



# Diets of European polecat *Mustela putorius* in Great Britain during fifty years of population recovery

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## Abstract

Following nineteenth-century declines, polecats *Mustela putorius* are recolonising Great Britain. Polecat diet relates to two potential risks to recovery. First, rabbits *Oryctolagus cuniculus*, which are important prey for polecats, have experienced extreme population fluctuations, with near extirpation due to myxomatosis in the 1950s, recovery in 1960s–1990s and declines in 1990s–2010s. Second, polecats are secondarily exposed to anticoagulant rodenticides by eating contaminated rodents, and the frequency of polecat exposure to rodenticides is increasing. We analysed stomach contents from 99 polecats collected in 2012–2016 and compared results with earlier studies. Lagomorphs were the most abundant prey (66% frequency of occurrence, 95% confidence interval 53–74%), followed by other mammals (12%, 4–18%), amphibians (10%, 3–16%) and birds (7%, 1–13%). Diet varied seasonally; lagomorph occurrence was highest in spring and summer and lowest in autumn. Dietary niche breadth was greater in the 1960s, when rabbits were scarce, than in other decades, but did not differ between the 1990s and 2010s, indicating that diets have not diversified with recent rabbit declines. This may be because rabbit abundance is not yet low enough to cause dietary diversification or because polecats were collected in areas where rabbits were still abundant. Rodents did not increase in diet between the 1990s and 2010s and still occur with < 10% frequency, indicating that rodents need not contribute much to diet to expose polecats to rodenticides. This potentially limits the effectiveness of management actions designed to minimise polecat exposure to contaminated rodent prey.

**Keywords** Polecat · *Mustela putorius* · Diet · Rabbits · Species recovery

## Introduction

Successfully colonizing species often demonstrate ecological flexibility during the process of population

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establishment and expansion (Rosecchi et al. 2001; Lee and Gelembiuk 2008). Some species have flexible foraging strategies that allow them to exploit variation in resource availability across temporal and spatial scales (Zhou et al. 2011). The European polecat *Mustela putorius* is a medium-sized mustelid carnivoran that is currently recolonising its former range in Great Britain, following catastrophic declines, mostly in the nineteenth century (Langley and Yalden 1977; Sainsbury et al. 2019). A significant element of this period of range expansion has coincided with extreme fluctuations in populations of rabbits *Oryctolagus cuniculus* (Sumption and Flowerdew 1985; Aebischer et al. 2011; Harris et al. 2018; Massimino et al. 2019), which are a major source of food for polecats in Great Britain (Birks and Kitchener 1999). It is unknown whether polecats in Britain may have altered their feeding strategy in response to recent rabbit declines and whether these declines may have impacted on their continued recolonisation.

Optimal foraging theory predicts that animals will maximise their net energy intake (MacArthur and Pianka 1966; Pyke et al. 1977; Perry and Pianka 1997), which may be maximised via different foraging strategies. For instance, specialist predators have narrow dietary niches and will forage for specific prey species, independent of their availability, whereas generalists have comparatively large dietary niches and consume prey in proportion to their availability (Futuyma and Moreno 1988; Amundsen 1995). Polecats eat a wide variety of food items across their European range and are usually described as generalist predators (Erlinge 1986; Lodé 1995; Baghli et al. 2002; Santos et al. 2009; Malecha and Antczak 2013). Rodents and amphibians are common food items in all regions (Lodé 1997). While polecats exhibit dietary diversity across their range, there is some evidence of regional specialisation. For example, polecats specialise on rabbits in the Mediterranean (Santos et al. 2009) and on amphibians in Switzerland (Weber 1989a) and Poland (Jędrzejewski et al. 1993; Jędrzejewska and Jędrzejewska 1998). In Białowieża National Park, Poland, frogs comprised 60% and 90% of polecat total food biomass in summer and winter, respectively, and although polecats ate rodents, this was only when frogs were not available (Jędrzejewska and Jędrzejewska 1998). In this example, polecats exhibited a very narrow dietary niche and clear preferential selection for frogs. In other localities, it has been found that impressions of specialisation by polecats simply reflect the local abundance of a given prey (Lodé 1995). Whilst more recent studies in Britain have found that polecats predominantly eat lagomorphs (Blandford 1986; Birks and Kitchener 1999), in periods when rabbit populations were severely reduced due to disease outbreaks, notably during the 1950s and 1960s due to myxomatosis (Sumption and Flowerdew 1985), mammals comprised a much smaller proportion of polecat diet (Walton 1968). There is also evidence of seasonal consumption of rodents (including brown rats *Rattus norvegicus* and field voles *Microtus agrestis*) particularly in the winter months (Birks 1998; Birks and Kitchener 1999). A flexible foraging strategy is thought to allow polecats to occupy and exploit diverse habitats, such as lowland, grassland, farmland and riparian habitats (Blandford 1987) and may play an important role in enabling population persistence (Lodé 1997).

