

1 **Running head: *Spatio-temporal segregation in giant petrels***

2

3 **Segregation in space and time explains the coexistence of two sympatric sub-Antarctic petrels**

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12 Biological communities are shaped by competition between and within species. Competition is
13 often reduced by inter- and intra-specific specialization on resources, such as differences in foraging
14 areas or time, allowing similar species to coexist and potentially contributing to reproductive
15 isolation. Here, we examine the simultaneous role of temporal and spatial foraging segregation
16 within and between two sympatric sister species of seabirds, Northern *Macronectes halli* and
17 Southern *M. giganteus* Giant Petrels. These species show marked sexual size dimorphism and
18 allochrony (with earlier breeding by Northern Giant Petrels), but this is the first study to test for
19 differences in foraging behaviours and areas across the entire breeding season both between the two
20 species and between the sexes. We tracked males and females of both species in all breeding stages
21 at Bird Island, South Georgia, to test how foraging distribution, behaviour and habitat use varies
22 between and within species in biological time (incubation, brood-guard or post-brood stages) and in
23 absolute time (calendar date). Within each breeding stage, both species took trips of comparable
24 duration to similar areas, but due to breeding allochrony they segregated temporally. Northern Giant
25 Petrels had a somewhat smaller foraging range than Southern Giant Petrels, reflecting their greater
26 exploitation of local carrion and probably contributing to their recent higher population growth.
27 Within species, segregation was spatial, with females generally taking longer, more pelagic trips
28 than males. Both sexes of both species showed unexpectedly plastic foraging behaviour, and there
29 was little evidence of inter-specific differences in habitat use. Thus, in giant petrels, temporal
30 segregation reduces inter-specific competition and sexual segregation reduces intra-specific
31 competition. These results demonstrate how both specialization and dynamic changes in foraging
32 strategies at different scales underpin resource division within a community.

33 **Keywords:** Allochrony, inter-specific competition, intra-specific competition, niche differentiation,
34 Procellariiformes, resource partitioning, sexual segregation, speciation, tracking

35 Competition in ecological communities, both between and within species, is a fundamental
36 ecological process, with important evolutionary consequences. Coexistence between community
37 members is promoted by a reduction in this competition, often because different members exploit
38 different parts of the resource space, such as particular habitats or prey (Schoener 1974). This
39 resource partitioning or niche differentiation can shape the structure of a community, driving
40 evolutionary change in its constituent organisms (Zupping-Dingley *et al.* 2014) and promoting
41 species coexistence in plants (Silvertown 2004), insects (Augustyn *et al.* 2016), fish (Bonin *et al.*
42 2015), mammals (Stewart *et al.* 2002; Nicholls & Racey 2006) and birds at local and regional scales
43 (Gotelli *et al.* 2010; Young *et al.* 2010).

44 Among morphologically similar competitors, partitioning of habitat or prey resources can
45 play an important role not only in maintaining coexistence but also in isolating diverging taxa
46 during incipient speciation (Svanback & Bolnick 2007). In such contexts, resources are often
47 partitioned through behavioural mechanisms (Nicholls & Racey 2006). Behaviour can change
48 rapidly in response to increasing competitive pressure, often in ways consistent with existing
49 morphological differences between sub-populations (Svanback & Bolnick 2007, Linnebjerg *et al.*
50 2013, Drago *et al.* 2015). Such behavioural divergence is frequently observed as segregation
51 between species or sub-populations either in space or in time, for example with sympatric species
52 reducing inter-specific competition by foraging in different areas during the breeding or
53 nonbreeding seasons (Phillips *et al.* 2005, Rayner *et al.* 2016), at differing heights or depths
54 (Nicholls & Racey 2006, Navarro *et al.* 2013), or at different times of day (Wilson 2010, Navarro *et al.*
55 2013). Particularly where it occurs on a variety of scales or in different foraging traits,
56 segregation may nonetheless be difficult to resolve, even in systems where competitors differ
57 morphologically (Connors *et al.* 2015).

58 Similar segregation patterns within species have also been well documented, with
59 behavioural differences resulting in partitioning of food resources according to sex or age class

60 (Lewis *et al.* 2001, Lecomte *et al.* 2010). The competition that drives such segregation is expected
61 to be particularly intense when adults of both sexes congregate to breed (Phillips *et al.* 2011,
62 Phillips *et al.* 2017, Drago *et al.* 2015), and sexual segregation in foraging, which can act to reduce
63 intra-specific competition, is widespread among vertebrates (Lewis *et al.* 2001, Ruckstuhl &
64 Neuhaus 2002, Catry *et al.* 2005, Breed *et al.* 2006). Males and females may forage in different
65 areas, at different times, or take different prey, in sexually monomorphic as well as dimorphic
66 species, although behavioural differentiation can be subtle even in species with extreme sexual
67 dimorphism (Lewis *et al.* 2002, Ruckstuhl & Neuhaus 2002, Phillips *et al.* 2011, 2017, Harris *et al.*
68 2013, Griffiths *et al.* 2014, Baylis *et al.* 2016). Competition between the sexes has been implicated
69 in the evolution of specialisation and segregation in many contexts, for example with males and
70 females exploiting different habitat types or taking different roles when rearing offspring, even
71 where segregation has proved difficult to resolve (Copello *et al.* 2011, Pinet *et al.* 2012, Ludynia *et al.*
72 2013, Cleasby *et al.* 2015, Baylis *et al.* 2016).

73 Both inter- and intra-specific differences in distribution and behaviour arise as dynamic
74 responses to the biotic and abiotic environment and may therefore vary over time. Such flexibility
75 in foraging strategy is particularly important when foraging is constrained, such as in central-place
76 foragers when habitat use is spatially restricted, and energy and time budgets are limited by the
77 changing demands of incubation and chick-rearing (Phillips *et al.* 2017). Breeding seabirds are
78 typically subject to these foraging constraints, making them an informative model for investigating
79 how the competition that such restrictions induce is resolved through variation in foraging
80 behaviour on different scales. These constraints may differ between the sexes and occur on different
81 schedules in otherwise similar species and habitats (Elliott *et al.* 2010, Beaulieu & Sockman 2012,
82 Pinet *et al.* 2012). Inter-specific segregation in timing of breeding (allochrony) may therefore be a
83 crucial mechanism enabling the coexistence of similar species, maintaining reproductive isolation in
84 related taxa, or potentially driving sympatric speciation (Wilson 2010, Brown *et al.* 2015). A

85 complete understanding of dynamic resource partitioning within a community requires
86 consideration of both inter- and intra-specific segregation patterns not only in space but also in time.
87 Moreover, while studies in a wide range of species have described certain components of the
88 structure of inter- and intra-specific spatial and temporal segregation in foraging areas, only recently
89 have such differences been linked explicitly to habitat variables (Pinet *et al.* 2012, Thiers *et al.*
90 2014, Cleasby *et al.* 2015). This is critical to understanding the mechanisms that give rise to
91 observed patterns in foraging behaviour and hence allow predictive modelling of foraging
92 distributions within a community.

