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Using ecological and field survey data to establish a national list of the wild bee pollinators of crops.

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Abstract

The importance of wild bees for crop pollination is well established, but less is known about which species contribute to service delivery to inform agricultural management, monitoring and conservation. Using sites in Great Britain as a case study, we use a novel qualitative approach combining ecological information and field survey data to establish a national list of crop pollinating bees for four economically important crops (apple, field bean, oilseed rape and strawberry). A traits data base was used to establish potential pollinators, and combined with field data to identify both dominant crop flower visiting bee species and other species that could be important crop pollinators, but which are not presently sampled in large numbers on crops flowers. Whilst we found evidence that a small number of common, generalist species make a disproportionate contribution to flower visits, many more species were identified as potential pollinators, including rare and specialist species. Furthermore, we found evidence of substantial variation in the bee communities of different crops. Establishing a national list of crop pollinators is important for practitioners and policy makers, allowing targeted management approaches for improved ecosystem services, conservation and species monitoring. Data can be used to make recommendations about how pollinator diversity could be promoted in agricultural landscapes. Our results suggest agri-environment schemes need to support a higher diversity of species than at present, notably of solitary bees. Management would also benefit from targeting specific species to enhance crop pollination services to particular crops. Whilst our study is focused upon Great Britain, our methodology can easily be applied to other countries, crops and groups of pollinating insects.

Key-words

Agri-environment Schemes, Apple, Biodiversity, Crop pollination, Dominant Pollinators, Ecosystem Services, Field Bean, Oilseed Rape, Rare Species, Strawberry.

1. Introduction

Insect pollination is key to global agricultural productivity (IPBES, 2016) due to growing demand for entomophilous crops (Godfray and Garnett 2014). The nutritional and economic importance of insect pollinated crops (Vanbergen et al., 2014), and the inability of managed pollinators (e.g.*, Apis mellifera*) to meet service demand, mean agriculture is highly dependent upon wild pollinators (Aizen and Harder 2009; Breeze et al., 2014). Yet conventional agricultural practices are a key driver of pollinator declines (Senapathi et al., 2015). Whilst agri-environment scheme options have had positive impacts (Tonietto et al., 2018), most benefit a limited suite of common species (Scheper et al., 2013) and homogeneous communities provide less reliable pollination services (Grab et al., 2019). Currently agri-environment schemes tend preferentially to benefit bumblebee populations (Wood et al., 2015a; Wood et al., 2015b, 2016a, b), yet solitary bee species are more important pollinators of some crops (Woodcock et al., 2013). As such, current agri-environment schemes may not be optimally designed to increase pollination services to many crops. Identifying key pollinating species to individual crops, and ones which may provide additional pollination and insurance against declines in other species, would help inform agricultural management for bee pollinators (Garratt et al., 2014a). Yet there is insufficient information on bee communities for many crops (Kremen and Chaplin-Kramer, 2007) and no studies have attempted to establish a ‘national list’ of crop pollinators to advise management or monitoring programmes.

Whilst the majority of crop flower visitation is attributed to a small proportion of bee species (Kleijn et al., 2015), species-rich communities have been shown to positively influence crop yields and pollination service stability (Hoehn et al., 2008; Garibaldi et al., 2011; Martins et al., 2015; Dainese et al., 2019; Woodcock et al., 2019). Biodiversity conservation and ecosystem service management are often seen as distinct objectives (Sutter et al., 2017), however management that only targets common crop pollinators will not safeguard production if it fails to encompass species that supplement service provision (Fijen et al., 2018). High species turnover means that diverse communities, including rare and specialist species, are required to maintain crop pollination service at regional scales (Winfree et al., 2018). With climate change reducing the occupancy and richness of some wild bee species (Soroye et al., 2020), supporting wider species diversity may be crucial for crop pollination service stability under the substantial future environmental change that is predicted (Oliver et al., 2015; Dainese et al., 2019). Additionally, different crops have distinct pollinator communities and it will be beneficial to identify the pollinating taxa of individual crops and target management accordingly (Garratt et al., 2014a). Furthermore, a national list of crop pollinators can inform monitoring schemes to ensure they include important crop pollinating species (Carvell et al., 2017; Garratt et al., 2019).

In order to inform pollinator management and monitoring, our study aimed to compile the bee species visiting four crops: apple (*Malus domestica*), field bean (*Vicia faba*), oilseed rape (*Brassica napus*) and strawberry (*Fragaria x ananassa*). Insect pollination has been shown to enhance yield quantity and quality in all four crops (Bartomeus et al., 2014; Garratt et al., 2014b). Additionally, they differ in flower phenology and morphology (Garibaldi et al., 2015) and likely show corresponding differences in their pollinator community composition (Garratt et al., 2014a). We use sites in Great Britain as a case study because its bee fauna is comprehensively described and their occupancy is well recorded over a long time period (Powney et al. 2019). We compiled a list of all British bee species and their available physiological and ecological traits, and combined it with field survey data in order to devise an approach to generate lists of (i) definite flower visitors to each crop (ii) likely flower visitors, which are expected to also contribute to crop pollination (iii) possible crop flower visitors whose contribution to pollination is not well understood and merits further investigation. Our aim was to compile these lists for reference purposes, but not to statistically compare pollinator communities between crops, due to the unstandardised nature of the datasets used to generate the lists of bee species. Additionally, we identify dominant crop pollinating species, and asses the contribution of wild bees compared to honey bees for crop flower visitation.

2. Materials and Methods

2.1 Potential crop pollinators.

First, a species database of all extant, resident wild bee species in Great Britain was established using the most recent checklist of UK species (Else et al., 2016). For each species, data on the following were collated: flight period (months); sociality (cleptoparasite, eusocial or solitary); lecty (oligolectic or polylectic, including if any of the target crop plant families are visited for pollen and/or nectar), tongue length (short/long), geographic coverage (distribution and habitat) (based on trait information compiled by Stuart Roberts for the EU- FP6 ALARM-project and BWARS, 2020) and conservation status (Webb et al., 2018). Potential crop pollinators, as defined here, are those bee species which, based upon these ecological traits, such as flight period, lecty, sociality and tongue length, could pollinate our target crops. Habitat specialists that are not coincident with cropland were initially excluded i.e., primarily coastal, heathland species. The known floral ecology of each species was then used to refine lists for each crop. Cleptoparasitic species, species that are oligolectic on plant families other than the target crop or polylectic, but not documented as foraging on the relevant plant family for pollen or nectar and species whose flight period does not overlap with the relevant crops flowering period were excluded. For field bean, only 'long-tongued' species (Michener, 2000) were considered as its flowers have deep corollas and most visits by 'short-tongued' species involve nectar robbing rather than legitimate visitation (Garratt et al., 2014a).

2.2 Field survey data

Field studies were sourced through literature searches in google scholar and existing datasets held by the authors. Fifty-seven datasets from across England, Scotland and eight other European countries were available to combine with the potential crop pollinator lists in order to establish shortlists of crop flower visitors (Figure 1 and Table S3).

