

1 **Indoor radon measurements in south west England explained by topsoil and stream**  
2 **sediment geochemistry, airborne gamma-ray spectroscopy and geology**

3 Antonio Ferreira<sup>1</sup>, Z. Daraktchieva<sup>2</sup>, D. Beamish<sup>1</sup>, C. Kirkwood<sup>1</sup>, T.R. Lister<sup>1</sup>, M. Cave<sup>1</sup>, J. Wragg<sup>1</sup>, K. Lee<sup>1</sup>

4 1 British Geological Survey, Keyworth, Nottingham, NG12 5GG, UK; 2 Public Health England, UK

5

6 **Corresponding author:**

7 Antonio Ferreira

8 British Geological Survey, Keyworth, Nottingham, NG12 5GG, UK

9 Email:[antonio@bgs.ac.uk](mailto:antonio@bgs.ac.uk)

10 Tel: +44(0)115 936 3465

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12 **KEYWORDS**

13 Indoor radon, Geology, Airborne gamma-ray, stream sediment geochemistry, topsoil geochemistry, analysis  
14 of variance, compositional data

15

16 **HIGHLIGHTS**

- 17 • Tested datasets (geology, airborne gamma-ray, stream sediment geochemistry and topsoil  
18 geochemistry) are useful for radon mapping purposes
- 19 • “Complete” topsoil geochemistry is preferable to the use of uranium concentrations only for radon  
20 prediction.
- 21 • For SW England the indoor radon variability is higher at short distances than at long distances

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6 **Corresponding author:**

7 Antonio Ferreira

8 British Geological Survey, Keyworth, Nottingham, NG12 5GG, UK

9 Email: [antonio@bgs.ac.uk](mailto:antonio@bgs.ac.uk)

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27 **ABSTRACT**

28 Predictive mapping of indoor radon potential often requires the use of additional datasets. A range of  
29 geological, geochemical and geophysical data may be considered, either individually or in combination.  
30 The present work is an evaluation of how much of the indoor radon variation in south west England can be  
31 explained by four different datasets: a) the geology (G), b) the airborne gamma-ray spectroscopy (AGR), c)  
32 the geochemistry of topsoil (TSG) and d) the geochemistry of stream sediments (SSG). The study area was  
33 chosen since it provides a large (197,464) indoor radon dataset in association with the above information.  
34 Geology provides information on the distribution of the materials that may contribute to radon release while  
35 the latter three items provide more direct observations on the distributions of the radionuclide elements  
36 uranium (U), thorium (Th) and potassium (K). In addition, (c) and (d) provide multi-element assessments of  
37 geochemistry which are also included in this study.  
38 The effectiveness of datasets for predicting the existing indoor radon data is assessed through the level (the  
39 higher the better) of explained variation (% of variance or ANOVA) obtained from the tested models. A  
40 multiple linear regression using a compositional data (CODA) approach is carried out to obtain the required  
41 measure of determination for each analysis.  
42 Results show that, amongst the four tested datasets, the soil geochemistry (TSG, i.e. including all the  
43 available 41 elements, 10 major – Al, Ca, Fe, K, Mg, Mn, Na, P, Si, Ti - plus 31 trace) provides the highest  
44 explained variation of indoor radon (about 40%); more than double the value provided by U alone (ca. 15%),  
45 or the sub composition U, Th, K (ca. 16%) from the same TSG data. The remaining three datasets provide  
46 values ranging from about 27% to 32.5%. The enhanced prediction of the AGR model relative to the U, Th, K  
47 in soils suggests that the AGR signal captures more than just the U, Th and K content in the soil.  
48 The best result is obtained by including the soil geochemistry with geology and AGR (TSG+G+AGR, ca. 47%).  
49 However, adding G and AGR to the TSG model only slightly improves the prediction (ca. +7%), suggesting  
50 that the geochemistry of soils already contain most of the information given by geology and airborne  
51 datasets together, at least with regard to the explanation of indoor radon.

52 From the present analysis performed in the SW of England, it may be concluded that each one of the four  
53 datasets is likely to be useful for radon mapping purposes, whether alone or in combination with others.  
54 The present work also suggest that the complete soil geochemistry dataset (TSG) is more effective for indoor  
55 radon modelling than using just the U (+ Th, K) concentration in soil.

56

## 57 **1. INTRODUCTION**

58 Radon ( $^{222}\text{Rn}$ ) is a geogenic radioactive gas permanently produced in all rocks and soils by radioactive decay  
59 of radium ( $^{226}\text{Ra}$ ), which primarily derives from uranium ( $^{238}\text{U}$ ). However uranium levels tend to vary with  
60 rock type, with higher Rn production being observed in uranium (and/or radium)-rich rocks, as expected.  
61 High concentrations of U (and/or Ra) can be found in rocks such as, uranium ores, some granites, black  
62 shales, some sandstones, phosphate rocks, some limestones and U (and/or Ra)-rich soils (e.g., over some  
63 limestones) (Appleton, 2013). In general, locations overlying rocks containing high levels of uranium show  
64 higher radon potential than those overlying rocks with low levels of U. Radon gas released from rocks and  
65 soils reaches open air, and rapidly disperses in the atmosphere to average levels of  $4 \text{ Bq.m}^{-3}$  (HPA, 2010). As  
66 a consequence, radon typically becomes a health risk only in confined areas, where it tends to accumulate,  
67 such as in underground caves and mines, as well as inside houses and other buildings. Radon from the  
68 ground can enter houses through gaps and cracks in the floor, walls and pipes mainly due to differential  
69 indoor - outdoor – soil air conditions (pressure and temperature) (Appleton, 2013). As radon is a heavy  
70 noble gas, it tends to sink and has the potential to concentrate further in poorly ventilated basements and  
71 cellars. In the UK, the average indoor radon concentration is  $20 \text{ Bq.m}^{-3}$  (ranging from less than  $10 \text{ Bq.m}^{-3}$  to  
72 over  $17000 \text{ Bq.m}^{-3}$ ) and thus is 10 times lower than the  $200 \text{ Bq.m}^{-3}$  Action Level established for indoor radon  
73 in the UK (NRPB, 1990). Radon concentration in the soil pore space is quite variable, ranging from less than  
74 1 to more than  $2500 \text{ kBq.m}^{-3}$  (Appleton, 2013), thus it is generally 1 to 3 orders of magnitude higher than in  
75 the air above the ground, both indoor and outdoor (Varley and Flowers, 1998). In addition to the rock and  
76 soil capacity for radon gas production, other ground characteristics, such as permeability, water content,

77 organic matter or proximity to faults, as well as the weather conditions, temperature, pressure and  
78 precipitation, are among the factors to be considered in order to account for the radon concentration in soil.  
79 Despite of other relevant factors (atmospheric conditions, construction of the building, lifestyle including  
80 heating and ventilation routines) contributing to radon levels inside houses, geology is the most important  
81 one, as it is the source. According to Appleton and Miles (2010) approximately 25% of the total variation of  
82 indoor radon in England and Wales can be explained by the geology. Hunter et al. (2009) found that  
83 geology, at 19.7%, is the most important factor explaining the variance of UK indoor radon, while less than  
84 half (8.9%) was attributed to the sum of seven house-related factors, including, house type (3.8%), double-  
85 glazing (2.3%), date of building (1.1%), floor level (1.0%), floor type (0.5%), ownership (0.1%) and draught  
86 proofing (0.1%).

