

Article (refereed) - postprint

This is the peer reviewed version of the following article:

Sabater, Ana M.; Ward, Helen C.; Hill, Timothy C.; Gornall, Jemma L.; Wade, Thomas J.; Evans, Jonathan G.  Prieto-Blanco, Ana; Disney, Mathias; Phoenix, Gareth K.; Williams, Mathew; Huntley, Brian; Baxter, Robert; Mencuccini, Maurizio; Poyatos, Rafael. 2020. **Transpiration from subarctic deciduous woodlands: environmental controls and contribution to ecosystem evapotranspiration.** *Ecohydrology*, 13 (3), e2190. 14, pp. [10.1002/eco.2190](https://doi.org/10.1002/eco.2190)

This article may be used for non-commercial purposes in accordance with Wiley Terms and Conditions for Use of Self-Archived Versions.

© 2020 John Wiley & Sons Ltd

This version is available at <https://nora.nerc.ac.uk/id/eprint/527281>

Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at <https://nora.nerc.ac.uk/policies.html#access>.

This document is the authors' final manuscript version of the journal article, incorporating any revisions agreed during the peer review process. There may be differences between this and the publisher's version. You are advised to consult the publisher's version if you wish to cite from this article.

The definitive version is available at <https://onlinelibrary.wiley.com/>

Contact UKCEH NORA team at
noraceh@ceh.ac.uk

Transpiration from subarctic deciduous woodlands: environmental controls and contribution to ecosystem evapotranspiration.

Ana M. Sabater ^{1,2,3,4}, Helen C. Ward ⁵, Timothy C. Hill ⁶, Jemma L. Gornall⁷, Thomas J. Wade ⁸, Jonathan G. Evans ⁹, Ana Prieto-Blanco ¹⁰, Mathias Disney ^{10, 11}, Gareth K. Phoenix ¹², Mathew Williams ⁸, Brian Huntley ¹³, Robert Baxter ¹³, Maurizio Mencuccini ^{2, 14}, Rafael Poyatos ^{1, 2, 15}

¹ Universitat Autònoma de Barcelona, Barcelona, Spain.

² CREAF, Cerdanyola del Vallès, Barcelona, Spain.

³ Fundación CEAM, Joint Research Unit University of Alicante-CEAM, Alicante, Spain.

⁴ Department of Ecology, University of Alicante, Alicante, Spain.

⁵ Department of Atmospheric and Cryospheric Sciences (ACINN), University of Innsbruck, Innsbruck, Austria.

⁶ College of Life and Environmental Sciences, Department of Geography, University of Exeter, Exeter, UK.

⁷ Met Office, Exeter, UK.

⁸ School of Geosciences, University of Edinburgh, Edinburgh, UK.

⁹ Centre for Ecology and Hydrology, Wallingford, UK.

¹⁰ University College London, Department of Geography, London, UK.

¹¹ NERC National Centre for Earth Observation (NCEO), Leicester, UK.

¹² Department of Animal and Plant Sciences, University of Sheffield, Sheffield, UK.

¹³ Department of Biosciences, Durham University, Durham, UK.

¹⁴ ICREA, Barcelona, Spain.

¹⁵ Laboratory of Plant Ecology, Faculty of Bioscience Engineering, Ghent University, Ghent, Belgium.

This article has been accepted for publication and undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process which may lead to differences between this version and the Version of Record. Please cite this article as doi: 10.1002/eco.2190

Correspondence. R. Poyatos, CREAM, Edifici C, Campus UAB Bellaterra, E08193
Barcelona, Spain. Email: r.poyatos@creaf.uab.es, Tel.+34 935814676, Fax: +34935814151.

Short title: Transpiration and evapotranspiration of subarctic deciduous woodlands

Accepted Article

Abstract

Potential land-climate feedbacks in subarctic regions, where rapid warming is driving forest expansion into the tundra, may be mediated by differences in transpiration of different plant functional types. Here we assess the environmental controls of overstorey transpiration and its relevance for ecosystem evapotranspiration in subarctic deciduous woodlands. We measured overstorey transpiration of mountain birch canopies and ecosystem evapotranspiration in two locations in northern Fennoscandia, having dense (Abisko) and sparse (Kevo) overstories. For Kevo, we also upscale chamber-measured understorey evapotranspiration from shrubs and lichen using a detailed land cover map. Sub-daily evaporative fluxes were not affected by soil moisture, and showed similar controls by vapour pressure deficit and radiation across sites. At the daily timescale, increases in evaporative demand led to proportionally higher contributions of overstorey transpiration to ecosystem evapotranspiration. For the entire growing season, the overstorey transpired 33% of ecosystem evapotranspiration in Abisko and only 16% in Kevo. At this latter site, the understorey had a higher leaf area index and contributed more to ecosystem evapotranspiration compared to the overstorey birch canopy. In Abisko, growing season evapotranspiration was 27% higher than precipitation, consistent with a gradual soil moisture depletion over the summer. Our results show that overstorey canopy transpiration in subarctic deciduous woodlands is not the dominant evaporative flux. However, given the observed environmental sensitivity of evapotranspiration components, the role of deciduous trees in driving ecosystem evapotranspiration may increase with the predicted increases in tree cover and evaporative demand across subarctic regions.

Keywords

Arctic, branch cuvettes, eddy covariance, evapotranspiration partitioning, mountain birch, tundra, understorey

Introduction

Northern high latitudes (boreal and arctic biomes) exert an important influence in global biosphere-atmosphere interactions involving water, energy and atmospheric composition. These interactions are globally relevant because of the large extent of these biomes (arctic tundra and boreal forest cover *ca.* $1.24 \cdot 10^8 \text{ km}^2$) and the intense and rapid warming occurring at northern high latitudes (0.5 K/decade since 1979; IPCC, 2013), which is partly driven by regional positive feedbacks (Chapin et al., 2000). Warmer temperatures and longer growing seasons are already inducing poleward and altitudinal treeline migration and shrub expansion in the tundra zone, which may in turn drive considerable land-atmosphere feedbacks in these latitudes (Kattsov et al., 2005; Swann, Fung, Levis, Bonan, & Doney, 2010; Zhang et al., 2013)

Treelines across the subarctic vegetation belt are largely dominated by conifers, although deciduous broadleaves occupy 18% of the forest area at latitudes above 60° across Eurasia (Krankina et al., 2010) and can form the tundra-to-forest transition in many subarctic regions with oceanic influence (Callaghan et al., 2005). The area of deciduous broadleaf woodlands is increasing throughout the subarctic region (Hofgaard, Tømmervik, Rees, & Hanssen, 2013; Rundqvist et al., 2011; Tømmervik et al., 2004; Wang et al., 2019), following a general trend of increasing deciduous vegetation at northern high latitudes (Myers-Smith et al., 2011). These vegetation changes are predicted to continue in the future (Mekonnen, Riley, Randerson, Grant, & Rogers, 2019) and may cause substantial land-climate feedbacks mediated by changes in albedo, in carbon sequestration and in evaporative fluxes (Bonan, 2008; Bonfils et al., 2012). Higher transpiration rates by deciduous broadleaf forests could lead to stronger evaporative cooling locally (Chapin et al., 2000), although, in a regional context, the effects of the expansion of deciduous broadleaf trees into the tundra zone can be more complex and actually enhance Arctic warming (Swann et al., 2010). Moreover, increased soil moisture uptake by deciduous trees could lead to faster depletion of snowmelt water during the shoulder season, triggering further hydrological changes (Young-Robertson, Bolton, Bhatt, Cristóbal, & Thoman, 2016). Therefore, a greater understanding of the magnitudes and controls of evapotranspiration in deciduous woodlands is needed to predict future changes in land-atmosphere interactions in subarctic forest-tundra ecotones.

Syntheses addressing magnitudes and drivers of ecosystem evapotranspiration (ET_{eco}) at northern high latitudes show a paucity of data for deciduous broadleaf forests from subarctic

locations (Brümmer et al., 2011; Kasurinen et al., 2014; McFadden, Eugster, & Chapin III, 2003). These syntheses show that leaf area index (LAI), meteorological conditions and physiological regulation by vegetation are the three major factors affecting ET_{eco} in northern high-latitude ecosystems. In these ecosystems, evapotranspiration is largely driven by vapour pressure deficit (VPD), radiation and temperature, with soil moisture often playing a minor role (Beringer, Chapin, Thompson, & McGuire, 2005; Brümmer et al., 2011). In deciduous forests, growing season duration also affects seasonal evapotranspiration through the influence on LAI phenology (Brümmer et al., 2012). Deciduous broadleaf forests from northern high latitudes show higher evapotranspiration rates compared to conifer forests in the same region (Brümmer et al., 2011; Kasurinen et al., 2014), but they may also display a stronger stomatal control with increasing VPD (Welp, Randerson, & Liu, 2007). However, to what extent do these patterns in the drivers of ET_{eco} from northern high-latitude deciduous forests reflect the transpiration regulation by the main canopy?

The partitioning of ET_{eco} into transpiration and evaporation and the factors controlling this partitioning are still poorly known (Schlesinger & Jasechko, 2014). Subarctic and northern boreal woodlands typically show a low LAI of the dominant canopy species, meaning that the contribution of understorey and soil evaporation to ecosystem evapotranspiration may be moderate to high (Blanken et al., 2001; Iida et al., 2009; Lafleur, 1992), although it will depend on vegetation structure (Beringer et al., 2005). This substantial contribution of the soil and understorey to ET_{eco} implies that eddy flux-based estimates of ET_{eco} in these forests may well represent the mix of physical and biological controls on evaporative fluxes and will only partially capture the physiological regulation exerted by the main canopy (Ikawa et al., 2015; Kasurinen et al., 2014). Evaporative fluxes of overstorey, understorey and the forest floor may have contrasting hydroclimatic responses (Iida et al., 2009) and a strong seasonal variation (Blanken et al., 2001). Although several studies have addressed the magnitudes and drivers of the different components of ET_{eco} in northern boreal and subarctic forests (Blanken et al., 2001; Grelle, Lundberg, Lindroth, Morén, & Cienciala, 1997; Iida et al., 2009; Ikawa et al., 2015), we are not aware of any study of these characteristics from subarctic deciduous woodlands.

