



Article (refereed) - postprint

Vogel, Christian; Sekine, Ryo; Steckenmesser, Daniel; Lombi, Enzo; Herzel, Hannes; Zuin, Lucia; Wang, Dongniu; Félix, Roberto; Adam, Christian. 2019. **Combining diffusive gradients in thin films (DGT) and spectroscopic techniques for the determination of phosphorus species in soils**.

© 2019 Elsevier B.V. This manuscript version is made available under the CC-BY-NC-ND 4.0 license <u>http://creativecommons.org/licenses/by-nc-nd/4.0/</u>

This version available http://nora.nerc.ac.uk/522696/

NERC has developed NORA to enable users to access research outputs wholly or partially funded by NERC. Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at <u>http://nora.nerc.ac.uk/policies.html#access</u>

NOTICE: this is the authors' version of a work that was accepted for publication in *Analytica Chimica Acta*. Changes resulting from the publishing process, such as peer review, editing, corrections, structural formatting, and other quality control mechanisms may not be reflected in this document. Changes may have been made to this work since it was submitted for publication. A definitive version was subsequently published in *Analytica Chimica Acta* (2019), 1057. 80-87. https://doi.org/10.1016/j.aca.2019.01.037

www.elsevier.com/

Contact CEH NORA team at <u>noraceh@ceh.ac.uk</u>

The NERC and CEH trademarks and logos ('the Trademarks') are registered trademarks of NERC in the UK and other countries, and may not be used without the prior written consent of the Trademark owner.



1	Combining Diffusive Gradients in Thin films
2	(DGT) and Spectroscopic Techniques for the
3	Determination of Phosphorus Species in Soils
4	
5	Christian Vogel ¹ *, Ryo Sekine ^{2,3} , Daniel Steckenmesser ⁴ , Enzo Lombi ³ , Hannes Herzel ¹ ,
6 7	Lucia Zuin ⁵ , Dongniu Wang ⁵ , Roberto Félix ⁶ and Christian Adam ¹
8	¹ Division 4.4 Thermochemical Residues Treatment and Resource Recovery, Bundesanstalt
9	für Materialforschung und -prüfung (BAM), Unter den Eichen 87, 12205 Berlin, Germany
10	² Centre for Ecology & Hydrology, Maclean Building, Benson Lane, Crowmarsh Gifford,
11	Wallingford, Oxfordshire OX10 8BB, United Kingdom
12	³ Future Industries Institute, University of South Australia, Building X, Mawson Lakes, SA
13	5095, Australia
14	⁴ Institute of Plant Nutrition, Research Center for Biosystems, Land Use and Nutrition, Justus-
15	Liebig University Giessen, Heinrich-Buff-Ring 26-32, 35392 Giessen, Germany
16	⁵ Canadian Light Source, 44 Innovation Boulevard, Saskatoon, SK S7N 2V3, Canada
17	⁶ Renewable Energy, Helmholtz-Zentrum Berlin für Materialien und Energie GmbH, Hahn-
18	Meitner-Platz 1, 14109 Berlin, Germany
19	
20	* Corresponding author: e-mail: cv.vogel@yahoo.de
21	
22	Keywords:
23	Phosphorus plant-availability, X-ray adsorption near-edge structure (XANES) spectroscopy,
24	infrared spectroscopy, Diffusive Gradients in Thin films (DGT)

25 Abstract:

26 A wide range of methods are used to estimate the plant-availability of soil phosphorus (P). Published research has shown that the diffusive gradients in thin films (DGT) technique 27 28 has a superior correlation to plant-available P in soils compared to standard chemical 29 extraction tests. In order to identify the plant-available soil P species, we combined DGT with infrared and P K- and L_{2.3}-edge X-ray adsorption near-edge structure (XANES) spectroscopy. 30 This was achieved by spectroscopically investigating the dried binding layer of DGT devices 31 32 after soil deployment. All three spectroscopic methods were able to distinguish between different kinds of phosphates (poly-, trimeta-, pyro- and orthophosphate) on the DGT binding 33 layer. However, infrared spectroscopy was most sensitive to distinguish between different 34 types of adsorbed inorganic and organic phosphates. Furthermore, intermediates of the time-35 36 resolved hydrolysis of trimetaphosphate in soil could be analyzed.

37

38

39 **1. Introduction**

Phosphorus (P) is an essential element for all forms of life. It is necessary for the
metabolic process (ADP/ATP) and it is an integral part of the DNA molecule and the cell
membrane. Therefore, P is termed a macronutrient and is applied in the form of P-fertilizers
in agriculture for crop production.

The plant-availability of soil P can strongly influence the yield of agricultural crops. Hence, several simple chemical extraction methods are used to estimate the plant-available P of soils [1,2]. More recently, several research groups [3-8] have shown that the Diffusive Gradients in Thin films (DGT) technique have a much better correlation to plant-available P in soils than standard chemical extraction tests (e.g. calcium-acetate-lactate (CAL), Colwell, Olsen, water) when soils with different characteristics are considered. The DGT device consists of a binding layer, a diffusion gel and a filter (to protect the gel) assembled in a

51 plastic holder [8,9]. The dissolved and labile P fraction of the soil from moist soil samples 52 diffuses through the filter and diffusion gel and is subsequently adsorbed to the binding layer 53 during deployment. The amount of adsorbed P on the binding layer, which is the quantity that 54 accounts for the resupply from the solid phase over time, is then used as indicator for P plant-55 availability of the soil.

56 Six et al. [6] and Mason et al. [10] discovered by P³³/P³² labelled P sources that the 57 DGT method accessed the same pool of labile soil P as maize and wheat plants, while 58 conventional P extraction tests also include non-available P pools in their measured quantity. 59 Thus, the soil P compounds which are responsible for high yields of maize and wheat diffused 60 and bound to the DGT binding layer.