When rabbit populations in Great Britain crashed by up to 95% as a result of a myxomatosis epizootic in the 1950s (Sumption and Flowerdew 1985), the diet composition and population dynamics of other mustelid carnivorans, such as stoats *Mustela erminea* and weasels *Mustela nivalis*, were affected (McDonald et al. 2000). Rabbit populations recovered to their pre-myxomatosis levels by the 1990s (Aebischer et al. 2011; Aebischer 2019), but since then, rabbit numbers have declined across Britain (England – 44%; Scotland – 82%; and Wales – 48%; Harris et al. 2018), possibly as the result of rabbit haemorrhagic disease (RHD), which has

devastated rabbit populations across mainland Europe (Lees and Bell 2008). Analysis of changes in rabbit records between 2011 and 2015 reveals spatial variation in rabbit declines, with the greatest reductions in central and southern England, along the Scottish borders and in north-east Scotland (Massimino et al. 2018).

In Spain, generalist carnivorans (such as red fox *Vulpes vulpes*, badger *Meles meles* and genet *Genetta genetta*) reduced their consumption of rabbits in response to declines in rabbit populations following RHD outbreaks (Ferrerias et al. 2011). In contrast, whilst rabbit consumption by Iberian lynx *Lynx pardinus*, which are near-obligate predators of rabbits, also reduced, lynx continued to preferentially select rabbits in spite of their reduced availability (Ferrerias et al. 2011). Given that polecats in Britain are known to eat rodents and amphibians as well as rabbits (Blandford 1986; Birks and Kitchener 1999), it is possible that reductions in rabbit populations would lead to polecats diversifying their diet.

Rodents are thought to be the major route by which polecats are exposed to second-generation anticoagulant rodenticides (SGARs) in Britain (Shore et al. 2003). Secondary exposure of polecats to SGARs increased 1.7-fold between 1993 and 2016, and the most recent study indicated that 79% of polecats had been exposed (Sainsbury et al. 2018). It may be that this increase has been a result of an increase in the proportion of rodents in polecat diet. Secondary exposure to SGARs may be lethal in sufficient concentration, or lead to a range of sub-lethal effects (Van den Brink et al. 2018). Increased rates of secondary exposure to SGARs have not prevented polecat expansion over the same time period (Sainsbury et al. 2019), but whether or not polecat abundance or the rate of population expansion have been affected by SGARs exposure is unknown (Sainsbury et al. 2018).

Between-sex dietary differences have been observed in some mustelids (McDonald 2002). Studies of polecat diet in Britain have previously found that female polecats tend to eat fewer rabbits and more birds than male polecats, though these differences were not statistically significant (Blandford 1986; Birks and Kitchener 1999). It is possible that as rabbit abundance has declined, increased competition for available rabbits may have led to more pronounced dietary differences between male and female polecat diets.

To explore dietary variation and niche breadth in polecats during the process of polecat population recovery and rabbit population variation, we analysed gut contents from polecats collected from 2012 to 2016 and compared our findings with earlier analyses of polecat diet in Britain in the 1960s (Walton 1968), 1980s (Blandford 1986) and 1990s (Birks and Kitchener 1999). We hypothesised that (i) rabbits will be reduced in importance in polecat diet compared with the 1990s in response to reduced rabbit abundance; (ii) rodent prey will have increased as a proportion of diet over the same period; (iii) polecat dietary niche breadth will have fluctuated over

time, in line with known long-term variations in rabbit abundance and (iv) rabbit consumption will have differed between the sexes.

## Materials and methods

Polecat carcasses, predominantly of animals killed on the road, were collected across Great Britain during the Vincent Wildlife Trust's national polecat survey in 2012–2016 (Croose 2016). Date of collection and location were recorded. Animals were stored frozen until necropsy examination, which was carried out at National Museums Scotland. Stomach contents were collected from 99 polecats (Fig. 1) and refrozen prior to dietary analysis. Stomach contents were soaked in biological detergent for 24 h, rinsed through a 53- $\mu\text{m}$  sieve, then stored in 70% ethanol. Identifiable macroscopic animal remains (undigested body parts, fragments of bone, feathers, fur, individual hair and insect remains) were separated from unidentifiable tissues. Ten samples were selected at random and were analysed for earthworm (Lumbricidae) chaetae and other microscopic remains (after Reynolds and Aebischer 1991). No identifiable microscopic remains were found. As there is no evidence from previous dietary studies to suggest that polecats eat earthworms (e.g. Birks and Kitchener 1999), and since variation in microscopic remains do not relate to the primary processes of interest in this study, we considered only macroscopic remains in the remaining 90% of gut samples. Plant debris was considered to have been ingested when catching prey (Walton 1968) and was not included in diet composition quantification.

Fur remains were identified using guard hair cuticle patterns after Teerink (1991). All loose hairs were collected and analysed. Cuticle patterns were examined under a microscope at  $\times 40$  magnification. Mammal remains were identified to species level, except for rabbits and brown hares *Lepus europaeus*, which were not separated and were classified as lagomorphs. Most bones were fragmented and unidentifiable, but those that were intact, together with teeth, were identified as insectivore, rodent, larger mammal or amphibian to species level using personal collections and appropriate keys (e.g. Thomas 2008; Inns 2011). Bird remains were identified to order using Day (1966). Amphibians were determined by skin texture and, where possible, by webbing on feet (Inns 2011). Fishes were identified by their bones and scales but were not identified to species, as fish were a rare item and not of primary interest.