93 The pattern of inter- and intra-specific (between-sex) segregation in foraging distributions
94 has received much attention in Northern and Southern Giant Petrels, *Macronectes halli* and *M.*
95 *giganteus*, two recently diverged (~500-200 kya, Techow *et al.* 2010) sister species of seabirds
96 distributed from subtropical to Antarctic waters that breed sympatrically across most of their range.
97 Both species are commonly thought of as scavengers, feeding predominantly on seal and penguin
98 carrion, and in the males, scavenging is often considered to be near-obligate (Hunter 1983,
99 González-Solís 2004, Forero *et al.* 2005). Indeed, several biologging, diet and isotope studies have
100 shown that the sexes employ different foraging strategies: males tend to feed closer to the colony,
101 presumed to be exploiting carrion on local beaches, while females undertake longer trips to capture
102 more pelagic, live prey and are less frequently observed at carcasses on shore (Hunter 1983,
103 González-Solís *et al.* 2000a, González-Solís *et al.* 2002, Forero *et al.* 2005, González-Solís *et al.*
104 2007, Copello *et al.* 2011, Thiers *et al.* 2014). This niche differentiation is considered to be linked
105 with sexual size dimorphism: males of both species are ~20% larger than females and have more
106 robust bills, which may equip them to more efficiently exploit carcasses on land (Hunter 1987,
107 González-Solís 2004). However, tracking with sufficient accuracy to separate coastal (likely
108 scavenging) trips from more pelagic trips has only been undertaken during a single breeding stage,
109 or in multiple stages but from only one species at that site (González-Solís *et al.* 2000b, González-

110 Solís *et al.* 2002, Thiers *et al.* 2014), and the relative roles of sexual and inter-specific segregation
111 across the entire breeding period therefore remain unclear. Crucially, the focus in previous work on
112 breeding stages overlooks a marked difference between the species in timing of breeding. Northern
113 Giant Petrels breed around 6 weeks earlier than Southern Giant Petrels (Hunter 1987, Brown *et al.*
114 2015), hence the breeding stages and associated constraints on foraging behaviour hardly overlap
115 between the species. This allochrony (difference in timing of breeding) is assumed to have been a
116 key factor in divergence of the two species (Hunter 1987 Friesen *et al.* 2007), but its importance in
117 terms of inter- and intra-specific differences in foraging distribution and ecology is much less clear.

118 Despite the similarity in their morphology, ecology and environment, Northern and Southern
119 Giant Petrels breeding sympatrically at Bird Island in South Georgia have displayed different
120 population trajectories over the last two decades, implying a role for behavioural differences in how
121 resources are divided between and within the species. Here, we provide a full analysis of inter- and
122 intra-specific segregation in foraging behaviour among giant petrels throughout breeding,
123 investigating the simultaneous role of different axes of segregation – temporal and spatial – in
124 supporting their coexistence. Previous work suggests that the two species segregate in both space
125 and time and the sexes in space, but these predictions do not account for the temporal offset of
126 breeding stages. We explicitly examine how segregation patterns vary both with competition in
127 absolute time and with the biological constraints of breeding stage, and assess the importance of
128 intra- and inter-specific differences in habitat use.

129

130

131 **METHODS**

132

133 **Study species and field data collection**

134 Tracking data were collected from male and female Northern and Southern Giant Petrels breeding

on Bird Island, South Georgia (54°00'S, 38°03'W) throughout the breeding season in the austral summer of 2005/06. Here, the two species breed sympatrically in the same areas and habitats, but in small groups that tend to consist of the same species. Two devices were deployed on each bird: a 3.6 g geolocator-immersion logger (Mk 4, Mk 5 or Mk 7; British Antarctic Survey, Cambridge, UK) attached with cable-ties to a plastic leg ring, and a 20 g or 30 g satellite tag (91 deployments: platform terminal transmitter, PTT-100; Microwave Telemetry) or a 68 g GPS logger (12 deployments: BGDL-II; Shizuoka University, Japan) to record locations, attached to mantle feathers using Tesa tape. The PTT transmitted at 90 s intervals, and an average of 25.0 locations were obtained per day (one location every 57 min) from each device using the ARGOS satellite system. The GPS devices were set to record a location every 60 min. The immersion loggers tested for saltwater immersion every 3 s, and then either recorded every change of state from wet-to-dry or *vice versa* that lasted ≥ 6 s (Mk 7) or binned these data into 10 min intervals, providing a value between 0 (completely dry) and 200 (completely wet) (Mk 4 and Mk 5). The combined mass of devices and attachments were always $<1.5\%$ of body mass, which is within the recommended 3% limit to avoid deleterious effects (Phillips *et al.* 2003) (heavier GPS devices deployed only on males; for 36 individuals with available data, mean mass \pm s.e. of male and female Northern Giant Petrels was 4.72 ± 0.10 kg and 3.51 ± 0.07 kg, and of male and female Southern Giant Petrels was 4.35 ± 0.10 kg and 3.64 ± 0.09 kg, respectively). All birds were sexed from bill dimensions, and had been fitted previously with a standard British Trust for Ornithology ring and a plastic ring with a unique combination of colour and 3-character alphanumeric code as part of a long-term study programme (Brown *et al.* 2015). Nests were visited daily during incubation and brood-guarding, ringed adults were identified, and the dates when the chick hatched and was first left unattended (i.e. end of brood-guarding) were recorded; thereafter, nests were visited every few days.

Locations provided by PTTs are accurate to 1-10 km (median across all but the poorest-quality ARGOS Location Classes (LC) B and Z (Douglas *et al.* 2012)) which is not sufficiently high

160 resolution to distinguish between attendance at the nest and foraging on land nearby. Separate
161 foraging trips to sea were therefore discriminated from the patterns of saltwater immersion. A trip
162 was defined as any period of intermittent wet records that included no more than 12 hours of
163 contiguous dry time, a conservative interval based on manual inspection of the immersion data in
164 relation to daily observations of nest attendance. However, Mk 7 loggers (55 out of 103
165 deployments) frequently showed likely erroneous wet events, commonly occurring as a series of
166 isolated, short (3-6 seconds) immersions, including from periods when visual records indicated that
167 the bird was at the nest. In order to identify trip start and end times consistently across all logger
168 types, we therefore disregarded all immersion events of ≤ 9 s if no immersion event of > 9 s
169 occurred within the previous or following hour. Trip assignments were confirmed from attendance
170 records for each individual (based on daily visits to nests during incubation and brood-guarding),
171 with only one of the 139 trips identified to that point requiring further splitting. For 13 of the 103
172 deployments, the immersion data did not allow individual trips to be discriminated. The remaining
173 deployments generated 127 trips from 71 individuals across all breeding stages (Table 1). Trips
174 were assigned to incubation, brood-guarding or post-guard chick-rearing, according to nest status
175 when the adult departed.

176

177 **Tracking data analysis**

178 All data preparation and analysis was done in R (R Core Team, 2015) using an azimuthal equal-area
179 projection centred on Bird Island on a WGS 84 coordinate system. Low quality PTT fixes (LC Z)
180 were removed and the retained fixes were iteratively speed-filtered to remove likely erroneous
181 locations, using McConnell *et al.*'s (1992) algorithm implemented in the *argosfilter* package (Freitas
182 2012). A maximum speed of 30 m/s was allowed between two subsequent fixes, based on the 99th
183 percentile of speeds between fixes (29.9 m/s, following González-Solís *et al.* (2000b)). Each filtered
184 location was classified as coastal if it fell within 10 km (conservative ARGOS error) of land at

185 South Georgia, or as pelagic if outside this zone. On this basis, we calculated the proportion of each
186 trip that was coastal, and if this exceeded 50% of fixes, the entire trip was classified as “coastal”; all
187 other trips were classified as “pelagic”. Locations in each trip were interpolated at 30 min intervals
188 using the *adehabitatLT* package (Calenge 2006). Each interpolated fix was classified as daylight or
189 darkness (daylight includes civil twilight, i.e. sun higher than 6° below the horizon). Travel
190 distance, trip duration and maximum distance from the colony (maximum range) were calculated
191 for each trip from the interpolated data. In addition, each trip was binned into early, middle and late
192 periods (hereafter “period”, a three-level factor) according to the calendar date at departure, with
193 cut-off points defined by thirds of the distribution of all trips across both species.