87

### 88 ***Radon mapping***

89 The dependence of indoor radon on the geology has led to the development of radon mapping methods that  
90 use geological maps together with indoor radon measurements. Other proxies, such as Rn concentration in  
91 the soil gas; U (and Ra) content measured in rock, soil or stream sediment; the U content estimated from  
92 ground or airborne Gamma-Ray spectroscopy; soil / rock permeability data, or proximity to faults are also  
93 used for the assignment of a radon potential classes (Appleton, 2013). Often, the proxies are used according  
94 to their availability. Relative to the other mentioned proxies, indoor radon measurements have the  
95 advantage of being directly related to the space where people spend most of their time (i.e. home and other  
96 buildings). Typically, the indoor radon measurements are spatially very unevenly distributed, clustered in  
97 highly populated areas and rare or absent in rural areas. This can be partially overcome by using the geology  
98 to extend the trends to areas with sparse or absent indoor data. Nevertheless, it may constitute a drawback  
99 for large areas with no human occupation at present, but which may be planned for the near future. In fact,  
100 the implementation of radon preventative measures in new buildings on radon prone areas (e.g., in the UK),  
101 will, over time, induce an artificial reduction of the estimated radon potential, in case this is based on indoor  
102 radon measurements.

103 In the UK, Public Health England (PHE) and the British Geological Survey (BGS) collaborate in the production  
104 of radon potential maps; these information have already been released for England and Wales (Miles et al.,  
105 2007), for Scotland (Miles et al., 2011) and for Northern Ireland with the collaboration of the Geological  
106 Survey Northern Ireland (Daraktchieva et al., 2015). Digital geological information (DiGMapGB-50k, mostly  
107 at 1:50 000 scale, [http://www.bgs.ac.uk/products/digitalmaps/DiGMapGB\\_50.html](http://www.bgs.ac.uk/products/digitalmaps/DiGMapGB_50.html)), is used together with  
108 the indoor radon measurements for mapping the radon risk. The bedrock and the superficial 1:50 000 scale  
109 units are simplified (by age and lithology, or permeability and genetic type respectively) to produce a new  
110 set of codes, which are then combined together in a bedrock-superficial parent material (**BS**, geological  
111 combination) polygon code. Finally, the **BS** polygons are intercepted with the British National Grid (BNG) 1-  
112 km grid squares, resulting in the **KM1BS** polygons which constitute the basic mapping units. The following  
113 steps of the UK radon mapping methodology are performed after allocating each indoor radon  
114 measurement to the **KM1BS** polygon underlying it. The radon potential is then computed for each basic  
115 polygon (**KM1BS**) using the nearest data and provided that it belongs to same geological combination (**BS**).  
116 After mapping each **BS** separately, these are assembled in a final map. This radon potential map provides an  
117 estimation of the probability of homes in the UK having radon concentrations above the UK Action Level  
118 ( $200 \text{ Bq.m}^{-3}$ ). Further explanation about the UK radon mapping method and main results can be found in  
119 Miles and Appleton (2005), Miles et al. (2007, 2011) and Daraktchieva et al. (2015).

120

121 This paper is a contribution to the European Geogenic Radon Potential / Natural Radiation mapping project,  
122 led by the Joint Research Centre (JRC) of the European Commission. JRC started to design a European map  
123 of the geogenic radon potential exploring different approaches used by different European countries for  
124 development of national radon risk maps. A general classification scheme applicable to all countries is under  
125 development, which reflects the different experiences and type of information used at national level.  
126 Datasets including key variables such as indoor radon, geology, soil gas radon, U (and Ra) concentrations in  
127 soil and bedrock, airborne gamma-ray data, terrestrial gamma dose rate and permeability, seem to be the  
128 most obvious and appropriate approach.

129 The main question addressed in the present work is to what extent **geology (G)**, **airborne gamma-ray**  
130 **spectroscopy (AGR)**, **topsoil geochemistry (TSG)** and the **stream sediment geochemistry (SSG)** are able to  
131 explain the **indoor radon** variation, and based on the results, to assess the usefulness of these proxies for  
132 geogenic radon mapping. The above parameters are tested on the radon prone area of SW England, which is  
133 one of the most tested areas for indoor radon in the world.

134

## 135 **2. MATERIALS AND METHODS**

### 136 **2.1. Geology and the KM1BS polygons**

137 The geology of SW England, as shown in Figure 1, mainly consists of a range of metasediments deposited in  
138 palaeo sedimentary basins during the Devonian and the Carboniferous, to which followed the granite  
139 intrusion of the prominent Cornubian batholith, emplaced in the later stages of the Variscan orogeny, Late  
140 Carboniferous to Early Permian (see Kirkwood et al., 2016 and references herein). Post-Variscan sediments  
141 (Permian and Triassic to Neogene, Figure 1) are also preserved namely in small extensions near the eastern  
142 limit of the study area. The Cornubian batholith, which is exposed onshore in five major granite outcrops  
143 (labelled in *italic* Figure 1) and other minor ones and extends offshore south-westwards, is U-enriched and  
144 one of the most radon-productive formations in the UK (Scheib et al., 2013). The intrusion of the granite  
145 caused dissemination of several elements into the intruded Devonian (and Devonian-Carboniferous)  
146 sedimentary systems, giving rise to local enrichments in the form of polymetallic and other mineralisations,  
147 including in Uranium (see mine locations in Figure 1).

148

149 FIGURE 1 HERE

150

151 Most of the area is not covered by superficial units (Figure 2, the geological units in greyscale), being  
152 alluvium and costal clays, silts and sands the most represented ones by far. Also, peat (1.5 % of the study  
153 area, the palest colour of the grayscale in Figure 2) covers some moorlands in upland areas, namely over the



154 *Dartmoor* granite and, in a lesser extent, over Devonian mudstones and sandstones in the northernmost of  
155 the study area and the *Bodmin* granite.

156 A simplified geological classification based on **BS** codes, derived from the BGS 1:50 000 scale geology  
157 (DiGMapGB-50k), as previously discussed, was developed by the BGS for radon mapping purposes in the UK.  
158 For the SW England study area (8,841 km<sup>2</sup>) a total of 134 **BS** codes were derived from combinations of 42  
159 simplified bedrock (BED) and 8 simplified superficial (SUP) codes (Figure 2).

160 From the interception of the **BS** polygons with the 1 km square grid framework, a total of 27,633 **KM1BS**  
161 polygons were derived. The **KM1BS** polygons are the basic mapping units to which the other datasets are  
162 allocated. Joining the datasets by the **KM1BS** polygons, ensures that both the spatial location and the  
163 geology (**BS**) are taken into account.

164

165 FIGURE 2 HERE

166

## 167 **2.2. Indoor radon measurements**

168 In addition to geological factors, indoor radon could be affected by changes in atmospheric factors, such as  
169 outdoor temperature, pressure and wind direction. To account for this variation, radon tests over a  
170 continuous three month period are offered to householders by Public Health England (PHE), with the data  
171 being stored in a database, and the associated house locations being accurately recorded (Miles and  
172 Appleton, 2005).

173 In the present work the available 197,464 indoor radon measurements (annual radon concentrations  
174 calculated as average of "living room" and "bedroom " results applying occupancy factors, and seasonally  
175 corrected) carried out by the Public Health England are distributed across 9,144 **KM1BS** of the total of  
176 27,633 **KM1BS** polygons, about 22 radon measurements per km<sup>2</sup> (Figure 3). This is one of the densest  
177 collections of indoor radon measurements in the world although it inevitably remains unevenly distributed,  
178 with the high number of measurements in densely populated areas, such as Plymouth, contrasting with the  
179 areas with few or no houses, such as the Dartmoor uplands (Figure 3). The spatial distribution of indoor

180 radon concentrations show high values over Carboniferous – Permian granites and, in a lesser extent, over  
 181 Devonian and Devonian – Carboniferous metasedimentary rocks, while the lower values are mainly related  
 182 to Namurian sandstones and mudstones, Permian sandstones and Meso - Cenozoic units (Figure 4). The  
 183 values shown in Figure 4 are the geometric mean (GM) of indoor radon measurements in each KM1BS  
 184 polygon.