In this article, we quantify the magnitude and seasonal controls on ET_{eco} and on the transpiration of the main canopy in two deciduous broadleaf woodlands dominated by mountain birch (*Betula pubescens* spp. *czerepanovii* (Orlova) Hamet- Ahti). This is a representative species of subarctic woodlands covering 600000 ha throughout northern

Fennoscandia (Haapanala et al., 2009). The Abisko site (N Sweden) displays a denser birch woodland compared to the sparser Kevo site (N Finland), which is also slightly colder and wetter. Therefore, the Abisko woodland would be representative of denser canopies which are becoming common across the subarctic in response to warming and reduced browsing (Callaghan et al., 2013). In both sites, we measured ET_{eco} and birch transpiration per leaf area (T_{leaf}), which was upscaled to the birch canopy level (T_{birch}). Our main goals were: (1) to identify the drivers of ET_{eco} and T_{leaf} , to understand the environmental controls between the two scales (ecosystem vs branch) and at sites, which differed substantially in stand structure (denser in Abisko, sparser in Kevo); and (2) to investigate how variation in canopy structure affects growing season values of ET_{eco} relative to growing season precipitation and to quantify the contribution of T_{birch} to ET_{eco} . To further understand this evapotranspiration partitioning in subarctic deciduous woodlands, at Kevo we also upscaled evaporative fluxes from birch and understorey ($ET_{upscaled}$) to explore how this variable compares to ET_{eco} .

2. Methodology

2.1. Study sites

Two mountain birch (*Betula pubescens* ssp. *czerepanovii*) forest sites within the northern Fennoscandia sub-Arctic vegetation belt were chosen for this study: Abisko (northern Sweden) and Kevo (northern Finland). Both sites were located near the mountain birch/tundra ecotone, where mountain birches are polycormic because of the harsh environmental conditions and the frequent defoliation by autumn and winter moths (*Epirrita autumnata* and *Operophtera brumata*). At both sites, we measured transpiration of mountain birch branches, ecosystem evapotranspiration and other environmental drivers during the mountain birch leaf-on period, hereby abbreviated as ‘growing season’, of 2007 (Abisko, DOY 153-241) and of 2008 (Kevo, DOY 171-257).

In Abisko (Figure 1a), measurements were undertaken at a location (68.326°N, 18.833°E, 519 m.a.s.l) ca. 3.2 km south-east of the Abisko Research Station. At the study site, mean annual temperature is -0.9°C and mean annual precipitation is 335 mm (1980-2010, temperature corrected assuming a lapse rate of 0.55 °C per 100 m of elevation). The predominant substrate is coarse glacial till and soils are typically micro-podzols, with no permafrost present (Hartley, Hopkins, Sommerkorn, & Wookey, 2010). The landscape presents a relatively complex topography, which results in highly variable forest cover

(Nyström, Holmgren, & Olsson, 2012) and stand structures (Table 1). Understorey vegetation is dominated by the dwarf shrubs *Empetrum nigrum* ssp *hermaphroditum*, *Vaccinium myrtillus* and *Vaccinium uliginosum* (Hartley et al., 2010; Poyatos, Gornall, Mencuccini, Huntley, & Baxter, 2012).

In Kevo (Figure 1b), measurements were undertaken at a location (69.492°N, 27.234°E, 260 m.a.s.l.) ca. 40 km south of the Kevo Subarctic Research Institute. Climate at the site (1978-2007, data from the the Kevo Institute station, corrected for lapse rate) is colder and wetter than in Abisko (-2.4°C and 422 mm mean annual temperature and precipitation, respectively) and the substrate is composed of gneiss covered by glacial till, and no permafrost is present at the forest site. Mountain birch forests in Kevo, located upon gentle slopes/ridges and surrounded by mires in topographically depressed areas, were sparser and showed a more homogeneous structure compared to Abisko (Table 1). Understorey vegetation showed a higher LAI compared to Abisko (Table 1); it consisted of *E. nigrum* below mountain birch canopies and distinct patches covered by *Betula nana* L. and *Cladonia* spp, lichens in the open areas (Poyatos et al., 2012).

One forest inventory was established in the vicinity of each of the branch bags sites to quantify stand structure at the plot level (a 10-m circular plot in Abisko and a 30 x 30 m plot in Kevo). Another set of 30 x 30 m plots was measured in Abisko (N = 5) and Kevo (N = 7) to quantify ecosystem-level stand structure and maximum leaf area index, LAI_{max} (m² leaf m⁻²ground). Forest inventory plots were at an average distance from the eddy flux tower of 105 m in Abisko and 450 m in Kevo. Diameters and heights of all stems with diameter at breast height DBH>12 mm within the plots were measured in 2007 at Abisko and in 2008 at Kevo. For Abisko, we used published allometric equations predicting leaf biomass from stem basal area and height (Dahlberg, Berge, Petersson, & Vencatasawmy, 2004) to convert leaf biomass supported by each stem into leaf area using site-specific leaf mass per area. For Kevo, we harvested N = 15 stems during the peak growing season in 2008, to measure their leaf area and we obtained site-specific allometries between stem diameter and leaf area (Table S1). Understorey LAI_{max} was obtained from 1 m² vegetation surveys (N = 5) in each of the sites, following Fletcher et al. (2012).

2.2. Branch-level transpiration measurements

At both sites, we selected eight mountain birch branches representative of low and mid-canopy conditions for branch transpiration measurements. Branch transpiration was measured using a multiplexed branch bag device based on the closed system approach (Rayment & Jarvis, 1999; Wingate, Seibt, Moncrieff, Jarvis, & Lloyd, 2007). This system measures water vapour concentration changes within eight 0.11 m³ ventilated cuvettes enclosing individual branches during 7.5 minutes. Branches were measured sequentially, and a measurement cycle of all eight branches was completed within an hour. During each measurement period, air temperature, T (°C), relative humidity, RH (%), and photosynthetically active radiation, PAR ($\mu\text{mol photons m}^{-2} \text{ s}^{-1}$), were recorded every 5 seconds by a datalogger. The system also recorded the value of environmental variables at the beginning of each transpiration observation (i.e. hourly). The subscript 'branch' was used to refer to branch-level meteorological variables ($\text{PAR}_{\text{branch}}$, $\text{VPD}_{\text{branch}}$). Further technical details of the branch bags system and of the calculation of branch-level transpiration can be found in the Supporting Information S2.

We quantified branch transpiration on a leaf area basis, T_{leaf} ($\text{l m}^{-2} \text{ hour}^{-1}$), by dividing whole-branch transpiration by the leaf area of the branch within the bag. To account for seasonal variation in branch leaf area, we periodically counted the number of leaves inside the bags during the growing season. We then multiplied the leaf counts by an estimation of the average leaf area obtained from a sample of leaves ($N = 10$) close to the measured branch, fitted a nonlinear response as a function of day of year and, if needed, corrected by differences in leaf size between inside and outside the bags (Poyatos et al., 2012). We expressed the seasonal variation in leaf area in relative terms between 0 and 1 (minimum and maximum leaf area, respectively) to use for the upscaling of branch transpiration fluxes.

2.3. Ecosystem evapotranspiration and environmental monitoring

At both sites, half-hourly ecosystem evapotranspiration, ET_{eco} (mm h^{-1}), was estimated from latent heat measurements using the eddy covariance (EC) technique in flux towers located above the mountain birch canopy (Aubinet, Vesala, & Papale, 2012). The three components of wind speed were measured with a sonic anemometer (R3, Gill Instruments, Lymington, UK) and water vapour concentrations were measured by an open-path infrared gas analyser (LI-7500, LI-COR Biosciences, Lincoln, USA). Raw data were logged at 20 Hz and processed to 30-minute statistics using FluxView (Centre for Ecology and Hydrology,

Wallingford, UK) and quality-controlled following standard procedures. These include correcting sonic data for angle-of-attack (Gash & Dolman, 2003), compensating for the lag time between sonic and gas analyser, rotating the co-ordinate system (so that the horizontal wind vector is aligned with the 30-min mean and the vertical component is forced to zero), correcting sonic temperature for humidity (Schotanus, Nieuwstadt, & De Bruin, 1983), correcting the fluxes for high- and low- frequency spectral losses and correcting gas fluxes for density effects (Webb, Pearman, & Leuning, 1980). Quality control involved despiking and removal of data outside physically reasonable limits, when instruments malfunctioned, when the windows of the gas analyser were wet or dirty, and during periods of heavy rain. Filtering of data during low turbulence conditions based on a friction velocity threshold was not applied. Energy balance closure was within the expected range (Stoy et al., 2013) and did not differ much across sites (Supplementary Information S3).

Meteorological stations installed at the flux towers measured half-hourly values of temperature, relative humidity, PAR and precipitation above the birch canopy and we refer to them using the subscript 'eco' (PAR_{eco} , VPD_{eco}). Soil volumetric water content in the upper 30 cm of the soil, SWC ($cm^3 cm^{-3}$), was measured with 1 or 2 frequency domain reflectometers (CS616, Campbell Scientific, UK) at each site. To account for site-specific differences in maximum and minimum water-holding capacity, we transformed SWC into soil moisture deficit (SMD), which ranged from 0 (maximum soil moisture) to 1 (minimum soil moisture) (Granier & Loustau, 1994).

2.4. Modelling environmental controls of evaporative fluxes

Firstly, T_{leaf} and ET_{eco} data were filtered ($PAR > 50 \mu mol photons m^{-2} s^{-1}$) to avoid noisy vapour concentration data in the branch bags and low turbulence conditions in the case of EC. For T_{leaf} , the values of the meteorological drivers were measured locally in each individual branch (VPD_{branch} , PAR_{branch}) and for ET_{eco} they were measured above the canopy (VPD_{eco} , PAR_{eco}).

All models were fitted using the nlme package (Pinheiro, Bates, DebRoy, Sarkar, & R Core Team, 2018) in R (R Core Team 2016). T_{leaf} was modelled using a linear mixed effects model (lme), with VPD_{branch} , PAR_{branch} and SMD as fixed factors and ET_{eco} was fitted as a function of VPD_{eco} , PAR_{eco} and SMD using a generalized least squares model (gls). In view of the residual distributions after preliminary analyses, we log-transformed the response variables, T_{leaf} and ET_{eco} , and the explanatory variables, except for the case of PAR in ET_{eco} modelling.

All models included a first-order autoregressive correlation structure for the residuals, specifying fractional day of year as a continuous time covariate. We applied model selection to include those terms which minimised the value of the Akaike Information Criterion (AIC) while checking that variance inflation factors were below 10 (Zuur, Ieno, & Elphick, 2010). Model selection was carried out with models fitted using maximum likelihood, but final models were fitted using restricted maximum likelihood (Pinheiro & Bates, 2000). Normality, linearity and homoscedasticity of residuals were visually inspected and temporal autocorrelation was analysed visually by autocorrelation plots using the *acf* function in R. We calculated marginal and conditional R^2 , the proportion of variance explained by fixed and by both fixed and random factors, respectively (Nakagawa & Schielzeth, 2013).