194 Habitat use was analysed for pelagic trips by females only, as the few pelagic trips by males
195 were unbalanced in terms of representing the different breeding stages. Using a 50 km grid across
196 the entire study area, we calculated time spent per grid cell using the package *trip* (Sumner 2015).
197 This was based only on fixes during daylight, as much of the night is spent resting in giant petrels
198 and other large Procellariiformes (González-Solís *et al.* 2002; Phalan *et al.* 2007; but see Connors *et*
199 *al.* 2015), and excluding cells that intersected with land at South Georgia. Habitat data were
200 extracted as 8-day composites for each used grid cell in the relevant temporal period. These were
201 remotely-sensed sea surface temperature (SST), chlorophyll a and non-directional wind speed
202 obtained from the National Oceanic and Atmospheric Administration (BloomWatch 360;
203 http://coastwatch.pfel.noaa.gov/browsers/cwbrowser_global360.html) and bathymetry data
204 (GEBCO) obtained from the British Oceanographic Data Centre. These rasterized data were
205 projected and averaged from their original resolutions (0.05 degrees for SST, 0.05 degrees for
206 chlorophyll a, 0.125 degrees for wind, and 0.5 degrees for bathymetry) onto the coarser 50 km grid
207 used for the calculation of habitat usage.

208 Kernel density contours (utilisation distributions) were calculated from all interpolated
209 locations (daylight and darkness) from both sexes, including pelagic and coastal trips, and also the

210 13 deployments in which individual trips could not be discriminated. We used a 10km grid with the
211 same smoothing parameter of 50 km for all species-sex-breeding stage combinations (the median
212 value of the least-squares cross-validated smoothing factors selected in each subgroup's kernel
213 analysis) to avoid bias when comparing usage distributions. Pairwise overlaps were calculated
214 between the distributions of each species-sex-breeding stage combination as the volume of
215 intersection between the 100% kernels in the package *adehabitatHR*, providing a value ranging
216 from 0 (no overlap) to 1 (complete overlap).

217

218 **Analysis**

219 We investigated how species, sex and trip timing affected trip characteristics and habitat use. Trip
220 characteristics comprised: trip duration; travel distance; maximum range; the proportion of the trip
221 that was coastal, as an indication of scavenging; and trip start time (time of day of departure), which
222 in other dimorphic species varies with sex-specific breeding roles or constraints (Harris *et al.* 2013).
223 Habitat variables were compared across those grid cells in which either species was present. For all
224 these response variables, we first tested a set of models that included breeding stage as the temporal
225 predictor variable, examining how ecological constraints shape foraging behaviour across the
226 species and sexes. We then compared these results to an analysis including period (reflecting
227 calendar date; see above) rather than breeding stage as the temporal predictor, examining whether
228 real-time competition between the species and sexes alters behaviour and habitat use. Northern
229 Giant Petrels lay around 6 weeks earlier (Brown *et al.* 2015; mean first lay date across all monitored
230 nests in study year and area: Northern Giant Petrels [$n = 242$], 01 October 2005; Southern Giant
231 Petrels [$n = 126$], 10 November 2005); among tracked birds, all Southern Giant Petrel incubation
232 fell within Northern Giant Petrels' brood-guarding, while later breeding stages overlapped by
233 around one week between the species. Due to this allochrony, only one individual Northern Giant
234 Petrel yielded discriminated trips in the late period. To prevent this restricted sample from skewing

235 model fits, all models examining period were fitted to tracks only from the early and middle
236 periods, except those for maximum range and trip start time, for which non-discriminated trips were
237 also used because locations at the nest would not affect these variables.

238 Activity patterns could not be reliably inferred from the immersion data due to the erroneous
239 wet events recorded by the Mk 7 loggers (proportion of trip wet for Mk 7 compared to other Mk 4
240 and Mk 5 loggers, in a binomial generalized linear mixed model controlling for species, sex and
241 breeding stage as fixed effects and individual as a random effect with a logit link: 1.54 ± 0.15 , $Z =$
242 10.40 , $P < 0.001$). Within any breeding stage, which is the main constraint on behaviour, there were
243 insufficient numbers of individuals fitted with Mk 4 or Mk 5 loggers to allow robust comparison of
244 species or sexes, and therefore we do not address activity patterns further. In the light of this
245 malfunctioning of the saltwater sensing – likely oversensitivity, responding to wet nests, vegetation
246 or salts washing off plumage in addition to full immersion – previous studies that have used BAS
247 Mk 7 loggers may benefit from a reassessment of their conclusions.

248 We used the same model selection approach for all analyses: an information theoretic
249 approach based on AIC to select the most informative variables from a set of 13 candidate models.
250 This metric assesses the fit of the entire model and thus avoids type I errors associated with multiple
251 testing in this extensive exploratory analysis, while penalizing over-fitting (Burnham & Anderson
252 2013). Our candidate models included 1, 2, or 3 main effects, and interactions between species and
253 sex, and between species and the temporal variable (breeding stage or period), to test whether these
254 variables influenced foraging behaviour differently in the two species. All models with an
255 interaction also contained both component variables as main effects (full list in captions of Tables 3-
256 4 and Supplementary tables 1-2) and all models were fitted using maximum likelihood. Models are
257 generally considered to have an equivalent fit (i.e. similarly informative) if their AIC is within 2
258 units ($\Delta AIC < 2$) of the best-fitting model (Burnham & Anderson 2013). To assess the relative
259 explanatory power of our two temporal variables, breeding stage or period, we compared the fits of

both model sets using a reduced dataset that allowed all models to be robustly fitted; removing the insufficient late period trips left the post-brooding stage insufficiently represented, so this comparison excludes both these temporal divisions. To help differentiate between models of similar fits, we also calculated the AIC weight (relative likelihood) of each model, normalized across the candidate model set to capture their relative fits. We considered a model to have “some” support if it was at least twice as likely relative to the next-best fit model, “good” support if its relative likelihood was over 0.5 (absolutely most likely), and “robust” support if both of these criteria applied. In model sets that did not yield a single, preferred model, we considered all models with likelihoods > 50% of that of the best-fit model, and considered those variables present in over half of these models to have “some” support. Models with a relative likelihood of < 50% of the best-fit model were not considered informative.

All models were linear or generalized linear mixed models ((G)LMMs) fitted using the *nlme* or *lme4* packages (Bates *et al.* 2015, Pinheiro *et al.* 2016). For trip characteristics, individual ID was fitted as a random intercept to account for multiple trips by the same individual. For habitat use, each response variable was modelled as its value in each grid square per trip with each value weighted by the time spent in that grid cell in that trip. As each trip covered many grid squares and each bird potentially undertook multiple trips, trip ID nested within individual ID was fitted as a random intercept. To ensure that all model assumptions were met, chlorophyll and wind were log-transformed to reduce skew in the raw data and normalize residuals.

RESULTS

Foraging areas

Northern and Southern Giant Petrels were both widely distributed in the south-west Atlantic in the

285 austral summer 2005/06, but with marked differences between sexes and breeding stages (Fig. 1).
286 Females tended to forage more widely in both species, and across both species, constraints on
287 distribution were strongest during brood-guarding (Fig. 1). Females of both species were recorded
288 from the Scotia Arc in the south and east, to the Patagonian Shelf and Antarctic Peninsula region in
289 the west. In contrast, male Northern Giant Petrels were always concentrated around the colony, and
290 male Southern Giant Petrels only exploited more distant areas south and south-west of the colony in
291 the post-brood chick-rearing period. Overall, Southern Giant Petrels travelled east and west of Bird
292 Island, whereas Northern Giant Petrels only travelled substantial distances to the west; neither
293 species made substantial use of pelagic waters to the north of the colony. The 50% kernels of
294 females of both species and male Southern Giant Petrels included the South Orkney Islands, and
295 these trips, though pelagic in range, may have included coastal scavenging. Southern Giant Petrels
296 of both sexes visited the South Sandwich Islands, and the females also consistently used a small
297 area to their west.