185 In the present study, as indoor radon data is allocated by **KM1BS**, the centroid coordinates of these polygons  
 186 are attributed to the indoor radon data instead of the original house coordinates. For modelling purposes,  
 187 the indoor radon concentrations are log-transformed (**lnRn**), as they follow a lognormal distribution  
 188 (histogram of Figure 4) as expected (Miles, 1998). The histogram of Figure 4 also shows that the shape and  
 189 the central values of the distribution of the indoor radon measurements are fairly similar before (histogram  
 190 bars of Figure 4, 197464 indoor radon measurements) and after averaging the indoor radon measurements  
 191 in each KM1BS (histogram curve of Figure 4, 9144 KM1BS GM indoor radon data polygons). This averaging  
 192 procedure, however, leads to a slight decreasing in the variability of the data (e.g., compare IQR,  
 193 interquartile range, in Table 1).

194

195 Table 1 – Summary statistics of indoor radon concentrations for the available 197,464 measurements (data  
 196 points) and for the geometric mean of indoor radon measurements in each KM1BS polygon (data polygons,  
 197 9,144 values, data (2) in table 5)

Data	N	Min	Q25	Mdn	Q75	Q95	Q99	Max	IQR	AM	GM	GSD
Data points	197464	1.1	37.0	77.5	162.4	523.2	1224	24343	125.4	154.3	<b>78.6</b>	<b>3.1</b>
Data polygons	9144	1.1	42.5	75.9	139.8	368.7	735.1	5014	97.2	121.8	<b>78.4</b>	<b>2.5</b>

198 **N**: number of samples; **Min**: minimum value; **Max**: maximum value; **Q##**: ##th quantile; **Mdn**: median; **IQR**: interquartile range; **AM**:  
 199 arithmetic mean; **GM**: geometric mean; **GSD**: geometric standard deviation. Values are in Bq.m<sup>-3</sup>, except for **N** and **GSD**  
 200 (dimensionless). The most used indoor radon statistics (GM and GSD) are in bold.

201

202 FIGURE 3 HERE

203 FIGURE 4 HERE

204

### 205 **2.3. Airborne Gamma-Ray**

206 Gamma radiation measured by airborne surveys relates to the surface layer, often < 50 cm, possibly deeper  
207 in less dense unconsolidated material such as some soils, to a maximum of a few metres in dry peat. Such  
208 data provide potassium abundance (%) directly, as gamma-rays are emitted from  $^{40}\text{K}$  decay to argon, while U  
209 and Th concentrations ( $\text{mg.kg}^{-1}$ ) are estimated from the gamma-ray emission peaks associated with  $^{214}\text{Bi}$  and  
210  $^{208}\text{Tl}$  respectively, and are expressed as equivalent uranium (eU) and equivalent thorium (eTh). Total counts  
211 (in counts per second, cps), across a wider energy range, is also frequently reported.

212 The airborne gamma-ray spectrometry data (**AGR**) available for the present study were acquired under the  
213 BGS Tellus South West airborne geophysical survey (<http://www.tellusgb.ac.uk/home.html>), carried out  
214 during the second half of 2013. The survey comprised a high resolution magnetic/magnetic gradient survey  
215 combined with a multichannel (256 channel) radiometric survey. The survey was carried out using 200 m  
216 (north–south) line separations at a mean elevation of 91 m. The survey provided 60,323 line-kilometres of  
217 data (Beamish et al., 2014). Calibration and processing of the airborne data was done following standard  
218 procedures described in IAEA (IAEA, 2003) and AGSO (Minty et al., 1997) publications.

219 The 684,384 **AGR** measurements covering the study area provide a sampling density of about 77 per  $\text{km}^2$ .  
220 The spatial distribution of equivalent uranium (eU, in  $\text{mg.kg}^{-1}$ ) from the **AGR** dataset shows the highest  
221 values over Carboniferous – Permian granites, except when these are covered by peat (Figure 5). The eU  
222 spatial distribution at KM1BS resolution, that is, after averaging the **AGR** data in each KM1BS (Figure 6,  
223 23,573 data polygons), shows the same patterns as Figure 5, although necessarily with less detail. From the  
224 27,633 **KM1BS** basic mapping unit polygons 23,573 include **AGR** data. The number of **AGR** data points in  
225 each KM1BS (ranging from 1 to 145) is mainly dependent on the size of the polygon, as the distribution of  
226 **AGR** points is quite regular. Only **KM1BS** polygons not fully covered by the survey (e.g., near the mouth of  
227 River Tamar, close to Plymouth), or reporting a very small area (thus behaving like a sliver), have no airborne  
228 data inside.

229 The statistics for all data points ( $N = 684,384$ ) and after allocating data to the KM1BS polygons ( $N = 23,573$ )  
230 are quite similar (Table 2, e.g., compare median and GM values for each of the four variables). This indicates

231 that the distribution after the averaging procedure keep the same structure of the original data, despite of a  
 232 very small decreasing in the variability of the data (Table 2, e.g., compare IQR and GSD).

233

234 FIGURE 5 HERE

235 FIGURE 6 HERE

236

237 Table 2 – Summary statistics for total counts (TC), potassium (K), equivalent uranium (eU) and  
 238 equivalent thorium (eTh) from the available airborne gamma ray (**AGR**) data points (N = 684,384) and  
 239 after averaging **AGR** data in each KM1BS polygon (data polygons, N = 23,573 values).

	Data points (N = 684,384, Figure 5)				Data polygons (N = 23,573, Figure 6)			
	TC (cps)	K (%)	eU (mg.kg <sup>-1</sup> )	eTh (mg.kg <sup>-1</sup> )	TC (cps)	K (%)	eU (mg.kg <sup>-1</sup> )	eTh (mg.kg <sup>-1</sup> )
<b>N%&gt;0</b>	99.93	99.75	99.83	99.92	99.96	99.89	99.84	99.93
<b>Min</b>	-471	-1.43	-7.10	-3.47	-150	-0.22	-2.03	-0.74
<b>Q25</b>	1215	0.99	1.41	7.20	1208	1.01	1.47	7.11
<b>Mdn</b>	<b>1428</b>	<b>1.28</b>	<b>1.73</b>	<b>8.45</b>	<b>1419</b>	<b>1.28</b>	<b>1.73</b>	<b>8.31</b>
<b>Q75</b>	1684	1.63	2.11	9.73	1667	1.61	2.08	9.47
<b>Q95</b>	2203	2.14	3.46	12.01	2170	2.05	3.43	11.42
<b>Q99</b>	2870	2.57	4.94	14.86	2724	2.42	4.68	14.12
<b>Max</b>	20157	4.79	84.20	25.76	4550	4.24	12.82	18.72
<b>IQR</b>	469	0.64	0.70	2.53	459	0.60	0.61	2.36
<b>AM</b>	1461	1.32	1.87	8.41	1444	1.31	1.88	8.18
<b>GM*</b>	1386	1.23	1.72	7.98	1370	1.22	1.74	7.77
<b>GSD*</b>	1.45	1.55	1.54	1.46	1.44	1.52	1.51	1.46

240 **N%>0**: number of samples (in %) with positive values; **Min**: minimum value; **Max**: maximum value; **Q##**: ##th quantile;  
 241 **Mdn**: median; **IQR**: interquartile range; **AM**: arithmetic mean; **GM\***: geometric mean (positive values only); **GSD\***:  
 242 geometric standard deviation (positive values only). Units (in brackets) are valid for all parameters except **N%>0** and **GSD**  
 243 (dimensionless).

244

#### 245 2.4. Topsoil Geochemistry and Stream Sediment Geochemistry

246 Top soil (TS) and stream sediments (SS) both derive mainly from the bedrock beneath. Soil is a complex  
 247 body developed in the bedrock-atmosphere interface, mainly from weathering of the bedrock, to which is  
 248 added organic matter, water and air due to interaction with living organisms (often considered a component

249 of the soil too) and with the atmosphere. The active stream sediment is mainly composed by downstream  
250 transportation of material comprising of rock, soil and random organic debris. Stream sediment chemistry  
251 reflects the original bedrock chemistry from which it is derived, however the sediment is sorted by gravity  
252 and also by the energy involved in the downstream dispersion of the parent materials. Soil chemistry should  
253 also reflect the original bedrock chemistry, but is modified by bio-physical-chemical processes that are  
254 behind its development. Furthermore, soil and close to source stream sediment can be seen, approximately,  
255 as two complementary products of the same bedrock; with soil retaining mostly the immobile parts of the  
256 elements and minerals whilst the sediment retains (at least in part) the elements and minerals mobilized  
257 from the soil. In practice, the content of an element in the soil and its ratio with other elements may be  
258 quite different from that in a stream sediment, as well as in the original bedrock, depending on the hosting  
259 minerals and the element's mobility at different environmental conditions (pH/Eh).