Two methods were used to assess accuracy when identifying guard hair cuticle pattern. First, 10% of samples were randomly selected for a second blind analysis by the same analyst; the correspondence in the results was 100%. A third analysis, again blind, was carried out by a second researcher, this time on 10% of samples that contained hard parts and

20% of the samples that relied on hair identification. There was a 100% match between analysts for samples containing hard parts, an 86% correspondence for hair samples and a calculated Cohen's kappa test of interrater agreement of 0.7, which is "substantial" according to Landis and Koch (1977). The level of overlap in identification indicated that identification of guard hair was sufficiently robust for inclusion in our data analysis.

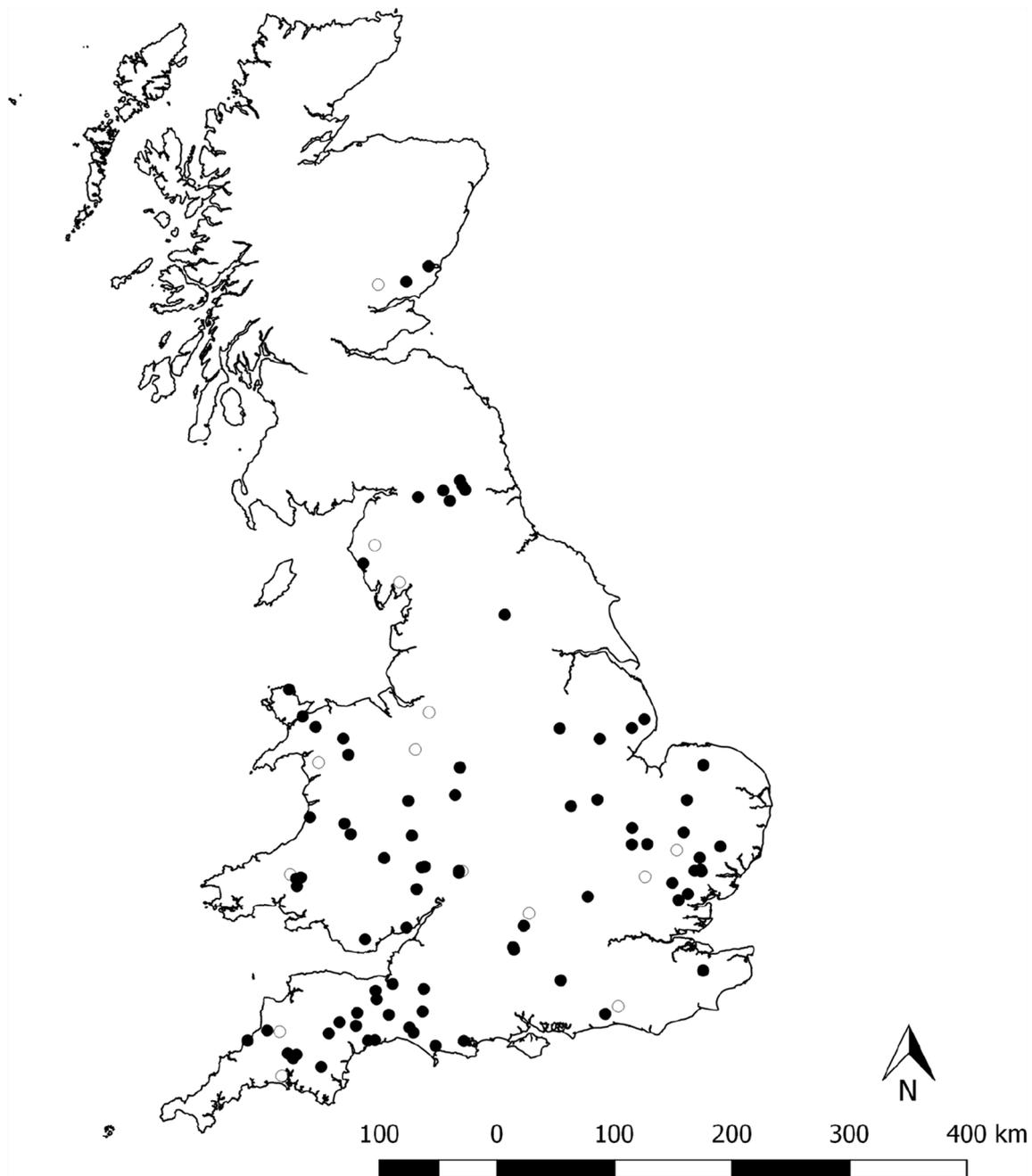
All analyses were carried out in R (R Core Team 2011). Diet was summarised as the percentage frequency of occurrence (% FO), calculated as the number of each type of dietary item as a percentage of the total number of identifiable prey items (e.g. Lodé 1994; McDonald et al. 2000). As our objective was to evaluate changes in polecat diet over time rather than assess differences in energetic requirements, frequency of occurrence was chosen as the best method for comparison as it was consistent with historical studies. Whilst frequency of occurrence may overestimate the importance of smaller food items in a carnivore's diet, and caution should be used applying it in isolation to understand the effect of predators on prey populations, it is still an appropriate method for exploring a carnivore's ecology (Klare et al. 2011). Polecats, in line with other small carnivores (e.g. McDonald et al. 2000), usually only have one prey item per stomach (e.g. Weber 1989a; Birks and Kitchener 1999). This means that the difference between frequency of occurrence calculated using total prey items or that using the number of stomachs is negligible (in this study, of the 99 polecat stomachs investigated and 79 that had identifiable contents, only three individual stomachs contained more than one item). As a result and for simplicity, we calculated frequency of occurrence per food item, expressed as a percentage of the number of occurrences of one food item of the total number of occurrences of all food items, to indicate the relative importance in diet (Klare et al. 2011). The frequency of occurrence matrix was replicated randomly 1000 times (bootstrapped with replacement 1000 times) to generate 95% confidence intervals following Reynolds and Aebischer (1991). Differences in FO of prey groups were compared using a chi-squared test.