The aim of our work was to identify the plant-available P compounds of soils by a 61 novel combination of DGT and spectroscopic techniques. Our approach was to analyze the 62 63 binding layers of DGT deployed in soils by Fourier Transform infrared (FTIR) spectroscopy, and by K-edge and L_{2.3}-edge X-ray adsorption near-edge structure (XANES) spectroscopies. 64 65 Previously, other groups have shown that X-ray adsorption spectroscopy was able to 66 distinguish between As(III) and As(V) compounds [11] and different mercury compounds 67 [12], respectively, on the DGT binding layer. However, this combination has not been applied to examine highly plant-available P as detected by DGT. Here, we demonstrate the strengths 68 69 of this approach by i) spectroscopically analyzing DGT deployed in various P solutions (as references), and ii) applying this to a time-resolved investigation of trimetaphosphate (TMP) 70 hydrolysis, a polyphosphate with a high plant-availability [13,14], in incubated P-71 72 fertilizer/soil mixtures.

73

74 **2. Materials and Methods**

75 2.1 DGT experiments

DGT devices [8] (window size: 2.54 cm²; 0.8 mm APA (polyacrylamide) diffusion 76 77 layer) with ferrihydrite (Fh; thickness 0.6 mm) and zirconium oxide (ZrO; thickness 0.4 mm) binding layer (DGT Research, Lancaster, UK), respectively, were loaded with 200 mL 78 79 solutions (50 mg P/L) of various inorganic and organic P compounds (KH₂PO₄, D-glucose-6-80 phosphate disodium-salt (both Carl Roth, Karlsruhe, Germany), Ca(H₂PO₄)₂·H₂O (ABCR, 81 Karlsruhe, Germany), (NH₄)₂HPO₄ (Merck, Darmstadt, Germany), Na₄P₂O₇·10H₂O, Na₅P₃O₁₀, (NaPO₃)₃, adenosine-5'-monophosphate Na-salt (AMP), adenosine-5'-diphosphate 82 83 adenosine-5´-triphosphate Na-salt (ATP), adenosine-3',5'-cyclic Na-salt (ADP), 84 monophosphate Na-salt (cAMP), L-a-phosphatidylchinoline, aminomethylphosphonic acid, 85 β-glycerophosphate Na-salt and creatine phosphate (all Alfa Aesar, Karlsruhe, Germany) and phytic acid Na-salt (Sigma-Aldrich, Steinheim, Germany). The DGT devices were deployed 86 87 for 24 h at 22°C in constantly agitated solutions. After deployment, the binding layers of the DGT devices were dried at room temperature and spectroscopically investigated as described 88 89 below. Finally, P from the binding layer was eluted with 1 M HNO₃ (for Fh binding layers) or 90 1 M NaOH (ZrO binding layers) and the P concentration was analyzed by inductively coupled 91 plasma - mass spectroscopy (ICP-MS; Thermo iCAP Q, Dreieich, Germany) to calculate the 92 P mass accumulated to the binding layer.

93 Furthermore, time-resolved DGT measurements (Fh binding layer) were performed 94 with trimetaphosphate (as sodium salt, $(NaPO_3)_3$) applied to a soil. Therefore, subsoil from a 95 loess-derived brown soil was mixed with quartz sand at 1:1 mass ratio to decrease the P mass fraction ($P_{Total} = 180 \text{ mg kg}^{-1}$, $P_{CAL} = 10 \text{ mg kg}^{-1}$, pH in 0.01 M CaCl₂ = 6.7 of mixture). Sixty 96 97 grams of soil/sand-mixture was mixed with 6 mg P of TMP (= 100 mg P per kg of soil, as in a 98 previous pot experiment [7,15]) in 100 ml plastic containers. The soil was maintained daily at 99 60% of the water holding capacity (WHC) at 22°C. DGT was deployed for 24 h at four-time 100 points: at the start of the incubation experiment (0 h) and after 6 hours, 2 and 7 days. Before 101 DGT deployment the water content was increased to 100% WHC.

102 Additionally, soil samples from a pot experiment [15] with P-fertilizers with novel 103 focus on recycled materials compared to triple super phosphate (TSP) were analyzed. The 104 recycled P materials included sewage sludge from a waste water treatment plant with 105 enhanced biological phosphorus removal (B_{bio}); and sewage sludge precipitated with FeCl₂ 106 and post-treated with sodium sulfate under reducing conditions (B_{chem}+Na). These soil 107 samples were taken after the growth experiment. Previous research of these soils showed P DGT results with a superior correlation to the P uptake by maize in the pot experiment [15]. 108 109 After a 24 h conditioning period of the soils at 60% of the WHC, they were brought to 100% 110 WHC, transferred onto the DGT devices (Fh and ZrO binding layer) and deployed for 24, 48 111 and 72 h, respectively, at 22°C.

112 The binding layers from the DGT experiments with the soils from the pot experiment 113 and with the TMP/soil mixture were dried at room temperature and spectroscopically 114 investigated, followed by P extraction with 1 M HNO₃ (Fh) and 1 M NaOH (ZrO), 115 respectively (see above).