260 Geochemistry of topsoils (**TSG**) and/or of stream sediments (**SSG**) are among the most commonly available  
261 geoscience dataset in many countries of the world. Unfortunately, radium, the direct precursor of radon  
262 gas, is not part of the routinely analysed list of elements due to analytical difficulties. On the contrary, U, Th  
263 and K, the three main sources of natural geogenic radiation, as well as an additional 30 to 50 elements, allow  
264 a very comprehensive understanding of the analysed sample.

265 Data from two systematic geochemical surveys, the UK-wide G-BASE (Geochemical Baseline Survey of the  
266 Environment, <http://www.bgs.ac.uk/gbase/home.html>) project and the England and Wales NSI (National  
267 Soils Inventory, <http://www.landis.org.uk/data/nsi.cfm>), are available for the present work. The 3382  
268 stream sediment samples, shown in Figure 7, were collected under the G-BASE project at a sampling density  
269 of approximately one site per 2.5 km<sup>2</sup>. From the available 987 soil samples, shown in Figure 8, 644 were  
270 collected during the 2013 season of the G-BASE project, at a sampling density ranging from about one  
271 sample per 10 km<sup>2</sup> to one sample per 30 km<sup>2</sup>, depending on the nature and variability of underlying  
272 geological units. The remaining 343 soil samples were collected under the NSI at a regular sampling density  
273 of 1 sample per 25 km<sup>2</sup>. The overall soil sample density is about 1 sample per 9 km<sup>2</sup>. All the processes from  
274 sampling to analysis, as well as the data, were subjected to quality control to ensure countrywide

275 consistency and continuity with other existing G-BASE and NSI geochemical data. XRF data for the < 150 µm  
276 size fraction of the stream sediment samples and for the < 2 mm size fraction of the soil samples are  
277 available for the present work for 41 elements, including 10 major (Al, Ca, Fe, K, Mg, Mn, Na, P, Si and Ti)  
278 and 31 trace (As, Ba, Br, Ce, Co, Cr, Cs, Cu, Ga, Ge, Hf, La, Mo, Nb, Nd, Ni, Pb, Rb, Sb, Sc, Se, Sm, Sn, Sr, Th, U,  
279 V, W, Y, Zn and Zr).

280 The most visible feature of the spatial distribution of uranium concentrations (U, in mg.kg<sup>-1</sup>) from both  
281 stream sediments (Figure 7) and topsoils (Figure 8) is that high values are generally overlying the  
282 Carboniferous – Permian granites.

283 After the 3,382 stream sediment data points were allocated to the underlying KM1BS, 3,027 data polygons  
284 was obtained. The number of stream sediment data points in each of these 3,027 KM1BS is only one in most  
285 cases (2,699) and range from two to four in the remaining 328 KM1BS polygons. The 3,027 stream sediment  
286 data polygons (i.e., after averaging SS data in each KM1BS polygon) show very similar statistics to the 3,382  
287 data points as can be seen for K, U and Th in Table 3. An extra table including all elements is shown in  
288 supplementary material (Extra Table A).

289 A number of data polygons (n=947) was obtained after allocating the 987 top soil data points to the  
290 underlying KM1BS. The number of top soil data points in each of these 947 KM1BS is only one in most cases  
291 (907) and two in the remaining 40 KM1BS polygons. The 947 top soil data polygons (i.e., after averaging TS  
292 data in each KM1BS polygon) show very similar statistics to the 987 data points as can be seen for K, U and  
293 Th in Table 4. An extra table including all elements is shown in supplementary material (Extra Table B).

294

295 FIGURE 7 HERE

296

297 Table 3 – Summary statistics for potassium (K), uranium (U) and thorium (Th) from stream  
298 sediment geochemistry of the SW England (N= 3,382 data points) and after taking the  
299 average of the data points in each KM1BS polygon (N = 3,027 data polygons).

Data points (N = 3,382, Figure 7)	Data polygons (N = 3,027)
-----------------------------------	---------------------------

	K <sub>2</sub> O (%)	U (mg.kg <sup>-1</sup> )	Th (mg.kg <sup>-1</sup> )	K <sub>2</sub> O (%)	U (mg.kg <sup>-1</sup> )	Th (mg.kg <sup>-1</sup> )
<b>Min</b>	0.53	1.00	2.60	0.53	1.00	2.60
<b>Q25</b>	2.33	2.50	9.80	2.33	2.50	9.88
<b>Mdn</b>	<b>2.75</b>	<b>2.90</b>	<b>11.30</b>	<b>2.75</b>	<b>2.90</b>	<b>11.30</b>
<b>Q75</b>	3.18	3.50	12.90	3.19	3.50	13.00
<b>Q95</b>	3.92	24.48	43.38	3.91	23.91	44.20
<b>Q99</b>	4.53	44.44	77.24	4.48	43.87	76.41
<b>Max</b>	6.78	206.9	321.7	6.17	206.9	321.7
<b>IQR</b>	0.85	1.00	3.10	0.86	1.00	3.13
<b>AM</b>	2.78	5.70	15.01	2.78	5.70	15.07
<b>GM</b>	2.69	3.64	12.53	2.70	3.66	12.58
<b>GSD</b>	1.29	2.10	1.64	1.28	2.11	1.64

300 **N**: number of samples; **Min**: minimum value; **Max**: maximum value; **Q##**: ##th quantile; **Mdn**: median; **IQR**:  
301 interquartile range; **AM**: arithmetic mean; **GM**: geometric mean; **GSD**: geometric standard deviation. Units (in  
302 brackets) are valid for all parameters except **GSD** (dimensionless).

303

304 FIGURE 8 HERE

305

306 Table 4 – Summary statistics for potassium (K), uranium (U) and thorium (Th) from top soil  
307 geochemistry of the SW England (N= 987 data points) and after taking the average of the  
308 data points in each KM1BS polygon (N = 947 data polygons).

	Data points (N = 987, Figure 8)			Data polygons (N = 947)		
	K <sub>2</sub> O (%)	U (mg.kg <sup>-1</sup> )	Th (mg.kg <sup>-1</sup> )	K <sub>2</sub> O (%)	U (mg.kg <sup>-1</sup> )	Th (mg.kg <sup>-1</sup> )
<b>Min</b>	0.01	0.70	0.30	0.01	0.66	0.30
<b>Q25</b>	1.96	2.50	8.80	1.95	2.50	8.78
<b>Mdn</b>	<b>2.47</b>	<b>2.90</b>	<b>10.30</b>	<b>2.47</b>	<b>2.90</b>	<b>10.25</b>
<b>Q75</b>	2.96	3.40	11.60	2.96	3.43	11.60
<b>Q95</b>	3.43	6.10	13.40	3.43	5.87	13.40
<b>Q99</b>	3.96	11.71	17.21	3.92	11.70	17.25
<b>Max</b>	4.94	45.70	22.20	4.94	35.35	22.20
<b>IQR</b>	1.01	0.90	2.80	1.01	0.93	2.82
<b>AM</b>	2.41	3.34	9.99	2.41	3.31	9.99
<b>GM</b>	2.20	3.01	9.43	2.19	3.00	9.42
<b>GSD</b>	1.84	1.49	1.51	1.86	1.48	1.51

309 **N**: number of samples; **Min**: minimum value; **Max**: maximum value; **Q##**: ##th quantile; **Mdn**: median; **IQR**:  
310 interquartile range; **AM**: arithmetic mean; **GM**: geometric mean; **GSD**: geometric standard deviation. Units (in  
311 brackets) are valid for all parameters except **GSD** (dimensionless).