2.5. Overstorey contributions to ecosystem evapotranspiration

Before upscaling, evaporative flux data were aggregated at the daily scale, using models obtained in section 2.4 to gap-fill missing hourly data and fitting daily models when meteorological data from the measurement systems were missing (Supporting Information S4). We obtained transpiration of the mountain birch canopy, T_{birch} (mm day⁻¹), by multiplying T_{leaf} by the LAI of mountain birch in each stand (Table 1), corrected for seasonal variation (see section 2.2). The calculation was done using mean and \pm standard error (SE) of the LAI values, to propagate the uncertainty of the LAI values at each site into the upscaled estimates of T_{birch} .

At both sites we calculated the mountain birch contribution to daily ecosystem evapotranspiration, T_{birch}/ET_{eco} (%). We analysed T_{birch}/ET_{eco} as a separate linear model of VPD, PAR (both log-transformed) and SMD, including a factor coding for site (Abisko and Kevo) which interacted with each of the environmental drivers. Model selection was carried out based on AIC, as described in section 2.4. We also tested for a possible influence of interception and subsequent canopy evaporation on T_{birch}/ET_{eco} by testing for differences between dry and wet days, using a gls model as described in the previous paragraph. We considered wet days as those within 2 days after a precipitation event > 1 mm, assuming all wet surfaces would have dried up during this period (Knauer, Werner, & Zaehle, 2015).

Growing season values (mm) of precipitation (P), T_{birch} , ET_{eco} and $ET_{upscaled}$ were calculated by aggregating daily values. We also quantified the overall growing season contribution of T_{birch} and $ET_{upscaled}$ to ET_{eco} and expressed growing season evaporative fluxes as a percentage of growing season precipitation.

2.6. Upscaling evapotranspiration components in Kevo

In Kevo, measurements of evapotranspiration were available for other ecosystem components, i.e., understorey shrubs and lichen (Table S3, Figure 1). These evapotranspiration measurements were representative of small patches and were obtained with an automated chamber system (Poyatos et al., 2014) operated during the 2008 growing season, in a forest-mire ecotone *ca.* 200 m from the flux tower (Figure 1). Hourly evapotranspiration of 12 tundra plots was calculated similarly to branch bags fluxes (Supporting Information S5). Because of microclimatic alterations, water vapour sorption in the tubing system and imperfect chamber sealing the automated chamber system used here has been reported to underestimate the evaporative fluxes (Cohen et al., 2015). Therefore, we applied a correction factor of 2.3, obtained in that study, which used a similar device under comparable environmental conditions (Cohen et al., 2015).

Shrub evapotranspiration (ET_{shrub}) was estimated as the mean of $N = 9$ plots (mean $LAI_{max} \pm SE = 0.77 \pm 0.2$) with dwarf tundra vegetation (mainly *Empetrum hermaphroditum*, *Calluna vulgaris* and *Vaccinium* spp.) while lichen evaporation (ET_{lichen}) was calculated as the mean of $N = 3$ lichen heath plots (Poyatos et al., 2014). We then combined evapotranspiration of the individual components with the fractional covers (f) of each component within the footprint of the flux tower. Fractional covers were obtained from aerial photography obtained in August 2008 and subsequent vegetation classification (Hartley et al., 2015). We used a dynamic footprint approach (Hartley et al., 2015) to obtain f values which varied with atmospheric conditions, although results were comparable to those using a simpler, fixed footprint approach (Figure S4). We calculated $ET_{upscaled}$ ($mm\ day^{-1}$) as the product of the time-variable f of each component and its corresponding T or ET value:

$$ET_{upscaled} = T_{birch} + f_{birch} \cdot ET_{shrub} + f_{shrub} \cdot ET_{shrub} + f_{lichen} \cdot ET_{lichen} \quad (1)$$

Where f_{birch} , f_{shrub} and f_{lichen} represent the fractional covers of birch forest, understorey shrubs and lichen, respectively. This equation assumes that shrubs were also typically present under the birch canopies (cf. section 2.1) and that components other than birch, shrubs and lichen (around 5% of fractional cover, Table S3) behave similarly to shrubs.

3. Results

3.1. Temporal variation of environmental variables and evaporative fluxes

Evaporative demand (Figure 2a-d) was higher in Abisko than in Kevo, as shown by higher mean growing season values (\pm standard deviation, SD) of air temperatures ($10.5\text{ }^{\circ}\text{C} \pm 3.8$ and $9.5\text{ }^{\circ}\text{C} \pm 3.6$, respectively), VPD_{eco} ($0.5 \pm 0.3\text{ kPa}$ and $0.3 \pm 0.2\text{ kPa}$) and PAR_{eco} ($407.0 \pm 170.0\text{ }\mu\text{mol m}^{-2}\text{ s}^{-1}$ and $260.4 \pm 130.3\text{ }\mu\text{mol m}^{-2}\text{ s}^{-1}$). Light transmission through the birch canopy was higher in Kevo: $\text{PAR}_{\text{branch}}/\text{PAR}_{\text{eco}}$ was 56% in Kevo compared to 30% in Abisko (Figure 2a,b). This was associated with the larger difference between $\text{VPD}_{\text{branch}}$ and VPD_{eco} (Figure 2c,d) in Kevo (average $\text{VPD}_{\text{branch}} - \text{VPD}_{\text{eco}} = 0.30\text{ kPa}$) compared to Abisko (average $\text{VPD}_{\text{branch}} - \text{VPD}_{\text{eco}} = 0.14\text{ kPa}$). Kevo also received heavier and more frequent precipitation (Figure 2e,f), resulting in higher total growing season precipitation (167.5 mm) compared to Abisko (126.6 mm).

Both T_{leaf} and ET_{eco} tended to be higher in Abisko than in Kevo, on average 50% higher for T_{leaf} and 62% higher for ET_{eco} . Their seasonal dynamics were similar and followed the course of evaporative demand (Figure 2g-j). However, some differences between T_{leaf} and ET_{eco} during the early growing season (before DOY 160) were apparent for Abisko. The diurnal cycles of evaporative fluxes and their drivers varied seasonally in both sites (Figure S2, S3), as expected due to the changing daylight hours at these latitudes. Abisko typically presented higher ET_{eco} and T_{leaf} except during the late season, when T_{leaf} was equal for the two sites.

3.2. Modelling environmental controls of evaporative fluxes

ET_{eco} and T_{leaf} increased with PAR and VPD but the relationship with VPD showed much less scatter (Figure 3). In general, T_{leaf} and ET_{eco} at a given value of PAR or VPD were higher for Abisko. Models of ET_{eco} and T_{leaf} showed a good predictive ability, with marginal R^2 values > 0.7 (Table 2,3). Model predictors included a negative interaction between PAR and VPD but did not include SMD (Table 2,3). The environmental responses of ET_{eco} did not vary across sites and we only detected site differences for the intercept and the PAR coefficient in the T_{leaf} model (Table 2,3). In both models, the interaction between VPD and PAR resulted in complex patterns in the variation of T_{leaf} and ET_{eco} (Figure 4). For example, for T_{leaf} , steeper relationships with $\text{VPD}_{\text{branch}}$ were predicted at low $\text{PAR}_{\text{branch}}$ in both sites. In Abisko, higher ET_{eco} was predicted under conditions of high PAR_{eco} and low VPD_{eco} values (Figure 4).

3.3. Overstorey and understorey contributions to ecosystem evapotranspiration

Higher spatial variability of LAI in Abisko (Table 1) translated into a much larger variability in T_{birch} , while T_{birch} was lower and less variable in Kevo (Figure 5). On average, the daily contribution of mean T_{birch} to ET_{eco} reached peak values of *ca.* 65% in Abisko and *ca.* 30% in Kevo. However, the highly variable LAI in Abisko (Table 1) resulted in the upper bound of T_{birch}/ET_{eco} occasionally approaching 100% at this location (Figure 5c).

The value of T_{birch}/ET_{eco} increased with VPD_{eco} and PAR_{eco} (both log-transformed; Table S4, Figure 6a,b). In both cases, model selection retained the interaction between site and the environmental variable, but it was not significant for either driver (Table S4). We did not detect any effect of SMD on T_{birch}/ET_{eco} (Figure 6c; Table S4). We did not find any difference in T_{birch}/ET_{eco} between dry and wet days ($p = 0.27$).

The mean growing season contribution of T_{birch} to ET_{eco} was relatively low in Abisko (*ca.* 33%) but it was even lower in Kevo (16%, Table 4). Daily evapotranspiration by understorey components in Kevo was generally lower compared to T_{birch} (Figure 5d). For the whole of the growing season, $ET_{upscaled}$ only amounted to *ca.* 40% of ET_{eco} in Kevo (Table 4).

Daily ET_{eco} was higher in Abisko (Figure 5a,b), which also showed higher growing season totals compared to Kevo (Table 4). Remarkably, in Abisko ET_{eco} was 27% higher than the precipitation in the same period, while in Kevo the ecosystem returned to the atmosphere only *ca.* 59% of precipitation (ET_{eco}/P , Table 4). Nevertheless, the relative role of mountain birch transpiration in recycling precipitation was much higher in Abisko than in Kevo (T_{birch}/P , Table 4).

4. Discussion

4.1. Differences in seasonal and environmental controls on transpiration and evapotranspiration between sites

Boreal and arctic regions are undergoing very rapid and pronounced climatic warming, which is expected to modify water and energy fluxes across much of the terrestrial biosphere of these northern regions. We find that controls of evaporative fluxes by mixed birch-tundra communities of Northern Fennoscandia largely consist of controls by VPD (which strongly depends on air and canopy temperature) and by PAR. The relative importance of these effects depended partly on specific site conditions and the scale (branch versus ecosystem) at which

they were considered. Predicted increases in air temperature can therefore be expected to increase the relative contribution of VPD relative to PAR in controlling evaporative fluxes.

Conversely, we find that the evaporative fluxes are not affected by temporal changes in soil moisture, suggesting that water supply is currently not a major limiting factor to evapotranspiration. Thus, there were no edaphic drought stress effects in T_{leaf} regulation by mountain birch, confirming results observed for other birch species (Gartner, Nadezhdina, Englisch, Čermak, & Leitgeb, 2009; Yan et al., 2018). Our results at the ecosystem level are consistent with field studies in forest-tundra systems (Beringer et al., 2005) and with a recent data synthesis, where no effect of soil moisture was reported for evapotranspiration at high latitudes (Kasurinen et al., 2014). Nevertheless, evaporative fluxes in boreal forests in more continental climates, with higher evaporative demands, may be influenced by soil moisture (Ohta et al., 2008).