116

117 2.2 Infrared spectroscopy

118 Fourier-transform infrared (FT-IR) spectra of the dried DGT binding layers were collected with a Bruker Alpha FT-IR spectrometer (Ettlingen, Germany) with a DTGS 119 120 detector. The Fh binding layers were measured in transmission mode (spectral resolution 8) cm⁻¹; 32 scans were coadded per spectrum) in a compression cell with diamond windows 121 122 (Micro Compression Cell II, Thermo Fisher Scientific, Madison, USA). The ZrO binding layers were measured with the eco-attenuated total reflection (ATR) module. The infrared 123 spectra were normalized (Min-Max to 1495-1382 cm⁻¹ region) and the peak positions and 124 125 second derivative analyzed with the software OPUS (Bruker, version 7.0).

126

127 2.3 P K-edge XANES spectroscopy

P K-edge XANES measurements of dried DGT Fh binding layers were carried out in 129 the High Kinetic Energy Photoelectron Spectrometer (HIKE) endstation [16] located at the 130 BESSY II KMC-1 beamline [17] at Helmholtz-Zentrum Berlin (HZB). The ring was operated 131 in top-up mode at a current of 280 mA. The beamline uses a Si (111) double-crystal 132 monochromator. P K-edge XANES spectra were measured in fluorescence mode using a silicon drift detector XFlash[®] 4010 (Bruker, Berlin, Germany) from 2130 eV to 2200 eV in 133 134 steps of 0.25 eV at room temperature. The data was analyzed using the freeware Demeter 135 Athena (version 0.9.24) [18]. The spectra were background corrected using a linear regression 136 fit through the pre-edge region $[-18 \text{ to } -8 \text{ eV} \text{ relative to } E_0]$ and a polynomial regression fit 137 through the post-edge region $[E_0+30 \text{ to } +47 \text{ eV}]$.

138

139 2.4 P L_{2.3}-edge XANES spectroscopy

140 P L_{2.3}-edge XANES analysis of dried DGT Fh binding layers were carried out at the Variable Line Spacing Plane Grating Monochromator (VLS-PGM) beamline [19] at the 141 142 Canadian Light Source. The electron storage ring was operated in decay mode with a current 143 range of 220-170 mA. All spectra were recorded at room temperature in the energy range 144 from 130 to 155 eV, with a step size of 0.1 eV and a dwell time of 4 s or 16 s. The entrance and exit slits were set to 200 µm. The spectra were collected in total fluorescence yield mode 145 146 (FLY), using a microchannel plate detector [20] and were normalized with respect to the incident photon flux (I_0) . I_0 was simultaneously recorded with the FLY by monitoring the 147 148 drain current emitted from a Nickel mesh (90% transmission) located in front of the samples. 149 The data were also analyzed using the freeware *Demeter Athena* (version 0.9.24) [18].

150

151

152 3. Results and discussion

3.1 DGT (Fh) of P-solutions 153

154 Table 1 shows the P mass accumulated to the Fh binding layer of DGT deployed in 155 solutions containing various P compounds (also relative to KH₂PO₄ (=100%)). The inorganic 156 ortho-, pyro- and polyphosphates show relatively high P adsorption values, best for $Ca(H_2PO_4)_2$ and $Na_4P_2O_7$. In contrast, the values of most organic P compounds (except ADP, 157 158 ATP and aminomethylphosphonic acid) are much lower, especially for phytic acid and L- α -159 phosphatidylcholine. This agrees with results of Van Moorleghem et al. [21] and illustrates that the diffusion coefficient for the high molecular weight organic P compounds is lower 160 161 than for inorganic species. Mohr et al. [22] experimentally determined the diffusion coefficient for AMP ($2.9 \times 10^{-6} \text{ cm}^2 \text{ s}^{-1}$; 20°C) and phytic acid ($1.0 \times 10^{-6} \text{ cm}^2 \text{ s}^{-1}$; 20°C), 162 which is significantly lower than for orthophosphate $(5.27 \times 10^{-6} \text{ cm}^2 \text{ s}^{-1}; 20^{\circ}\text{C})$. Afterwards, 163 164 these binding layers were used as references for the spectroscopic measurements.

165

166

3.2 Spectroscopy of DGT (Fh) binding layers exposed to various P species

Figure 1 shows the normalized FT-IR spectra of the pure Fh binding layer and from 167 168 the DGT deployed with different P species. In contrast to the blank, the P loaded binding layers show additional absorption bands between 1300 cm⁻¹ and 850 cm⁻¹. After subtraction of 169 170 the pure Fh binding layer spectrum, the absorption bands of the different P compounds adsorbed to the Fh binding layer become clearly visible (Fig. 2 – spectra and second 171 172 derivative). The orthophosphates show two absorption bands of the P-O stretching vibrations ca. 1100 cm⁻¹ and 1000 cm⁻¹. This is in agreement to previous studies on sorption of 173 174 phosphates onto ferrihydrite [23-25]. The pyrophosphate $(Na_4P_2O_7)$ shows two additional bands at *ca*. 1160 cm⁻¹ and 910 cm⁻¹, the tripolyphosphate (Na₅P₃O₁₀) at around 1220 cm⁻¹. 175 1160 cm⁻¹ and 910 cm⁻¹ and TMP ((NaPO₃)₃) at around 1270 cm⁻¹, 1160 cm⁻¹ and a small 176 band ca. 910 cm⁻¹. The band at around 910 cm⁻¹ is the P-O-P stretching vibration of the 177 condensed phosphates, the band at around 1160 cm⁻¹ belongs to the stretching vibration of the 178 PO₃-group and the bands at 1220 cm⁻¹ and 1270 cm⁻¹ are the stretching vibrations of the 179

bridging PO₂ [26,27]. The bridging PO₂ stretching vibrations occur for polyphosphates \ge P₃ only and the frequency of the band is chain length dependent [26].