312

## 313 **2.5. Integration of Datasets and models produced**

314 As explained in the previous subsections, the point based datasets (**Indoor Radon, AGR, TSG, SSG**) were  
315 allocated to the **KM1BS** polygons, through joining by spatial location (in ArcGIS). **KM1BSs** are polygons  
316 corresponding to the interception of a 1km<sup>2</sup> grid with the geology (**BS**) polygons. If two or more data points  
317 of a certain dataset are within the same **KM1BS** polygon, the resulting record will correspond to the *average*  
318 value. After grouping the available 197,464 log<sub>e</sub>-transformed indoor radon measurements (data points)  
319 according to the **KM1BS**, a dataset of 9,144 records (data polygons) was generated. By the same procedure,  
320 the available 684,384 airborne data points were grouped into 23,573 data polygons; the available 3382  
321 stream sediment data points were grouped into 3027 data polygons and the 987 top soil data points were  
322 grouped into 947 data polygons.

323

324 Finally the information in these new files was integrated in *ACCESS*, according to the data requirements and  
325 availability for each specific model. Table 5 reports the final data files used in the present work and the  
326 respective number of records. Indoor radon modelled by **AGR** can be carried out by using data (3), (4), (5) or  
327 (6), as shown in Table 5, however with a decreasing number of available records. The models including **SSG**  
328 can be performed either with 1182 (data (4)) or with 83 (data (6)) records; while models with **TSG** are based  
329 either on 600 (data (5)) or on 83 (data (6)) records (Table 5). Indoor radon models exclusively with geology  
330 can be carried out by any of the data ((1) to (6)) referred to in Table 5, however the value obtained with data  
331 (1) cannot be directly compared with those from the remaining data ((2) to (6)). While these last datasets  
332 only account with the variance *between* **KM1BS** polygons, models with data (1) will account with both the  
333 variance *within* and *between* **KM1BS** polygons.

334 The spatial join by **KM1BS** ensures that data is joined over the same geology, contrary to a simple join based  
335 exclusively on distance. It should also be noted that **KM1BS** polygons, which may be sized up to a maximum  
336 of 1 km<sup>2</sup>, constitute the maximum detail shown in UK radon maps, thus, most likely providing enough detail  
337 also for the European radon mapping. Further explanation about the variances can be found under section  
338 *2.6.2. Variance Explained.*



339

340

Table 5: number of records used for **InRn** modelling according to data required.

Data file used	Number of records
(1) Rn	197464
(2) Rn (averaged by KM1BS)	9144
(3) Rn + AGR	8905
(4) Rn + AGR + SSG	1182
(5) Rn + AGR + TSG	600
(6) Rn + AGR + SSG + TSG	83

341

Rn:  $\log_e$  of indoor radon measurements; **AGR**: airborne gamma-ray data; **SSG**: Stream sediment

342

geochemistry data; **TSG**: Topsoil geochemistry data.

343

344

The resulting **InRn** ( $\log_e$  of indoor radon) models produced for the present work are summarised in section 3

345

(Results).

346

Models using only the geology (**G**) as predictor variable were the first set to be run. As these can be

347

produced using all the data files (i.e., (1) to (6) of Table 5), and using 3 types of geological classification (**GPE**:

348

simplified geological period classes, **BED**: simplified bedrock classes and **BS**: simplified bedrock/superficial

349

geological combinations), a total of 18 models have been produced (Table 6). Running all these models

350

allows comparison of those obtained with the three different geological classification types (i.e., values along

351

the same line in Table 6) as well as checking for consistency of results as the number of records decreases

352

from data file (2) to data file (6) of table 5 (i.e., along the same column in Table 6).

353

Models using airborne gamma ray (**AGR**) data as predictor follow those with geology only, as the spatial

354

cover of **AGR** data is much more complete than that of **SSG** and **TSG**. A total of  $4 \times 6 = 24$  models have been

355

produced (Table 7), from four data files (from (3) to (6) of Table 5) and using six different predictor

356

combinations, namely total counts (**AGR[TC]**); eU concentrations (**AGR[U]**); eU, eTh and K concentrations

357

(**[AGR]**); eU, eTh, and K concentrations and total counts (**[AGR] + AGR[TC]**); eU concentration and geology

358

(**AGR[U]+G**, with **BS** classes) and finally eU, eTh and K concentrations and geology (**[AGR] + G**, with **BS**

359

classes). This set of models provides an understanding of how good the **AGR** data is (i.e., eU, eTh, K) in

360

predicting indoor radon, weighting and comparing uranium relative to the whole **AGR**, concluding about the

361 usefulness of total counts in radon prediction and measuring the improvement of models as a result of  
362 adding **AGR** to **G** in the set of predictor variables.

363 Finally models including stream sediment geochemistry (**SSG**) or top soil geochemistry (**TSG**) as predictor  
364 variables are shown in Table 8 and Table 9 respectively. Data files (4) and (5) of Table 5 are used for models  
365 with **SSG** and **TSG** respectively. A total of  $3 \times 4 = 12$  models including **SSG** data (Table 8) or **TSG** data (Table 9)  
366 have been produced. Three possible sets of chemical elements are used, namely uranium concentrations  
367 only (**SSG[U]**, **TSG[U]**); uranium, thorium and potassium concentrations (**SSG[UThK]**, **TSG[UThK]**) and the  
368 entire set of the 41 available elements (**[SSG]**, **[TSG]**). Four different groups of variables are used, namely  
369 the geochemistry only (first column of results in Table 8 and Table 9); geochemistry and geology (second  
370 column of results in Table 8 and Table 9); geochemistry and airborne gamma ray (third column of results in  
371 Table 8 and Table 9); and geochemistry, airborne gamma ray and geology (last column of Table 8 and Table  
372 9). These two sets of models allow comparison of the performance of the geochemistry of the two media in  
373 predicting radon, weighting and comparing the two subsets of elements (uranium and U, Th, K) with each  
374 other and relative to the whole composition, and measuring the improvement of models as a result of  
375 adding the geochemistry to **AGR** and **G** in the set of predictor variables. Models including both **SSG** and **TSG**  
376 (data (6) of Table 5) were not considered after less consistent results obtained from the same data file (6) for  
377 models with **AGR** and **G** (Table 7).

378 Before going into detail on the results a brief introduction to compositional data, compositional regression  
379 and the tools used are presented.

380

## 381 **2.6. Compositional data**

382 Geochemical datasets, including the **TSG** and **SSG** used in the present work, are typical examples of  
383 *compositional (closed) data*, *CODA*, (Aitchison, 1986), as the sum of all elements (parts) in a sample  
384 necessarily sum up to a constant (often 1, 100 % or  $1\,000\,000\text{ mg.kg}^{-1}$ ). This implies that the concentration  
385 of one element does not vary independently from the others, and thus that the interpretation and statistical  
386 evaluation of the observed element concentrations is only meaningful if the relationship to the

387 concentration of the remaining elements is taken into account (Aitchison, 1986; Filzmoser et al., 2010;  
388 Reimann et al., 2012).