At the seasonal time scale, fluxes were primarily controlled by LAI dynamics at both sites (cf. Poyatos et al., 2012). Seasonal courses of T_{leaf} and ET_{eco} mirrored each other, except during the start of the growing season in Abisko, when the discrepancy between T_{leaf} and ET_{eco} may have been caused by combined errors in the quantification of low fluxes and leaf area during early leaf development. Alternatively, this temporal mismatch between T_{leaf} and ET_{eco} may have been driven by substantial evaporation from moist soils after snowmelt and/or spatial variability in the phenology of greening up between the measured branches and the rest of the forest.

Both T_{leaf} and ET_{eco} were higher in Abisko than in Kevo because of the generally higher evaporative demand in Abisko (Figure 2). Environmental controls on T_{leaf} were very similar across sites. The only significant difference in the response of T_{leaf} to PAR may be due to differences in stand structure at the plot level (Table 1). The responses of evaporative fluxes to PAR and VPD differed between the two sites more clearly for T_{leaf} than for ET_{eco} , suggesting a higher sensitivity to VPD of the birch canopy compared to other ecosystem components (see also section 4.2). The negative interaction between VPD and PAR produced complex response surfaces of evaporative fluxes to environmental conditions. Model responses during conditions of high evaporative demand were reasonable, apart from those by ET_{eco} at Abisko, where the model showed a decrease of ET_{eco} with VPD at high PAR. The more extreme responses were found for unrealistic combinations of environmental

conditions, which are not usually found in the field. (i.e. high VPD and low PAR), and when the model's predictions of the interaction effects are less reliable.

4.2. Contribution of mountain birch transpiration to ecosystem evapotranspiration across sites and environmental conditions

The mean daily contribution of birch transpiration to ecosystem evapotranspiration (i.e. T_{birch}/ET_{eco}) was much higher in Abisko than in Kevo. In Abisko, the higher variability in LAI at the landscape level propagates to a larger range of T_{birch}/ET_{eco} values compared to Kevo. When explaining seasonal variability in T_{birch}/ET_{eco} , we found that T_{birch}/ET_{eco} strongly depended on VPD and PAR, with T_{birch}/ET_{eco} saturating at high VPD, but this environmental control on T_{birch}/ET_{eco} was stronger in Abisko. Therefore, our results show an increased relative role of mountain birch in controlling ecosystem evapotranspiration as evaporative demand increases, especially in denser forests, in contrast with studies on waterlogged peatlands where understorey contribution increases with VPD (Ikawa et al., 2015). In our sites, mountain birch roots possibly access soil moisture at greater depths (Hunziker, Sigurdsson, Halldorsson, Schwanghart, & Kuhn, 2014), supplying water to meet the increasing evaporative demand and causing the increase in T_{birch}/ET_{eco} .

At the growing season level, birch transpiration contributed *ca.* 33% of total ecosystem evapotranspiration in Abisko but the contribution was only *ca.* 16% in Kevo (Table 4). These differences were attributable not only to a higher birch LAI in Abisko (Table 1), but also to the higher T_{leaf} values at this site (Figure 2). Lower T_{birch}/ET_{eco} values in Kevo could also result from a disproportionately higher contribution from the understorey in a sparser woodland (i.e. higher below-canopy incident radiation compared to Abisko). The values of T_{birch}/ET_{eco} at the two sites are consistent with the generally low contribution of overstorey to total evapotranspiration in subarctic and northern boreal forests (Iida et al., 2009; Ikawa et al., 2015; Kelliher et al., 1997; Lafleur, 1992; Warren et al., 2019). However, in Kevo, our estimates of upscaled evapotranspiration from individual ecosystem components (i.e. mountain birch, understorey shrubs and lichen heath) yielded growing season values, which were still far from total ecosystem evapotranspiration measured by eddy covariance (Table 4, cf. section 4.3). In the following section, we discuss potential methodological artefacts and unmeasured processes that could explain this discrepancy.

4.3. Methodological considerations

This study jointly analyses a multi-scale dataset of evaporative fluxes from subarctic forest communities. Comparing evaporative fluxes across scales is hindered by the numerous potential errors associated with measurement techniques and upscaling procedures. Transpiration measurements from closed chambers could have been affected by radiation-driven overheating (Poyatos et al., 2012), by raising VPD_{branch} above VPD_{eco} and causing an overestimation of T_{leaf} . However, the relatively low values of T_{birch} and $ET_{upscaled}$, both based on closed chamber measurements, do not suggest that the conclusions of this study could have been affected by this artefact.

The upscaling procedure also has a number of potential limitations that warrant consideration. Due to the sparseness of the forest in Kevo (i.e. little shading effects on understorey vegetation), we assumed that the magnitude and regulation of understorey evapotranspiration was similar to that shown by patches with similar composition in the forest-tundra transition (Poyatos et al., 2014). However, LAI of the patches measured with automated chambers in the forest-mire transition (see section 2.5) was *ca.* 50% of the LAI actually measured in survey plots located within the forest (Table 1). Rescaling the understorey fluxes according to this understorey LAI, evapotranspiration from understorey components at the ecosystem level would be amount to 23.6 mm, an evaporative flux 55% larger than T_{birch} . Scaling-up evapotranspiration estimated from canopy and understorey components, accounting for their land cover fractions and applying the LAI correction outlined above to understorey measurements would increase growing season $ET_{upscaled}$ values to 44.4 mm, or *ca.* 45% of ET_{eco} .

4.4. Differences in growing season water balance across sites

Even accounting for this likely underestimation of ET_{shrub} and ET_{lichen} , there is still a fraction of ET_{eco} that cannot be explained by upscaled gas exchange measurements from individual ecosystem components. Taking into account that T_{birch} obtained from branch-bag measurements excludes evaporation of intercepted water, we showed that T_{birch}/ET_{eco} does not vary between dry and wet days. This may suggest that evaporation of intercepted water may not be captured by eddy covariance measurements, otherwise T_{birch}/ET_{eco} would have been lower on wet days than dry days. Potentially high evaporation rates after precipitation may be partially missed from ET_{eco} and $ET_{upscaled}$, because data from open-path gas analysers are removed when the sensor windows are wet and subsequent gap-filling would not account for

the missed evaporation of intercepted water (Oishi, Oren, & Stoy, 2008). Combined interception by overstorey canopies and mosses in northern boreal forests may amount up to 40% of bulk precipitation (Price, Dunham, Carleton, & Band, 1997), and we are not currently accounting for this substantial contribution.

We found stark differences between sites in the percentage of precipitation returned to the atmosphere as evapotranspiration; the mountain birch woodland in Abisko evaporated more water than it received during the growing season, as observed in other deciduous boreal forests (Blanken et al., 2001; Kelliher et al., 1997). In contrast, Kevo showed a substantial water surplus (Table 4). Our measurements did not include the snowmelt period, but these sites can reach snowpack depths of > 1 m (data for Kevo, 2009) and tree water uptake during this period, especially from deciduous species, can progressively deplete soil water sources (Young-Robertson et al., 2016). This decline in soil water content after snowmelt is very clear in the seasonal course of SMD measured in Abisko in 2008 and 2009 (outside our measurement period in Abisko, Fig S5). Therefore, these differences in the role of the mountain birch canopy between Abisko and Kevo, mediated by their different stand structure, can illustrate the potential changes in the hydrological regime that can result from the expansion and densification of subarctic deciduous woodlands.

4.5. Concluding remarks

We have shown that the dominant mountain birch canopy plays only a partial role in driving ecosystem evapotranspiration in both subarctic sites, and this may be a general feature of low-LAI subarctic and northern boreal forests (Saugier, Granier, Pontailier, Dufrene, & Baldocchi, 1997). Our results also show that both increased woodland cover and increased woodland density under climate change conditions (Rundqvist et al., 2011) will result in larger controls of the water fluxes by the canopies of deciduous trees as opposed to the understorey vegetation. However, our upscaling exercise also shows that adequately accounting for understorey components (and transpiration vs evaporation processes; Stoy et al., 2019) may be necessary to constrain future hydrological changes in these areas. The highly variable and patchy nature of subarctic vegetation may require flux upscaling approaches considering spatial variation not only of land cover (Hartley et al., 2015), but also of LAI (Stoy et al., 2013).

In the longer term, shifts towards deciduous-dominated communities in subarctic regions and an increased land cover by forest as opposed to tundra are expected to induce large hydro-

climatic effects. These effects are expected to be mediated by higher transpiration rates, inducing complex land-climate feedbacks (Bonfils et al., 2012; Swann et al., 2010), which need to be considered together with carbon- and energy-related feedbacks (Wit et al., 2014). Overall, combining several flux datasets and land cover information we provide, for the poorly studied subarctic deciduous woodlands, highly valuable results that will help to calibrate and validate evapotranspiration processes in ecosystem models.

Acknowledgements

AM. Sabater is supported by European Social Fund and Generalitat Valenciana (GVA) under a PhD contract (ACIF – 2017/9830). This study was funded by the following grants: ABACUS NE/D005795/1 (NERC, UK), SAPFLUXNET CGL2014-JIN-55583 (MINECO, Spain), VERSUS CGL2015-67466-R (MINECO/FEDER), SGR-2017-1001 (AGAUR, Generalitat de Catalunya) and IMAGINA PROMETEU/2019/110 (Conselleria de Cultura, GVA). We are grateful for the support by the staff at the Kevo Subarctic Research Institute and at the Abisko Research Station. We would also like to acknowledge the help in the field by T. August, A. Robertson, K. Leslie, D. Sayer and J.R.M. Allen.