298 In pairwise comparisons, the mean overlap between the utilisation distributions of each
299 species-sex-breeding stage combination was 0.487 (Table 2). This reflected a mixture of generally
300 high overlaps between the distribution of certain groups – males of both species, male and female
301 Northern Giant Petrels in general, and between both species within the brood-guarding period – and
302 generally low overlaps between other groups – female Northern and Southern Giant Petrels during
303 incubation, and female Southern Giant Petrels in incubation or post-brood and males of both species
304 (Table 2).

305

306 **Trip characteristics**

307 In the model sets testing the effect of breeding stage on the characteristics of foraging trips, there
308 was weak support that trips of Southern Giant Petrels were longer in terms of both travel distance
309 and duration than those of Northern Giant Petrels, although the effect size of the species term in the

310 best-fit model overlapped zero (mean travel distance \pm s.e. for Northern Giant Petrels, 1775 ± 426
 311 km and for Southern Giant Petrels, 3113 ± 461 km, Fig. 2; mean trip duration 79.7 ± 16.1 h and
 312 154.8 ± 21.5 h, respectively; Table 3, Supporting Information Table S1). In keeping with this, the
 313 maximum range of foraging trips – closely correlated with both travel distance and trip duration
 314 (Spearman's r^2 of 0.861 and 0.707, respectively, both $P < 0.001$), but tested with the larger sample
 315 that included non-discriminated trips it did not differ between the two species (overall mean
 316 maximum range, 399 ± 43 km). Regardless of species, the sex and breeding stage affected trip
 317 length: males travelled less far and for less time (Fig. 2; mean trip duration for males, 99.6 ± 17.4 h;
 318 for females 160.6 ± 25.6 h) and for both sexes, compared to incubation, trips during brood-guarding
 319 were somewhat shorter and trips in post-brood substantially longer (Fig. 2, Table 3, Supporting
 320 Information Table S1). These differences were reflected in the smaller maximum range of males
 321 than females, and the maximum range of both sexes during brood-guarding and post-brood were
 322 shorter and longer, respectively (Table 3, Supporting Information Table S1). There was some
 323 evidence that trips started progressively earlier in the day through the breeding stages, as breeding
 324 stage was included in all top-fitting models, but none had strong support, and mean trip start time
 325 did not vary between the species or sexes (Table 3, Supporting Information Table S1). There was no
 326 evidence that sex or breeding stage affected characteristics of foraging trips in different ways in the
 327 two species (i.e. no well-supported interactions with species, Table 3).

328 Clearer differences were seen in the proportion of the trip that birds spent within 10 km of
 329 South Georgia (coastal foraging). Males spent more time near the coast, but this effect was less
 330 pronounced in Northern than in Southern Giant Petrels (Table 3, Supporting Information Table S1,
 331 Fig. 3). In addition, birds spent more time near the coast during brood guarding than in incubation
 332 or post-brood (mean proportion of trip coastal in incubation, 0.402 ± 0.076 , in brood-guarding,
 333 0.473 ± 0.048 , and in post-brood, 0.423 ± 0.056 ; Table 3, Supporting Information Table S1). Across
 334 both species and sexes and all breeding stages, coastal trips were shorter in duration than pelagic

trips (across all trips, mean duration of coastal trips 71.3 ± 11.2 h; of pelagic trips, 178.3 ± 25.6 h; in LMM accounting for species, sex and breeding stage, effect of trip type (pelagic vs. coastal) 88.3 ± 22.8 h, $t = 3.87$, $P = 0.0003$). Among coastal trips, there was no strong evidence that species, sex, or breeding stage affected maximum range (null model best fit, AIC weight 0.407, next-best 0.182), suggesting that none of these factors influenced whether birds stayed close to the colony or foraged coastally elsewhere at South Georgia.

In model sets testing the effect of period instead of breeding stage, there were marked differences in the characteristics of foraging trips of Northern and Southern Giant Petrels, reflecting the interspecific difference in laying dates and therefore the relative (calendar) timing of breeding stages. Southern Giant Petrels made longer trips in the early period and shorter trips in the late period, whereas the reverse was true for Northern Giant Petrels (for Northern Giant Petrels, mean travel distance in early period 1250 ± 307 km, middle period 3827 ± 1582 km; for Southern Giant Petrels, early period 5970 ± 1134 km, middle period 1439 ± 310 km; Table 3, Supporting Information Table S1). Mirroring the patterns by breeding stage, Southern Giant Petrels had a greater maximum range overall, and across both species, later trips were to waters further from the colony, with weak support for an interaction between species and period. Also similar to the breeding stage models, the effect of period on the proportion of a trip that was coastal did not differ between the species, and the start time of trips was not explained by any of the factors examined alongside period (Table 3, Supporting Information Table S1). Overall, breeding stage explained more variation in all trip characteristics than period, with the best-fit breeding stage model having a ΔAIC of > -7 compared to the best-fit period model (ΔAIC between best-fit breeding stage and period models fitted to the same reduced dataset: for travel distance -7.8 , for trip duration -9.9 , for maximum range -10.8 , for trip start time -7.1 , for proportion of the trip that was coastal -28.3).

Habitat use

360 By comparison with the analyses of trip characteristics, there was limited evidence that females of
361 the two species differed in their habitat use on pelagic trips. In models examining the effect of
362 breeding stage, used grid cells did not differ notably in either wind or bathymetry between either
363 species or breeding stage (Table 4). Sea surface temperature of used grid cells was similar in both
364 species, although Northern Giant Petrels used areas of warmer SST across breeding stages whereas
365 SST used by Southern Giant Petrels was more consistent across the breeding season (mean SST \pm
366 s.e. used by Northern Giant Petrels in incubation, brood-guarding and post-brooding 1.32 ± 0.08 °C,
367 1.82 ± 0.07 °C, 4.05 ± 0.12 °C, respectively; and similarly by Southern Giant Petrels 2.50 ± 0.09 °C,
368 2.76 ± 0.06 °C, 2.07 ± 0.06 °C; Table 4, Supporting Information Table S2). The species also differed
369 in chlorophyll a concentrations in their foraging areas, which tended to be lower in grid cells used
370 by Southern than Northern Giant Petrels (mean chlorophyll a \pm s.e. in cells used by Northern Giant
371 Petrels 0.86 ± 0.05 mg/m³, and by Southern Giant Petrels 0.64 ± 0.02 mg/m³; Table 4, Supporting
372 Information Table S2).

373 Models examining the effect of period provided little evidence for interspecific differences
374 in habitat use. Consistent with seasonal environmental changes, period affected both temperature
375 and winds experienced by tracked birds (Table 4), with all birds using areas of warmer water and
376 lower winds in the middle than early period (mean SST \pm s.e. in used cells in early and middle
377 periods, 2.77 ± 0.09 °C and 3.45 ± 0.09 °C respectively; mean wind speed \pm s.e., 8.76 ± 0.04 m/s
378 and 7.55 ± 0.04 m/s respectively; Supporting Information Table S2). In addition, birds foraged in
379 deeper waters in the early than the middle period (mean depth \pm s.e. in used cells in early and
380 middle periods, 3062 ± 48 m and 2476 ± 54 m, respectively; Table 4, Supporting Information Table
381 S2). As in the analysis by breeding stage above, Southern Giant Petrels foraged in areas of
382 somewhat lower chlorophyll a concentrations than Northern Giant Petrels (Table 4, Supporting
383 Information Table S2).