389 Aitchison (1986) developed a geometry on the simplex on which compositional data fully adjust.  
390 Compositional data must be transformed into the correct geometry first, after which classical methods can  
391 be fully applied. In CODA context the statistics are performed after opening the data through one of three  
392 possible log-ratio (logarithm of a ratio) transformations. The basic idea behind these transformations is that  
393 the meaningful information for an element is not in the reported concentration, but only achievable if the  
394 other elements are taken into account. The *alr* - additive log-ratio transformation, proposed by Aitchison  
395 (1986), uses one of the elements as divisor, against which each one of the other elements is presented after  
396 taking the logarithm; when three elements ( $D=3$ ,  $x_1$ ,  $x_2$ ,  $x_3$ ) are reported in a compositional dataset  $x$ , the  
397 respective *alr* will be composed by 2 log-ratios (coordinates):  $alr(x) = [\ln(x_1/x_3); \ln(x_2/x_3)]$ . The *clr* – centred  
398 log-ratio transformation, proposed by Aitchison (1986), use the geometric mean ( $g(x)$ ) of all measured  
399 elements as the divisor against which each element is “normalized”; when three elements ( $D=3$ ,  $x_1$ ,  $x_2$ ,  $x_3$ )  
400 are reported in a compositional dataset  $x$ , the respective *clr* will contain 3 log-ratios (coordinates):  $clr(x) =$   
401  $[\ln(x_1/g(x)); \ln(x_2/g(x)); \ln(x_3/g(x))]$ , where  $g(x)$  is the geometric mean of  $x_1$ ,  $x_2$ ,  $x_3$ . This is perhaps the most  
402 popular log-ratio transformation as there is a direct relation to the original element. However working with  
403 this transformation requires some care as the covariance and correlation matrices are singular because the  
404 *clr* coefficients necessarily sum to zero (Buccianti, et al., 2006). The *ilr* – isometric log-ratio transformation,  
405 proposed by Egozcue et al. (2003), have the best geometrical properties for multivariate data analysis  
406 (Buccianti, et al., 2006) and linear regression (van den Boogaart & Tolosana-Delgado, 2013), but the results  
407 are difficult to interpret as there is no direct relation with the original elements (parts). The *ilr*  
408 transformation is given by  $ilr(x) := clr(x) \cdot V$ , where  $V$  is a matrix with  $D$  rows and  $D-1$  columns (with  $D$  the  
409 number of original parts) which columns form an orthonormal basis of the *clr*-plane. Thus *ilr* relates with the  
410 *clr* transformation, allowing to reconstruct the names of the original parts through an inverse  
411 transformation; when three elements ( $D=3$ ,  $x_1$ ,  $x_2$ ,  $x_3$ ) are reported in a compositional dataset  $x$ , the  
412 respective *ilr* will contain 2 log-ratios (coordinates):  $ilr(x) = [\sqrt{1/2} \cdot \ln(x_1/x_2); \sqrt{1/6} \cdot \ln(x_1 \cdot x_2 / x_3 \cdot x_3)]$

413 (Pawlowsky-Glahn & Egozcue, 2006). Further explanation on how to analyse compositional data can be  
414 found in e.g. Buccianti et al. (2006), Pawlowsky-Glahn & Buccianti (2011), Tolosana-Delgado (2012), van den  
415 Boogaart & Tolosana-Delgado (2013), among other publications. Further discussion and practical  
416 applications of CODA principles to a European geochemical mapping project (GEMAS) can be found in  
417 Reimann et al. (2014) and Reimann et al. (2012). The latest discussions about compositional data mapping  
418 are found in McKinley et al. (2015), Buccianti et al. (2015), Pawlowsky-Glahn and Egozcue (2015).

419

### 420 **2.6.1. Compositional Regression**

421 The indoor radon modelling using the four datasets described above was carried out in R (**compositions**  
422 package), and using a compositional approach. The main reason is rooted in the compositional nature of  
423 some of the datasets, namely the stream sediment geochemistry, soil geochemistry, and eventually the  
424 airborne gamma-ray dataset.

425 The **compositions** is an R package, developed by Gerald van den Boogaart, Raimon Tolosana and Matevz  
426 Bren (<https://cran.r-project.org/web/packages/compositions/compositions.pdf>), providing functions for the  
427 consistent analysis of compositional data. Both R (programming language) and **compositions** are available as  
428 free software. R (Rx64 3.1.0) was downloaded from <http://www.cran.r-project.org>, while **compositions**  
429 package (version 1.40) was downloaded and installed using an R friendly interface, the RStudio (version  
430 0.98.1102 of the open source edition), which is available in <https://www.rstudio.com/products/rstudio/>.  
431 More information about **compositions** package can be found in K.G. van den Boogaart and R. Tolosana-  
432 Delgado (2013). Other open source packages, such as the CoDaPack (<http://ima.udg.edu/codapack/>), are  
433 available for Compositional Data Analysis.

434 According to van den Boogaart & Tolosana-Delgado (2013), compositional regression is analogous to  
435 multiple linear regression, although more complex in the details. For the present case, the *dependent*  
436 variable (**Y**) is the indoor radon ( $\text{Bq}\cdot\text{m}^{-3}$ ) transformed by the natural logarithm ( $\mathbf{Y} = \log_n(\text{indoor radon})$ ) or  
437 simply  $\mathbf{Y} = \ln Rn$ ) as the available indoor radon data follows a lognormal distribution as expected (Miles,

438 1998). Each one of the datasets **G**, **AGR**, **SSG**, **TSG** will be part of the *independent* variables (**X**), with **G** as a  
439 *categorical (factor)* variable, and **AGR**, **SSG**, **TSG** as *continuous compositional* variables.

440 According to van den Boogaart & Tolosana-Delgado (2013), the *ilr* isometric logratio transformation is the  
441 only one that completely fulfils the requirements in linear modelling involving compositions, namely when  
442 metric variances are required. This is true for the present work, as **SSG** and **TSG** (and eventually **AGR**) are  
443 compositions and the  $R^2$ , a measure of determination will be used to compare the different models.

444

445 Following procedures found in van den Boogaart & Tolosana-Delgado (2013), the indoor radon prediction  
446 was performed using ANOVA and a compositional regression model of the general form  $Y_i = \mathbf{a} +$   
447  $(\mathbf{b}, X_i)_A + \varepsilon_i$  (**Equation 1**), where **Y** is a real random variable, **X** is an independent compositional variable,  
448 and the parameter **a**, **b** and  $\varepsilon$  are, respectively, a real-valued intercept, a regression “slope” composition  
449 (belonging to a D-part *Simplex*), and a real-valued zero-mean error (typically modelled as normally  
450 distributed).

451 According to van den Boogaart & Tolosana-Delgado (2013) the estimation of the regression parameters will  
452 be a multiple linear regression problem of the type  $Y_i = \mathbf{a} + \sum_{k=1}^{D-1} \beta_k \text{ilr}_k(X_i) + \varepsilon_i$  (**Equation 2**), with  $\beta =$   
453  $\beta_k$  a vector of slope parameters.

454 The multiple linear regression are carried out using a simple script line under the *compositions* R package,  
455 **(model = lm (Y ~ ilr(X<sub>A</sub>)))** (**Script line 1**), where **X<sub>A</sub>** is a compositional dataset (van den Boogaart & Tolosana-  
456 Delgado, 2013). This can be applied to the entire composition (e.g., a dataset including contents of “all”  
457 elements in soil samples) as well as to a sub-composition (e.g., a subset including the U, Th, K contents in soil  
458 samples). If two compositional datasets (**X<sub>A</sub>**, **X<sub>B</sub>**) are used together with the geology categories (**G**), the script  
459 line will be **(model = lm (Y ~ ilr(X<sub>A</sub>) + ilr(X<sub>B</sub>) + G))** (**Script line 2**).