Levins' (1968) index of niche breadth was calculated following the formula:

$$N_b = 1/\sum p_i^2$$

where  $p_i$  is the proportion of records for each species in each group. The proportion of prey records for each group was bootstrapped with replacement 1000 times to generate 95% confidence intervals for Levins' index (Reynolds and Aebischer 1991).

To analyse variation in the occurrence of lagomorphs in polecat diet in more depth using the 2010s data, a binomial logistic regression model of presence/absence was fitted to sex, season (where spring is March to May, summer is June



**Fig. 1** Map showing the collection locations of polecat carcasses collected between 2012 and 2016. Shaded circles indicate polecats with items in their stomach. White circles indicate stomachs that were empty. Number of dots = 97, as two polecats came from unverified locations

to August, autumn is September to November and winter is December to February) and region (north, south, east and west based on British government regions) as explanatory variables. Backwards stepwise model selection using the “drop1” function in R (with  $P = 0.05$  used as the significance level to assess whether or not variables should be retained) was carried out to find the most parsimonious model.

Our results were compared with earlier polecat dietary studies by Walton (1968), Blandford (1986) and Birks and Kitchener (1999). Some of the data collected historically

was from stomach contents (Walton 1968; Birks and Kitchener 1999) and some from scat analyses (Blandford 1986), which created a potential source of sampling bias. Scats may be considered similar to intestinal or rectal contents and therefore a separate meal to that found in the stomach (Day 1968). Comparisons between stomach and intestinal content of stoats (*Mustela erminea*: Day 1968; *M. e. cicognanii*: Aldous and Manweiler 1942) and common weasels *Mustela nivalis* (Day 1968) found no difference between stomach and intestinal content analysis in terms of dietary

reconstruction, and therefore, it was considered acceptable to compare polecat diets between all of the historical studies.

To compare changes in frequency of occurrence of prey groups over time, binomial logistic regressions were run for the prey groups: all mammals, birds and amphibians for the 1960s–2010s datasets using decade as an explanatory variable. In addition, binomial logistic regressions were run for the prey groups: all mammals, lagomorphs, other mammals, birds and amphibians using the 1980s, 1990s, and 2010s datasets with decade and sex as explanatory variables. As backwards step-wise deletion (with  $P = 0.05$  as the significance level used to assess whether or not variables should be retained in models) found that sex was not significant in any of the models, it was excluded from the analysis and only the results for the prey groups lagomorphs and other mammals are reported (as the models that included the 1960s datasets already include the other main prey groups). Models were not run for the prey groups fish and invertebrates due to small sample sizes.

Levins' niche breadth was calculated using five categories (all mammals, birds, amphibians, fishes and invertebrates) for comparisons between 1960s, 1980s, 1990s and 2010s as the 1960s data did not distinguish mammal species. For the male and female calculations by decade between the 1980s and 2010s (the original 1960s data did not distinguish between the sexes), Levins' niche breadth was calculated based on six categories (lagomorphs, other mammals, birds, amphibians, fishes and invertebrates) as data from the 1960s did make this distinction. Differences in niche breadth over time (1960s–2010s) and between the sexes (1980s–2010s) were compared.

## Results

Of the 99 stomachs that contained some remains (65 male, 32 female, 2 sex unknown), 14 contained only liquid and six contained unidentifiable remains, such as undigested flesh. This left 79 stomachs containing identifiable prey items (54 male, 24 female, 1 unknown). Sixteen stomachs contained polecat hair, but this was excluded from the dietary analysis as it was assumed to be the result of grooming (Rysava-Novakova and Koubek 2009). Ten stomachs contained plant matter.

The observed frequencies of occurrence of the prey groups differed significantly from even ( $\chi^2_4 = 114.8$ ,  $P < 0.001$ ). Mammals were the most frequently identified prey group (78% FO), and lagomorphs comprised 66% FO of prey items (Table 1). Eight of the 54 lagomorph samples (15% FO) were neonates or juveniles, identified on the basis of intact ears, tails or feet. "Other mammals" comprised 12% FO of dietary items. Amphibians were the second most frequently identified prey group (10% FO) and mostly comprised frogs. Birds, fishes and invertebrates comprised approximately 7%, 4%

and 1% FO of all items, respectively. Only one instance of carrion was found (identified based on the presence of maggots with the flesh), which was classified as lagomorph.