384

385

386 **DISCUSSION**

387

388 The two closely-related species of sexually dimorphic giant petrels at South Georgia provide an
389 opportunity to test the simultaneous roles of spatial and temporal segregation in enabling
390 coexistence between and within species. By tracking both sexes over the entire breeding season, we
391 have shown that foraging Northern and Southern Giant Petrels segregate in time, and that males and
392 females within each species largely segregate in space. Both species employed similar strategies,
393 showing broadly matching trip characteristics and habitat use under the same biological constraints
394 (associated with breeding stage), but direct competition between the species is reduced by the
395 considerably earlier breeding of Northern Giant Petrels (by six weeks; Brown *et al.* 2015). This is
396 rare evidence supporting the importance of breeding allochrony in the coexistence of sister species.
397 Tracked Northern Giant Petrels had a slightly smaller overall foraging range, and a much smaller
398 range during incubation and brood-guarding than Southern Giant Petrels. The smaller range earlier
399 in the season likely reflects their exploitation of carrion, associated with the recovery of the local
400 Antarctic Fur Seal *Arctocephalus gazella* population, which may have contributed to the faster
401 growth in numbers of breeding Northern than Southern Giant Petrels in recent decades (González-
402 Solís *et al.* 2000a; Brown *et al.* 2015). Broadly, males of both species made shorter and more
403 coastal trips (presumably scavenging) while females were more pelagic, yet both sexes were
404 capable of long, pelagic trips. Thus the common perception that giant petrels, especially the males,
405 are near-obligate scavengers through the breeding season needs to be reconsidered. Brood-guarding
406 birds of both sexes tended to forage more coastally and to take shorter trips, as expected given that
407 young chicks require frequent meals, and constant parental care to aid thermoregulation and reduce
408 the risk of predation, as in other surface-breeding Procellariiformes (Catry *et al.* 2006). Overall, our
409 results demonstrate the role of segregation along different axes of variation (temporal, in terms of

410 both breeding stage and absolute time, and spatial) in maintaining coexistence among similar
411 species.

412 Although the tracked Northern and Southern Giant Petrels differed to some extent in
413 distribution and behaviour, sex and breeding stage had a stronger influence than species or period
414 (calendar date) on their foraging ecology, with little evidence that these effects differed between the
415 two species. The similarity in the foraging strategies of these sister taxa implies that intra-specific
416 segregation patterns have changed little in the 500-200ky since the species diverged (Techow *et al.*
417 2010) and hence suggests that they are ancestral to both species (Hunter 1987). Despite their similar
418 strategies in biological time (breeding stage), in absolute time (i.e. in terms of calendar date), many
419 aspects of foraging behaviour differed between the species. Thus, rather than partitioning
420 contemporaneous foraging resources through changes in a complex suite of behaviours, competition
421 between these ecologically and morphologically similar species appears instead to be resolved
422 through a shift in breeding phenology, a trait that in many species has the potential for rapid
423 evolution (Franks *et al.* 2007, Friesen *et al.* 2007, Tarka *et al.* 2015). In other systems, inter-specific
424 competition may be resolved through dynamic differentiation in other behavioural traits: Rock
425 Shags *Phalacrocorax magellanicus* undertake longer trips when breeding in sympatry with the Red-
426 legged Cormorant *P. gaimardi* than when breeding in single-species colonies (Frere *et al.* 2008).
427 Indeed, it has been suggested that behavioural modification is the most likely immediate response
428 of seabirds to environmental change (Lewis *et al.* 2006). In giant petrels, our results indicate that
429 several aspects of behaviour, most notably breeding phenology but also foraging strategies, appear
430 to be labile traits that may have allowed initial rapid responses to reduce competition through
431 resource partitioning. In addition to the temporal segregation, Southern Giant Petrels tended to
432 forage slightly further afield and exploit more westerly areas than Northern Giant Petrels. This
433 difference in foraging ranges is likely to reduce costs of reproduction to some extent in Northern
434 Giant Petrels, and hence may underlie the difference between the species in population growth rate.

435 The sexual segregation we observed in both species, with females typically foraging further
436 afield than males, particularly in early breeding stages, is consistent with previous findings from a
437 range of biologging, stable isotope and diet studies (Hunter 1983, González-Solís *et al.* 2000a,
438 González-Solís *et al.* 2000b, González-Solís *et al.* 2002, Forero *et al.* 2005, Phillips *et al.* 2011,
439 Raya Rey *et al.* 2012, Thiers *et al.* 2014). These behavioural differences are associated with
440 pronounced sexual size dimorphism: size differences have been assumed to constrain the larger
441 males, which presumably have higher costs of flying, to specialise on the exploitation of seal and
442 penguin carrion on local beaches, excluding the slighter females from these resources and forcing
443 them to forage further afield (Hunter 1983, González-Solís *et al.* 2000b). However, our finding that
444 both sexes of both species undertook substantial (> 4 days and > 1800 km) pelagic trips as well as
445 coastal trips shows that foraging distributions in the giant petrels are more plastic than commonly
446 appreciated during the breeding season, and not tightly constrained by physiological limits such as
447 wing loading or competitiveness. This supports previous findings from the non-breeding season and
448 diet studies (Hunter 1983, González-Solís *et al.* 2007, but see Thiers *et al.* 2014). Although females
449 in particular undertook long foraging trips that resembled other medium-to-large, wide-ranging
450 Procellariiformes, such as albatrosses and White-chinned Petrels *Procellaria aequinoctialis* (Phillips
451 *et al.* 2004, Phillips *et al.* 2005, Phillips *et al.* 2006), males were far from exclusively coastal when
452 not constrained by chick demands. Similarly, consistent with the morphological specializations of
453 both sexes (but particularly males) for exploiting carrion (Hunter 1983, González-Solís 2004,
454 Forero *et al.* 2005), coastal foraging by both males and females indicates that scavenging is less
455 male-dominated than commonly assumed. Overall, this variability suggests that giant petrels,
456 broadly considered to be specialist scavengers (Hunter 1983, González-Solís 2004, Forero *et al.*
457 2005), should more rightly be considered as generalists, with both sexes able to exploit both coastal
458 and pelagic areas and hence consume both carrion and oceanic prey. Indeed, the high level of
459 opportunism would explain why their population trends are not correlated with carrion availability

460 at another sub-Antarctic colony, Marion Island (De Bruyn *et al.* 2007).

461 The range of and plasticity in foraging strategies among giant petrels indicates that
462 segregation patterns are less shaped by adult physiology, and more by the demands of reproduction
463 (incubating the egg or feeding the growing chick). Indeed, breeding stage explained more variation
464 in foraging behaviour than calendar date, and moreover the range of both species was most
465 restricted around the colony during the brood-guard stage of chick-rearing, when the parents
466 alternate at the nest to provide food, warmth and protection from predators, which may be
467 facilitated by the shorter duration of the coastal trips favoured in this breeding stage. This change in
468 behaviour in response to biotic conditions indicates that both male and female giant petrels make
469 dynamic foraging decisions throughout the breeding season, and hence probably also throughout the
470 rest of year, to meet changing demands. This behavioural diversity suggests that divergent selection
471 on foraging strategies between males and females is not strong, and hence that morphological
472 differences related to foraging are unlikely to completely explain intra-specific foraging segregation
473 in this system. Indeed, rather than morphology driving segregation, as often assumed, it is not
474 currently possible to rule out that intra-specific behavioural differences may have arisen first – as
475 with the inter-specific differences above, a labile initial response to reduce competition – with the
476 striking morphological differences emerging in consequence as further specialisation.