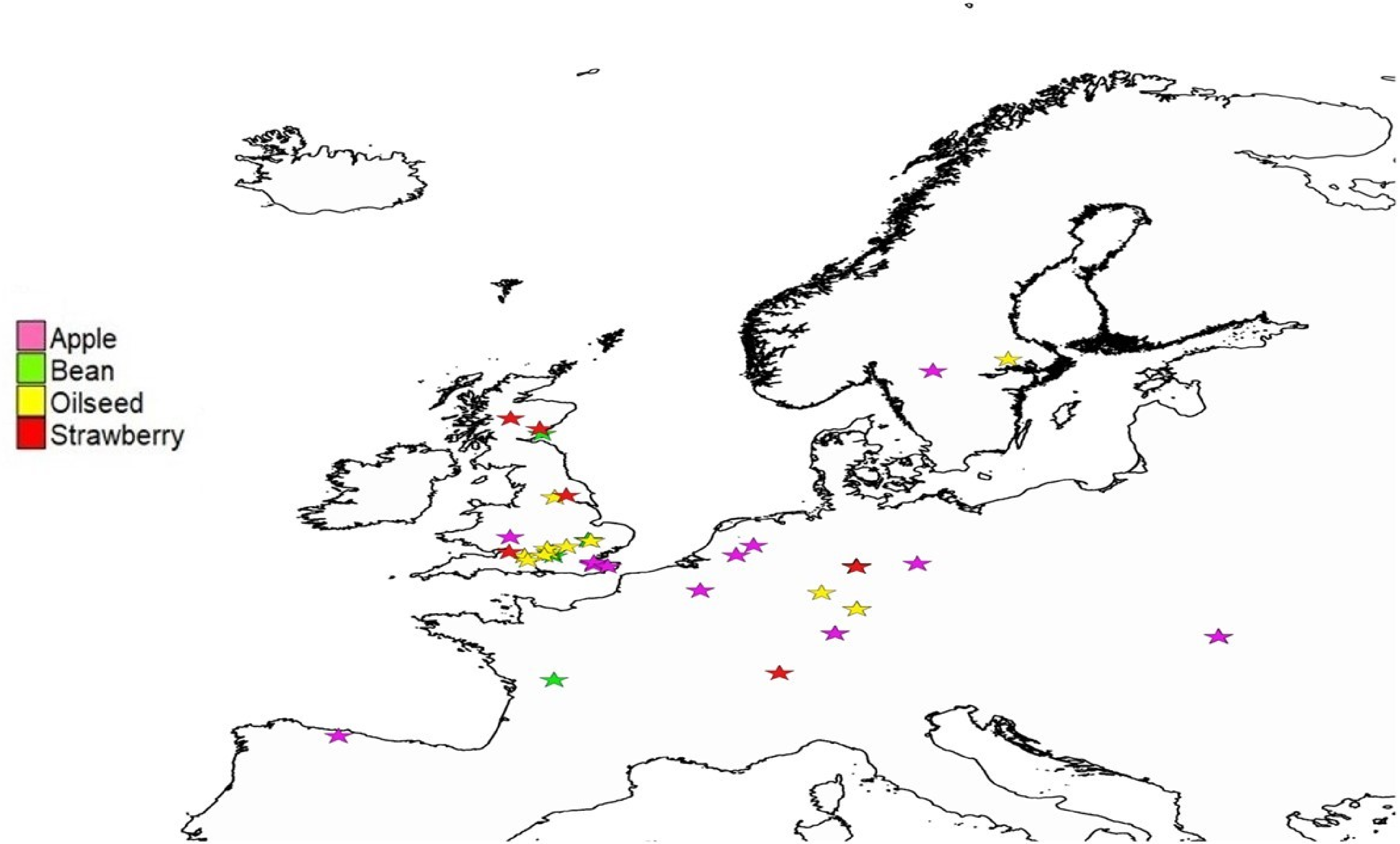


Figure 1: Map of Europe, showing the countries from which field studies were sourced for each crop.

Lists of bee species recorded in crop fields were compiled using three types of survey data:

i) British flower visitation studies (e.g. transect walks, observation plots).

ii) British pan trap studies in crop fields.

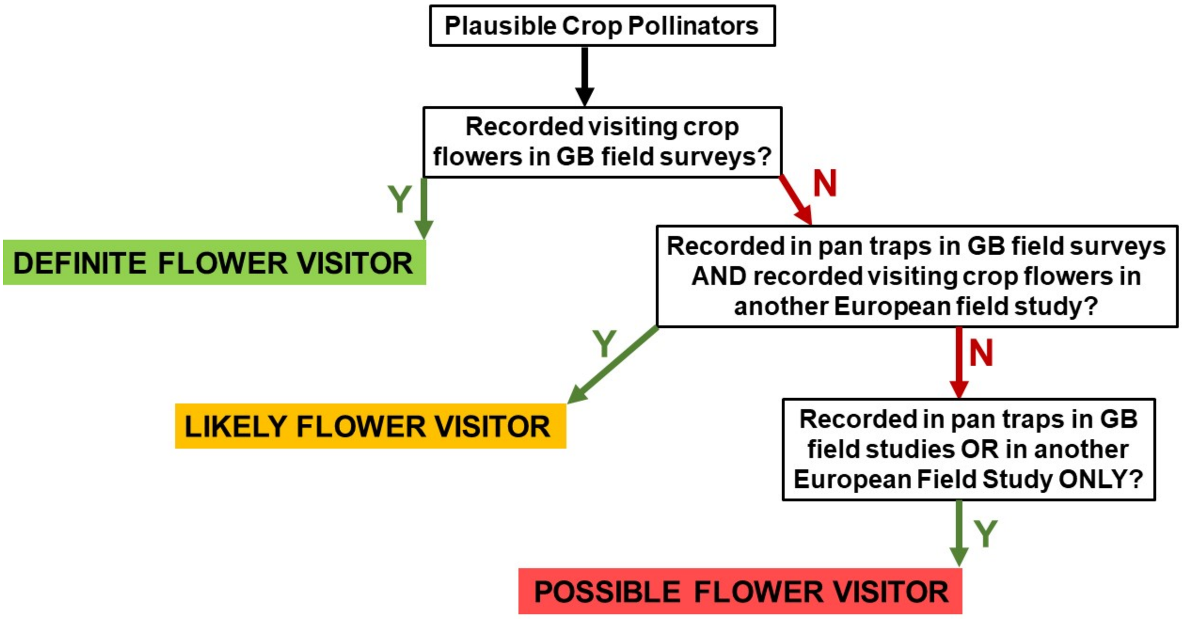
iii) Other European flower visitation studies (used to validate crop flower visitation for species sampled in British pan traps only).

For every bee species the total number of reported legitimate flower visits and number of studies recorded in were calculated for each crop. If studies did not include quantitative data then a conservative approach was taken whereby each bee species listed was taken as representing a single crop flower visit. As pan trap catches do not provide information on floral associations (Westphal et al., 2008), these data were only used, in combination with trait data, to generate the list of possible pollinators.