The % FO of all prey items in male and female polecat stomachs was similar across the board, and there was no difference in niche breadth between male (Levins' index = 2.2, 95% CI 1.6–3.0) and female polecats (2.2, 95% CI 1.4–3.2).

The model analysing the factors influencing polecat diet in the 2010s found that season was the only factor that significantly affected the occurrence of lagomorphs (Fig. 2). Sex and region were removed from the model. The FO of lagomorphs was lowest in the autumn (Table 2) and significantly greater in spring (coefficient estimate 2.19, standard error 0.69,  $z = 3.18$ ,  $P = 0.001$ ) and summer (coefficient = 0.60, SE = 0.70,  $z = 0.85$ ,  $P = 0.396$ ) than in the autumn, but FO in winter (coefficient = 2.19, SE = 0.69,  $z = 3.18$ ,  $P = 0.001$ ) was not statistically significantly greater than in the autumn. Niche breadth was highest in autumn and lowest in the spring (Levins' index, 95% CI spring 1.1, 1.0–1.3; summer 1.9, 1.2–3.1; autumn 4.1, 2.5–5.0; winter 2.6, 1.4–4.3; Fig. 2).

Analysis of changes in polecat diets over time indicated that the occurrence of all mammals more than doubled between the 1960s (35%) and the 1980s (74%), after which the occurrence of mammals stabilised as a proportion of diet (Table 3). The change in FO of mammals was statistically significant between the 1960s and all other decades, but the FO of mammals was the same between the 1980s and 1990s–2010s and the 1990s and 2010s (Table 4). Birds decreased significantly as a proportion of diet between the 1980s and 1990s and the 1980s and 2010s, but was similar in the 2010s compared with the 1990s (Table 4). Amphibians decreased significantly in frequency between the 1960s and 1980s but increased again in the 2010s (Table 3). Amphibian frequency was significantly lower in all other decades compared with the 1960s, but there was no significant difference in amphibians between any other decades (Table 4). Invertebrates decreased as a proportion of diet after the 1960s and were found only at low levels since then (1.1–2.0% of diet; Table 3). Lagomorph occurrence increased significantly between the 1980s (25%) and 1990s (69%) but did not differ between samples collected in the 1990s and the 2010s (66%; Tables 3 and 4). Other mammals increased in polecat diet between the 1980s and 1990s and 2010s. There was no significant difference between males and female polecats in FO of all mammals (including lagomorphs and other mammals), birds or amphibians. Polecat dietary niche breadth was greatest in the 1960s and lowest in the 1990s and 2010s (Table 3).

## Discussion

Polecat diet composition in Great Britain was dominated by lagomorphs in all the samples analysed from the 1980s to

**Table 1** Summary of composition of stomach contents of polecats *Mustela putorius* collected in Great Britain between 2012 and 2016

