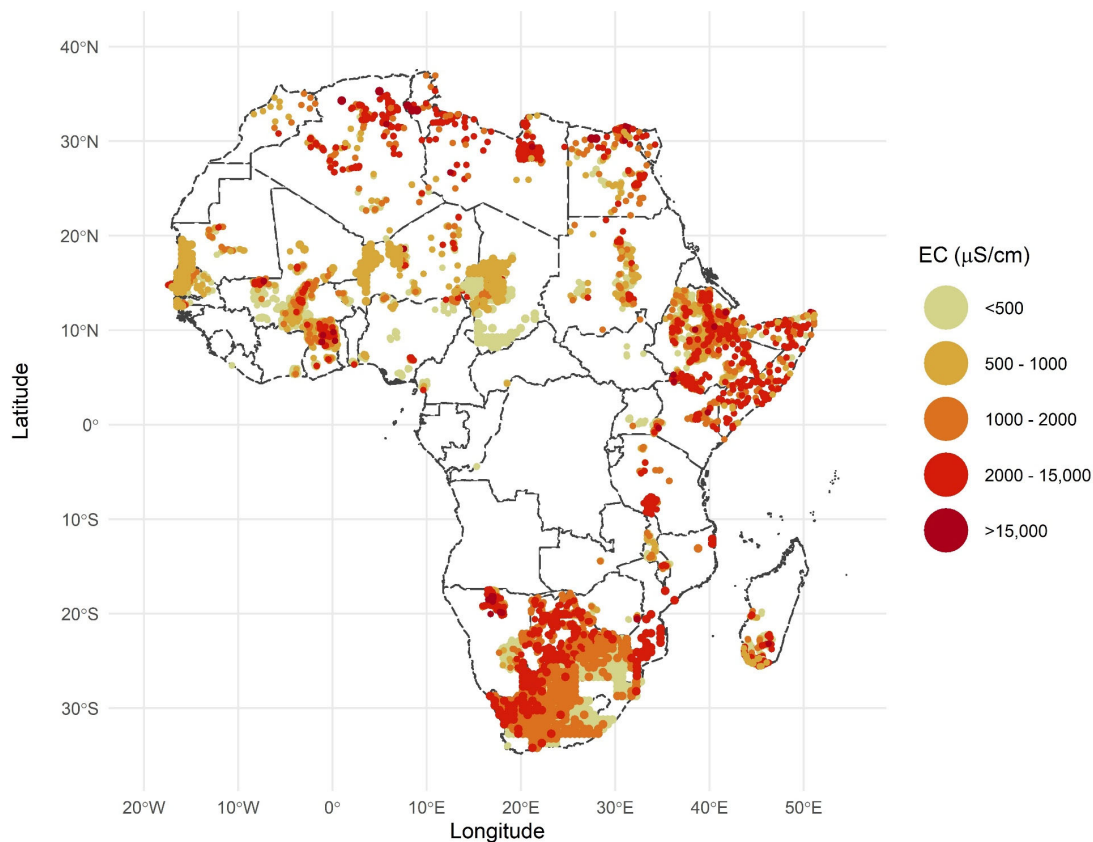
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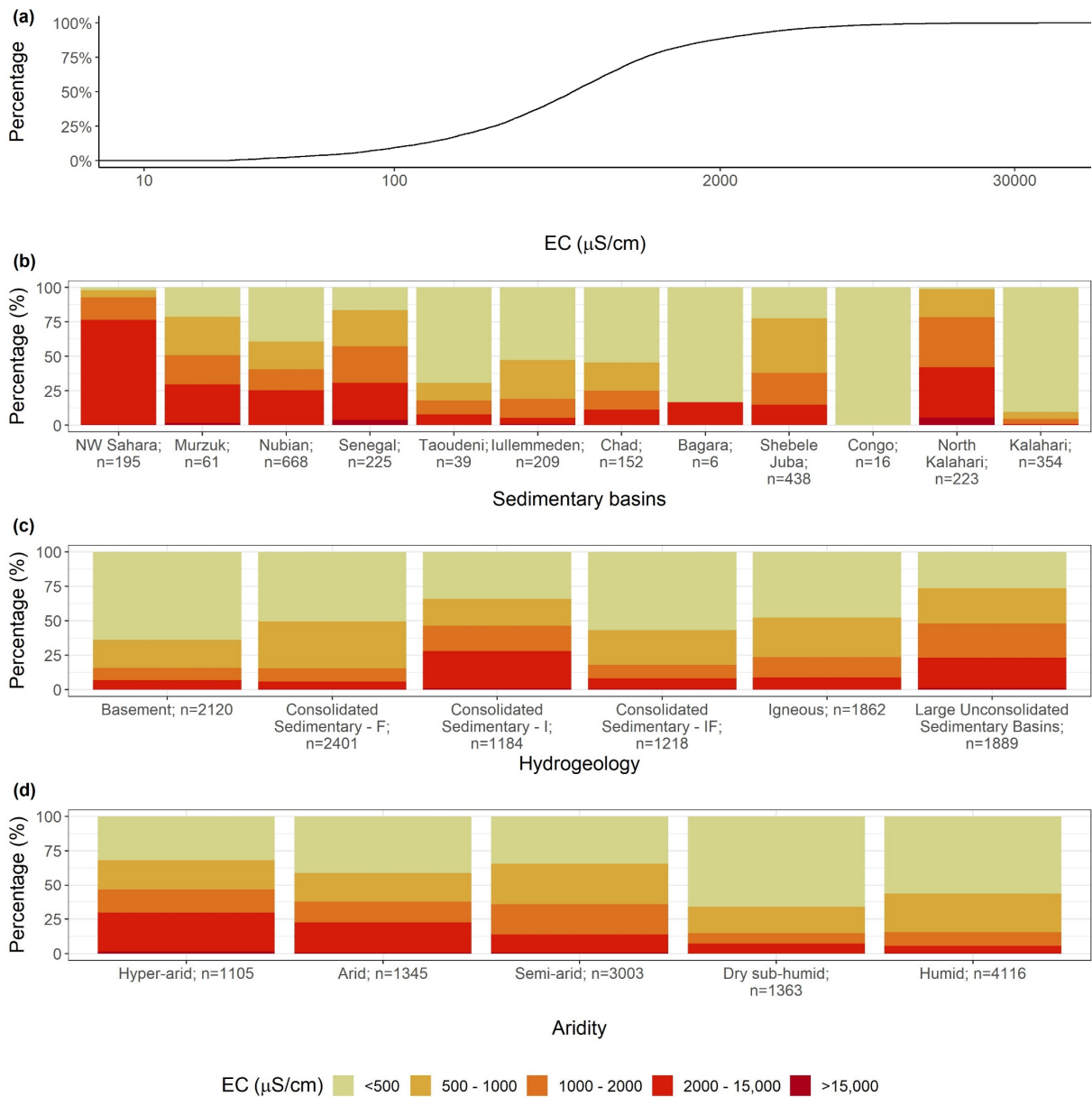
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177 *Figure 2. Distribution of EC in groundwaters across Africa.*