2.3 Crop flower visitors

The lists of potential crop pollinators were combined with the field survey data to categorize bee species into one of three flower visitor categories (Figure 2; Full details in Supplementary Methods 1):

i) Definite Flower Visitors – Species recorded visiting crop flowers in British flower visitation studies.
ii) Likely Flower Visitors - Species recorded in British pan trap crop studies and recorded as making at least two flower visits in other European studies.
iii) Possible Flower Visitors - Species only recorded in British pan trap studies, or in other European flower visitor studies only, and classified as a potential crop flower visitor.



2.4 Dominant crop flowers visitors

As visitation rate to crop flowers is a good proxy of relative contribution to pollination service delivery (Vazquez et al., 2005), we identified the dominant British flower visiting bee species per crop by approximating the species attributed with a combined total of 80% of flower visits, the proportion identified as corresponding to the dominant flower visitors by Kleijn et al. (2015). Only British flower visitation datasets where bee species were either all identified to species or genus were included in the analysis (Supplementary Methods 2). Additionally, we calculated the average proportion of visits to crop flowers attributed to wild bees compared to honey bees for all crops (Supplementary Methods 2).

3. Results

3.1 Potential crop pollinators

A preliminary list of 229 extant, resident British wild bee species was compiled. Of those 132 species were excluded due to ecological and lecty traits that were deemed incompatible with these bees being present in crop fields and/or crop flower visitors (Table S1). Four species were treated as an aggregate – *Bombus terrestris* aggregate – due to the difficulties of separating their workers in the field (Wolf et al., 2010; Bossert, 2015). Therefore, a total of 97 species were initially identified as potential crop pollinators. Accounting for their documented foraging ecology and flight period, the following number of species were considered as potential pollinators per crop: apple- 83, bean- 30, oilseed- 60, and strawberry – 90 (Table S2).

3.2 Field survey data

The total number of studies sourced per crop were as follows: apple – 17; bean – 10; oilseed – 19; strawberry – 11. The number of studies per survey type for each crop is provided in Figure S1.

3.3 Crop flower visitors

Seventy-three species from ten genera where categorised as flower visitors of one or more crops, 63 of which were recorded in British crop field studies (Table 1, Figure 3). Fourteen species were included in flower visitor categories that were not initially identified as potential crop pollinators. Ten of those were widely polylectic *Bombus* or *Lasioglossum* species, all recorded in oilseed datasets, but not documented in the literature as foraging on *Brassicaceae*. The remaining species were three short-tongued *Andrena* species recorded visiting bean flowers, two of which are oligolectic on Fabaceae and a *Colletes* species, recorded in a single strawberry dataset, that is documented as being oligolectic on another plant family. The majority of species identified as potential pollinators, but not recorded in crop field surveys were either rare species or polylectic species documented as having distinct preferences for plant families other than the target crop. The remaining species were overwhelmingly smaller species from the genera *Hylaeus* and *Lasioglossum* or cavity nesting *Megachilidae*. Most species identified as crop flower visitors were geographically widespread (BWARS, 2020) and polylectic species. However, a quarter (n=18) of species included in flower visitor categories, currently have a designated conservation status in Britain. Full details of all species in crop flower visitor categories are given in tables S4a-d and S5a – S8d.

Table 1: Number of bee species, based upon field datasets and trait information, that were assigned to each category of flower visitor per crop

	Flower Visitor Category			
Crop	**Definite**	**Likely**	**Possible**	**Total**
Apple	19	13	25	57
Field Bean	11	0	3	14
Oilseed Rape	37	11	3	51
Strawberry	9	6	18	33

Apple

All five British apple flower visitor studies recorded every bee to species level. *Andrena* were the most speciose genus of flower visitor, both overall (n=22) and in the definite flower visitor category (n=10). *Bombus* species were the next most commonly represented genus in the latter category (n=6), but were less frequent overall (n=9) than *Lasioglossum* species (n=16). Within the definite flower visitor category 80% of flower visits were attributed to eight species, only half of which were recorded in all studies. Most likely and possible flower visitors were *Andrena* or *Lasioglossum* species.

Bean

Three of the five British bean flower visitor studies recorded all bee to species level, the remainder only recorded *Bombus* to species, which was both the most common genus overall (n=9) and in the definite flower visitor category (n=7). Three short-tongued *Andrena* sp. were identified as definite flower visitors, but all were recorded as very low numbers of flower visits

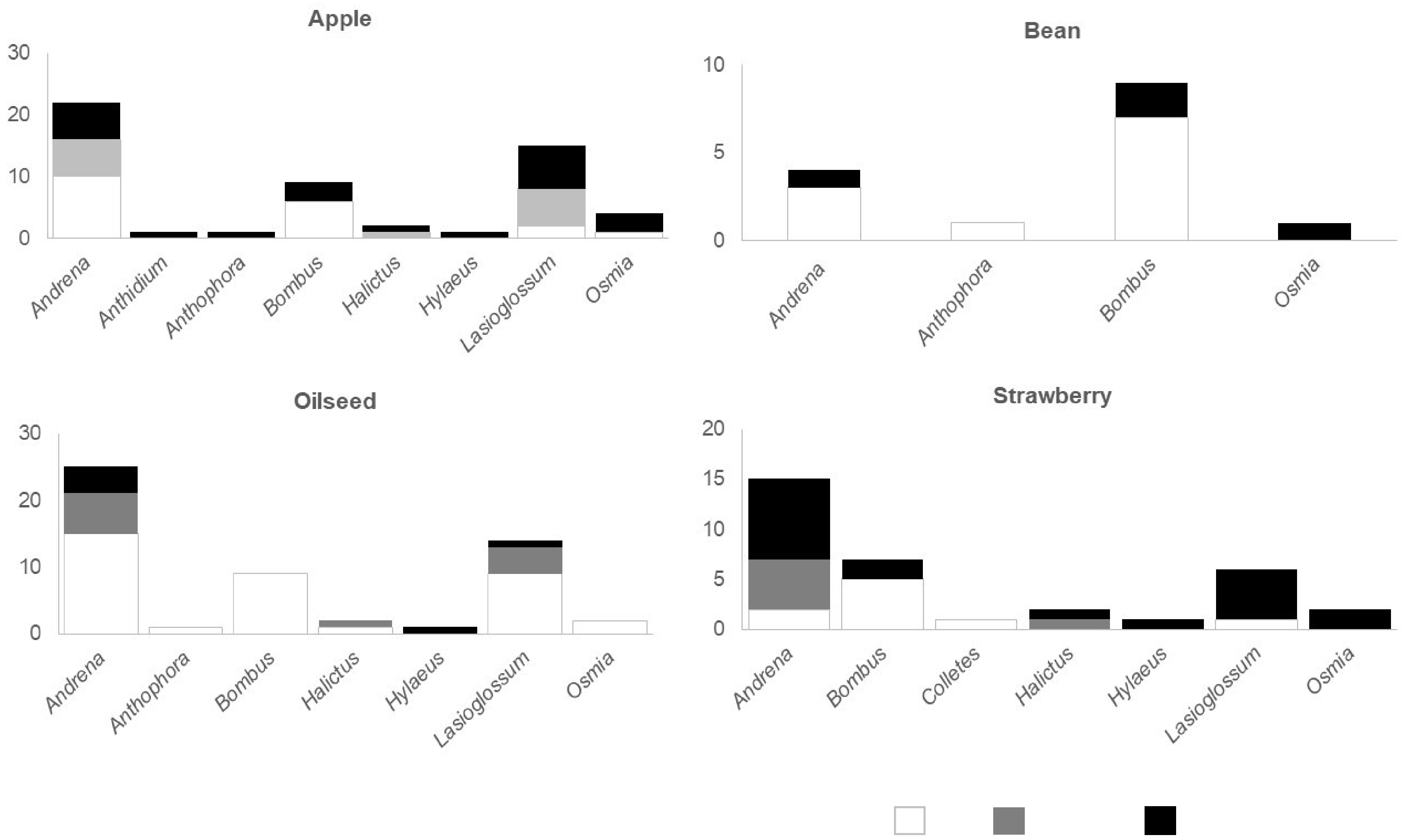
(≤10). Four *Bombus* species and *Anthophora plumipes* accounted for 95% of all visits recorded in British flower visitation studies. However, all the *A. plumipes* records derived from one study (Bond and Kirby, 1999) carried out at a single site. The four *Bombus* were the only species recorded in four or more studies. No species met the criteria for the likely flower visitor category. The possible flower visitor category included two *Bombus* and one *Osmia* species.