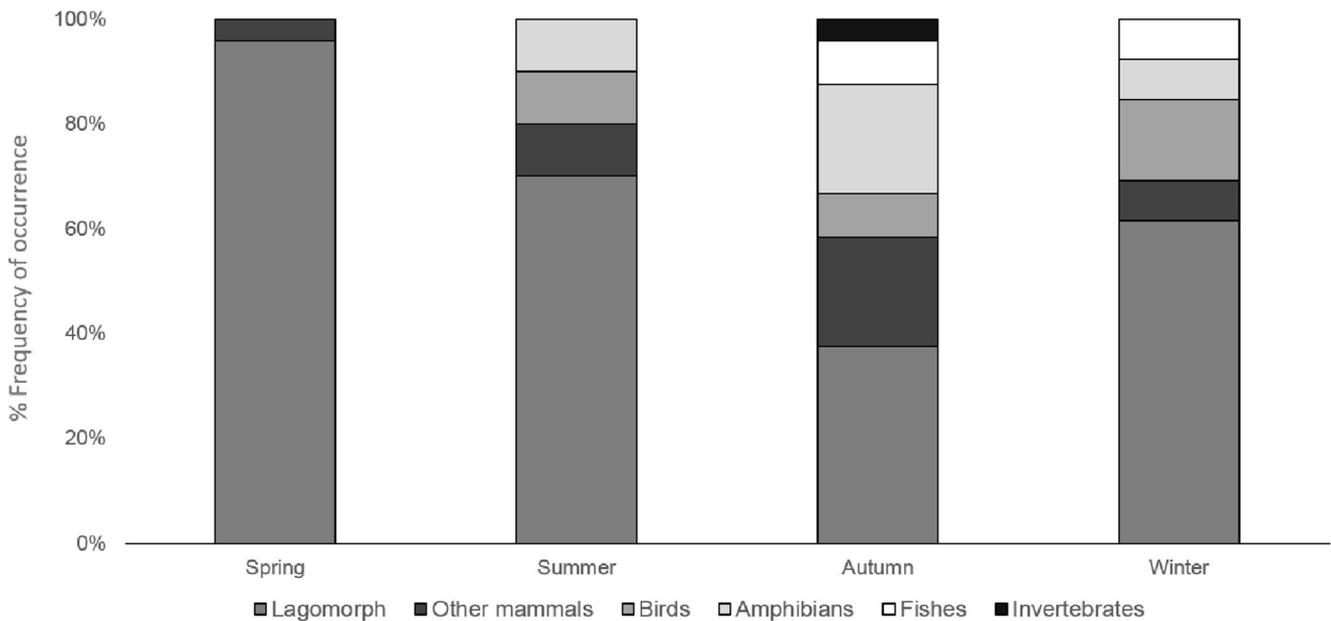
Prey	Total			Males			Females		
	FO	% FO	(95% CI)	FO	% FO	(95% CI)	FO	% FO	(95% CI)
All mammals	64	78.0	(64.7–84.7)	43	78.1	(62.9–86.1)	20	76.9	(53.9–94.1)
Lagomorphs	54	65.9	(53.7–74.0)	36	65.5	(50.8–76.5)	17	65.4	(44.4–81.5)
Other mammals	10	12.2	(4.2–17.5)	7	7.3	(2.5–19.3)	3	11.5	(0.0–23.0)
<i>Microtus agrestis</i>	1	1.2	(0.0–3.5)	1	1.8	(0.0–5.4)			
<i>Myodes glareolus</i>	1	1.2	(0.0–3.6)				1	3.8	(0.0–10.8)
<i>Sciurus</i> spp.	3	3.7	(0.0–7.7)	3	5.5	(0.0–11.3)			
<i>Rattus norvegicus</i>	1	1.2	(0.0–3.6)				1	3.8	(0.0–11.2)
<i>Sorex</i> spp.	1	1.2	(0.0–3.7)				1	3.8	(0.0–11.0)
Unidentified small mammals	3	3.7	(0.0–5.8)	3	5.5	(0.0–8.7)			
Birds	6	7.3	(1.2–12.9)	4	7.3	(0.4–14.2)	2	7.7	(0.0–17.2)
Galliformes	1	1.2	(0.0–3.6)	1	1.8	(0.0–5.4)			
Columbiformes	1	1.2	(0.0–3.6)				1	3.8	(0.0–11.2)
Passeriformes	2	2.4	(0.0–5.9)	1	1.8	(0.0–5.4)	1	3.8	(0.0–11.0)
Unidentified bird	2	2.4	(0.0–5.5)	2	3.6	(0.0–8.7)			
Amphibians	8	9.8	(3.2–16.1)	4	7.2	(0.2–14.3)	4	15.4	(0.5–29.1)
<i>Rana temporaria</i>	7	8.5	(2.2–14.7)	3	5.5	(0.0–11.5)	4	15.4	(1.0–28.6)
<i>Bufo bufo</i>	1	1.2	(0.0–3.7)	1	1.8	(0.0–5.4)			
Fish	3	3.7	(0.0–7.5)	3	5.5	(0.0–11.5)			
Invertebrates	1	1.2	(0.0–3.6)	1	1.8	(0.0–5.3)			
Number of identifiable prey items	82	100%		55	100%		26	100%	
Stomachs containing identifiable prey items	79 <sup>a</sup>			54			24		
Stomachs with two identifiable items	3			1			2		
Non-prey items									
<i>Mustela putorius</i>	16			11			5		
Vegetation	10			8			2		
Stomachs containing remains	99 <sup>a</sup>			65			32		

95% percentile range confidence intervals are bootstrapped estimates of the variability associated with sampling. Italicized entries indicate the main prey groups used in analyses

<sup>a</sup> Sex was not recorded for two polecats; the stomach of one contained lagomorph remains and the other contained unidentifiable remains

2010s, although there was seasonal variation in the 2010s, which reflects the opportunistic foraging strategy of this species. Contrary to our expectations, there was no evidence of a reduction in rabbit consumption by polecats since the 1990s, despite declines in rabbit records over this period (Harris et al. 2018). Instead, we found an increase in the frequency of occurrence of mammals in polecat diet since the 1960s and an increase in lagomorphs between the 1980s and 1990s, which is consistent with similar variation in the importance of lagomorphs in the diet of stoats over the same time period (McDonald et al. 2000). We also found that lagomorphs were equally important as prey for both male and female polecats in the 1990s and 2010s (Table 1) and so there was no evidence of resource partitioning between the sexes, consistent with McDonald (2002). We found evidence of polecat predation on young rabbits, which was also found in the 1990s study (Kitchener, unpublished data).