477 Although foraging behaviour displayed multi-faceted and dynamic variability, we found
478 little evidence for consistent use of particular habitats among females on pelagic trips. Southern
479 Giant Petrels appeared to forage in areas of lower chlorophyll (at the time of the trip) than Northern
480 Giant Petrels, and males and females in later breeding stages, especially Northern Giant Petrels,
481 used warmer waters than birds in earlier breeding stages. However, we cannot confidently discern
482 whether these are active habitat choices to optimise nutrient intake or unavoidable consequences of
483 the spatial and temporal segregation patterns. In particular, breeding allochrony in giant petrels is
484 likely associated with variation in the habitats available to each species at each breeding stage; the

485 extent to which habitat use reflects real preferences for certain oceanographic features would
486 require further investigation. Moreover, as the limited number of pelagic trips undertaken by males
487 precluded a robust analysis of their habitat choices, we were not able to investigate the population-
488 level importance of habitat preferences. Further, some trips that we classified as pelagic included
489 time spent close or on land at the South Orkney and South Sandwich islands, and other parts of the
490 Scotia Arc, where giant petrels may have been exploiting carrion on beaches rather than oceanic
491 prey. Lastly, as with the trip characteristics discussed above, foraging habitat varied substantially
492 across a relatively small number of trips (max. 22 trips per species/sex/temporal variable group);
493 tracking more individuals may serve to reduce this variation and hence clarify real differences in
494 foraging strategies between the groups. Due to these difficulties in interpreting the observed
495 patterns in habitat use, the implications of interspecific differences in habitat preference for
496 population trajectories remain unclear. Studies of habitat use at other colonies where the two species
497 breed sympatrically but show different population growth rates (Delord *et al.* 2008; Ryan *et al.*
498 2009) would provide an informative natural experiment in which to fully assess the role of habitat
499 availability in population trajectories. This would be of substantial value in moving from a
500 descriptive to a predictive understanding of inter- and intra-specific resource partitioning in seabirds
501 (Pinet *et al.* 2012, Thiers *et al.* 2014, Cleasby *et al.* 2015).

502 Our single-colony study has highlighted the importance of segregation along both spatial
503 and temporal axes for the coexistence of two ecologically and morphologically similar seabird
504 species. Sexual segregation was mainly spatial, constrained by breeding stage, whereas segregation
505 between the species arose from breeding allochrony. Intra-specific competition had a stronger
506 influence on distribution and behaviour than inter-specific competition, mirroring patterns in other
507 sympatric vertebrates (seabirds: Weimerskirch *et al.* 2009, Young *et al.* 2010; reef fish: Forrester *et*
508 *al.* 2006). These behavioural differences are not only of ecological and evolutionary importance but
509 may also have conservation implications if they expose different parts of the population or

community to different threats (van Toor *et al.* 2011). However, while wider-ranging female giant
petrels have historically been proposed to be more at risk of incidental mortality in longline
fisheries than males (González-Solís *et al.* 2000a, Otley *et al.* 2006) and the same could apply to
wider-ranging Southern compared to Northern Giant Petrels, current management of demersal
fisheries at South Georgia and around the Falkland Islands includes strict bycatch mitigation
methods and monitoring of compliance (Tancell *et al.* 2016), and so these range differences are
unlikely to affect bycatch likelihood. Finally, while our investigation focussed on segregation of two
species within the seabird community, foraging strategies within a multi-species assemblage also
depend on those of other taxa, segregation at smaller scales (Navarro *et al.* 2013), and individual
habitat preferences (Phillips *et al.* 2017). Given the structured segregation patterns of giant petrels
at South Georgia that we have described, both within and between species, this species pair could
be an informative model in which to extend our understanding of how simultaneous competition on
many different scales shapes foraging strategies and habitat use.

523

524

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534 **SUPPORTING INFORMATION**

535

536 Additional Supporting Information can be found in the online version of this article:

537 **Table S1:** Full model outputs of the best-fit models describing trip characteristics of

538 foraging Northern and Southern Giant Petrels tracked at Bird Island, South Georgia.

539 **Table S2:** Full model outputs of the best-fit models describing habitat use of foraging

540 Northern and Southern Giant Petrels tracked at Bird Island, South Georgia.

541 **Tables**

542

Breeding stage	Sex	Northern Giant Petrels		Southern Giant Petrels	
		Trips	Individuals	Trips	Individuals
For trip characteristics					
Incubation	Females	3 (0)	3	10 (0)	10
	Males	4 (0)	4	9 (0)	9
Brood-guarding	Females	10 (2)	7	22 (0)	13
	Males	19 (3)	11	18 (1)	11
Post-brood	Females	11 (2)	9	10 (0)	8
	Males	8 (5)	6	16 (0)	12
For habitat use					
Incubation	Females	3	3	10	10
	Males	0	0	1	1
Brood-guarding	Females	5	3	19	10
	Males	7	5	1	1
Post-brood	Females	8	6	7	6
	Males	0	0	7	6

543

544 **Table 1.** Sample sizes by species, sex and breeding stage in analyses of trip characteristics for all
545 deployments and habitat use for pelagic trips only of giant petrels tracked from Bird Island, South
546 Georgia, in austral summer 2005/06. The number of deployments in which separate trips could not
547 confidently be discriminated are indicated in parentheses. Note that very few males undertook
548 pelagic trips and that these were unevenly distributed between breeding stages; analysis of habitat
549 use were therefore carried out on females only.

			Northern Giant Petrels						Southern Giant Petrels					
			Incubation		Brood-guard		Post-brood		Incubation		Brood-guard		Post-brood	
			Females	Males	Females	Males	Females	Males	Females	Males	Females	Males	Females	Males
Northern Giant Petrels	Incubation	Females		0.459	0.569	0.488	0.463	0.448	0.167	0.450	0.438	0.454	0.284	0.366
		Males			0.821	0.956	0.360	0.944	0.137	0.821	0.359	0.925	0.219	0.432
	Brood-guard	Females				0.850	0.510	0.815	0.213	0.799	0.496	0.821	0.336	0.457
		Males					0.379	0.939	0.142	0.839	0.374	0.931	0.228	0.436
	Post-brood	Females						0.353	0.326	0.379	0.602	0.368	0.462	0.470
		Males							0.134	0.852	0.349	0.974	0.215	0.432
Southern Giant Petrels	Incubation	Females								0.203	0.424	0.139	0.339	0.337
		Males									0.432	0.867	0.261	0.498
	Brood-guard	Females										0.356	0.497	0.579
		Males											0.221	0.437
	Post-brood	Females												0.496
		Males												

550

551 **Table 2.** Overlap in utilisation distributions by species, sex and breeding stage of giant petrels
552 tracked from Bird Island, South Georgia, in austral summer 2005/06. Values range from 0 (no
553 overlap) to 1 (kernels completely intersect). For ease of interpretation, high overlaps (> 0.75) are
554 shown in bold and low overlaps (< 0.25) in italics.