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179 Comparing salinity levels with aridity (figure 3d) reveals that aridity is the best predictor of groundwater
 180 salinity with the most arid environments hosting higher salinity groundwaters compared to humid
 181 environments. There is a progressive increase in salinity with the increase in the degree of aridity (Figure
 182 3d). In hyper arid, arid and semi-arid environments, the median EC value exceeds 1000 $\mu\text{S}/\text{cm}$, while in
 183 dry sub-humid and humid environments the median groundwater EC value is <1000 $\mu\text{S}/\text{cm}$ (see annex ii,
 184 Supplementary Materials). Outlier, high EC values, however, exist in all aridity regimes. Aquifers which
 185 contain higher EC values (>1000 $\mu\text{S}/\text{cm}$) are unconsolidated sediments and consolidated inter-granular
 186 porosity aquifers located in arid, semi-arid and hyper arid zones. Hyper-arid environments also show an
 187 abundance of moderate- to non-saline groundwaters; the most likely cause is the paleo climate regime

188 (wetter and humid) under which recharge took place. The vast majority of the groundwater in the Sahel and
 189 Northern Africa have been recharged under the humid and wet climate regime of the Pleistocene or
 190 Holocene (Jasechko, 2019).
 191



192

193

194 *Figure 3. Frequency plot of salinity distribution, b) salinity statistics for the major sedimentary*
195 *basins, c) salinity distribution by geology (I- inter-granular, F-Fracture) and d) salinity*
196 *distribution by aridity.*

197

198

199 Although we cannot examine the variation in salinity with depth using the dataset we compiled, a review
200 of the literature shows that there is no clear systematic relation between salinity and depth to the
201 groundwater bodies. For example, the deeper section of the sedimentary formation in Northern Africa
202 contains fresh groundwater bodies (e.g., the Continental Intercalair), while aquifers at a shallow depth such
203 as the continental complex aquifer and the Djifara aquifer in coastal North Africa contain locally
204 mineralized waters (Trabelsi et al., 2009). The complex salinity pattern in Tunisia, Algeria and Libya
205 (Figure 2) reveals groundwater wells tapping into these different aquifers bearing different salinities.
206 Additionally, in the transboundary Lower and Upper Kalahari aquifers in Namibia, Angola, and Botswana,
207 the deep aquifers (Lower Kalahari Aquifer) hold fresh groundwater compared to the unconsolidated shallow
208 aquifers as also revealed by Bäumle et al., (2019). Contrary to the two examples given above, in the
209 Shebelle-Juba basin, the deeper sedimentary layers contain more saline groundwaters compared to the
210 shallow aquifers in Ethiopia as revealed by Hadwen et al., (1973). An increase in salinity with depth is also
211 a common feature in volcanic terrains in the East Africa region (Kebede et al., 2005).

212

213

214 *Groundwater salinity and water security:*

215

216 There is a clear inverse correlation between groundwater recharge and salinity (Figure 4a, c) and thus
217 groundwater salinity adds to the water insecurity in environments with low groundwater recharge. While
218 most of the high recharge areas (Western Africa, Central Africa, East Africa highlands, Central African

219 Republic) show a low level of groundwater salinity (Figure 4a, c), the low to very low recharge regions in
220 North Africa are characterized by variable salinities with more frequent saline groundwater wells.
221 Generally, groundwaters in arid and semi-arid areas with minor recharge (Botswana, Namibia, Somalia,
222 the Rift Valley in Ethiopia, Somalia, Kenya, Tanzania, and Malawi, as well as arid coastal areas in Northern
223 Africa) contain complex salinity patterns with saline groundwaters occurring side by side with fresh
224 groundwaters (Figure 4a).

225

226 Salinity adds to groundwater insecurity in both high and low groundwater storage environments. There is
227 a weak positive correlation between groundwater storage and salinity (Figure 4b, d). Groundwater salinity
228 increases with storage volume likely because the high groundwater residence times in aquifers with high
229 storage lead to a longer time for rock water interactions to occur and minerals equilibrate between
230 groundwater and host rocks (e.g. Edmunds et al., 2003).

231

232 Combining the groundwater salinity with the long-term annual groundwater recharge and storage (Figure
233 4d) provides a basis to assess water security on a country-by-country basis. Figure 4e illustrates on average
234 how each groundwater salinity class corresponds to specific groundwater recharge and storage conditions.
235 Non-saline ($EC < 500 \mu S/cm$) and moderately saline ($500 > EC > 1000 \mu S/cm$) groundwater environments are
236 characterized by high recharge and low storage, while high salinity corresponds to low recharge and high
237 storage areas. While regions with large groundwater storage, high recharge, and low salinity are the most
238 water secure, regions with low groundwater storage, low recharge, and high groundwater mineralization
239 are fragile with respect to their groundwater resources. In this regard, countries that have been identified as
240 the most water-secure based on groundwater recharge and storage alone (Nigeria, Guinea Bissau, Congo,
241 Angola, Democratic Republic of Congo) (MacDonald et al., 2021) remain water-secure when the
242 groundwater salinity dimension is considered. Except in some specific areas (e.g, Southern Angola, coastal
243 regions of Nigeria, Figure 2), groundwaters in these countries are predominantly low salinity ($EC < 1000$
244 $\mu S/cm$).