182 Additionally, the P K-edge XANES spectra of the Fh binding layers from the DGT 183 experiments with different P solutions were measured. The XANES spectra of the analyzed P 184 standards are very similar (see Fig. 3 top left) and show adsorbed P. A limitation of the P K-185 edge XANES technique is the inability to reliably distinguish among different phosphate 186 adsorption complexes which result in identical XANES spectra [28]. Furthermore, most iron-187 P compounds show a minor pre-peak [29], which is not detectable for the Fh binding layers 188 from the DGT experiment. This is probably because of the adsorption complex with Fe, 189 which has only a slight pre-edge feature [28]. However, the zoomed-in edge region and the 190 first derivate of these spectra (Fig. 3 top, middle and right, respectively) show a little shift of 191 the K-edge inflection point ((NaPO₃)₃: 2152.5 eV; Na₄P₂O₇: 2152.75 eV; and KH₂PO₄: 2153.0 eV). It should be noted that this spectral shift of 0.25 eV is almost similar to the 192 spectral resolution of the beamline (0.2 eV). Thus, FT-IR spectroscopy is much more 193 194 sensitive to distinguish among different kinds of phosphates absorbed to the Fh binding layer 195 than P K-edge XANES spectroscopy.

196 In contrast, P L_{2,3}-edge XANES spectroscopy provides better resolved spectral 197 features than P K-edge XANES spectroscopy [30]. The features of the P L_{2.3}-edge XANES 198 spectra of the Fh binding layers from the DGT experiments with different P solutions come 199 close to the applied poly- and pyrophosphate but with less intensity and are more blurred (see 200 spectra in Fig. 4). The polyphosphate TMP (DGT-(NaPO₃)₃) shows two shoulders at around 201 136.1 eV and 137.1 eV and the pyrophosphate (DGT-Na₄P₂O₇) two shoulders at ca. 136.5 eV 202 and 137.5 eV. In contrast, the orthophosphate (DGT-KH₂PO₄) absorbed on the DGT Fh 203 binding layer shows no obvious shoulders in the L_{2.3}-edge possibly due to more disordered 204 structure compared to standards. Only a slight pre-peak at 136-137 eV was detectable which 205 is characteristic for orthophosphate adsorbed to ferrihydrite [31]. All these phosphates show

206 characteristic features which could be clearly distinguished from each other. However, it is 207 worth to mention that the sensitivity (fluorescence yields) is much lower for P $L_{2,3}$ -edge 208 XANES spectroscopy than for P K-edge XANES spectroscopy, thereby, only Fh binding 209 layers with >43 µg of accumulated P could be analyzed with this technique.

210 Finally, various organic P compounds on the Fh binding layers were also analyzed by 211 FT-IR spectroscopy (see Fig. S1 and Table 2). The organic orthophosphate monoesters show the P-O stretching vibrations bands around 1080 cm⁻¹ and 980 cm⁻¹ and the PO₃ stretching 212 vibration at 1160 cm⁻¹ in the FT-IR spectra. In comparison to the inorganic orthophosphates 213 the P-O stretching vibrations are shifted (approx. 20 cm⁻¹) to lower wavenumbers. However, it 214 215 is not possible to differentiate between the different organic orthophosphate monoesters due 216 to the very similar bonding of the phosphate group (R-O-PO₃). In contrast, the organic pyroand polyphosphate monoesters ADP and ATP show, similarly to the inorganic pyro-217 /polyphosphates, the P-O-P stretching vibration (926 and 915 cm⁻¹, respectively) and the 218 219 stretching vibrations of the bridging PO₂ (1216 and 1234 cm⁻¹, respectively) (see Fig. S1; 220 Table 2). ADP shows the bridging PO_2 band, in contrast to the inorganic pyrophosphate, 221 because of the ester bond to the carbon (R-O-PO₂-O-PO₃). Furthermore, this bridging PO₂ 222 band is also detectable for the orthophosphate diesters cAMP (cyclic) and L-aphosphatidylcholine (linear), respectively, (R-O-PO₂-O-R bond; 1238 cm⁻¹ and 1265 cm⁻¹). 223 224 Additionally, these orthophosphate diesters show also a band at ca. 850 cm⁻¹. Moreover, the 225 phosphonates aminomethylphosphonic acid and creatine phosphate show three absorption 226 bands very similar to the orthophosphate monoesters due to the P-O and PO₃-group stretching vibrations of the R-PO₃-group. 227

228

229 3.3 Incubated TMP/soil-mixtures

Figure 5 shows the FT-IR spectra and second derivative of the Fh binding layers fromDGT experiments with the incubated TMP/soil-mixtures. At the start of the experiment

absorption bands of ortho-, pyro- and polyphosphates were detected (0 min). Surprisingly, the absorption band of added TMP at 1270 cm⁻¹ is not visible. After 6 h of incubation an almost similar spectrum was observed, but after 2 days of incubation the polyphosphate band at around 1220 cm⁻¹ is no longer detectable, while the pyrophosphate bands at *ca*. 1160 cm⁻¹ and 910 cm⁻¹ are still observed. Finally, seven days of incubation led to an almost complete hydrolysis of TMP in the soil to orthophosphates and/or orthophosphate monoesters.

238 Additionally, P K-edge XANES spectra of the Fh binding layers from the DGT 239 experiments with incubated TMP/soil-mixture were measured (see Fig. 3 bottom). Similar to 240 the XANES spectra of the P standards (Fig. 3 top) the XANES spectra from the incubated TMP/soil-mixture are also very similar. The zoomed-in edge (Fig. 3 bottom middle) and first 241 242 derivate of these XANES spectra (Fig. 3 bottom right) displays a shift of the edge to higher 243 energy from the beginning of the experiment (0 min. and 6 h) to two and seven days of 244 incubation. These results support the FT-IR data, and together they indicate that TMP is first hydrolyzed in the soil over time to a linear polyphosphate and then to pyro- and 245 246 orthophosphates, which is in agreement with previous literature [32-34].