There are several possible explanations for this lack of dietary shift in response to declining resources. One is that the more recent rabbit declines have been uneven across the landscape (Massimino et al. 2018), unlike the 1950s myxomatosis epizootic (Sumption and Flowerdew 1985), and so, polecats have still been able to find and take rabbits as their major prey item. When rabbit numbers were still low in Wales in the 1980s (Aebischer et al. 2011) and polecats were predominantly found only in that country (Sainsbury et al. 2019), lagomorphs comprised a lower proportion of polecat diet than in the 2010s and niche breadth was correspondingly greater than in the 1990s or 2010s (Table 3; Blandford 1986). There was also some evidence of differences in resource use between males and females in the 1980s (lower lagomorph consumption by females) when rabbits were limited in availability (Table 3; Blandford 1986); this resource partitioning was not evident in the 2010s.



**Fig. 2** Percentage frequency of occurrence of five categories of prey identified in polecat stomachs from animals collected between 2012 and 2016, presented by season and as a percentage of prey items collected. Total  $n$  polecats = 79. Total  $n$  prey items = 82. Seasons are spring: March–May ( $n$  items identified = 24), summer: June–August ( $n$  = 20), autumn:

September–November ( $n$  = 24) and winter: December–February ( $n$  = 13). One individual not shown in the graph but included in the main analyses did not have the date recorded. One polecat in summer and two polecats in autumn had two identifiable items in their stomachs

The continued high prevalence of rabbits in polecat diets means that other prey items, notably rodents, were consumed less frequently than might have been expected. The relatively high occurrence of amphibians is in line with polecat diet studies in Switzerland (Weber 1989a), Poland (Jędrzejewski et al. 1993; Jędrzejewska and Jędrzejewska 1998), France (Lodé 2000), and Denmark (Hammershøj et al. 2004). Whilst some dietary studies have correlated amphibian consumption to periods of abundance (Lodé 2000), others have found that polecats eat them preferentially (Weber 1989b). Data on nationwide trends in amphibians in Britain are limited, but available evidence that common toad *Bufo bufo* have been declining over the last 40 years (Petrovan and Schmidt 2016). Unlike in France, where amphibians are more

commonly eaten in spring when amphibians are more active at the beginning of their breeding season (Lodé 2000), amphibians were most commonly caught in autumn in this study. The skin of the common toad was consumed intact in spite of its dermal secretions (Sidorovich and Pikulik 1997).

Given recent increases in the rates of exposure of polecats to anticoagulant rodenticides (Sainsbury et al. 2018), it is perhaps surprising that we did not observe an increase in the proportion of rodents, especially brown rats, consumed by polecats over this same time period. Furthermore, our data indicate that the current high proportion of polecats exposed to SGARs is associated with an intake of rodents that comprises less than 10% of total diet. This suggests that even relatively low rates of rodent consumption can result in high rates of secondary exposure. This means that our ability to minimise exposure of polecats to SGARs may be limited, as rodents comprise only a small component of their diet.

The recent occurrence of birds in the diet of polecats was consistent with that observed in previous decades. However, the importance of bird remains is often underestimated in analysing stomach contents (Reynolds and Aebischer 1991). Furthermore, polecats eat eggs, but we did not detect any evidence of this, possibly because polecats tend to break them open and lick out the contents (Weber 1989a). In the 1980s, Blandford (1986) found that Galliformes comprised 5% FO of polecat diet. This study was carried out before polecats had expanded beyond the Welsh border counties, an area where game management is less widespread (Tapper 1992). Birks

**Table 2** Results from a generalised linear model of the effects of season on the presence or absence of lagomorph remains in polecat *Mustela putorius* stomach samples

	Coefficient estimate	Standard error	Z value	P value
Intercept	-0.44	0.43	-1.03	
Spring	2.19	0.69	3.18	0.001
Summer	1.29	0.65	1.99	0.047
Winter	0.60	0.70	0.85	0.396

$n$  = 79 stomachs and 82 prey items. Spring is March–May, summer is June–August, autumn is September–November and winter is December–February. The reference level in the model is autumn as this was the time period with the smallest effect sizes. Italicized entries indicate where effects are significantly different from autumn