460 Models including 3 or more elements of the same dataset (i.e., **[AGR]**, **SSG[UThK]**, **[SSG]**, **TSG[UThK]** and  
461 **[TSG]**) were processed including a residual (**R**) in addition to the respective elements, and using a  
462 compositional (CODA) approach, according to the following procedures. First, all elements were converted  
463 into mg.kg<sup>-1</sup>. Secondly, the data was closed (using the *acomp()* function in *compositions* R package) taking

464 into account the residual part (**R**) with respect to the total (1,000,000 mg.kg<sup>-1</sup>) computed as **R = 1,000,000 –**  
465 **(sum of elements)** (Equation 3), where **sum of elements** is K + eU + eTh for **[AGR]**, K + U + Th for both  
466 **SSG[UThK]** and **TSG[UThK]**, and the sum of the 41 elements available (see list in section 2.4.) for both **[SSG]**  
467 and **[TSG]**, with concentrations in mg.kg<sup>-1</sup> for all mentioned cases. Thus, in short, each one of the models  
468 **[AGR]**, **SSG[UThK]** and **TSG[UThK]** is based in a four-parts composition (D = 4, i.e., 3 elements and the  
469 respective residual **R**), while each one of the models **[SSG]** and **[TSG]** is based in a 42-parts composition (D =  
470 42, i.e., 41 elements and the respective residual **R**). Finally, the isometric logratios for each composition are  
471 computed directly, using the *ilr()* function in **compositions** R package, while running the required model (see  
472 script lines 1 or 2 above). It should be emphasized, thus, that the predictor variables for these models are a  
473 set of dimensionless **D-1 logratios** (logarithm of a ratio of 2 or more elements), the *ilr* coordinates, not a set  
474 of elements (see section 2.6. Compositional data for further explanation). Also, the significance for each of  
475 the D-1 *ilr* coordinates can be obtained from the output of the model, but not the significance for each  
476 element. However, the focus of the present work is on the significance of each model as a whole, rather  
477 than evaluating the significance of the *ilr* coordinates. Models with sub-compositions (e.g., **SSG[UThK]**), or  
478 with one element only (e.g., **SSG[U]**) are mainly used to frame the results of the whole relative to a set of  
479 elements commonly used in indoor radon modelling.

480 To note that the use of the residual R as one of the parts in a composition is not compulsory in CODA  
481 context. However it was used in the present work as it may contain valid information (Kirkwood et al.,  
482 submitted). As an example, peat soil samples from SW England always have a high residual R, which is  
483 related to the Loss on Ignition (Kirkwood et al., submitted) of the samples.

484

### 485 **2.6.2. Variance Explained**

486 Two measures of determination **R<sup>2</sup>** and **adjusted R<sup>2</sup> (adj R<sup>2</sup>)** can be used to numerically evaluate the strength  
487 of the relationship being modelled and be interpreted as an estimate of the portion of the total variability  
488 that is explained by the model. **Adjusted R<sup>2</sup>** will be preferably shown instead of **R<sup>2</sup>**, as it takes into  
489 consideration the sample size and the number of predictor variables, contrary to **R<sup>2</sup>**, which thus tend to

490 increase with increasing number of parameters. These two and other model parameters can be retrieved by  
491 calling for a model summary (script line 3: *summary(model)*) and an ANOVA table (script line 4:  
492 *anova(model)*) in **compositions** R package (van den Boogaart & Tolosana-Delgado, 2013). The ANOVA table  
493 provides an overall significance test for each dataset included in the model, thus allowing to check whether  
494 the information provided by each dataset as a whole influences or not the indoor radon. The model  
495 summary provides a significance test for each of the individual factors of the geology, and for the *individual*  
496 coordinates (that is, each one of the log-ratios) of the *ilr*-transformed composition(s) (van den Boogaart &  
497 Tolosana-Delgado, 2013). This allows assessment of the relevant *ilr* coordinates, but not the relevant  
498 original elements directly, for the indoor radon explanation.

499

500 The **total**  $\log_e$  indoor radon variability ( $V_t$ ), can only be computed from data (1) in Table 5, that is, directly  
501 using all indoor radon measurements. In models computed with data (1) the mentioned explained variation  
502 ( $V_e$ , **adj. R<sup>2</sup>** in %) is a **portion** ( $V_p$ ) of the **total**  $\log_e$  indoor radon variability  $V_t$  ( $V_e=100 \times V_p/V_t$ ). This  $V_t$  can be  
503 decomposed in two main sources of variability:  $V_t = V_b + V_w$ , where  $V_b$  is the variability **between KM1BS**  
504 polygons (or **lateral** variability) and  $V_w$  is the variability **within KM1BS** polygons (or **local** variability). The  $V_w$   
505 can only be computed from data (1) in Table 5, as well. This is because the **local** variability  $V_w$  is  
506 automatically removed after grouping and averaging data per **KM1BS**, as happen with data (2) to data (6) in  
507 Table 5. Also, the overall **lateral**  $\log_e$  indoor radon variability ( $V_b$ ) can be computed from data (1) as the  $\log_e$   
508 indoor radon variation explained by **KM1BS**, 32.3% (Table 6). Thus, for models computed from data (2) to  
509 data (6) the mentioned explained variation ( $V_e$ , **adj. R<sup>2</sup>** in %) is a **portion** ( $V_p$ ) of the overall **lateral**  $\log_e$  indoor  
510 radon variability  $V_b$  ( $V_e=100 \times V_p/V_b$ , with  $V_b=32.3\%$ ), not of the **total**  $\log_e$  indoor radon variability ( $V_t$ ). As an  
511 example, the **InRn** variation explained by **BSs** in data (1), 10.1%, is relative to  $V_t$ , while the value reported for  
512 data (2), 29%, is relative to  $V_b$  (Table 6). Note that 29% is similar to 10.1% if computed relative to 32.3%  
513 ( $29 \times 32.3/100=9.4\%$ ), being the difference between them ( $10.1\%-9.4\% = 0.7\%$ ) an effect of the highly variable  
514 number of indoor radon measurements available per **KM1BS** (from 1 to 933). The effect explained for data

515 (2) also applies to data (3), (4), (5) and (6). Thus, for models with data (2) to data (6), the explained variance  
516 ( $V_e$ ) is a portion ( $V_p$ ) of the overall *lateral*  $\log_e$  indoor radon variability  $V_b$  ( $V_e=100 \times V_p/V_b$ , with  $V_b=32.3\%$ ).

517

518 The **adjusted  $R^2$**  is the statistical parameter chosen to evaluate and compare models, and thus shown in the  
519 four tables reporting the results for the 67 models (Table 6 to Table 9). To note that all models presented  
520 report a **p-value < 0.001** except when indicated otherwise. The models shown in Figure 9 to Figure 12  
521 include information about the number of records, the significance level (p value), the Pearson R, the  $R^2$  and  
522 the adj.  $R^2$ . All these parameters have been retrieved from the model outputs after following instructions  
523 above.

524

### 525 **3. RESULTS**

#### 526 **3.1. Indoor radon explained by Geology**

527 The  $\log_e$  indoor radon (**InRn**) variance explained (adjusted  $R^2$ ) by geology after regression analysis is shown in  
528 Table 6 for models from three different geological classifications: as Geological Period (**GPE**), as bedrock  
529 (**BED**) or as bedrock-superficial (**BSs**).

530 For SW England, the proportion of  $\log_e$  transformed indoor radon (**InRn**) variation explained by geology,  
531 whether bedrock (**BED**) or bedrock-superficial (**BS**), is about 10% (Table 6 and Figure 9 A). The *superficial*  
532 units only add about 1% of explained variability to that from *bedrock* units alone ( $10.1-9.2=0.9\%$ ). If data is  
533 classified according to Geological Period (GPEs), the variation of **InRn** explained by geology is only 1.4%  
534 (Table 6). Thus, not surprisingly, the variation of indoor radon explained by geology increases with the level  
535 of detail of the geology. Also, for the present case, a classification based on geological periods is not  
536 effective for indoor radon analysis, contrary to a classification based on bedrock. SW England shows a low  
537 proportion of **InRn** variation explained by geology, in comparison to the average value found for England and  
538 Wales (24.6%, Appleton and Miles, 2010), Scotland (21%, Scheib et al. 2009) or in a lesser extent, to values  
539 found in Belgium (15.4% - 17.7%, Tondeur et al., 2014), and similar to those found in Austria (11.2%, Bossew  
540 et al., 2008), Northern Ireland (12.1%, Appleton et al., 2015). The low value is a result of 1) the short



541 geological time range, most of the area belongs to two successive periods (Devonian and Carboniferous),  
 542 thus constraining the explanation of indoor radon variability given by the geology, and 2) the uranium  
 543 mineralization (Appleton and Miles, 2010), which are spatially and genetically dependent on granitic rocks,  
 544 but scattered throughout other surrounding geological units, several Devonian and Devonian-Carboniferous  
 545 sedimentary basins. This leads to a general increasing of the **InRn** variation within geological units, thus  
 546 decreasing contrast between them.