Oilseed

Six of the nine British oilseed flower visitor studies recorded bees to species level, but only two included quantitative data on all bee species. *Andrena* was the most speciose genus of bee, both overall (n=27) and within the definite flower visitor category (n=15). *Bombus* and *Lasioglossum* species were equally represented in the definite flower visitor category (n=9), but *Lasioglossum* were more frequent overall (n=14). Within the definite flower visitor category 80% of recorded flower visits were attributed to six species, only two of which were recorded in all nine studies, with the remainder only recorded in between five and eight studies, despite all being large *Andrena* or *Bombus* species, generally identified and quantified in all field studies. The likely and possible visitor categories were entirely comprised of *Andrena* or *Halictidae* species, two of which are oligolectic on *Brassicaceae.*

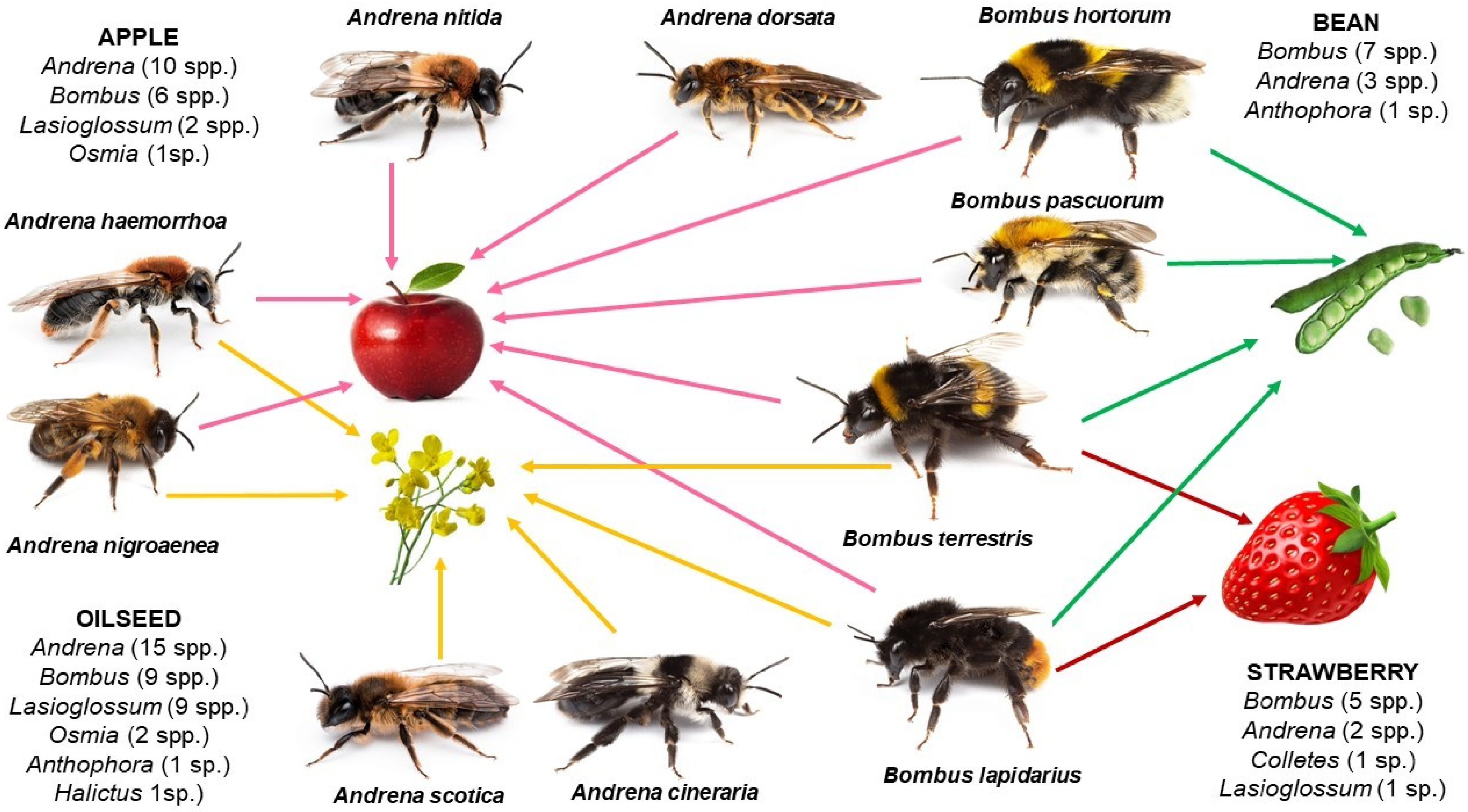
Strawberry

Two British strawberry flower visitor studies recorded all bees to species level. The remaining three only recorded a group of large *Andrena* and *Bombus* to species. *Bombus* species were the most common genus of bee within the definite flower visitor category (n=5), but joint second as the most frequent genus overall, alongside *Lasioglossum* (n=7), with *Andrena* species being the most prevalent genus across all categories (n=14). Within the definite flower visitor category 80% of recorded flower visits were attributed to just two *Bombus* species, which along with two other *Bombus*, were the only species recorded in more than two studies. The likely visitor category was almost exclusively represented by *Andrena* species. The possible visitor category was largely comprised of solitary bees from five different genera.