**Table 3** Summary of polecat dietary studies in Great Britain from the 1960s to the 2010s

	1960s		1980s		1990s		2010s	
	% FO (95% CI)		% FO (95% CI)		% FO (95% CI)		% FO (95% CI)	
<i>n</i> (males and females)	38	Stomachs	754	Scats	87	Prey items	82	Prey items
All mammals	35.1	(19.6–48.8)	73.7	(70.6–76.9)	79.3	(70.8–87.8)	78.0	(64.7–84.7)
Lagomorphs	–		24.9	(21.8–28.0)	69.0	(54.5–78.4)	65.9	(53.7–74.0)
Other mammals	–		48.8	(45.1–52.5)	10.3	(3.6–17.1)	12.2	(4.2–17.5)
Birds	14.0	(2.2–24.1)	16.0	(13.4–18.7)	5.7	(0.7–10.8)	7.3	(1.2–12.9)
Amphibians	26.2	(11.7–40.9)	8.2	(6.1–10.3)	8.0	(2.2–13.9)	9.8	(3.2–16.1)
Fish	–		–		1.1	(0.0–3.6)	3.7	(0.0–7.5)
Invertebrates	24.6	(9.8–37.5)	2.0	(0.9–3.0)	10.3	(3.9–16.8)	1.2	(0.0–3.6)
<i>Levins' niche breadth</i>	3.9	(3.0–4.2)	1.7	(1.6–1.9)	1.6	(1.3–1.9)	1.6	(1.3–2.0)
<i>n</i> (males)			411		56		55	
All mammals			73.2	(69.1–77.3)	82.1	(71.9–92.4)	78.1	(62.9–86.1)
Lagomorphs			26.3	(22.1–30.5)	75.0	(63.3–86.7)	65.5	(50.8–76.5)
Other mammals			47.0	(42.0–51.9)	7.1	(0.3–14.0)	7.3	(2.5–19.3)
Birds			14.4	(10.8–17.9)	7.1	(0.3–14.0)	7.3	(0.4–14.2)
Amphibians			8.8	(6.0–11.5)	8.9	(1.3–16.5)	7.2	(0.2–14.3)
Fish			–		–		5.5	(0.0–11.5)
Invertebrates			3.6	(1.8–5.5)	1.8	(0.0–5.3)	1.8	(0.0–5.3)
<i>Levins' niche breadth</i>			3.1	(2.9–3.4)	1.7	(1.4–2.3)	2.2	(1.6–3.0)
<i>n</i> (females)			343		31		26	
All mammals			74.3	(69.4–79.3)	74.2	(58.6–89.8)	76.9	(53.9–94.1)
Lagomorphs			23.6	(18.9–28.4)	58.1	(41.0–75.1)	65.4	(44.4–81.5)
Other mammals			50.7	(45.4–56.1)	16.1	(2.9–29.4)	11.5	(0.0–23.0)
Birds			18.1	(13.9–22.2)	12.9	(1.2–24.6)	7.7	(0.0–17.2)
Amphibians			7.6	(4.8–10.4)	6.5	(0.0–15.6)	15.4	(0.5–29.1)
Fish			–		3.2	(0.0–9.3)	–	
Invertebrates			–		3.2	(0.0–9.7)	–	
<i>Levins' niche breadth</i>			2.9	(2.6–3.1)	2.7	(1.8–4.2)	2.2	(1.4–3.2)

Sources are Walton (1968), Blandford (1986), Birks and Kitchener (1999) and the current study. 1960 data are calculated using FO in stomachs. 1980 data are based on FO in scats. 1990 and 2010 data are based on the total prey items identified. Levins' niche breadth was calculated using four categories (all mammals, birds, amphibians/fishes and invertebrates) for the 1960s, 1980s, 1990s and 2010s. For the male and female calculations by decade, Levins' niche breadth was calculated based on five categories (lagomorphs, other mammals, birds, amphibians/fishes and invertebrates). 95% percentile range confidence intervals are a bootstrapped estimate of the variability associated with sampling errors. Sex of animals was not recorded for the 1960 sample

and Kitchener (1999), who collected roadkill carcasses from across Wales and the English Midlands, did not identify any Galliformes in polecat stomachs, and our results are broadly consistent with this. Polecats spend time on game estates in Britain (Packer and Birks 1999), and in other countries, polecats are known to consume wounded or dead gamebirds (Rysava-Novakova and Koubek 2009). Given this and the difficulties associated with detecting birds in stomach contents, it is possible that birds might be eaten more frequently than our results suggest.

We have highlighted long-term increases in the proportion of lagomorphs in polecat diet in Britain during a period of polecat population recovery. When rabbits were almost extirpated from Britain in the 1950s, polecat diet

was significantly more diverse. Polecat niche breadth has declined as rabbit populations have recovered. Our results, which show how niche breadth and diet composition has varied over time, demonstrate the importance of long-term studies for determining whether species are generalist or specialists. Overall, rabbits are an important prey item for polecats in Great Britain. Although rabbit populations are once again in decline, there is no evidence of a concomitant reduction of consumption in polecat diet, but it is possible that such effects may only be apparent from studies conducted at a finer spatial scale. Finally, our study shows that rodents do not have to constitute a high proportion of polecat diet, to lead to frequent secondary exposure to anticoagulant rodenticides.

**Table 4** Summary of statistically significant changes in the frequencies of occurrence of prey groups in polecat *Mustela putorius* dietary samples between the 1960s, 1980s, 1990s and 2010s

		1960s	1980s	1990s	2010s
All mammals	1960s	–	< 0.001	< 0.001	< 0.001
	1980s	–	–	0.262	0.278
	1990s	–	–	–	0.995
Lagomorphs	1960s	NA	NA	NA	NA
	1980s	NA	–	< 0.001	< 0.001
	1990s	NA	–	–	0.735
Other mammals	1960s	NA	NA	NA	NA
	1980s	NA	–	< 0.001	< 0.001
	1990s	NA	–	–	0.673
Birds	1960s	–	0.635	0.171	0.309
	1980s	–	–	0.015	0.042
	1990s	–	–	–	0.680
Amphibians	1960s	–	< 0.001	0.009	0.022
	1980s	–	–	0.955	0.634
	1990s	–	–	–	0.696

Data presented are *P* values calculated from binomial logistic regressions of the presence/absence of each prey group by decade. Frequency of occurrence data are presented in Table 3. Statistically significant effects are italicised.

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