References

- Aubinet, M., Vesala, T., & Papale, D. (2012). *Eddy Covariance: A Practical Guide to Measurement and Data Analysis*. Springer Science & Business Media.
- Beringer, J., Chapin, F. S., Thompson, C. C., & McGuire, A. D. (2005). Surface energy exchanges along a tundra-forest transition and feedbacks to climate. *Agricultural and Forest Meteorology*, 131(3–4), 143–161.
- Blanken, P. D., Black, T. A., Neumann, H. H., den Hartog, G., Yang, P. C., Nesic, Z., & Lee, X. (2001). The seasonal water and energy exchange above and within a boreal aspen forest. *Journal of Hydrology*, 245(1–4), 118–136. [https://doi.org/10.1016/S0022-1694\(01\)00343-2](https://doi.org/10.1016/S0022-1694(01)00343-2)
- Bonan, G. B. (2008). Forests and Climate Change: Forcings, Feedbacks, and the Climate Benefits of Forests. *Science*, 320, 1444–1449. <https://doi.org/10.1126/science.1155121>
- Bonfils, C. J. W., Phillips, T. J., Lawrence, D. M., Cameron-Smith, P., Riley, W. J., & Subin, Z. M. (2012). On the influence of shrub height and expansion on northern high latitude climate. *Environmental Research Letters*, 7(1), 015503. <https://doi.org/10.1088/1748-9326/7/1/015503>
- Brümmer, C., Black, T. A., Jassal, R. S., Grant, N. J., Spittlehouse, D. L., Chen, B., ... Wofsy, S. C. (2011). How climate and vegetation type influence evapotranspiration and water use efficiency in Canadian forest, peatland and grassland ecosystems. *Agricultural and Forest Meteorology*. <https://doi.org/10.1016/j.agrformet.2011.04.008>

- Callaghan, T. V., Björn, L. O., Chapin Iii, F., Chernov, Y., Christensen, T. R., Huntley, B., ... Shaver, G. R. (2005). Arctic tundra and polar desert ecosystems. In *Arctic climate impact assessment* (Vol. 1, pp. 243–352).
- Callaghan, T. V., Jonasson, C., Thierfelder, T., Yang, Z., Hedenås, H., Johansson, M., ... Sloan, V. L. (2013). Ecosystem change and stability over multiple decades in the Swedish subarctic: Complex processes and multiple drivers. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 368(1624), 20120488. <https://doi.org/10.1098/rstb.2012.0488>
- Chapin, F. S., Mcguire, A. D., Randerson, J., Pielke, R., Baldocchi, D., Hobbie, S. E., ... Running, S. W. (2000). Arctic and boreal ecosystems of western North America as components of the climate system. *Global Change Biology*, 6(S1), 211–223. <https://doi.org/10.1046/j.1365-2486.2000.06022.x>
- Cohen, L. R., Raz-Yaseef, N., Curtis, J. B., Young, J. M., Rahn, T. A., Wilson, C. J., ... Newman, B. D. (2015). Measuring diurnal cycles of evapotranspiration in the Arctic with an automated chamber system. *Ecohydrology*, 8(4), 652–659. <https://doi.org/10.1002/eco.1532>
- Dahlberg, U., Berge, T. W., Petersson, H., & Vencatasawmy, C. P. (2004). Modelling biomass and leaf area index in a sub-arctic Scandinavian mountain area. *Scandinavian Journal of Forest Research*, 19, 60–71. <https://doi.org/10.1080/02827580310019266>
- Fletcher, B. J., Gornall, J. L., Poyatos, R., Press, M. C., Stoy, P. C., Huntley, B., ... Phoenix, G. K. (2012). Photosynthesis and productivity in heterogeneous arctic tundra: Consequences for ecosystem function of mixing vegetation types at stand edges. *Journal of Ecology*, 100(2), 441–451. <https://doi.org/10.1111/j.1365-2745.2011.01913.x>
- Gartner, K., Nadezhdina, N., Englisch, M., Čermak, J., & Leitgeb, E. (2009). Sap flow of birch and Norway spruce during the European heat and drought in summer 2003. *Forest Ecology and Management*, 258(5), 590–599. <https://doi.org/10.1016/j.foreco.2009.04.028>
- Gash, J. H. C., & Dolman, A. J. (2003). Sonic anemometer (co)sine response and flux measurement: I. The potential for (co)sine error to affect sonic anemometer-based flux measurements. *Agricultural and Forest Meteorology*, 119(3), 195–207. [https://doi.org/10.1016/S0168-1923\(03\)00137-0](https://doi.org/10.1016/S0168-1923(03)00137-0)
- Granier, A., & Loustau, D. (1994). Measuring and modelling the transpiration of a maritime pine canopy from sap-flow data. *Agricultural and Forest Meteorology*, 71, 61–81.
- Grelle, A., Lundberg, A., Lindroth, A., Morén, A.-S., & Cienciala, E. (1997). Evaporation components of a boreal forest: Variations during the growing season. *Journal of Hydrology*, 197(1), 70–87. [https://doi.org/10.1016/S0022-1694\(96\)03267-2](https://doi.org/10.1016/S0022-1694(96)03267-2)
- Haapanala, S., Ekberg, A., Hakola, H., Tarvainen, V., Rinne, J., Hellén, H., & Arneth, A. (2009). Mountain birch—potentially large source of sesquiterpenes into high latitude atmosphere. *Biogeosciences*, 6, 2709–2718.
- Hartley, I. P., Hopkins, D. W., Sommerkorn, M., & Wookey, P. A. (2010). The response of organic matter mineralisation to nutrient and substrate additions in sub-arctic soils. *Soil Biology and Biochemistry*, 42(1), 92–100.
- Hartley, Iain. P., Hill, Timothy. C., Wade, Thomas. J., Clement, Robert. J., Moncrieff, John. B., Prieto-Blanco, Ana., ... Baxter, Robert. (2015). Quantifying landscape-level methane fluxes in subarctic Finland using a multiscale approach. *Global Change Biology*, 21(10), 3712–3725. <https://doi.org/10.1111/gcb.12975>
- Hofgaard, A., Tømmervik, H., Rees, G., & Hanssen, F. (2013). Latitudinal forest advance in northernmost Norway since the early 20th century. *Journal of Biogeography*, 40(5), 938–949. <https://doi.org/10.1111/jbi.12053>

- Hunziker, M., Sigurdsson, B. D., Halldorsson, G., Schwanghart, W., & Kuhn, N. (2014). Biomass allometries and coarse root biomass distribution of mountain birch in southern Iceland. *Icelandic Agricultural Sciences*, 27, 111–125.
- Iida, S., Ohta, T., Matsumoto, K., Nakai, T., Kuwada, T., Kononov, A. V., ... Yabuki, H. (2009). Evapotranspiration from understory vegetation in an eastern Siberian boreal larch forest. *Agricultural and Forest Meteorology*, 149(6), 1129–1139. <https://doi.org/10.1016/j.agrformet.2009.02.003>
- Ikawa, H., Nakai, T., Busey, R. C., Kim, Y., Kobayashi, H., Nagai, S., ... Hinzman, L. (2015). Understory CO₂, sensible heat, and latent heat fluxes in a black spruce forest in interior Alaska. *Agricultural and Forest Meteorology*, 214–215, 80–90. <https://doi.org/10.1016/j.agrformet.2015.08.247>
- IPCC. (2013). *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. <https://doi.org/10.1017/CBO9781107415324>
- Kasurinen, V., Alfredsen, K., Kolari, P., Mammarella, I., Alekseychik, P., Rinne, J., ... Berninger, F. (2014). Latent heat exchange in the boreal and arctic biomes. *Global Change Biology*, 20(11), 3439–3456. <https://doi.org/10.1111/gcb.12640>
- Kattsov, V. M., Källén, E., Cattle, H. P., Christensen, J., Drange, H., Hanssen-Bauer, I., ... others. (2005). *Future climate change: Modeling and scenarios for the Arctic*.
- Kelliher, F. M., Hollinger, D. Y., Schulze, E.-D., Vygodskaya, N. N., Byers, J. N., Hunt, J. E., ... Bauer, G. (1997). Evaporation from an eastern Siberian larch forest. *Agricultural and Forest Meteorology*, 85(3), 135–147. [https://doi.org/10.1016/S0168-1923\(96\)02424-0](https://doi.org/10.1016/S0168-1923(96)02424-0)
- Knauer, J., Werner, C., & Zaehle, S. (2015). Evaluating stomatal models and their atmospheric drought response in a land surface scheme: A multibiome analysis. *Journal of Geophysical Research: Biogeosciences*, 120(10), 2015JG003114. <https://doi.org/10.1002/2015JG003114>
- Krankina, O. N., Pflugmacher, D., Hayes, D. J., McGuire, A. D., Hansen, M. C., Häme, T., ... Nelson, P. (2010). Vegetation cover in the eurasian arctic: Distribution, monitoring, and role in carbon cycling. In *Eurasian arctic land cover and land use in a changing climate* (pp. 79–108). Springer.
- Lafleur, P. M. (1992). Energy balance and evapotranspiration from a subarctic forest. *Agricultural and Forest Meteorology*, 58(3–4), 163–175. [https://doi.org/10.1016/0168-1923\(92\)90059-D](https://doi.org/10.1016/0168-1923(92)90059-D)
- McFadden, J. P., Eugster, W., & Chapin III, F. S. (2003). A regional study of the controls on water vapor and CO₂ exchange in arctic tundra. *Ecology*, 84(10), 2762–2776.
- Mekonnen, Z. A., Riley, W. J., Randerson, J. T., Grant, R. F., & Rogers, B. M. (2019). Expansion of high-latitude deciduous forests driven by interactions between climate warming and fire. *Nature Plants*, 1–7. <https://doi.org/10.1038/s41477-019-0495-8>
- Myers-Smith, I. H., Forbes, B. C., Wilmking, M., Hallinger, M., Lantz, T., Blok, D., ... Hik, D. S. (2011). Shrub expansion in tundra ecosystems: Dynamics, impacts and research priorities. *Environmental Research Letters*, 6(4), 045509. <https://doi.org/10.1088/1748-9326/6/4/045509>
- Nakagawa, S., & Schielzeth, H. (2013). A general and simple method for obtaining R² from generalized linear mixed-effects models. *Methods in Ecology and Evolution*, 4(2), 133–142. <https://doi.org/10.1111/j.2041-210x.2012.00261.x>
- Nyström, M., Holmgren, J., & Olsson, H. (2012). Prediction of tree biomass in the forest–tundra ecotone using airborne laser scanning. *Remote Sensing of Environment*, 123, 271–279. <https://doi.org/10.1016/j.rse.2012.03.008>