Apple
30
20
10
0
Andrena
Anthidium
Anthophora
Bombus
Halictus
Hylaeus
Lasioglossum
Osmia
Bean
10
5
0
Andrena
Anthophora
Bombus
Osmia
Oilseed
30
20
10
0
Andrena
Anthophora
Bombus
Halictus
Hylaeus
Lasioglossum
Osmia
Strawberry
20
15
10
5
0
Andrena
Bombus
Colletes
Halictus
Hylaeus
Lasioglossum
Osmia

3.4 Dominant crop flower visitors

Ten bee species were attributed with 80% of flower visits across the four crops (Figure S2; Figure 4). There were differences however in the number and composition of those species making up the 80% of flower visits on a per crop basis. Differences in crop communities were even more distinct when considering the entire suite of bee species included in the characterisation of each crops' total flower visiting community (Figure 3; Figure 4). Wild bees were attributed with an average of between 63 and 83 percent of crop flower visits compared to honey bees (Apple: solitary bee visits = 68%; Bean: solitary bee visits = 83%; Oilseed: solitary bee visits = 63%; Strawberry: solitary bee = 77%).


APPLE
Andrena (10 spp.)
Bombus (6 spp.)
Lasioglossum (2 spp.)
Osmia (1sp.)
Andrena nitida
Andrena dorsata
Bombus hortorum
BEAN
Bombus (7 spp.)
Andrena (3 spp.)
Anthophora (1 sp.)
Andrena haemorrhoa
Bombus pascuorum
Andrena nigroaenea
Bombus terrestris
OILSEED
Andrena (15 spp.)
Bombus (9 spp.)
Lasioglossum (9 spp.)
Osmia (2 spp.)
Anthophora (1 sp.)
Halictus 1sp.)
Andrena scotica
Andrena cineraria
Bombus lapidarius
STRAWBERRY
Bombus (5 spp.)
Andrena (2 spp.)
Colletes (1 sp.)
Lasioglossum (1 sp.)

4. Discussion

4.1 Crop pollinator species

Our study is the one of the first to evaluate the entire wild bee community of multiple crops on a national basis and can be used as model approach for other countries, crops and pollinators. With the identification of bee species important for pollinating crops we build the basis to better sustainably manage services with changing climate and land use. Whilst in accordance with other studies (Rader et al., 2012; Kleijn et al., 2015) our results indicate that a small proportion of common, generalist bee species do make the majority of crop flower visits, many more species were evidenced as crop flower visitors. Additionally, our results suggest that the contribution of wild bee species to crop flower visitation may be even greater than previously thought. Whereas previous estimates indicate that wild bees make a similar overall contribution to honey bees (Kleijn et al. 2015), when considering the entire suite of flower visiting species our results indicate that wild bees make on average between 63 and 83% of flower visits to our target crops. Given the benefits of biodiverse communities for current and future crop pollination services (Kremen et al., 2002; Hoehn et al., 2008; Garibaldi et al., 2011; Rader et al., 2012), interventions to support crop pollinators should target a more significant proportion of the bee fauna than at present (Wood et al., 2015b, 2016a; Gresty et al., 2018). Establishing a list of currently important, but also potentially relevant crop pollinators, is necessary to help target monitoring and conservation (Carvell et al., 2017).

Our results also support prior evidence of distinct differences in individual crop pollinator communities (Garratt et al., 2014a). The overwhelming majority of field bean and strawberry flower visits were attributed to bumblebees. However, whereas field bean was visited by the three longest tongued species in Britain, strawberry crops were almost exclusively visited by two other bumblebee species, with relatively shorter tongues. This supports a link between trait matching of bees and flowers in crop pollination (Garibaldi et al., 2015). *Bombus* species were also recorded visiting apple and oilseed rape. However, due to their low abundance in early spring during apple flowering (Martins et al., 2015), and lower rate of pollen transfer when visiting oilseed flowers (Woodcock et al., 2013) they are less important pollinators of these crops compared to solitary species. *Andrena* and *Lasioglossum* species were prevalent across both apple and oilseed flower visitor categories. *Andrena* are known to be highly efficient pollinators of both crops (Martins et al., 2015; Woodcock et al., 2013), especially apple (Russo et al., 2017). Most *Lasioglossum,* species however*,* generally emerge later than many *Andrena* species, and peak after apple flowering, whereas oilseed tends to flower later and longer, and *Lasioglossum* are likely important pollinators of this crop (Perrot et al., 2018; Catarino et al., 2019). Furthermore, we almost certainly significantly underestimated the diversity and abundance of *Lasioglossum* bees visiting oilseed rape, given that many studies did not include detailed quantitative data on this genus.

Our datasets also indicate that rare and specialist species may visit crop flowers when they are locally abundant or are especially attracted to crop flowers (MacLeod et al., 2020). Several rare species recorded in apple orchards are most common in south-east England, Britain's principal apple growing region, and bee species that are oligolectic on Brassicaceae were recorded in oilseed rape studies. Given that biodiversity benefits pollination (Dainese et al., 2019), strategies to support biodiverse crop communities may prove critical to sustain ecosystem service provision. Yet current agri-environment schemes options rarely consider rare species (Senapathi et al., 2015). There is however, a significant overlap in the floral resources used by common and rare crop pollinators (Sutter et al., 2017; MacLeod et al., 2020), and thus there are opportunities to promote both biodiversity and conservation in agricultural landscapes.

Our findings also offer an opportunity to anticipate potentially important future crop pollinators. For example, whilst a number of European crop flower visitors not presently recorded in British crop fields are currently geographically restricted, should they expand their range in the future, they could ameliorate the threat of ecological mismatches between current pollinators and crops due to climate change (Polce et al., 2013; Polce et al., 2014; Settele et al., 2016). Taken further, this information could be used to refine existing models of bee populations used to project pollinator populations at large spatial scales (e.g. Gardner et al., 2020), which can assist in larger scale planning of pollinator management.

Identifying specific bee crop pollinating species, as we have done, can inform refinements to agri-environment schemes to promote more biodiverse communities in agricultural landscapes. For example, *Andrena* were the most speciose genus of bees identified across flower visitor categories in three of the crops. Currently European agri-environment measures to boost pollinator populations have focused on the creation of flower-rich habitats, including wildflower buffer strips (Wratten et al., 2012). Yet evidence suggests these are primarily visited by bumblebees, with solitary bees preferring non-sown, wild plants (Wood et al., 2015). In apple orchards for example, early-flying *Andrena* species have been positively associated with dandelions (Taraxacum agg.) rather than sown species, which often bloom later than apple flowers (Campbell et al., 2017). Reduced mowing regimes in orchards, and other crop areas, particularly in early spring could boost *Andrena* numbers and hence pollination. Such interventions are also likely to benefit early flying *Lasioglossum*, many species of which are known be attracted to yellow flowers in the family *Asteraceae*. *Osmia* species have also been demonstrated as efficient pollinators of apple, oilseed and strawberry crops (Abel et al., 2003; Garratt et al., 2016; Horth and Campbell, 2018), but as in this study, are frequently recorded in low numbers, likely due to a lack of suitable nesting and floral resources in agricultural landscapes for cavity nesting species (Gardner and Ascher, 2006; Blitzer et al., 2016). Incorporating hedgerow species such as Dog Rose and Bramble, alongside, areas of old and dead wood, around crop areas would provide both forage and nesting resources (Else and Edwards 2018; Gresty et al., 2018) for these and other cavity nesting bees. Future management to support long-tongued solitary bees could benefit field bean pollination. *Anthophora plumipes*, for example, prefers to nest in vertical soil profiles, which are not currently a common feature in agricultural landscapes.