Model terms	d.f.	ΔAIC	AIC weight	Support	Model terms	d.f.	ΔAIC	AIC weight	Support
By breeding stage					By calendar date (period)				
<i>Travel distance</i>					<i>Travel distance</i>				
Species + <u>Sex</u> + <u>Breeding stage</u>	7	0.00	0.402	*	<u>Sex</u> + <u>Species*Period</u>	7	0.00	0.985	**
<u>Sex</u> + <u>Breeding stage</u>	6	1.35	0.204	*	Species*Period	6	8.51	0.014	
Species*Sex + Breeding stage	8	1.68	0.173		Sex	4	16.62	0.000	
Sex + Species*Breeding stage	9	1.84	0.160		Sex + Period	5	18.15	0.000	
Species + Breeding stage	6	5.19	0.030		Species + Sex	5	18.15	0.000	
<i>Trip duration</i>					<i>Trip duration</i>				
<u>Sex</u> + <u>Species*Breeding stage</u>	9	0.00	0.397	*	<u>Sex</u> + <u>Species*Period</u>	7	0.00	0.896	**
<u>Species</u> + <u>Sex</u> + <u>Breeding stage</u>	7	1.19	0.219	*	Species*Period	6	4.33	0.103	
Species*Sex + Breeding stage	8	2.79	0.098		Sex	4	14.66	0.001	
Species*Breeding stage	8	2.85	0.095		Species + Sex	5	15.92	0.000	
Species + Breeding stage	6	2.91	0.092		Sex + Period	5	16.63	0.000	
<i>Max. range</i>					<i>Max. range</i>				
<u>Sex</u> + <u>Breeding stage</u>	6	0.00	0.425	*	<u>Sex</u> + <u>Species*Period</u>	9	0.00	0.427	*
Species + <u>Sex</u> + <u>Breeding stage</u>	7	0.49	0.332	*	<u>Sex</u> + <u>Period</u>	6	1.14	0.241	*
Species*Sex + Breeding stage	8	2.42	0.126		Species + Sex + Period	7	2.91	0.100	
Sex + Species*Breeding stage	9	2.61	0.115		Sex	4	3.00	0.095	
Breeding stage	5	13.68	0.000		Species + Sex	5	3.70	0.067	
<i>Proportion of trip coastal</i>					<i>Proportion of trip coastal</i>				
<u>Species*Sex</u> + <u>Breeding stage</u>	7	0.00	0.791	**	<u>Species*Sex</u> + <u>Period</u>	6	0.00	0.495	*
Species + Sex + Breeding stage	6	4.11	0.101		<u>Species</u> + <u>Sex</u> + <u>Period</u>	5	0.77	0.338	*
Sex + Breeding stage	5	4.52	0.083		Sex + Species*Period	6	2.49	0.143	
Sex + Species*Breeding stage	8	6.94	0.025		Sex + Period	4	6.00	0.025	
Species *Sex	5	52.57	0.000		Species *Sex	5	28.63	0.000	
<i>Trip start time</i>					<i>Trip start time</i>				
Species* <u>Breeding stage</u>	8	0.00	0.361	*	(intercept only)	3	0.00	0.370	**
<u>Breeding stage</u>	5	1.08	0.210	*	Species	4	1.68	0.160	
Sex + Species*Breeding stage	9	1.48	0.172		Sex	4	1.98	0.137	
Species + Breeding stage	6	2.92	0.084		Period	5	3.23	0.074	
Sex + Breeding stage	6	3.03	0.079		Species + Sex	5	3.66	0.059	

557 **Table 3.** The five best-fitting models describing trip characteristics of giant petrels tracked from
558 Bird Island, South Georgia, in austral summer 2005/06, in relation to either breeding stage (left
559 model set) or period (reflecting calendar date; right model set). The full candidate model set for

560 each response variable was: Species; Sex; Time (breeding stage or period); Species + Sex; Sex +
561 Time; Species + Time; Species + Sex+ Species:Sex; Species + Time + Species:Time; Species + Sex
562 + Time; Species + Sex+ Time + Species:Sex; Species + Sex + Time + Species:Time; null model
563 (intercept only). Important variables (present in over half of models within 50% of the relative
564 likelihood of the best-fit model) are underlined. In the “Support” column, ** indicates a single best-
565 fit model and * indicates one of several models considered informative (full model selection criteria
566 in main text).

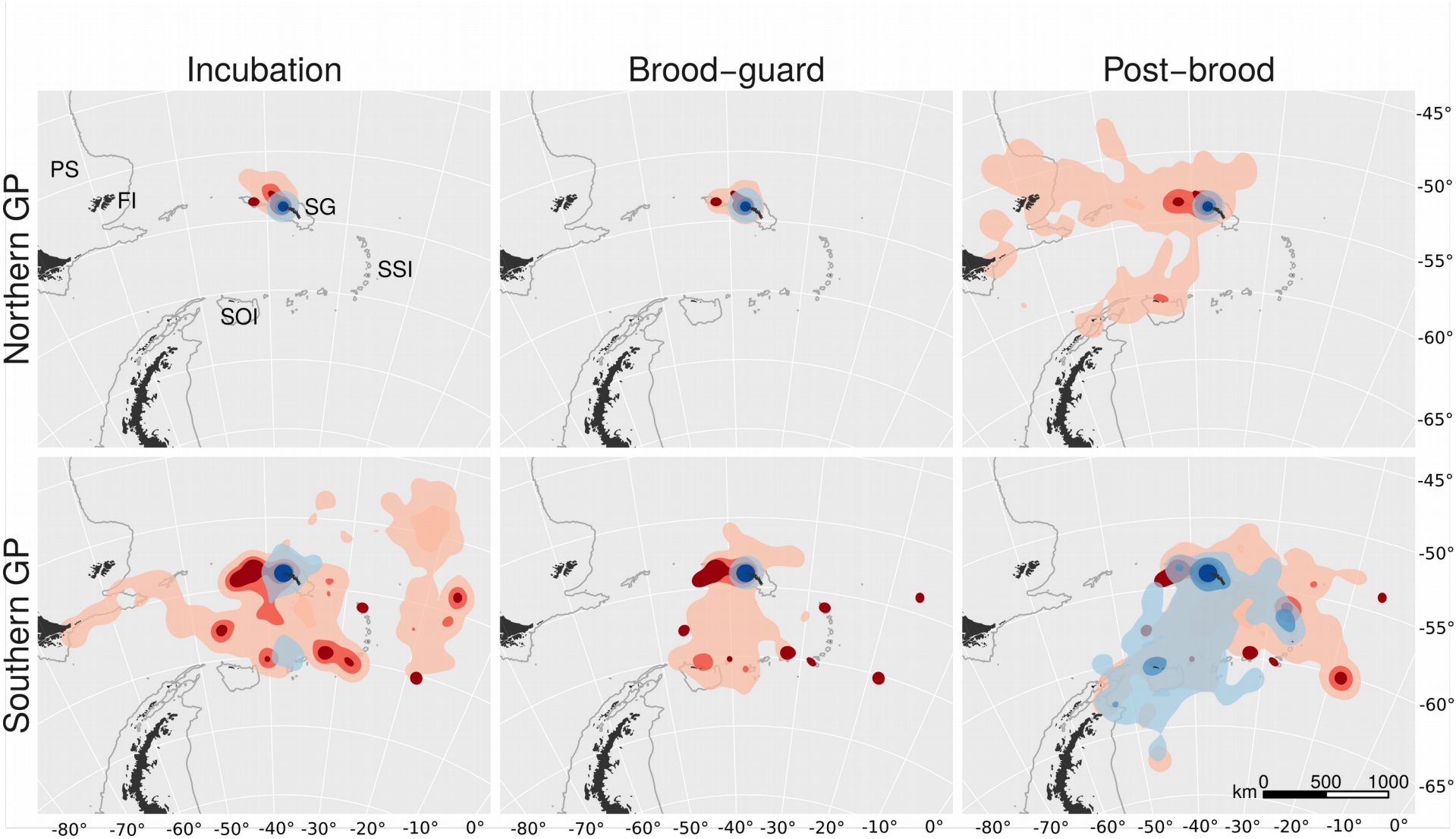
Model terms	d.f.	ΔAIC	AIC weight	Support	Model terms	d.f.	ΔAIC	AIC weight	Support
By breeding stage					By calendar date (period)				
<i>Temperature</i>					<i>Temperature</i>				
<u>Species * Breeding stage</u>	9	0.00	0.561	**	<u>Period</u>	5	0.00	0.318	*
Breeding stage	6	2.28	0.180		Species + <u>Period</u>	6	0.60	0.235	*
(intercept only)	4	2.80	0.138		Species * <u>Period</u>	7	0.82	0.211	*
Species + Breeding stage	7	4.25	0.067		(intercept only)	5	3.03	0.167	*
Species	5	4.66	0.055		Species	5	3.03	0.070	
<i>Chlorophyll</i>					<i>Chlorophyll</i>				
<u>Species</u> + Breeding stage	7	0.00	0.466	*	<u>Species</u> + Period	6	0.00	0.375	*
<u>Species</u>	5	0.38	0.386	*	<u>Species</u>	5	0.20	0.339	*
Species * Breeding stage	9	3.21	0.094		Species * Period	7	1.99	0.139	
(intercept only)	4	4.65	0.045		(intercept only)	4	2.52	0.106	
Breeding stage	6	7.97	0.009		Period	5	4.39	0.042	
<i>Wind</i>					<i>Wind</i>				
(intercept only)	4	0.00	0.458	*	Species + <u>Period</u>	6	0.00	0.460	*
Species * Breeding stage	9	1.38	0.230	*	<u>Period</u>	5	0.52	0.355	*
Species	5	1.91	0.177		Species * Period	7	2.00	0.169	
Breeding stage	6	3.08	0.098		(intercept only)	4	7.52	0.011	
Species + Breeding stage	7	4.99	0.038		Species	5	9.23	0.005	
<i>Bathymetry</i>					<i>Bathymetry</i>				
(intercept only)	4	0.00	0.412	*	<u>Period</u>	5	0.00	0.479	*
Breeding stage	6	0.70	0.290	*	Species + <u>Period</u>	6	0.50	0.373	*
Species	5	2.00	0.152		Species * Period	7	2.41	0.144	
Species + Breeding stage	7	2.70	0.107		(intercept only)	4	9.79	0.004	
Species * Breeding stage	9	4.67	0.040		Species	5	11.78	0.001	