- Ohta, T., Maximov, T. C., Dolman, A. J., Nakai, T., van der Molen, M. K., Kononov, A. V., ... Yabuki, H. (2008). Interannual variation of water balance and summer evapotranspiration in an eastern Siberian larch forest over a 7-year period (1998–2006). *Agricultural and Forest Meteorology*, 148(12), 1941–1953. <https://doi.org/10.1016/j.agrformet.2008.04.012>
- Oishi, A. C., Oren, R., & Stoy, P. C. (2008). Estimating components of forest evapotranspiration: A footprint approach for scaling sap flux measurements. *Agricultural and Forest Meteorology*, 148(11), 1719–1732.
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team. (2018). *nlme: Linear and Nonlinear Mixed Effects Models*. Retrieved from <https://CRAN.R-project.org/package=nlme>
- Poyatos, R., Gornall, J., Mencuccini, M., Huntley, B., & Baxter, R. (2012). Seasonal controls on net branch CO₂ assimilation in sub-Arctic Mountain Birch (*Betula pubescens* ssp. *Czerepanovii* (Orlova) Hamet-Ahti). *Agricultural and Forest Meteorology*, 158–159, 90–100. <https://doi.org/10.1016/j.agrformet.2012.02.009>
- Poyatos, R., Heinemeyer, A., Ineson, P., Evans, J. G., Ward, H. C., Huntley, B., & Baxter, R. (2014). Environmental and Vegetation Drivers of Seasonal CO₂ Fluxes in a Sub-arctic Forest–Mire Ecotone. *Ecosystems*, 17(3), 377–393. <https://doi.org/10.1007/s10021-013-9728-2>
- Price, A. G., Dunham, K., Carleton, T., & Band, L. (1997). Variability of water fluxes through the black spruce (*Picea mariana*) canopy and feather moss (*Pleurozium schreberi*) carpet in the boreal forest of Northern Manitoba. *Journal of Hydrology*, 196(1), 310–323. [https://doi.org/10.1016/S0022-1694\(96\)03233-7](https://doi.org/10.1016/S0022-1694(96)03233-7)
- Rayment, M. B., & Jarvis, P. G. (1999). Seasonal gas exchange of black spruce using an automatic branch bag system. *Canadian Journal of Forest Research*, 29, 1528–1538.
- Rundqvist, S., Hedenås, H., Sandström, A., Emanuelsson, U., Eriksson, H., Jonasson, C., & Callaghan, T. V. (2011). Tree and Shrub Expansion Over the Past 34 Years at the Tree-Line Near Abisko, Sweden. *AMBIO: A Journal of the Human Environment*, 40(6), 683–692. <https://doi.org/10.1007/s13280-011-0174-0>
- Saugier, B., Granier, A., Pontailier, J. Y., Dufrene, E., & Baldocchi, D. D. (1997). Transpiration of a boreal pine forest measured by branch bag, sap flow and micrometeorological methods. *Tree Physiology*, 17, 511–519.
- Schlesinger, W. H., & Jasechko, S. (2014). Transpiration in the global water cycle. *Agricultural and Forest Meteorology*, 189–190, 115–117. <https://doi.org/10.1016/j.agrformet.2014.01.011>
- Schotanus, P., Nieuwstadt, F. T. M., & De Bruin, H. A. R. (1983). Temperature measurement with a sonic anemometer and its application to heat and moisture fluxes. *Boundary-Layer Meteorology*, 26(1), 81–93. <https://doi.org/10.1007/BF00164332>
- Stoy, P. C., Williams, M., Evans, J. G., Prieto-Blanco, A., Disney, M., Hill, T. C., ... Street, L. E. (2013). Upscaling tundra CO₂ exchange from chamber to eddy covariance tower. *Arctic, Antarctic, and Alpine Research*, 45(2), 275–284.
- Stoy, Paul C., El-Madany, T. S., Fisher, J. B., Gentile, P., Gerken, T., Good, S. P., ... Wolf, S. (2019). Reviews and syntheses: Turning the challenges of partitioning ecosystem evaporation and transpiration into opportunities. *Biogeosciences*, 16(19), 3747–3775. <https://doi.org/10.5194/bg-16-3747-2019>
- Stoy, Paul C., Mauder, M., Foken, T., Marcolla, B., Boegh, E., Ibrom, A., ... Varlagin, A. (2013). A data-driven analysis of energy balance closure across FLUXNET research sites: The role of landscape scale heterogeneity. *Agricultural and Forest Meteorology*, 171–172, 137–152. <https://doi.org/10.1016/j.agrformet.2012.11.004>

- Swann, A. L., Fung, I. Y., Levis, S., Bonan, G. B., & Doney, S. C. (2010). Changes in Arctic vegetation amplify high-latitude warming through the greenhouse effect. *Proceedings of the National Academy of Sciences*, 107(4), 1295–1300. <https://doi.org/10.1073/pnas.0913846107>
- Tømmervik, H., Johansen, B., Tombre, I., Thannheiser, D., Høgda, K. A., Gaare, E., & Wielgolaski, F. E. (2004). Vegetation Changes in the Nordic Mountain Birch Forest: The Influence of Grazing and Climate Change. *Arctic, Antarctic, and Alpine Research*, 36(3), 323–332.
- Wang, J. A., Sulla-Menashe, D., Woodcock, C. E., Sonnentag, O., Keeling, R. F., & Friedl, M. A. (2019). Extensive land cover change across Arctic–Boreal Northwestern North America from disturbance and climate forcing. *Global Change Biology*, 0(0). <https://doi.org/10.1111/gcb.14804>
- Warren, R. K., Pappas, C., Helbig, M., Chasmer, L. E., Berg, A. A., Baltzer, J. L., ... Sonnentag, O. (2019). Minor contribution of overstorey transpiration to landscape evapotranspiration in boreal permafrost peatlands. *Ecohydrology*, 11(5), e1975. <https://doi.org/10.1002/eco.1975>
- Webb, E. K., Pearman, G. I., & Leuning, R. (1980). Correction of flux measurements for density effects due to heat and water vapour transfer. *Quarterly Journal of the Royal Meteorological Society*, 106(447), 85–100. <https://doi.org/10.1002/qj.49710644707>
- Welp, L. R., Randerson, J. T., & Liu, H. P. (2007). The sensitivity of carbon fluxes to spring warming and summer drought depends on plant functional type in boreal forest ecosystems. *Agricultural & Forest Meteorology*, 147(3–4), 172–185.
- Wingate, L., Seibt, U., Moncrieff, J. B., Jarvis, P. G., & Lloyd, J. (2007). Variations in $\delta^{13}C$ discrimination during CO₂ exchange by *Picea sitchensis* branches in the field. *Plant, Cell & Environment*, 30(5), 600–616. <https://doi.org/10.1111/j.1365-3040.2007.01647.x>
- Wit, H. A. de, Bryn, A., Hofgaard, A., Karstensen, J., Kvilevåg, M. M., & Peters, G. P. (2014). Climate warming feedback from mountain birch forest expansion: Reduced albedo dominates carbon uptake. *Global Change Biology*, 20(7), 2344–2355. <https://doi.org/10.1111/gcb.12483>
- Yan, C., Wang, B., Zhang, Y., Zhang, X., Takeuchi, S., & Qiu, G. (2018). Responses of Sap Flow of Deciduous and Conifer Trees to Soil Drying in a Subalpine Forest. *Forests*, 9(1), 32. <https://doi.org/10.3390/f9010032>
- Young-Robertson, J. M., Bolton, W. R., Bhatt, U. S., Cristóbal, J., & Thoman, R. (2016). Deciduous trees are a large and overlooked sink for snowmelt water in the boreal forest. *Scientific Reports*, 6, srep29504. <https://doi.org/10.1038/srep29504>
- Zhang, W., Miller, P. A., Smith, B., Wania, R., Koenigk, T., & Döscher, R. (2013). Tundra shrubification and tree-line advance amplify arctic climate warming: Results from an individual-based dynamic vegetation model. *Environmental Research Letters*, 8(3), 034023. <https://doi.org/10.1088/1748-9326/8/3/034023>
- Zuur, A. F., Ieno, E. N., & Elphick, C. S. (2010). A protocol for data exploration to avoid common statistical problems. *Methods in Ecology and Evolution*, 1(1), 3–14. <https://doi.org/10.1111/j.2041-210X.2009.00001.x>

Table 1. Stand characteristics of mountain birch forests in Abisko and Kevo. Values labelled as ‘Site’ represent the site mean (\pm SE) of all inventory plots in Abisko (N=6) and Kevo (N=8). Values labelled as ‘BB’ are the values of the plots in the vicinity of the branch bags measuring sites. Tree density refers to polycormic individuals, with multiple stems per tree.

	Tree density (trees ha ⁻¹)	Stems per tree	Basal area (m ² ha ⁻¹)	DBH (mm)	Height (m)	Overstorey LAI _{max} (m ² m ⁻²)	Understorey LAI _{max} (m ² m ⁻²)
Abisko							
Site	1260 \pm 80	3.7 \pm 0.2	6.5 \pm 0.2	36.9 \pm 1.0	3.9 \pm 0.2	1.2 \pm 0.3	1.0 \pm 0.2
BB	1146	4.4	6.7	35.7	-	1.8	-
Kevo							
Site	876 \pm 85	3.3 \pm 0.2	3.8 \pm 0.4	37.2 \pm 2.0	3.8 \pm 0.0	0.7 \pm 0.1	1.5 \pm 0.1
BB	833	3.8	3.0	30.6	3.8	0.6	-

Table 2. Summary statistics of the linear mixed model of log-transformed T_{leaf} as a function of environmental variables (VPD_{branch} , PAR_{branch} and SMD) for Abisko and Kevo. Asterisks denote significant differences from zero (* $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$). Statistical differences in model coefficients ($p < 0.05$) between Abisko and Kevo were marked in bold. SD: Standard deviation. Interactions between variables are denoted by colon (:) and variables not included after model selection are denoted by ‘n.i.’.

	Abisko	Kevo
Fixed effects		
Intercept	-2.98 ± 0.09*	-4.00 ± 0.07***
log(VPD_{branch})	1.26 ± 0.01***	1.27 ± 0.01***
PAR_{branch}	4.1·10⁻⁴ ± 0.4·10⁻⁴***	7.5·10⁻⁴ ± 0.4·10⁻⁴***
log(VPD_{branch}): PAR_{branch}	-8.4·10 ⁻⁴ ± 0.5·10 ⁻⁴ ***	-9.4·10 ⁻⁴ ± 0.4·10 ⁻⁴ ***
SMD	n.i.	n.i.
Random effects (branch)		
SD (Intercept)	0.26	0.20
Residual error	0.40	0.48
Correlation structure (ϕ)	4.40·10 ⁻⁸	7.23·10 ⁻⁷
R ² marginal (R ² conditional)	0.78 (0.84)	0.77 (0.80)

Table 3. Summary statistics of the generalised least squares model of ET_{eco} as a function of environmental variables (VPD_{eco} , PAR_{eco} and SMD) for Abisko and Kevo. Asterisks denote significant differences from zero (* $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$). No significant differences ($p < 0.05$) were found between model coefficients between Abisko and Kevo. Interactions between variables are denoted by colon (:) and variables not included after model selection are denoted by ‘n.i.’.