4.2 Data constraints and limitations

There are caveats to using foraging ecology to identify potential bee pollinators, as done here and elsewhere (Ahrenfeldt et al., 2015). There is a lack of published data for many bee species and others visit a wider range of flowers than can be realistically documented (Else and Edwards, 2018). As such, determining the status of bee species as crop flower visitors requires field survey data for confirmation. Yet comprehensive crop pollinator data is currently lacking as sampling is irregular, undertaken almost exclusively as part of bespoke research projects rather than systematic monitoring (Breeze et al., 2020). Furthermore, whilst census methods can provide information on floral associations, they require experienced surveyors to comprehensively record species richness (O'Connor et al., 2019). Across all four crops the only bees which were consistently identified to species level were large, conspicuous ones from the genera *Bombus* and *Andrena*. Small and inconspicuous species, particularly from the genus *Lasioglossum*, were often only extensively sampled in the pan trap surveys. Additionally, whilst the visitation rate of dominant species is strongly correlated to pollination service delivery (Winfree et al., 2015; Fijen et al., 2018), the assumption here and elsewhere that quantitative visitation data can be used to infer pollination (Kleijn et al., 2015), neglects to factor in that flower visitation alone is not a perfect proxy for pollination (King et al., 2013; Senapathi et al., 2015; Ollerton, 2017). Certain physiological and behavioural traits also influence pollination service delivery (Martins et al., 2015). Further detailed data and research is required before any definitive conclusions can be made about the contributions of individual bee species to crop pollination.

5. Conclusions

Given the importance of wild pollinators and the detrimental impacts of conventional agriculture on their populations it is unsurprising that the management of wild and managed pollinating insects is considered a critical step for future food security (Garibaldi et al., 2019; Kleijn et al., 2019; Rollin and Garibaldi et al., 2019; Reilly et al., 2020). Yet information on which species contribute most to ecosystem service delivery has long been elusive (Kremen and Chaplin-Kramer, 2007) despite its critical importance for both monitoring and conservation measures. Here we combine ecological and field data to provide a uniquely comprehensive overview of the crop pollinating bees of a single region, Great Britain. Whilst we have focused on Great Britain, a similar approach would be applicable across Europe, and could also be applied to non-bee species that have been identified as important crop pollinators (Rader et al., 2016). Our research bolsters evidence that many wild bee species, including rare and specialised ones, may contribute to crop pollination (Klein et al., 2003; Sutter et al., 2017; Winfree et al., 2018; MacLeod et al., 2020), thus it can be argued that agri-environment scheme options should not focus solely on dominant crop pollinators.

Future climatic changes threaten to further deplete already impoverished bee populations (Soroye et al., 2020) and create spatial mismatches between crops and their pollinators, which could exacerbate existing pollination deficits (Polce et al., 2014). To that end, the species identified as possible crop pollinators could represent an as yet untapped pollinator resource. Whilst some species may not currently visit crops due to ecological or environmental constraints, they could be assisted to expand by dedicated conservation measures in agricultural landscapes, allowing them to compensate for any declines in current crop pollinating species. Many such species are solitary, which presently benefit much less from agri-environment schemes than social species (Wood et al., 2015b, 2016a, 2016b; Gresty et al., 2018). As such land managers may need to re-evaluate existing pollinator management interventions and consider a broader range of species to safeguard the ecosystem service of crop pollination in an uncertain future.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Authors' contributions

LH conceived the ideas, analysed the data and wrote the manuscript. MG, TB and TO contributed to the conceptual development and manuscript revisions. All other authors provided data and contributed to manuscript revisions.

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Figure 4 – Bee photographs courtesy of Nicolas J. Vereecken and Stéphane De Greef.

References

Abel, C.A., Wilson, R.L. and Luhman, R.L., 2003. Pollinating efficacy of Osmia cornifrons and Osmia lignaria subsp. lignaria (Hymenoptera: Megachilidae) on three Brassicaceae species grown under field cages. Journal of Entomological Science, 38(4), 545-552.

Ahrenfeldt, E.J., Klatt, B.K., Arildsen, J., Trandem, N., Andersson, G.K.S., Tscharntke, T., … and Sigsgaard, L., 2015. Pollinator communities in strawberry crops–variation at multiple spatial scales. Bulletin of Entomological Research, 105(4), 497-506.

Aizen, M. A. and Harder, L. D., 2009. The global stock of domesticated honey bees is growing slower than agricultural demand for pollination. Current Biology, 19, 915–918.

Bartomeus, I., Potts, S.G., Steffan-Dewenter, I., Vaissiere, B.E., Woyciechowski, M., Krewenka, K.M., … and Bommarco, R., 2014. Contribution of insect pollinators to crop yield and quality varies with agricultural intensification. PeerJ, 2, 328.

Blitzer, E.J., Gibbs, J., Park, M.G. and Danforth, B.N., 2016. Pollination services for apple are dependent on diverse wild bee communities. Agriculture, Ecosystems & Environment, 221, 1-7.

Bond, D.A. and Kirby, E.J.M., 1999. Anthophora plumipes (Hymenoptera: Anthophoridae) as a pollinator of broad bean (Vicia faba major). Journal of Apicultural Research, 38(3-4), 199-203.

Bossert, S., 2015. Recognition and identification of species in the Bombus lucorum-complex-A review and outlook. bioRxiv, 011379.

Breeze, T.D., Vaissière, B.E., Bommarco, R., Petanidou, T., Seraphides, N., Kozák, L., … and Moretti, M., 2014. Agricultural policies exacerbate honeybee pollination service supply-demand mismatches across Europe. PloS one, e82996.

Breeze T.D., Bailey A.P., Balcombe K.G., Brereton T., Comont R., Edwards M., … and Carvell C., 2020. Pollinator Monitoring More than Pays for Itself. Journal of Applied Ecology, In Press

BWARS., 2020. Bees, Wasps & Ants Recording Society. https://www.bwars.com/home

Campbell, A. J., Wilby, A., Sutton, P. and Wäckers, F. L., 2017. Do sown flower strips boost wild pollinator abundance and pollination services in a spring-flowering crop? A case study from UK cider apple orchards. Agriculture, Ecosystems & Environment, 239, 20-29.

Carvell, C., Isaac, N., Jitlal, M., Peyton, J., Powney, G., Roy, D., … and Roy, H., 2017. Design and testing of a national pollinator and pollination monitoring framework. Technical Report. Department for Environment, Food and Rural Affairs.

Catarino, R., Bretagnolle, V., Perrot, T., Vialloux, F. and Gaba, S., 2019. Bee pollination outperforms pesticides for oilseed crop production and profitability. Proceedings of the Royal SocietyB, 286 (1912), 20191550.

Dainese, M., Martin, E.A., Aizen, M., Albrecht, M., Bartomeus, I., Bommarco, R., … and Ghazoul, J., 2019. A global synthesis reveals biodiversity-mediated benefits for crop production. bioRxiv, 554170.