245

246 Out of the five countries (Eritrea, eSwatini, Lesotho, Zambia, and Zimbabwe) with the lowest water security
247 by the measure of groundwater recharge and storage, eSwatini, Zambia, Zimbabwe and Lesotho, water
248 security issues are not compounded by high groundwater salinity. With the exception of recorded local high
249 groundwater salinity, the aquifers underlying the four countries hold generally low salinity groundwaters.

250

251 In Eritrea on the other hand, high groundwater salinity is prevalent in coastal areas in the east as well as in
252 the western lowlands. Low salinity groundwaters are restricted to the crystalline aquifers in the highlands
253 of Eritrea. By this measure, this makes Eritrea the most groundwater insecure country on the continent. In
254 Zimbabwe, high groundwater salinity is restricted to the arid south and southern part of the country
255 associated with alluvial sediments.

256

257 Five of the countries with the highest groundwater recharge and low storage (Burundi, Liberia, Guinea,
258 Equatorial Guinea, Ivory Coast), i.e. countries underlain by aquifers vulnerable to drought but capable of
259 withstanding long-term depletion because of sustained recharge, are characterized by aquifers bearing low
260 salinity groundwaters. Thus, groundwater salinity does not exacerbate groundwater security in these
261 countries.

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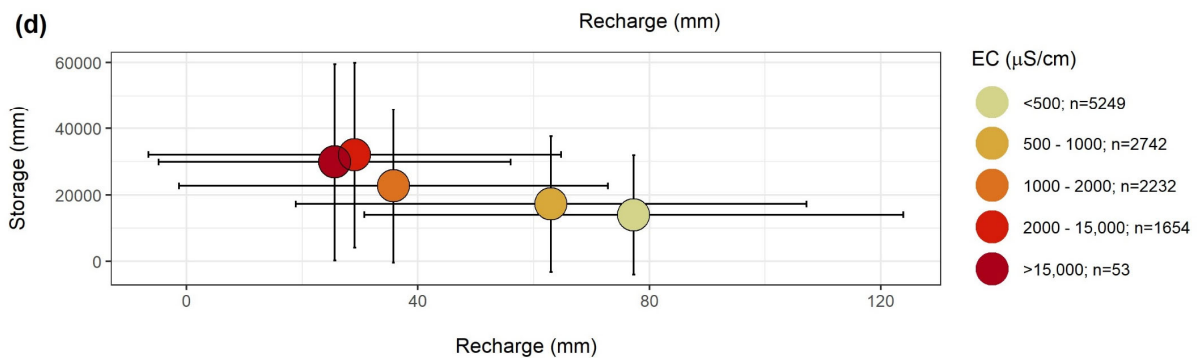
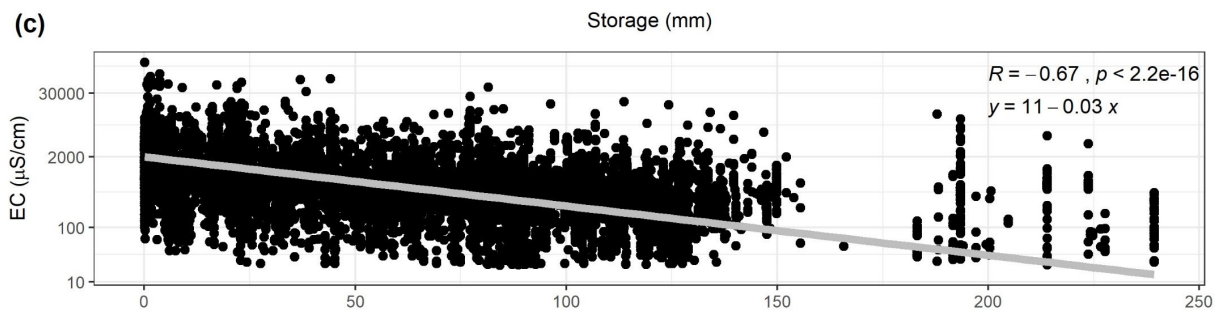
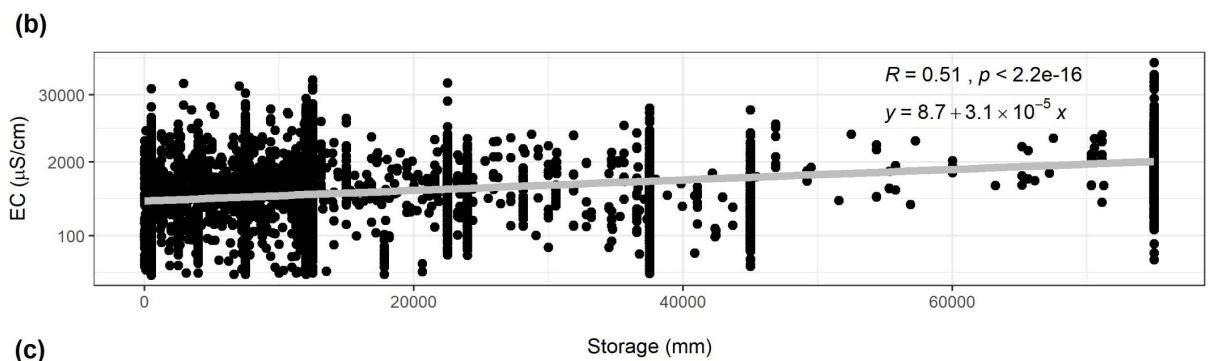
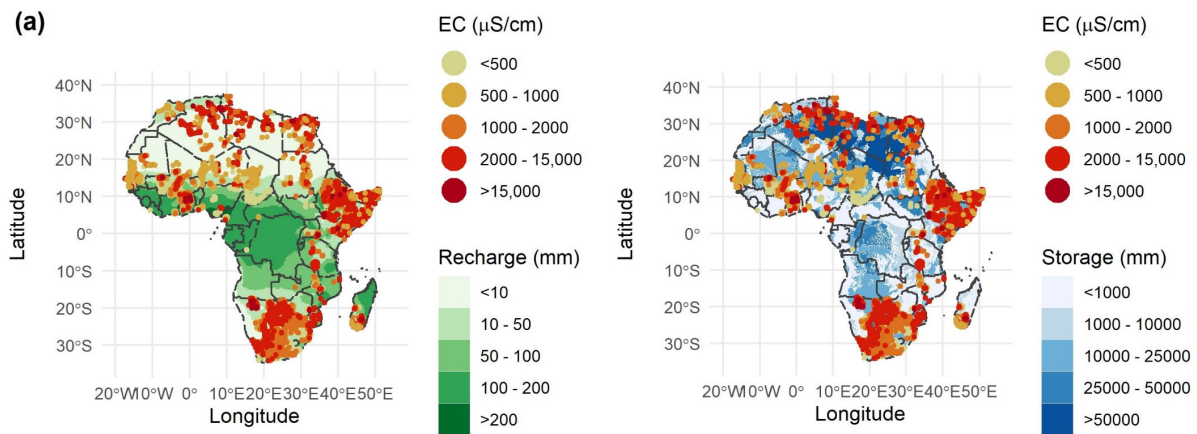
263 The situation is more complex for countries with the highest groundwater storage and low recharge (Egypt,
264 Algeria, Tunisia, Niger, and Botswana), i.e. countries underlain by aquifers vulnerable to long-term
265 depletion but capable of withstanding short-term droughts. Groundwater salinity adds an additional layer
266 of water insecurity in Tunisia, Egypt, Algeria, and Botswana as these countries are underlain by saline
267 groundwaters associated with fresh groundwater zones. In contrast, Niger is generally exempt from the
268 compounded effect of salinity induced groundwater insecurity as groundwater salinity in Niger appears
269 only a localized problem. Countries with intermediate storage and recharge characteristics (<50 mm) but
270 sustained episodic recharge (e.g. Somalia, Botswana, Namibia, South Africa) face the additional challenge