247 The P mass accumulated on the binding layer for the incubation experiment with TMP 248 (see Fig. 5) is very high (40-35 µg P; maximum approx. 58 µg) at the beginning (first day) 249 and rapidly decreased after two days of incubation (9 µg P). However, these amounts of 250 accumulated P on the binding layer were still below the detection limit for P L₂ 3-edge 251 XANES spectroscopy. Based on the FT-IR results of the TMP incubation time series, it is 252 likely that the plant-available P in the incubated TMP/soil-mixture is also decreasing. 253 Previously, Torres-Dorrante et al. [35] showed that the polyphosphate concentration in the 254 soil solution of incubated TMP/soil-mixtures dropped rapidly in the first few days, which is 255 consistent with our findings. Blanchar and Hossner [32] found that TMP, in contrast to other phosphates has a lower sorption rate in soil. Thus, TMP stays in the soil solution and can be 256 easily accessed by DGT. After hydrolysis of the TMP, the orthophosphates can be absorbed to 257

258 the soil. This, together with the ageing process of the phosphate, might be an explanation for 259 the high value of accumulated P on the binding layer at the beginning and the decrease after 260 two days of incubation. These experiments were also done with a reduced deployment time 261 for the DGT devices of 3 h (see Fig. S2 in supporting information). However, the results 262 obtained were similar to the 24 h deployment.

- 263
- 264

3.4 Use of different DGT binding layers

265 The DGT Fh binding layer can adsorb only up to approximately 58 µg of P and larger amounts of P in fertilizer/soil-mixtures could possibly saturate the binding layer. Therefore, 266 binding layers with higher capacities may be used, such as titanium oxide (TiO₂) [36] or 267 zirconium oxide (ZrO) [37]. DGT experiments showed that the dried TiO₂ binding layers 268 269 were impossible to analyze by FT-IR spectroscopy because after drying they became stiff and 270 fragile and could not be pressed in the diamond compression cell. The ZrO binding layers 271 were also a bit fragile after drying, but remained sufficiently intact to allow for their analysis 272 by ATR/FT-IR spectroscopy.

Figure 6 shows a comparison of the normalized FT-IR spectra of DGT experiments 273 274 with ZrO and Fh binding layers, respectively, with different P solutions. In contrast to the Fh 275 binding layer much stronger adsorption bands were detected due to the higher capacity of the 276 ZrO binding layer. Small, adsorbent-dependent shifts in the IR absorption frequencies can be 277 expected since the adsorption process would alter the bond strengths of adsorbing groups (e.g. 278 P-O), similar to the differences observed between pure vs Fh-adsorbed phosphates.

279

280 3.5 DGT of soils from pot experiment

281 Soil from a pot experiment with triple superphosphate [15] was analyzed with 282 different deployment times of the DGT devices (ZrO binding layer). The ATR/FT-IR spectra showed that for all deployment times (24, 48 and 72 h) orthophosphates were adsorbed onto 283

the ZrO binding layer (see Fig. S3 in the Supporting Information). It is also clear that longer deployment times lead to stronger orthophosphate absorption bands as the adsorbed amount increases.

287 Analysis of this and other soils from a pot experiment by DGT showed bands from orthophosphates (around 1100 and 1000 cm⁻¹, respectively) and organic orthophosphate 288 289 monoester (additional band at approx. 1160 cm⁻¹) only on the DGT binding layers (ZrO and Fh; see Fig. S4 and S5, respectively). Notably, the absorption band at 1160 cm⁻¹ was also 290 291 detected for the untreated soil (see Fig. S3 top). Therefore, the orthophosphates appear to 292 originate only from the applied fertilizers. However, the presence of the organic orthophosphate monoester on the binding layer shows that the use of the single 293 294 orthophosphate diffusion coefficient, as is conventionally used, may not be strictly correct in 295 analyzing DGT P. In this case, (semi-)quantitative P speciation information on the binding 296 layer as demonstrated in this study has the potential to improve the P plant-availability investigations of DGT by allowing the use of multiple diffusion coefficients from the multiple 297 298 detectable P species.

299

300

301 **4. Conclusions**

302 In this paper, we showed the potential for the combination of the DGT technique with 303 spectroscopic methods. Different kinds of phosphates in solutions and soils can be distinguished on the DGT binding layer by infrared and P K- and L_{2,3}-edge XANES 304 305 spectroscopy, respectively (see a summary of all analyzed samples in table S1). However, 306 various orthophosphates adsorbed to the binding layer show very similar FT-IR and XANES 307 spectra, respectively. The organic orthophosphate monoesters also show very similar FT-IR 308 adsorption bands. For the here investigated sample series, the infrared spectra show 309 comparatively more features and thus more information about the adsorbed inorganic and

310 organic P-species. Additionally, infrared microspectroscopy [38] make it also possible to analyze P compounds on the binding layer with a lateral resolution down to $5 \,\mu m^2$. Therefore, 311 312 P species of a spatial soil segment (e.g. rhizosphere) [39] can be mapped and analyzed. Analysis of P hotspots from soil segments could possibly also be done by P L₂ 3-edge XANES 313 314 microspectroscopy [40] which will be available from autumn of 2019 at the Canadian Light 315 Source. Furthermore, the hydrolysis of TMP in soil shows a further benefit of this 316 combination. Intermediates of the time-resolved hydrolysis were absorbed on the DGT 317 binding layer and could be analyzed afterwards.