	Abisko	Kevo
Intercept	$-4.93 \pm 0.38^{***}$	$-4.84 \pm 0.40^{***}$
$\log(VPD)$	$2.58 \pm 0.30^{***}$	$2.13 \pm 0.27^{***}$
$\log(PAR_{eco})$	$0.47 \pm 0.06^{***}$	$0.47 \pm 0.06^{***}$
$\log(VPD_{eco}):\log(PAR_{eco})$	$-0.39 \pm 0.05^{***}$	$-0.26 \pm 0.05^{***}$
SMD	n.i.	n.i.
Correlation structure (φ)	$8.23 \cdot 10^{-3}$	$1.84 \cdot 10^{-2}$
R^2 marginal	0.71	0.69

Table 4. Growing season values of precipitation (P), birch transpiration (T_{birch}) and ecosystem evapotranspiration (ET_{eco}) in Abisko and Kevo. Percentage of evaporative fluxes as a fraction of ET_{eco} and P are also shown for growing season values. Values with an uncertainty measure represent means \pm standard error.

	Abisko	Kevo
T_{birch} (mm)	52.5 ± 13.0	15.2 ± 1.5
ET_{eco} (mm)	160.5	98.5
$ET_{upscaled}$ (mm)	-	39.4 ± 1.5
T_{birch} / ET_{eco} (%)	32.7 ± 8.1	15.5 ± 1.5
T_{birch} / P (%)	41.4 ± 10.2	9.1 ± 0.9
ET_{eco} / P (%)	126.6	58.8
$ET_{upscaled} / ET_{eco}$ (%)	-	40.0 ± 1.5

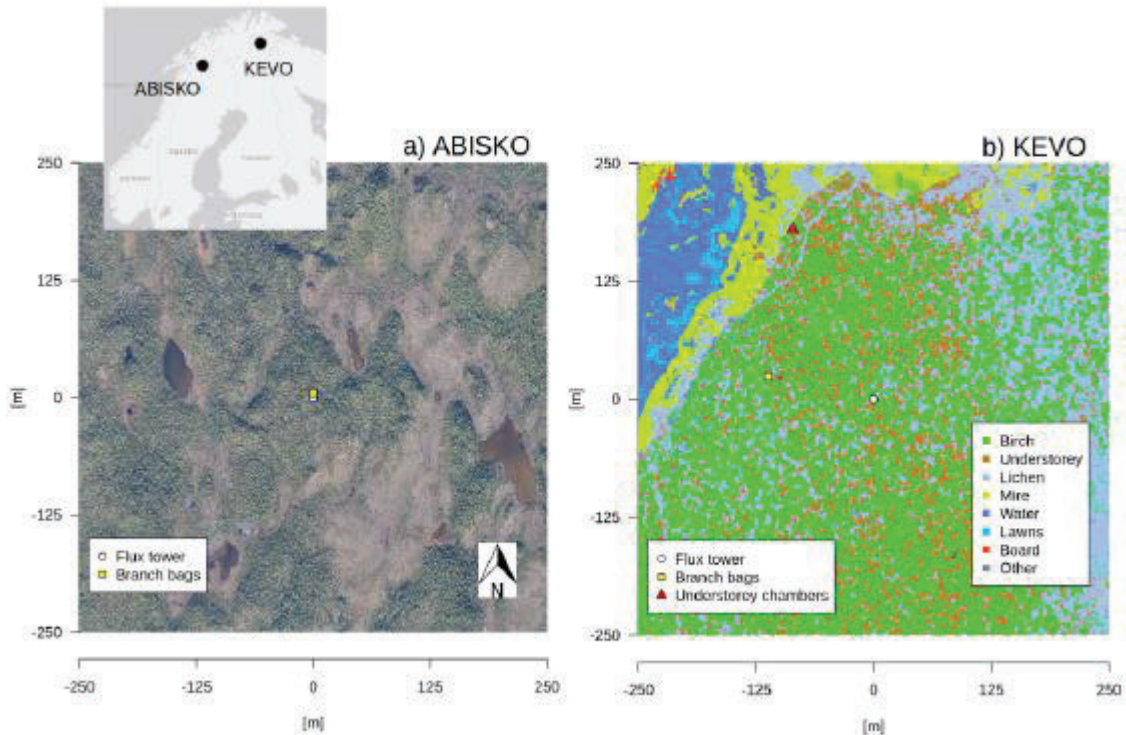


Figure 1. Study sites at Abisko (a) and Kevo (b), showing the locations of the branch bags systems, the eddy flux towers and the understorey automated chambers at Kevo. Panel (a) shows the aerial photography obtained in Abisko and (b) shows the land classification at Kevo obtained from aerial photography (cf. Hartley et al., 2015). *Birch* : mountain birch woodland; *Understorey*: low- and dwarf-shrubs; *Lichen*: lichen heath; *Mire*: organic hummocks and interhummocks with shrubs and *Sphagnum*; *Water*: open water; *Lawns*: graminoid lawns; *Board*: boardwalks; *Other*: other land cover.

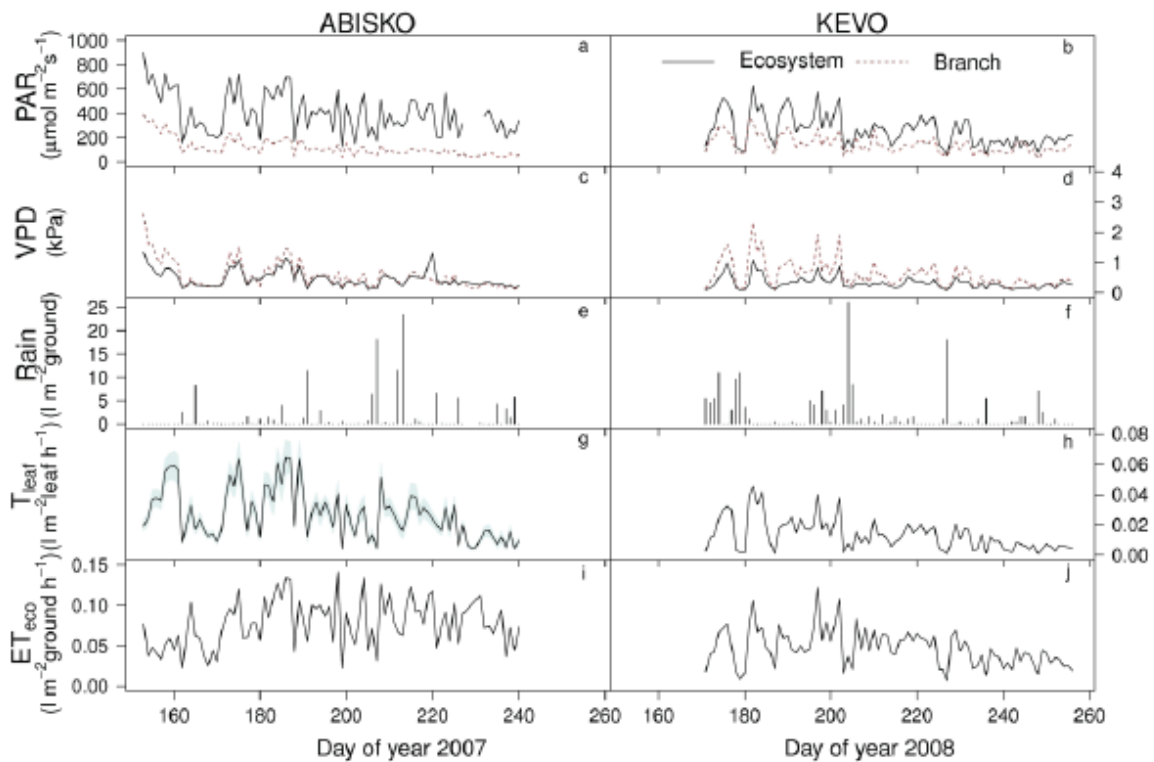


Figure 2. Seasonal course of environmental variables and evaporative fluxes (daily means) in Abisko and Kevo. Environmental variables include photosynthetically active radiation (a-b, PAR), vapour pressure deficit (c-d, VPD) and rainfall (e, f). Environmental variables were measured at the ecosystem (black lines) and at the branch level (red lines). Mountain birch transpiration per unit leaf area (g-h, T_{leaf}) and ecosystem evapotranspiration (i-j, ET_{eco}) are also shown. Standard error is shown as shaded grey.

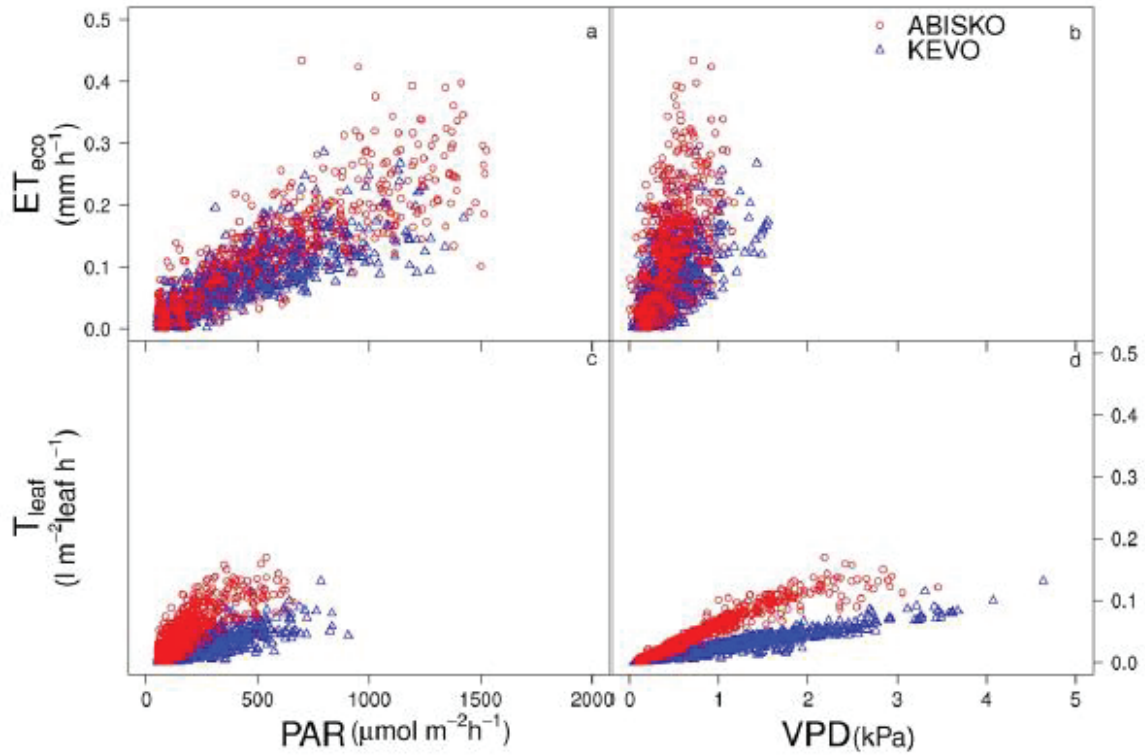


Figure 3. Sub-daily responses of ecosystem evapotranspiration (ET_{eco}) and mountain birch transpiration per unit leaf area (T_{leaf}) to PAR (panels a,c) and VPD (panels b,d), measured at the corresponding ecological scale (i.e. ‘branch’ for T_{leaf} and ‘eco’ for ET_{eco}) in Abisko (red) and Kevo (blue).