271 of groundwater insecurity when groundwater salinity is taken into consideration. These countries are

272 underlain by aquifers with high salinity and complex salinity patterns.

273

274



276 *Figure 4: a) salinity vs recharge and storage spatial plot, b) salinity vs storage correlation, c)*
277 *salinity recharge correlation, d) storage-recharge-salinity relation ($R^2=0.94$, $p=0.016$)*

278

279 In summary, groundwater salinity does not affect the water security status of most countries in Africa.
280 However, salinity compounds water insecurity issues in arid low groundwater recharge environments such
281 as Eritrea, Djibouti, Somalia, Botswana, Namibia, Chad, Ethiopia, Senegal, Mauritania, parts of Kenya and
282 Tanzania, and South Africa as well as in arid large groundwater storage environments such as Egypt, Libya,
283 Tunisia, and Algeria.

284

285 **3.2 Impacts of groundwater salinization on local water supply**

286

287 Despite the general finding that 80% of the water wells show EC value less than 2000 $\mu\text{S}/\text{cm}$ and may be
288 acceptable for drinking use, groundwater salinity is an issue in many localized aquifers and has implications
289 for primary users of groundwater. For example, high groundwater salinity has been shown to be responsible
290 for the abandonment of water wells (Pavelic et al., 2012) (UNICEF, 2008) limiting access to safe drinking
291 water for the primary users. However, the lack of systematic data on boreholes and abandoned boreholes
292 limits our understanding of the scale of impact of salinity on the abandonment of boreholes at national and
293 continental scales. Several studies report the more than 30% abandonment in local environments including,
294 for example, Mozambique (UNICEF, 2008) and Sudan (Fragaszy and Closas, 2016), Abandonment of
295 boreholes caused by the saline groundwaters is most severe in alluvial aquifers and coastal aquifers. For
296 example, in the Shire Valley, southern Malawi, persistent non-functionality or abandonment of boreholes
297 tapping alluvial aquifer was typically ascribed to salinity (Rivett et al., 2020), (Lapworth et al., 2020)
298 reported that for upscaled country estimates >10% of handpump borehole sources in Malawi TDS exceeded
299 1000 mg/L and these were focused in southern Malawi. High salinity resulted in the abandonment of up to
300 70% of drilled boreholes in the vast sedimentary aquifers of the Horn of Africa countries such as Somalia,

301 Ethiopia, and Kenya (Kebede and Taye, 2021). In Libya, 75% of boreholes have been lost to saltwater
302 intrusion in the coastal region (Steyl and Dennis, 2010). In the coastal plains of Comoros island, fewer than
303 30% of water wells provide groundwater of acceptable quality (Comte et al., 2016) when measured by the
304 salinity levels. In Kenya, (Foster et al., 2018) show for every 100 $\mu\text{S}/\text{cm}$ EC increase there is a 3% rise in
305 the risk of failure of rural water wells.

306

307

308 **3.3. Typologies of groundwater salinization in Africa**

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310 To effectively manage the challenge of groundwater salinity in Africa, there is a need to understand the
311 underlying drivers and causes which can be broadly categorized as anthropogenic, hydrologic, geologic,
312 and saline water encroachment, all of which we discuss further below. Since most of the high salinity
313 environments are local in extent, not all these zones may be visible in Figure 2.