318

319 Acknowledgments:

CV and DS thank the German Federal Ministry for Food and Agriculture for financial 320 support (2811NA022/2811NA023). CV thanks the German Research Foundation (VO 321 322 1794/4-1) for financial support. Collaboration between BAM (CV, CA) and the University of South Australia (RS, EL) was supported by the Australian Technology Network – DAAD 323 324 Researcher Exchange Scheme (2014). RF acknowledges financial support by the Impuls- und Vernetzungsfonds of the Helmholtz-Association (VH-NG-423). Research described in this 325 326 paper was performed at the Canadian Light Source, which is supported by the Canada 327 Foundation for Innovation, Natural Sciences and Engineering Research Council of Canada, 328 the University of Saskatchewan, the Government of Saskatchewan, Western Economic 329 Diversification Canada, the National Research Council Canada, and the Canadian Institutes of 330 Health Research. We thank HZB for the allocation of synchrotron radiation beamtime.

331

332

333 References

- R. Wuenscher, W. Unterfrauner, R. Peticzka, F. Zehetner, A comparison of 14 soil
 phosphorus extraction methods applied to 50 agricultural soils from Central Europe.
 Plant Soil Environ. 61 (2015) 86-96.
- M. Yli-Halla, J. Schick, S. Kratz, E. Schnug, Determination of plant available P in
 soil, in: E. Schnug, L.J. De Kok (Eds.), Phosphorus in agriculture: 100 % zero,
 Springer, Dordrecht, 2016, pp. 63-93.
- N.W. Menzies, B. Kusumo, P.W. Moody, Assessment of P availability in heavily
 fertilized soils using the diffusive gradient in thin films (DGT) technique. Plant Soil
 269 (2005) 1-9.
- S. Mason, A. McNeill, M.J. McLaughlin, H. Zhang, Prediction of wheat response to
 an application of phosphorus under field conditions using diffusive gradients in thinfilms (DGT) and extraction methods. Plant Soil 337 (2010) 243-258.
- 346 [5] S. Tandy, S. Mundus, J. Yngvesson, T.C. de Bang, E. Lombi, J.K. Schjoerring, S.
 347 Husted, The use of DGT for prediction of plant available copper, zinc and
 348 phosphorus in agricultural soils. Plant Soil 346 (2011) 167-180.
- L. Six, P. Pypers, F. Degryse, E. Smolders, R. Merckx, The performance of DGT
 versus conventional soil phosphorus tests in tropical soils An isotope dilution study.
 Plant Soil 359 (2012) 267-279.
- C. Vogel, R. Sekine, D. Steckenmesser, E. Lombi, D. Steffens, C. Adam,
 Phosphorus availability of sewage sludge-based fertilizers determined by the diffusive
 gradients in thin films (DGT) technique. J. Plant Nutr. Soil Sci. 180 (2017) 594-601
- 355 [8] H. Zhang, W. Davison, R. Gadi, T. Kobayashi, In situ measurement of dissolved
 356 phosphorus in natural waters using DGT, Anal. Chim. Acta 370 (1998) 29-38
- W. Davison, Diffusive Gradients in Thin-films for environmental measurements,
 Cambridge University Press, Cambridge, 2016.

- 359 [10] S. Mason, M.J. McLaughlin, C. Johnston, A. McNeill, Soil test measures of
 available P (Colwell, resin and DGT) compared with plant P uptake using isotope
 dilution. Plant Soil 373 (2013) 711-722.
- 362 [11] T. Huynh, H.H. Harris, H. Zhang, B.N. Noller, Measurement of labile arsenic
 363 speciation in water and soil using diffusive gradients in thin films (DGT) and X-ray
 364 absorption near edge spectroscopy (XANES). Environ. Chem. 12 (2015) 102-111.
- A.L. Pham, C. Johnson, D. Manley, H. Hsu-Kim, Influence of sulfide nanoparticles
 on dissolved mercury and zinc quantification by diffusive gradient in thin-film
 passive samplers. Environ. Sci. Technol. 49 (2015) 12897–12903.
- 368 [13] L.O. Torres-Dorante, N. Claassen, B. Steingrobe, H.W. Olfs, Fertilizer-use
 369 efficiency of different inorganic polyphosphate sources: effects on soil P availability
 370 and plant P acquisition during early growth of corn. J. Plant Nutr. Soil Sci. 169 (2006)
 371 509-515.
- T.M. McBeath, E. Lombi, M.J. McLaughlin, E.K. Bünemann, Polyphosphate-fertilizer
 solution stability with time, temperature, and pH. J. Plant Nutr. Soil Sci. 170 (2007)
 374 387-391.
- 375 [15] D. Steckenmesser, C. Vogel, C. Adam, D. Steffens, Effect of various types of
 376 thermochemical processing of sewage sludges on phosphorus speciation, solubility,
 377 and fertilization performance. Waste Manage. 62 (2017) 194-203.
- M. Gorgoi, S. Svensson, F. Schäfers, G. Öhrwall, M. Mertin, P. Bressler, O. Karis,
 H. Siegbahn, A. Sandell, H. Rensmo, et al. The high kinetic energy photoelectron
 spectroscopy facility at BESSY progress and first results. Nucl. Instr. Meth. Phys. Res.
 A 601 (2009) 48 53.
- F. Schaefers, M. Mertin, M. Gorgoi, KMC-1: a high resolution and high flux soft xray beamline at BESSY. Rev. Sci. Instrum. 78 (2007) 123102.