Else, G.R., Bolton, B. and Broad, G.R., 2016. Checklist of British and Irish Hymenoptera-aculeates (Apoidea, Chrysidoidea and Vespoidea). Biodiversity Data Journal, 4.

Else, G.R. & Edwards, M., 2018. Handbook of the Bees of the British Isles: Volume 2. Ray Society.

Fijen, T.P., Scheper, J.A., Boom, T.M., Janssen, N., Raemakers, I. and Kleijn, D., 2018. Insect pollination is at least as important for marketable crop yield as plant quality in a seed crop. Ecology Letters, 21 (11), 1704-1713.

Gardner, K.E. and Ascher, J.S., 2006. Notes on the native bee pollinators in New York apple orchards. Journal of the New York Entomological Society, 114 (1), 86-91.

Gardner E., Breeze T.D., Clough Y., Smith H., Baldock K., Campbell A., … and Oliver T., 2020. Reliably Predicting Pollinator Abundance: Challenges of Process Based Ecological Models; Methods in Ecology and Evolution, In Press

Garibaldi, L.A., Steffan-Dewenter, I., Kremen, C., Morales, J.M., Bommarco, R., Cunningham, S.A., … and Holzschuh, A., 2011. Stability of pollination services decreases with isolation from natural areas despite honey bee visits. Ecology Letters, 14 (10), 1062-1072.

Garibaldi, L.A., Bartomeus, I., Bommarco, R., Klein, A.M., Cunningham, S.A., Aizen, M.A., … and Morales, C.L., 2015. Trait matching of flower visitors and crops predicts fruit set better than trait diversity. Journal of Applied Ecology, 52, 1436-1444.

Garibaldi, L.A., Pérez-Méndez, N., Garratt, M.P., Gemmill-Herren, B., Miguez, F.E. and Dicks, L.V., 2019. Policies for ecological intensification of crop production. Trends in ecology & evolution, 34(4), pp.282-286.

Garratt, M.P., Coston, D.J., Truslove, C.L., Lappage, M.G., Polce, C., Dean, R., … and Potts, S.G., 2014a. The identity of crop pollinators helps target conservation for improved ecosystem services. Biological Conservation, 169, 128-135.

Garratt, M.P., Breeze, T.D., Jenner, N., Polce, C., Biesmeijer, J.C. and Potts, S.G., 2014b. Avoiding a bad apple: Insect pollination enhances fruit quality and economic value. Agriculture, ecosystems & environment, 184, pp.34-40.

Garratt, M.P.D., Breeze, T.D., Boreux, V., Fountain, M.T., Mckerchar, M., Webber, S.M., … and Biesmeijer, J.C., 2016. Apple pollination: demand depends on variety and supply depends on pollinator identity. PloS one, 11(5), e0153889.

Garratt, M.P.D., Potts, S.G., Banks, G., Hawes, C., Breeze, T.D., O'Connor, R.S. and Carvell, C., 2019. Capacity and willingness of farmers and citizen scientists to monitor crop pollinators and pollination services. Global Ecology and Conservation, 20, e00781.

Godfray, H. C. J. and Garnett, T., 2014. Food security and sustainable intensification. Philosophical Transactions of the Royal Society B: Biological Sciences, 369, 20120273.

Grab, H., Branstetter, M.G., Amon, N., Urban-Mead, K.R., Park, M.G., Gibbs, J., … and Danforth, B.N., 2019. Agriculturally dominated landscapes reduce bee phylogenetic diversity and pollination services. Science, 363(6424), 282-284.

Gresty, C.E., Clare, E., Devey, D.S., Cowan, R.S., Csiba, L., Malakasi, P., … and Willis, K.J., 2018. Flower preferences and pollen transport networks for cavity-nesting solitary bees: Implications for the design of agri-environment schemes. Ecology and evolution, 8(5), 7574-7587.

Hoehn, P., Tscharntke, T., Tylianakis, J.M. and Steffan-Dewenter, I., 2008. Functional group diversity of bee pollinators increases crop yield. Proceedings of the Royal Society of London B: Biological Sciences, 275(1648), 2283-2291.

Horth, L. and Campbell, L.A., 2018. Supplementing small farms with native mason bees increases strawberry size and growth rate. Journal of Applied Ecology, 55(2), 591-599.

IPBES., 2016. Deliverable 3a: Thematic assessment of pollinators, pollination and food production. http://www.ipbes.net/sites/default/files/downloads/pdf/3a_pollination_individual_chapters_20161124.pdf

King, C., Ballantyne, G. and Willmer, P.G., 2013. Why flower visitation is a poor proxy for pollination: measuring single-visit pollen deposition, with implications for pollination networks and conservation. Methods in Ecology and Evolution, 4(9), 811-818.

Klein, A.M., Steffan–Dewenter, I. and Tscharntke, T., 2003. Fruit set of highland coffee increases with the diversity of pollinating bees. Proceedings of the Royal Society of London. Series B: Biological Sciences, 270(1518), 955-961.

Kleijn, D., Winfree, R., Bartomeus, I., Carvalheiro, L.G., Henry, M., Isaacs, R., … and Ricketts, T.H., 2015. Delivery of crop pollination services is an insufficient argument for wild pollinator conservation. Nature Communications, 6, 7414.

Kleijn, D., Bommarco, R., Fijen, T.P., Garibaldi, L.A., Potts, S.G. and van der Putten, W.H., 2019. Ecological intensification: bridging the gap between science and practice. Trends in ecology & evolution, 34 (2), pp.154-166.

Kremen, C., Williams, N.M. and Thorp, R.W., 2002. Crop pollination from native bees at risk from agricultural intensification. Proceedings of the National Academy of Sciences, 99(26), 16812-16816.

Kremen, C. and Chaplin-Kramer, R., 2007. Insects as providers of ecosystem services: crop pollination and pest control. In: Insect conservation biology: proceedings of the royal entomological society's 23rd symposium, 349-382, Wallingford, UK: CABI Publishing.

MacLeod, M., Reilly, J., Cariveau, D.P., Genung, M.A., Roswell, M., Gibbs, J. and Winfree, R., 2020. How much do rare and crop-pollinating bees overlap in identity and flower preferences? Journal of Applied Ecology, 57(2), 413-423.

Martins, K.T., Gonzalez, A. and Lechowicz, M.J., 2015. Pollination services are mediated by bee functional diversity and landscape context. Agriculture, Ecosystems & Environment, 200, 12-20.

Michener, C.D., 2000. The bees of the world (Vol. 1). JHU press.

O'Connor, R.S., Kunin, W.E., Garratt, M.P., Potts, S.G., Roy, H.E., Andrews, C., … and Morris, R.K., 2019. Monitoring insect pollinators and flower visitation: the effectiveness and feasibility of different survey methods. Methods in Ecology and Evolution, 10(12), 2129-2140.

Oliver, T.H., Heard, M.S., Isaac, N.J., Roy, D.B., Procter, D., Eigenbrod, F., … and Proença, V., 2015. Biodiversity and resilience of ecosystem functions. Trends in Ecology & Evolution, 30(11), 673-684.