314 **Anthropogenic related:** The major causes of anthropogenic salinization include irrigation practices,
315 mining, and land-use changes. Human induced salinization of groundwater has been reported in some river
316 basins and aquifers in Africa. Given the fact that river basin irrigation development in Africa is scarce in
317 the large river basins such as the Congo and Niger, irrigation-induced salinization of groundwater is
318 restricted to those intensively developed river and aquifer basins such as the Nile, Senegal, Orange, Shire,
319 Zambezi, Volta, Vaal and Awash basins. Substantial portions of groundwaters in the Nile Valley, the
320 Senegal River Delta (Gning et al., 2017), and the Val River in South Africa are now seriously threatened
321 by salinization that has led to the abandonment of many irrigated areas and loss of water wells to salinization
322 (Fragaszy and Closas, 2016). In Tunisia and Algeria, irrigation return water led to the high salinity of the
323 shallow groundwater (Yangui et al., 2012). The mechanisms that lead to irrigation induced groundwater
324 salinity involve multiple processes (Foster et al., 2018)(Mirzavand et al., 2020). Firstly, successive
325 evaporative concentration of water in the irrigation ponds, primary and secondary channels leads to the

326 increase in salinity of irrigation water. Subsequent downward flux of salt causes an increase in the salinity
327 of groundwaters. Secondly, the salinity of groundwater increases because of the addition of fertilizers and
328 plant protection products. Thirdly, rising water tables can result in the dissolution and mobilization of salt
329 that accumulates in the shallow soils to the groundwater and increase the salinity further. In coastal areas,
330 upstream damming of rivers discharging into the coastal plains has caused saltwater intrusion in Kenya and
331 Tanzania (Steyl and Dennis, 2010).

332

333 Extractive resource industries, such as coal mining, have also caused aquifer and surface water salinization.
334 One of the mechanisms of salinization is acid and non-acid mine drainage causing mobilization of ions and
335 metals, leading to waters of very high salinity. This is a widespread problem in coal mining areas in South
336 Africa (Nephalama and Muzerengi, 2016). Mixing of petroleum mine waste water with groundwater during
337 petroleum extraction is reported to have caused high groundwater salinities in South Sudan (Rueskamp et
338 al., 2014). Local salinization of aquifers is reported to occur in Egypt from disposal of brine from
339 desalination plants into the pumped aquifer (Jahnke et al., 2019).

340

341 Other human induced land-use changes are reported to have impacted groundwater quality. For instance,
342 groundwater salinity is reported to have increased accompanying the rise in recharge owing to conversion
343 of bushlands to agriculture land in Sahel Africa (Favreau et al., 2009). Subsequent to the increase in
344 groundwater recharge, the water table migrates towards the surface causing ponding of water, the
345 evaporative concentration of the water, and thereby an increase in salinity of the groundwaters. The high
346 groundwater salinity in the capital city of Mauritania, Nouakchott, results from the inter basin water transfer
347 from the Senegal River meant to supply the city. The excess water from the newly diverted water supply
348 resulted in water table rise of the pre-existing saline groundwater body leading to water logging and very
349 high salinity levels in the shallow coastal aquifer (Mohamed et al., 2017). More generally, increases in
350 groundwater salinity can also be the direct result of a range of urban land use pressures, including waste

351 water recharge from various industries as well as domestic waste water and sanitation sources(Lapworth et
352 al., 2017).

353

354 **Paleo hydrology related:** Groundwater salinity observed in several parts of Africa relates to paleo
355 hydrological processes. There are a number of mechanisms by which past hydrological processes govern
356 present day salinity patterns in groundwater.

357

358 Firstly, the direct entrapment or intrusion of paleo saline waters (e.g. paleo marine water, connate water)
359 causes groundwater salinization observed in many coastal regions observed in Figure 2 (see Supplementary
360 Material, annex ii for details for example for Senegal). Saline water entrapment/intrusion in coastal aquifers
361 during the paleo regression-transgression cycles is a notable source of groundwater salinity in many places
362 (Kafri, 1984) including the Senegal River Delta (Gning et al., 2017); the coastal aquifer of Western Africa
363 (Akouvi et al., 2008); the Danakil Depression, Ethiopia (Kafri, 1984); the Nile Delta, Egypt (van Engelen
364 et al., 2019); in complex volcano sedimentary aquifers, Djibouti (Awaleh et al., 2017); in the coastal
365 aquifers in Togo (Akouvi et al., 2008), in coastal aquifer of Mozambique (Stigter and Zhou, 2018), the
366 coastal aquifers of Tanzania (Walraevens et al., 2015); and in some portion of the North Western Saharan
367 Aquifer System in Tunisia (Edmunds et al., 2003) and explains the observed high groundwater salinities
368 that extend far inland in the coastal region such as in Senegal, Tunisia, Egypt and Libya (van Engelen et
369 al., 2019).

370

371 Secondly, paleo hydrological conditions can lead to the formation of hydrogeologic environments favoring
372 groundwater salinization. For instance, the aridification of the North and Horn of Africa since the mid-
373 Holocene has resulted in the deposition of many large paleolake sediments. Paleolakes, by their nature,
374 contain sediments of different grain sizes and permeability conditions. Differential permeability leads to
375 different groundwater flow velocities, causing stagnation of groundwater in some zones and faster flows in
376 other zones and eventually leading to a complex pattern of salinity as well as exceptionally high salinities.

377 Fine grained sediments often contain high salinity groundwaters when compared with coarse grained
378 sediments with higher permeability. Paleolakes explain the salinity pattern observed in unconsolidated
379 aquifers in many parts of Africa (Figure 3c), including for example the complex salinity patterns in the
380 alluvial-lacustrine sediments in the Kalahari region of Zambia (Banda et al., 2019), in the Shire River Basin
381 in Malawi (Rivett et al., 2020), and the alluvial lacustrine sediments of the East African Rift Valley (Ligate
382 et al., 2021). Salinity caused by paleolake sediments is common in arid environments since there has been
383 insignificant flushing to remove salts from the layers since the time of aquifer recharge.