- B. Ravel, M. Newville, ATHENA, ARTEMIS, HEPHAESTUS: data analysis for Xray absorption spectroscopy using IFEFFIT. J. Synchrotron Radiat. 12 (2005) 537–
 541.
- 387 [19] Y.F. Hu, L. Zuin, G. Wright, R. Igarashi, M. McKibben, T. Wilson, S.Y. Chen, T.
- Johnson, D. Maxwell, B.W. Yates, T.K. Sham, R. Reininger, Commissioning and
 performance of the variable line spacing plane grating monochromator beamline at the
 Canadian Light Source. Rev. Sci. Instrum. 78 (2007) 083109.
- 391 [20] M. Kasrai, Z.F. Yin, G.M. Bancroft, K.H. Tan, X-ray fluorescence measurements of
 392 XANES at the Si, P and S L-edges. J. Vac. Sci. Technol. A 11 (1993) 2694-2699.
- 393 [21] C. Van Moorleghem, L. Six, F. Degryse, E. Smolders, R. Merckx, Effect of
 394 organic P forms and P present in inorganic colloids on the determination of
 395 dissolved P in environmental samples by the diffusive gradient in thin films
 396 technique, ion chromatography, and colorimetry. Anal. Chem. 83 (2011) 5317-5323.
- 397 [22] C.W. Mohr, R.D. Vogt, O. Royset, T. Andersen, N.A. Parekh, An in-depth
 398 assessment into simultaneous monitoring of dissolved reactive phosphorus (DRP) and
 399 low-molecular-weight organic phosphorus (LMWOP) in aquatic environments using
 400 diffusive gradients in thin films (DGT). Environ. Sci.: Processes Impacts 17 (2015)
 401 711 707
- 401 711-727.
- 402 [23] M. Nanzyo, Infrared spectra of phosphate sorbed on iron hydroxide gel and the
 403 sorption products. Soil Sci. Plant Nutr. 32 (1986) 51-58.
- 404 [24] M.I. Tejedor-Tejedor, M.A. Anderson, Protonation of Phosphate on the Surface of
 405 Goethite As Studied by CIR-FTIR and Electrophoretic Mobility Langmuir 6 (1990)
 406 602-611.
- 407 [25] Y. Arai, D.L. Sparks, ATR–FTIR spectroscopic investigation on phosphate
 408 adsorption mechanisms at the ferrihydrite–water interface J. Colloid Interface Sci.
 409 241 (2001) 317-326.

- 410 [26] W. Gong, A real time in situ ATR-FTIR spectroscopic study of linear phosphate
 411 adsorption on titania surfaces. Int. J. Miner. Process. 63 (2001) 147-165.
- 412 [27] X.H. Guan, Q. Liu, G.H. Chen, C. Shang, Surface complexation of condensed
 413 phosphate to aluminum hydroxide: An ATR-FTIR spectroscopic investigation *J*.
 414 Colloid Interface Sci. 289 (2005) 319-327.
- 415 [28] J.G. Hamilton, G. Hilger, D.J. Peak, Mechanisms of tripolyphosphate adsorption and
 416 hydrolysis on goethite. Colloid Interface Sci. 491 (2017) 190–198.
- 417 [29] N. Khare, D. Hesterberg, J.D. Martin, XANES investigation of phosphate sorption in
 418 single and binary systems of iron and aluminum oxide minerals. Environ. Sci.
 419 Technol. 39 (2005) 2152-2160.
- 420 [30] J. Kruse, P. Leinweber, K.W. Eckhardt, F. Godlinski, Y. Hu, L. Zuin, Phosphorus
 421 L_{2,3}-edge XANES: overview of reference compounds. J. Synchrotron Radiat. 16
 422 (2009) 247-259.
- 423 [31] W. Xiong, J. Peng, Y. Hu, Use of X-ray adsorption near edge structure (XANES) to
 424 identify physiorption and chemisorption of phosphate onto ferrihydrite-modified
 425 diatomite. J. Colloid Interface Sci. 368 (2012) 528-532.
- 426 [32] R.W. Blanchar, L.R. Hossner, Hydrolysis and sorption reactions of orthophosphate,
 427 pyrophosphate, tripolyphosphate, and trimetaphosphate anions added to an Elliot soil.
 428 Soil Sci. Soc. Amer. Proc. 33 (1969) 141-144.
- 429 [33] L.M. Busman, M.A. Tabatabai, Hydrolysis of trimetaphosphate in soils. Soil Sci. Soc.
 430 Amer. Proc. 49 (1985) 630-636.
- 431 [34] R.P. Dick, M.A. Tabatabi, Factors affecting hydrolysis of polyphosphates in soils. Soil
 432 Sci. 143 (1987) 97-104.
- L.O. Torres-Dorante, N. Claassen, B. Steingrobe, H.W. Olfs, Hydrolysis rates of
 inorganic polyphosphates in aqueous solution as well as in soils and effects on P
 availability. J. Plant Nutr. Soil Sci. 168 (2005) 352-358.