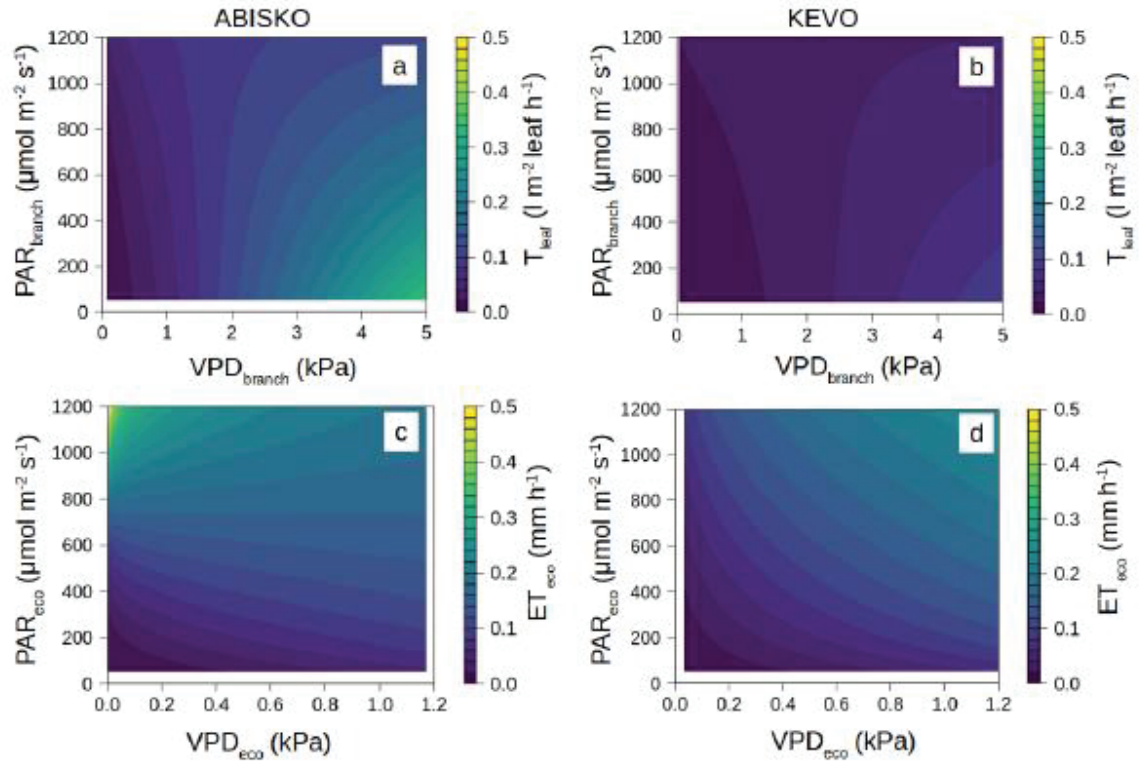


Figure 4. Response surfaces of modelled T_{leaf} (panels a, b) and ET_{eco} (panels c, d) as a function of VPD and PAR, in Abisko (panels a, c) and Kevo (panels b, d). Please note the different scales in the VPD axes in panels a and b compared to panels c and d.

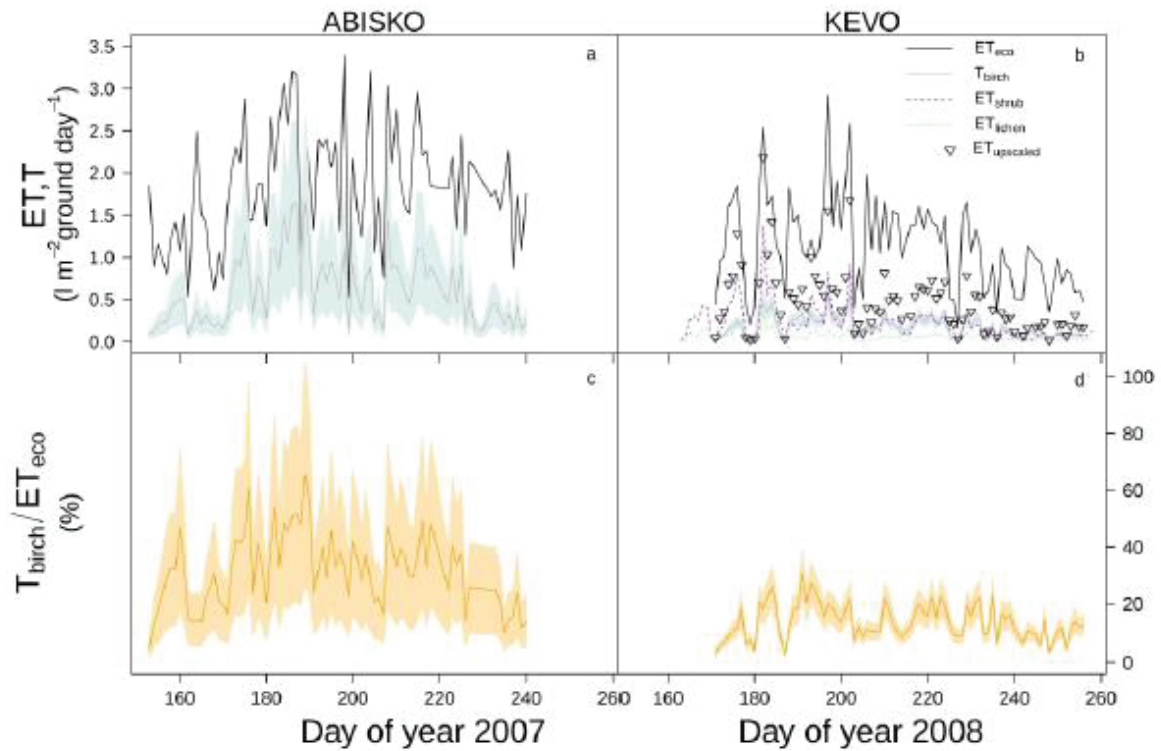


Figure 5. Seasonal course of daily ecosystem evapotranspiration (ET_{eco} , black lines) and upscaled birch transpiration (T_{birch} , grey lines), for Abisko (a) and Kevo (b). The shaded regions in panels a and b depict upscaled T_{birch} using mean \pm SE values of LAI (Table 1). Daily percentage of T_{birch}/ET_{eco} for Abisko (c) and Kevo (d). Panel (f) shows evapotranspiration components and their upscaled values for Kevo only: ET_{eco} (black line), T_{birch} (grey line), ET_{shrub} (purple line), ET_{lichen} (green line), $ET_{upscaled}$ (asterisk).

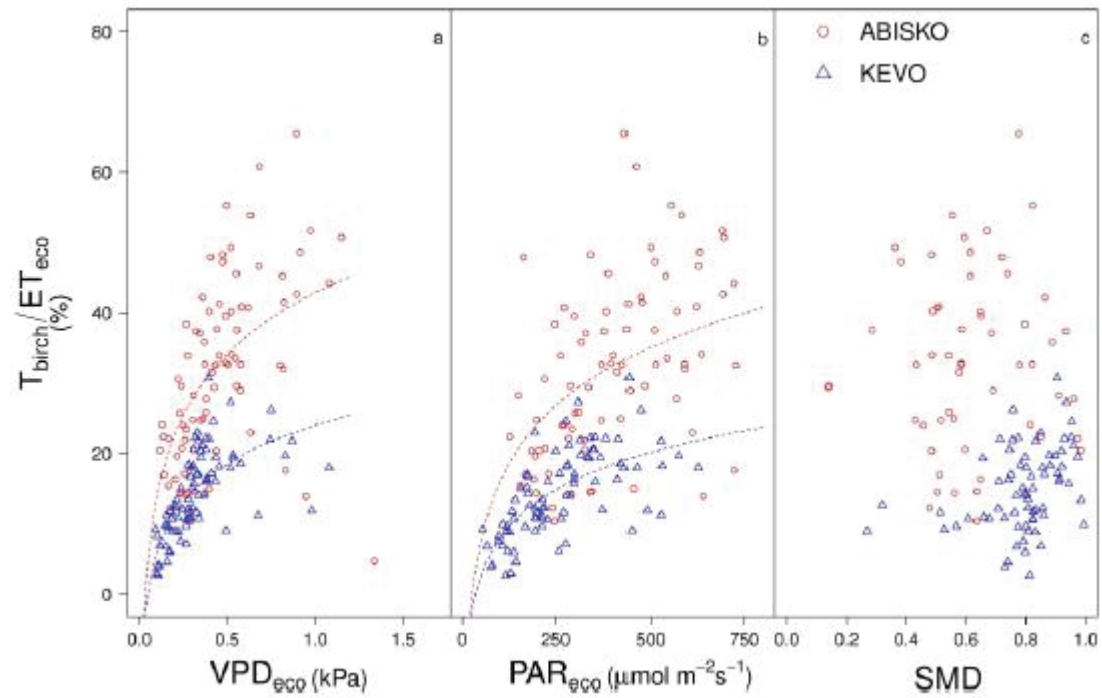


Figure 6. Variation of daily T_{birch} / ET_{eco} in response to VPD_{eco} (a), PAR_{eco} (b) and SMD (c), for Abisko (red) and Kevo (blue). Models summary are shown in Table S3. Significant interaction between site and environmental value is shown in solid line and no-significant interaction in dashed line.