384

385 Extensive deposits of paleolake sediments explain the complex salinity pattern observed in Somalia, Chad,
386 Namibia, Botswana, Ethiopia, and Kenya for example. Salinity may also be imparted owing to the shallow
387 depth of most of the alluvio-lacustrine sediments and the groundwaters hosted in these aquifers. The
388 shallow water table allows direct evapotranspiration from aquifers leading to the formation of groundwater
389 troughs with pockets of saline groundwater. The high salinity groundwaters in Sudan in 'Khors'
390 (groundwater troughs adjacent to the Nile River) is a typical example of this type of salinization (Kebede et
391 al., 2017; Abdalla, 2009). The high salinities seen in the Okavango delta, northern Botswana is caused by
392 direct groundwater evapotranspiration leading to a build-up of TDS in groundwaters (Zimmermann et al.,
393 2006). When paleo sediments are located in arid and semi-arid environments with the deeper water table,
394 the episodic recharge mobilizes the salt from the shallow soil zone to the water table enhancing the salinity
395 of groundwaters (Sami, 1992). However, such localized, evaporation induced salinity also occurs in wet
396 environments associated with loose sediments and wetlands and explains high localized salinities in Lake
397 Tana basin- Ethiopia (Kebede et al., 2011), and in the Kisumu aquifer around the eastern fringe of Lake
398 Victoria (Olago, 2019). Groundwater salinity associated with wetlands may be temporal and shifting in
399 time depending on the hydrography/hydrology of the wetland environment (Mabidi et al., 2018). The
400 salinization induced by differential permeability, direct evapotranspiration from shallow aquifers, and
401 successive evaporation-dissolution cycles of paleo alluvio-lacustrine sediments is exacerbated in endorheic

402 basins which are found in places such as the Lake Chad, Lake Chilwa, the East African Rift (Ethiopia,
403 Kenya, Tanzania).

404

405 **Rock dissolution:** Aquifer matrix dissolution adds ions to groundwaters. Aquifers containing intercalations
406 of readily soluble minerals such as gypsum, anhydrite, and halite if present only in small quantities results
407 in high salinity groundwater. Such soluble rock materials are often times associated with paleo lake
408 sediments, marine deposits including for example shales and carbonates. Because of the high solubility of
409 evaporites, even groundwaters with short residence time may acquire high salinities. Rock dissolution
410 explains the high groundwater salinity (Figure 2) encountered in large sedimentary aquifer basins such as
411 the Juba Shebelle basin in Somalia, the North Western Saharan Aquifer basin in Tunisia and Algeria (Pina
412 et al., 2017; Yangui et al., 2012), and Karoo sediments in South Africa. However, other factors play a role
413 in exacerbating rock dissolution induced salinization. For example, in the greater rift valley region, in
414 Ethiopia, Kenya and Tanzania heat from geothermal activity and carbon dioxide influx from deeper earth
415 accelerate the dissolution of the otherwise less soluble volcanic rocks and result in high salinity
416 groundwaters (Darling et al., 1996). Geothermal induced salinization of groundwater is commonly
417 observed, for example, in Ethiopia (Darling et al., 1996), Kenya (Darling et al., 1996), Tanzania, Djibouti
418 (Awaleh et al., 2017) and Eritrea (Lowenstern et al., 1999). Groundwater residence time controls the degree
419 to which groundwater salinity increases due to rock dissolution. Oftentimes deep circulating groundwaters
420 with longer residence time are more saline (Edmunds et al., 2003) than more shallow circulating short
421 residence time groundwaters. Groundwater residence time control on groundwater salinity is evident in the
422 volcanic aquifers of Eastern Africa (Ethiopia, Kenya, Tanzania) whereby shallow aquifers generally show
423 low salinity compared to the deep circulating long residence time groundwaters (Kebede et al., 2005)
424 (Banks et al., 2021).

425

426 **Modern saline water encroachment:** Saline water encroachment is the process whereby saline water from
427 surface or groundwater sources intrudes the freshwater aquifer, as a result of sea water intrusion, down or

428 up welling of groundwater from saline aquifers into the fresh water zones, and ingress of saline sea water
429 from tidal waves and storm surges, and to a lesser extent due to geothermal water ingress into fresh
430 groundwater bodies.

431
432 Of all the salinization processes so far discussed, sea water intrusion is the most widely researched
433 salinization mechanism in Africa. Firstly, sea water intrusion into coastal groundwater results from
434 hydraulic perturbation owing to groundwater pumping from coastal aquifers or because of the action of tides
435 or sea level rise. In coastal environments, with geological time, a fragile hydraulic equilibrium is established
436 between fresh and saline water bodies. Groundwater abstraction can affect this delicate hydrological
437 balance leading to rapid salinization of groundwaters (Vengosh, 2013).

438
439 Because of its diverse geologic setting, the 40,000 km long coast of Africa is comprised of a myriad of
440 aquifer systems (Steyl and Dennis, 2010) with remarkable differences in the state of seawater intrusion.
441 The high groundwater salinity observed in Northern Libya, Egypt, and Tunisia (Figure 2) is partly attributed
442 to this process. Seawater intrusion induced by overexploitation of coastal aquifers has led to a salinity
443 increase in a number of aquifers. In Northeastern Tunisia, as the result of overexploitation of the coastal
444 Ras el Jabel aquifer the range of groundwater salinity has increased from 1.5–4 g/L in 1969 to 5–10 g/L in
445 the 1990s (Khater, 2019). Well fields in coastal southern Kenya show a gradual increase from less than 400
446 $\mu\text{S}/\text{Cm}$ in the 1970s to 1100 $\mu\text{S}/\text{cm}$ in 2017 (Oiro and Comte, 2019). In Djibouti a similar increase in the
447 salinity of coastal aquifers has been observed (Houssein and Jalludin, 1996). Pumping induced salinization
448 is reported in coastal plains of Libya (Al Farrah et al., 2011), Egypt (Eissa et al., 2016), and Tunisia (Ziadi
449 et al., 2019). Seawater intrusion owing to pumping of the coastal aquifers has been noted in a few other
450 coastal aquifers in East Africa (Comte et al., 2016), South Africa, and West African coasts (Hounsino,
451 2020)(Nlend et al., 2018). Saline water encroachment is more pronounced in arid coastal areas of North
452 Africa whereby there is limited recharge to sustain the delicate hydraulic balance between fresh
453 groundwater and saline seawater. However, in more humid coastal regions such as in West Africa,

454 combined seawater intrusion and dissolution of aquifer matrix impact salinization of coastal aquifers as
455 observed in the 1000 km stretch of the coastal regions. In the coastal region of western Africa, the saline
456 groundwaters are often overlain by lenses of fresh groundwaters, and the extent of the saline zone is limited
457 to the coastal strips (Steyl and Dennis, 2010). In some coastal cities (e.g., Abidjan, Douala) salinization of
458 the coastal aquifer is more often related to rock dissolution or pollution than to seawater intrusion
459 (Ramatlapeng et al., 2021) (Nlend et al., 2018).