- 436 [36] J.G. Panther, P.R. Teasdale, W.W. Bennett, D.T. Welsh, H. Zhao, Titanium dioxide437 based DGT technique for in situ measurement of dissolved reactive phosphorus in
 438 fresh and marine waters. Environ. Sci. Technol. 44 (2010) 9419-9424
- 439 [37] S. Ding, D. Xu, Q. Sun, H. Yin, C. Zhang, Measurement of dissolved reactive
 440 phosphorus using the diffusive gradient in thin films technique with a high-capacity
 441 binding phase. Environ. Sci. Technol. 44 (2010) 8169-8174.
- E. Grotheer, C. Vogel, O. Kolomiets, U. Hoffmann, M. Unger, H.W. Siesler, FT-IR
 and NIR spectroscopic imaging: Principles, practical aspects, and applications in
 Material and pharmaceutical science, in: R. Salzer, H.W. Siesler (Eds.), Infrared and
 Raman Spectroscopic Imaging, 2nd Edition, Wiley-VCH, Weinheim, 2014, pp. 341396.
- 447 [39] A. Kreuzeder, J. Santner, T. Prohaska, W.W. Wenzel, Gel for simultaneous chemical
 448 imaging of anionic and cationic solutes using diffusive gradients in thin films. Anal.
 449 Chem. 85 (2013) 12028–12036.
- 450 [40] C. Vogel, C. Rivard, I. Tanabe, C. Adam, Microspectroscopy Promising technique to
- 451 characterize phosphorus in soil. Comm. Soil Sci. Plant Anal. 47 (2016) 2088-2102.
- 452

453

454 **Table and Figure Caption:**

- 455 **Table 1:** DGT-measurable P over 24 h to the Fh binding layer of DGT devices from P456 solutions (50 mg P/L) of various P compounds
- 457 **Table 2:** Detected IR absorption bands (cm⁻¹) of various inorganic and organic P 458 compounds adsorbed to the Fh binding layer. (v = stretching vibration)
- 459 Figure 1: FT-IR spectra of Fh binding layers from DGT experiments with solutions of460 different phosphates
- 461 Figure 2: Blank subtracted FT-IR spectra (left) and second derivative (right) of Fh
 462 binding layers from DGT experiments with solutions of different inorganic
 463 phosphates
- 464 Figure 3: P K-edge XANES spectra of Fh binding layers from DGT experiments with
 465 solutions of different phosphates (top left) and time-resolved incubated TMP in
 466 soil (below left), zoomed in edge-region (middle) and their first derivative of
 467 the related edge region (right); the vertical lines show the edge and first
 468 derivative of the edge for the polyphosphate (blue), pyrophosphate (green) and
 469 orthophosphate (red).
- 470 Figure 4: P L_{2,3}-edge XANES spectra of Fh binding layers from DGT experiments with
 471 solutions of different phosphates in comparison to spectra of the applied
 472 phosphates
- 473 Figure 5: Blank subtracted FT-IR spectra (left) and second derivative (right) of Fh
 474 binding layers from DGT experiments of time-resolved incubated TMP in soil
- 475 Figure 6: Comparison between FT-IR spectra of ZrO (black) and Fh (red) binding layers
 476 from DGT experiments with solutions of different phosphates
- 477
- 478

480 Table 1:

	phosphate type	molecular weight	P mass accumulated to binding layer	Accumulated P relative to KH ₂ PO ₄
KH ₂ PO ₄	orthophosphate	136 g/mol	58.5 ± 0.4 µg	100% ± 1%
$Ca(H_2PO_4)_2$	orthophosphate	252 g/mol	58.5 ± 4.7 µg	100% ± 8%
Na ₄ P ₂ O ₇	pyrophosphate	266 g/mol	49.7 ± 5.3 µg	85% ± 9%
Na ₅ P ₃ O ₁₀	polyphosphate	368 g/mol	34.5 ± 0.6 µg	59% ± 1%
(NaPO ₃) ₃	polyphosphate	306 g/mol	43.3 ± 4.7 μg	74% ± 8%
β-Glycero phosphate	orthophosphate monoester	216 g/mol	29.8 ± 2.9 µg	51% ± 5%
Phytic acid	orthophosphate monoester	660 g/mol	6.4 ± 0.6 μg	11% ± 1%
Glucose-6-phosphate	orthophosphate monoester	308 g/mol	29.3 ± 4.7 µg	50% ± 8%
AMP	orthophosphate monoester	391 g/mol	31.6 ± 0.6 µg	54% ± 1%
ADP	pyrophosphate monoester	471 g/mol	48.6 ± 1.2 µg	83% ± 2%
ATP	polyphosphate monoester	533 g/mol	69.6 ± 0.8 µg	119% ± 2%
cAMP	cyclic phosphate diester	351 g/mol	10.5 ± 4.1 µg	18% ± 7%
L-α-Phosphatidylcholine	phosphate diester	644 g/mol	2.9 ± 0.5 μg	5% ± 1%
Aminomethylphosphonic acid	phosphonate	111 g/mol	47.4 ± 1.2 μg	81% ± 2%
Creatine phosphate	phosphonate	263 g/mol	23.4 ± 2.3 µg	40% ± 4%

481 Table 2:

482

	vPO ₂ bridging	vPO₃	<i>v</i> ₁ P-O	<i>v</i> ₂ P-O	<i>v</i> (P-O-P)	<i>v</i> (O-P-O) diester
Inorg. orthophosphate			1087	1000		
Inorg. pyrophosphate		1159	1088	1019	910	(
Inorg. tripolyphosphate	1220	1161	1084	978	906	
Inorg. trimetaphosphate	1270	1158	1087	988	903	7
AMP		1153	1079	981		
сАМР	1238	1159	1068	1010	909	849
ADP	1216	1161	1088	1001	926	
АТР	1234	1164	1087	989	915	
β-Glycero phosphate		1155	1070	976		
Phytic acid		1166	1070	984		
Glucose-6-phosphate		1153	1082	981		
L-α-Phosphatidylcholine	1267	1160		1003		854
Aminomethylphosphonic acid		1132	1053	985		
Creatine phosphate		1158	1095	979		

484

483

- 21 -







Service







Highlights:

- Combination of DGT and spectroscopic techniques determines plant-available phosphorus species in soils
- Intermediates of time-resolved soil phosphorus reactions can be analyzed
- Spectroscopic mapping techniques can detect phosphorus species of a spatial soil segment

Declarations of interest: none