Ollerton, J., 2017. Pollinator diversity: distribution, ecological function, and conservation. Annual Review of Ecology, Evolution, and Systematics, 48, 353-376.

Perrot, T., Gaba, S., Roncoroni, M., Gautier, J.L. and Bretagnolle, V., 2018. Bees increase oilseed rape yield under real field conditions. Agriculture, Ecosystems & Environment, 266, 39-48.

Polce, C., Termansen, M., Aguirre-Gutiérrez, J., Boatman, N.D., Budge, G.E., Crowe, A., … and Somerwill, K.E., 2013. Species distribution models for crop pollination: a modelling framework applied to Great Britain. PloS one, 8(10), e76308.

Polce, C., Garratt, M.P., Termansen, M., Ramirez-Villegas, J., Challinor, A.J., Lappage, M.G., … and Somerwill, K.E., 2014. Climate-driven spatial mismatches between British orchards and their pollinators: increased risks of pollination deficits. Global Change Biology, 20(9), 2815-2828.

Powney, G.D., Carvell, C., Edwards, M., Morris, R.K., Roy, H.E., Woodcock, B.A. and Isaac, N.J., 2019. Widespread losses of pollinating insects in Britain. Nature Communications, 10(1), 1018.

Rader, R., Howlett, B.G., Cunningham, S.A., Westcott, D.A. and Edwards, W., 2012. Spatial and temporal variation in pollinator effectiveness: do unmanaged insects provide consistent pollination services to mass flowering crops? Journal of Applied Ecology, 49(1), 126-134.

Rader, R., Bartomeus, I., Garibaldi, L.A., Garratt, M.P., Howlett, B.G., Winfree, R., … and Bommarco, R., 2016. Non-bee insects are important contributors to global crop pollination. Proceedings of the National Academy of Sciences, 113(1), pp.146-151.

Reilly, J.R., Artz, D.R., Biddinger, D., Bobiwash, K., Boyle, N.K., Brittain, C., … and Ellis, J.D., 2020. Crop production in the USA is frequently limited by a lack of pollinators. Proceedings of the Royal Society B, 287(1931), p.20200922.

Russo, L., Park, M.G., Blitzer, E.J. and Danforth, B.N., 2017. Flower handling behavior and abundance determine the relative contribution of pollinators to seed set in apple orchards. Agriculture, Ecosystems & Environment, 246, 102-108.

Scheper, J., Holzschuh, A., Kuussaari, M., Potts, S.G., Rundlöf, M., Smith, H.G. and Kleijn, D., 2013. Environmental factors driving the effectiveness of European agri-environmental measures in mitigating pollinator loss–a meta-analysis. Ecology Letters, 16(7), 912-920.

Senapathi, D., Biesmeijer, J.C., Breeze, T.D., Kleijn, D., Potts, S.G. and Carvalheiro, L.G., 2015. Pollinator conservation—the difference between managing for pollination services and preserving pollinator diversity. Current Opinion in Insect Science, 12, 93-101.

Settele, J., Bishop, J. and Potts, S.G., 2016. Climate change impacts on pollination. Nature Plants, 2(7),16092.

Soroye, P., Newbold, T. and Kerr, J., 2020. Climate change contributes to widespread declines among bumble bees across continents. Science, 367(6478), 685-688.

Sutter, L., Jeanneret, P., Bartual, A.M., Bocci, G. and Albrecht, M., 2017. Enhancing plant diversity in agricultural landscapes promotes both rare bees and dominant crop-pollinating bees through complementary increase in key floral resources. Journal of Applied Ecology, 54(6),1856-1864.

Tonietto, R.K. and Larkin, D.J., 2018. Habitat restoration benefits wild bees: A meta-analysis. Journal of Applied Ecology, 55(2), 582-590.

Vanbergen, A.J., Heard, M.S., Breeze, T., Potts, S.G. and Hanley, N., 2014. Status and value of pollinators and pollination services. http://nora.nerc.ac.uk/id/eprint/505259/

Vázquez, D.P., Morris, W.F. and Jordano, P., 2005. Interaction frequency as a surrogate for the total effect of animal mutualists on plants. Ecology Letters, 8(10), 1088-1094.

Webb, J., Heaver, D., Lott, D., Dean, H.J., van Breda, J., Curson, … and Foster, G., 2018. Pantheon - database version 3.7.6. https://www.brc.ac.uk/pantheon/

Westphal, C., Bommarco, R., Carré, G., Lamborn, E., Morison, N., Petanidou, T., … and Vaissière, B.E., 2008. Measuring bee diversity in different European habitats and biogeographical regions. Ecological Monographs, 78(4), 653-671.

Winfree, R., W. Fox, J., Williams, N.M., Reilly, J.R. and Cariveau, D.P., 2015. Abundance of common species, not species richness, drives delivery of a real-world ecosystem service. Ecology letters, 18(7), 626-635.

Winfree, R., Reilly, J.R., Bartomeus, I., Cariveau, D.P., Williams, N.M. and Gibbs, J., 2018. Species turnover promotes the importance of bee diversity for crop pollination at regional scales. Science, 359(6377), 791-793.

Wolf, S., Rohde, M. and Moritz, R.F., 2010. The reliability of morphological traits in the differentiation of Bombus terrestris and B. lucorum (Hymenoptera: Apidae). Apidologie, 41(1), 45-53.

Wood, T.J., Holland, J.M., Hughes, W.O. and Goulson, D., 2015a. Targeted agri-environment schemes significantly improve the population size of common farmland bumblebee species. Molecular Ecology, 24(8), 1668-1680.

Wood, T.J., Holland, J.M. and Goulson, D., 2015b. Pollinator-friendly management does not increase the diversity of farmland bees and wasps. Biological Conservation, **187**, pp.120-126.

Wood, T.J., Holland, J.M. and Goulson, D., 2016a. Providing foraging resources for solitary bees on farmland: current schemes for pollinators benefit a limited suite of species. Journal of Applied Ecology, 54(1),323-333.

Wood, T.J., Holland, J.M. and Goulson, D., 2016b. Diet characterisation of solitary bees on farmland: dietary specialisation predicts rarity. Biodiversity and Conservation, 25(13), 2655-2671.

Woodcock, B.A., Edwards, M., Redhead, J., Meek, W.R., Nuttall, P., Falk, S., … and Pywell, R.F., 2013. Crop flower visitation by honeybees, bumblebees and solitary bees: Behavioural differences and diversity responses to landscape. Agriculture, Ecosystems & Environment, 171, 1-8.

Woodcock, B.A., Garratt, M.P.D., Powney, G.D., Shaw, R.F., Osborne, J.L., Soroka, J., … and Jauker, F., 2019. Meta-analysis reveals that pollinator functional diversity and abundance enhance crop pollination and yield. Nature Communications, 10(1) 1-10.

Wratten, S.D., Gillespie, M., Decourtye, A., Mader, E. and Desneux, N., 2012. Pollinator habitat enhancement: benefits to other ecosystem services. Agriculture, Ecosystems & Environment, 159, pp.112-122.