460

461 Coastal aquifer pumping is not the only process that perturbs the hydraulic gradient leading to salinization
462 of coastal aquifers. The natural seasonal and diurnal variations of sea level and associated tidal fluctuations
463 have an important impact on coastal groundwater level and salinity. Tidal effects, including those
464 exacerbated by storm surges, explain groundwater salinity patterns observed in coastal aquifers in Morocco
465 (Fadili et al., 2018), Gambia (Bagbohouna et al., 2018), and Senegal (Ngom et al., 2016), although there is
466 no agreement on such salinization mechanism in western Africa (Ramatlapeng et al., 2021) (Fantong et al.,
467 2016). The other mechanism of seawater related salinization is through sea spray and dry deposition in
468 coastal environments. Sea spray and dry deposition are reported to have increased groundwater salinity in
469 many locations including in South Africa (van Gend et al., 2021) and West Africa (Osiakwan et al., 2021).

470

471 Salinization of fresh groundwater bodies by encroachment of salt waters also occurs in continental setting
472 whereby saline groundwater leaks up or downwards into fresh groundwater aquifers owing to natural
473 hydraulic gradient or pumping induced perturbations of the hydraulic gradients. The perturbation of the
474 hydraulic equilibrium can cause previously isolated saline groundwaters to mix with fresh groundwater as
475 a result of pumping. Such processes occur predominantly in the North Western Saharan Aquifer system
476 (Tunisia, Libya, Algeria) where intensive groundwater pumping is taking place (Meyer, 2011). In Tunisia,
477 the Miocene sand aquifer is salinized by the capture of saline water from the shallow aquifers (Trabelsi et
478 al., 2009). Pumping induced salinization is evoked in a number of places including, for example, the Merti
479 Aquifer in Kenya (Oord et al., 2014) and, the Djibouti Aquifer in Djibouti (Houssain and Jalludin, 1996).

480

481 **Groundwater-surface water salinity exchange:** In addition to the processes described above, many of
482 which explain the large-scale salinity patterns observed across Africa, localized connectivity between
483 surface water and groundwater has been noted to impact groundwater salinity patterns in many locations.
484 Groundwater found in the proximity of streams bodies typically shows low salinity compared to
485 groundwater bodies at furthest distance from streams. In other instances, particularly in arid and semi-arid
486 environments, it can be the surface water discharge into the riverbed aquifers that drives the salinization.
487 For instance, in Okavango delta, groundwater flow directed from river into adjacent aquifer and subsequent
488 evaporative concentration of the dissolved salt in the aquifer by evapotranspiration accumulate salt in the
489 aquifer leading to localized groundwater salinization adjacent to the streams (Zimmermann et al., 2006).
490 This type of salinity pattern has been observed in a number of locations other locations including along the
491 banks of the Nile River in Sudan (Kebede et al., 2017) and in Omo River Delta region.

492

493 Groundwater discharge to streams explain high salinities in streams waters in several river basins. Saline
494 groundwater discharge into surface waters causes river water salinization in Awash Valley in Ethiopia
495 (Kebede et al., 2021), in Hombolo dam in Tanzania (Shemsanga et al., 2017), in Molo Basin in Kenyan
496 Rift Valley (Chebet et al., 2020). In the Berg River Basin in South Africa, stream water salinization has
497 been noted because of saline groundwater discharge from the underlying geologic formations (Demlie et
498 al., 2011).

499

500 **4. Emerging management strategies in Africa**

501

502 There is a lack of a systematic approach to the management of groundwater salinity in Africa. Existing
503 policies or guidelines, instead of finding technological or management approaches to meet the drinking
504 water quality requirements (e.g. WHO, 2011), often focus on relaxation of the WHO guidelines. For

505 example, in the Comoros islands, where fewer than 30% of the wells provide groundwater of acceptable
506 quality, the local drinking water salinity guideline is usually taken as 3000 mg/L instead of 1000 mg/L as
507 recommended by the World Health Organization (Comte et al., 2016). High concentrations of dissolved
508 solids, which are generally present throughout the Horn of Africa, for example, have necessitated the
509 introduction of higher water quality standards. In Somalia, the electrical conductivity (EC) of up to 3,500
510 $\mu\text{S}/\text{cm}$ is considered safe for human consumption (Idowu and Lasisi, 2020). The republic of Djibouti
511 enacted policies to abandon boreholes that return saline water and drill new boreholes until fresh
512 groundwater is encountered. South Africa implements controlled release of saline mine waters into streams
513 during floods when dilution capacities of the streams are high (Idowu, 2007).

514

515 To overcome the water insecurity risk posed by high groundwater salinity, concerted management strategies
516 are required. The first and the most cost-effective strategy is to identify the scale of the salinity issues by
517 more concerted monitoring. Currently, there are no national guidelines on the management of groundwater
518 salinity in Africa. However, below we discuss a number of technical measures that can be taken to reduce
519 the salinity risks and describe some existing initiatives in operation across Africa.

520

521 **Salinity safe sourcing:** this technique involves finding low salinity groundwaters in otherwise high salinity
522 groundwater environments by mapping salinity micro variations in alluvio-lacustrine sediments or by
523 mapping vertical stratification in salinity in multi-layered aquifer systems. For example, in Malawi, shallow
524 large-diameter wells are shown to be less saline and more durable than deeper small-diameter boreholes
525 (Comte et al., 2016). In multi-layered aquifers it may be possible to find fresh groundwater layers at deeper
526 levels as observed in the Kalahari (Bäumle et al., 2019) and North Western Saharan Aquifer System
527 asserting that salinity does not always increases with depth. Basement and volcanic bedrock aquifers may
528 contain low salinity waters compared to overlying loose alluvio-lacustrine sediments. This may be
529 particularly true in arid environments whereby the deeper aquifers were recharged under different climate

530 conditions that allow fresh groundwaters. Deeper paleo-river channels may also provide low salinity
531 groundwater (Rivett et al., 2020).