Table 4. The five best-fitting models describing habitat variables in cells used by female giant
 petrels on pelagic trips tracked from Bird Island, South Georgia, in austral summer 2005/06, in
 relation to either breeding stage (left model set) or period (reflecting calendar date; right model set).
 As sex was not investigated in this analysis, the candidate model set was: Species; Time (breeding
 stage or time slot); Species + Time; Species + Time+ Species:Time; null model (intercept only).
 Important variables (present in over half of models within 50% of the relative likelihood of the best-
 fit model) are underlined. In the “Support” column, ** indicates a single best-fit model and *

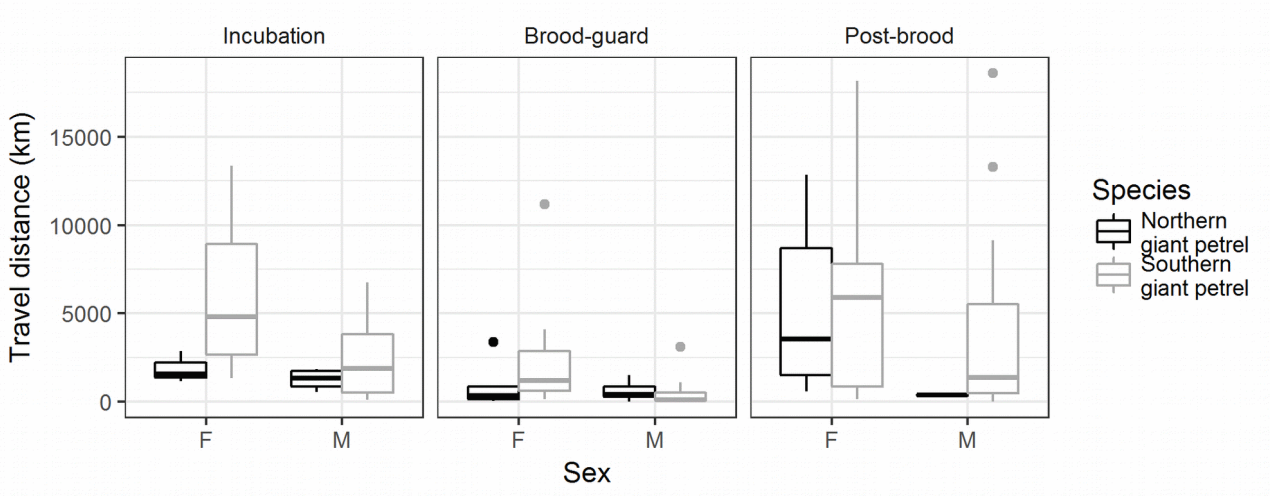
576 indicates one of several models considered informative (full model selection criteria in main text).

578

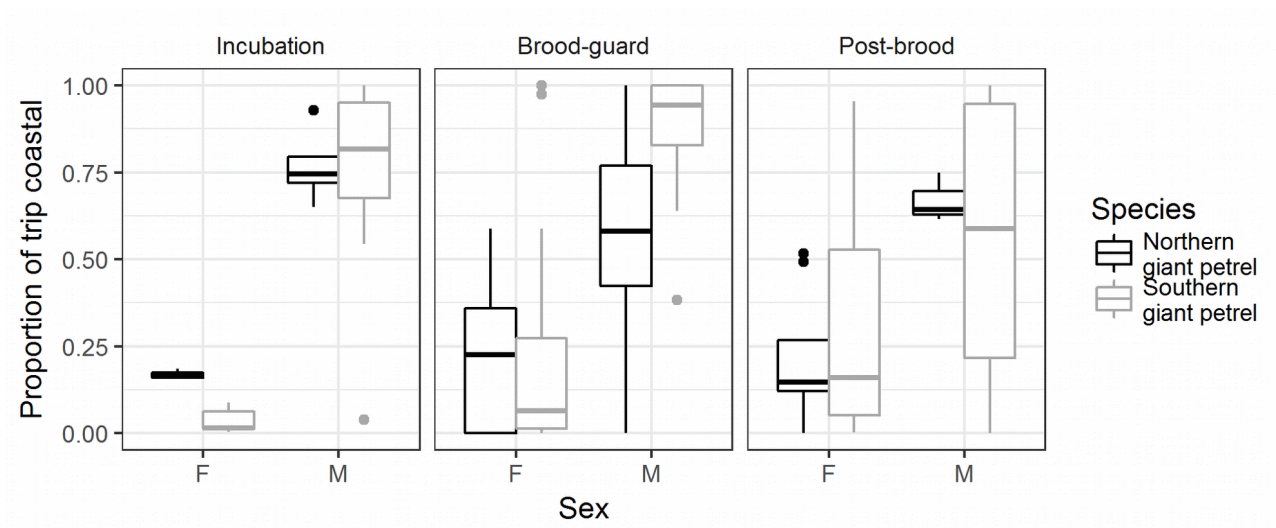
579



580 **Figure 1.** Distributions of Northern and Southern Giant Petrels tracked from Bird Island, South Georgia, in 2005/06 by sex and breeding stage
581 (calendar date differs between species), shown as 95%, 50% and 25% utilisation distributions. Females are shown in red and males in blue, with
582 progressively higher-probability kernels in darker colours. Landmasses are shown in black, with the tip of South America on the left hand map edge
583 and the Antarctic Peninsula on the bottom edge. Relevant island groups and features are marked on the first panel: the Patagonian Shelf (PS), Falkland
584 Islands (FI), South Georgia (SG), South Sandwich Islands (SSI) and South Orkneys (SOI); the latter three island groups illustrate the line of the Scotia
585 Arc. Pale grey lines show bathymetric contours at 1000 m. The map is in an azimuthal equal area projection centred on the colony at Bird Island, on
586 the western tip of South Georgia.



588 **Figure 2.** Travel distance of Giant Petrels tracked from Bird Island, South Georgia, in austral summer 2005/06 for each species, sex and breeding
589 stage, shown as box plots of the raw data for each subgroup. Boxes show the median with upper and lower quartiles. Whiskers show the data range
590 excluding observations more than 1.5 times the inter-quartile range from the median; if any subgroup contains such outliers, they are shown as filled
591 circles.



592 **Figure 3.** The proportion of each trip spent within 10 km of South Georgia (coastal foraging) by Giant Petrels tracked from Bird Island, South Georgia,
 593 in austral summer 2005/06, shown for each species, sex and breeding stage as boxplots of the raw data. Boxes show the median with upper and lower
 594 quartiles. Whiskers show the data range excluding observations more than 1.5 times the inter-quartile range from the median; if any subgroup contains
 595 such outliers, they are shown as filled circles.

596