532

533 **Pumping management:** There are a number of ways managed pumping of groundwater lowers
534 groundwater salinity. In drylands, intentional groundwater pumping can lead to lowering of the water table
535 preventing progression of dryland salinity (Dogramaci, 2004). In low lying coastal environments where a
536 thin lens of fresh groundwater rests on deeper saline water, controlled pumping of the deeper saline water
537 can lead to deepening of the interface between the fresh and salt water. The newly created space can be
538 used to as a storage space in the shallow aquifer (Mainuddin et al., 2021). Despite various models and field
539 practice based demonstrations in other parts of the world (Stein et al., 2019)(Mainuddin et al., 2021) the
540 use of groundwater pumping to manage salinity is not widely reported in Africa. One exception may be
541 the exploitation of the shallow fresh water in the Djibouti aquifer by limiting pumping from the shallow
542 fresh water aquifer so as to limit the encroachment from deeper salt water (Houssein and Jalludin, 1996).

543

544 **Managed Aquifer Recharge:** Managed aquifer recharge to augment water supply and simultaneously
545 preventing seawater intrusion in coastal aquifers is already practiced in South Africa. One such example is
546 the Atlantis scheme, which separates the domestic and industrial wastewater and storm water runoff based
547 on salinity, recharges the low salinity water into the inland aquifer used for domestic supply, and recharges
548 the more brackish water into the coastal aquifer to prevent seawater intrusion (Horriche and Benabdallah,
549 2020). Managed aquifer recharge is a proven approach in reducing the salinity of coastal aquifers and in
550 overcoming the challenges of seawater intrusion in a few places such as in Tunisia, (Horriche and
551 Benabdallah, 2020), Namibia (Tredoux et al., 2009), and South Africa (Bugan et al., 2016). As well as the
552 large scale managed aquifer recharge, salinity management is also being done on a smaller scale through
553 rainwater harvesting and river recharge schemes in dryland systems in South Africa and small island aquifer
554 systems (Hohne et al., 2021).

555

556 **Inter-basin water transfer:** This involves the transfer of water from water secure environments to water
557 insecure environments using artificial infrastructure such as channel and pipes but these are often
558 controversial schemes. The Great Man Made River in Libya to channel the fresh groundwater from the
559 Kufra (Nubian sandstone) in the south to the salinity affected region in the coastal plains is an example.
560 Several projects of inter basin groundwater transfer have been initiated in Africa but the sustainability of
561 the infrastructure and the financial feasibility of such projects often present insurmountable barriers to
562 implementation (Luedeling et al., 2015)(Oord et al., 2014)(Kebede and Taye, 2021). Nevertheless, inter
563 basin transfer is practiced or has been initiated in a few locations in Africa, including the transfer of the
564 Merti Aquifer waters to supply the City of Wajir (Luedeling et al., 2015). The transfer of lake water from
565 the Lac de Guier in the North of Senegal to supply Dakar is another example although it is not a groundwater
566 transfer. The transfer was needed because of the dire water quality conditions that have arisen locally in
567 Dakar in part due to pumping induced salinization (Re et al., 2011). Although published literature is lacking,
568 there are dozens of long-distance rural water transfer schemes in Ethiopia meant to address water salinity
569 challenges and known to the authors.

570

571

572 **5. Conclusions and future outlook**

573

574 Our analysis demonstrates that groundwater salinity across Africa is generally low ($<2000 \mu\text{S}/\text{cm}$) and is
575 suitable for irrigation and drinking water uses. The major sedimentary basins which account for 80% of the
576 groundwater storage contain predominantly fresh groundwater resources with high salinities localized to
577 specific zones and layers. The basement aquifers, which supplies 70% of Africa's rural population with
578 drinking water sources, contain predominantly fresh groundwater resources ($<500 \mu\text{S}/\text{cm}$) with high
579 salinities ($> 2000 \text{uS}/\text{cm}$) restricted to certain lithologies or minor alluvial sediments overlying the basement

580 aquifers. Groundwater salinity is extremely variable in sediments of paleo hydrological systems that contain
581 inter-fingering of fresh and saline groundwater bodies.

582

583 Groundwater salinity strongly correlates with groundwater recharge. Groundwater salinity increases with
584 the degree of aridity, with higher salinities found in hyper-arid environments as compared to humid
585 environments. Regions with high groundwater recharge generally have lower salinities than regions with
586 low groundwater recharge. Conversely, in regions with high groundwater storage, for example in arid
587 environments groundwater salinity is generally higher. The consideration of groundwater salinity as a water
588 security indicator will not alter the groundwater security status of the countries in humid high recharge
589 environments. Nigeria, Guinea Bissau, Congo, Angola, Democratic Republic of Congo Burundi, Liberia,
590 Guinea, Equatorial Guinea, Ivory Coast remain groundwater secure even when groundwater salinity is
591 considered. On the other hand, groundwater salinity adds to groundwater insecurity in arid countries with
592 low groundwater recharge including Namibia, Botswana, Somalia, Egypt, Libya, Tunisia, Morocco,
593 Senegal, and Algeria.

594

595 Since data on very high salinity groundwaters are not often the target of groundwater development or use,
596 the dataset compiled for this study may be biased towards low salinity groundwaters. The dataset lacks
597 depth information, therefore, a concerted effort is needed to produce quantitative 3D maps of groundwaters
598 based on information from boreholes as well as other tools such as geophysics, or other sources of data
599 including those from commercial drinks companies, petroleum/oil wells, and abandoned wells. Concerted
600 effort must be made to quantify, delineate and monitor salinization of groundwaters in Africa. A significant
601 challenge however remains the paucity of accessible data, particularly in the vast humid regions such as in
602 the Congo basin and Central Africa. Organizations that hold groundwater quality data are often located
603 outside Africa. Groundwater quality data is dispersed across local and international institutions that do not
604 often have data sharing protocols. Accessibility of such datasets would further enhance our knowledge of
605 the scale of groundwater salinization and its drivers. Our study provides the basis for planning future

606 groundwater salinity monitoring plans and design management strategies, and formulating policies at the
607 continental level.

608

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613

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