547 The overall **InRn** variability observed **between KM1BSs** ( $V_b$ ) is only 32.3% of the *total InRn* variability ( $V_t$ , last  
 548 column of Table 6). This implies that, in SW England, the overall **InRn** variability **within KM1BSs** is about two  
 549 thirds ( $100-32.3= 67.7\%$ ) of the **total InRn** variation ( $V_t$ ), that is, most of the **InRn** variability is observed at  
 550 **local** level.

551 To note that the adj  $R^2$  values reported in Table 6 for data (2) to (6) with **BSs** as geology are quite similar,  
 552 with an average value of 29.1% and a standard deviation of 2.5%. Three from the five values are between  
 553 29% and 30% (Figure 9 B), while the other 2 are one underestimation (25.1%) and one overestimation  
 554 (32.1%, Figure 9 C) relative to the average, which can be seen as a reference value. Furthermore, the  
 555 average coincide with the value for data (2), 29.0%, which is based on the largest number of **KM1BS**  
 556 polygons.

557 Values for geological units as **BSs** (shaded values in table 6) are those being considered in the remaining of  
 558 the present work.

559

560 Table 6 – Proportion of the variation (adjusted  $R^2$  in %) of natural log transformed indoor radon (**InRn**)  
 561 explained by Geology, based on three simplified geological classifications, geological period (GPEs), bedrock  
 562 (BEDs) and bedrock-superficial (BSs) geology polygons, derived from the BGS DiGMapGB-50k.

Data file used	Number of records	Geology as GPEs		Geology as BEDs		Geology as BSs		KM1BS	
		Number of units	Adj. $R^2$	Number of units	Adj. $R^2$	Number of units	Adj. $R^2$	Number of units	Adj. $R^2$
(1) Rn	197464	6	1.4	40	9.2	90	10.1	9144	32.3

<b>(2) Rn (joined by KM1BS)</b>	9144	6	0.6	40	27.9	90	29.0	9144	100
<b>(3) Rn + AGR</b>	8905	6	0.6	37	28.5	82	29.8	-	-
<b>(4) Rn + AGR + SSG</b>	1182	4	1.7	20	24.7	41	25.1	-	-
<b>(5) Rn + AGR + TSG</b>	600	4	<del>0.4</del>	19	30.7	29	32.1	-	-
<b>(6) Rn + AGR + SSG + TSG</b>	83	3	<del>-0.2</del>	12	31.9	15	29.5	-	-

563 Records in data (2) to (6) are data polygons, that is, average values per KM1BS polygon. The significance for the adjusted  $R^2$  is  $p <$   
564 0.001, except for strikethrough values. Columns in grey highlight the geological classification type (**BS**) that will be used in the  
565 following models (Table 7 to Table 9).

566

567 FIGURE 9 HERE

568

### 569 **3.2. Indoor radon explained by AGR**

570 The  $\log_e$  indoor radon (**InRn**) variance explained (adjusted  $R^2$ ) by airborne data after regression analysis is  
571 shown in Table 7 for six models, either using  $\log_{10}(eU)$  (**AGR[U]**, Figure 10 A),  $\log_{10}$  of total counts (**AGR[TC]**),  
572 or the set of elements available, eU, eTh, K, in a compositional (CODA) approach (**[AGR]**, Figure 10 B). For  
573 the last model, the isometric logratios of the four part composition are used. Two of the models also include  
574 geology (G), one of them shown in Figure 10 C.

575 For SW England, the **InRn** variation *between* **KM1BSs** explained by airborne gamma-ray using a CODA  
576 approach (**[AGR]**) is about **29%** (Table 7), as this is the value reported from the model with data (3), which  
577 have the largest number of **KM1BSs**. The value for data (4) is slightly underestimated (-1.8%) while for data  
578 (5) is slightly overestimated (+3.5%, Figure 10 B) relative to the reference value (29%), but both acceptable.  
579 The value from data (6) is clearly overestimated (about +17%), indicating that its 83 available records are not  
580 statistically representative of the 8905 data (3) polygons which include both indoor radon and airborne  
581 information. Consequently, models including data (6) will not be shown further and /or subject to further  
582 discussion.

583 The models including **AGR[U]** (e.g., Figure 10 A), explain a larger portion (+3.1% to +6.6%) of **InRn** variation  
584 than the models with **AGR[TC]**, and a similar but slightly lower (-1.9% to -4.3%) portion than that of **[AGR]**

585 (e.g., Figure 10 B). These values can be checked by comparing the respective columns in Table 7. Also, a  
 586 maximum of 0.1% is added to the explained variance by appending **AGR[TC]** to the **[AGR]** model (Table 7).  
 587 When geology (**G**, the **BSs**) is added to the model with CODA airborne data (**[AGR]**), the explained **lnRn**  
 588 *lateral* variation increase from about 5% to 9% (Table 7, Figure 10 C).  
 589 These results suggest that the inclusion of Total Counts (**AGR[TC]**) does not enhance equivalent U (**AGR[U]**)  
 590 or CODA eU, eTh, K (**[AGR]**) models, and that the last is (slightly) preferred to that with eU only (compare  
 591 Figure 10 A and Figure 10 B). Also, the **[AGR]+G** (Figure 10 C) provide a better model than geology (Figure 9  
 592 C) or airborne gamma-ray alone (Figure 10 B), suggesting that a (small) part of the information provided by  
 593 these two datasets is complementary to each other in respect to indoor radon.  
 594 Note that the **AGR[U]** model provides a similar to slightly higher level of **lnRn** explained variation to that  
 595 obtained by Appleton et al. (2008) in an area of the Northern Ireland ( $R^2=21\%$ ) and slightly lower to that  
 596 obtained by Scheib et al. (2006) for an area in Central England ( $R^2=31.4\%$ ).  
 597 **[AGR]** and **[AGR]+G**, with **BSs** used as the geological classes (shaded values in table 7) are those considered  
 598 in the remaining of the present work.

600 Table 7 – Proportion of the variation (adjusted  $R^2$  in %) of natural log transformed indoor radon  
 601 concentrations (**lnRn**) explained by airborne gamma-ray (**AGR**), and by **AGR** combined with Geology (**G**).

Data file used	Number of records	N. of BSs	Adjusted $R^2$					
			AGR[TC]	AGR[U]	AGR[U] + G	[AGR]	[AGR] + AGR[TC]	[AGR]+ G
(3) Rn + AGR	8905	82	21.5	25.7	37.8	29.0	29.1	37.9
(4) Rn + AGR + SSG	1182	41	19.8	22.9	32.3	27.2	27.3	32.6
(5) Rn + AGR + TSG	600	29	24.0	30.5	39.0	32.5	32.4	38.8
(6) Rn + AGR + SSG + TSG	83	15	31.5	40.6	42.3	45.7	45.5	49.4

602 **[AGR]**: the isometric logratio transformation of the composition eU, eTh, K and the residual R (Equation 3); **AGR[TC]**:  $\log_{10}$  of Total  
 603 Counts; **AGR[U]**:  $\log_{10}$  of equivalent U. The significance for the adjusted  $R^2$  is  $p < 0.001$  for all models. Models from data file (6) are

604 strikethrough indicating that they are excluded from further discussion. Columns in grey highlight the predictor combinations (**[AGR]**  
605 and **[AGR]+G**) that will be used in the following models (Table 8 and Table 9).

606

607 FIGURE 10 HERE

608

### 609 **3.3. Indoor radon explained by stream sediment geochemistry**

610 Table 8 summarises the  $\log_e$  indoor radon (**InRn**) variance explained (adjusted  $R^2$ ) by stream sediment  
611 geochemistry after regression analysis with examples of 3 models shown in Figure 11. The **InRn** models have  
612 been developed using stream sediment geochemical data from data (4) (referred to in Table 5, Table 6 and  
613 Table 7), and are separated in three groups. A first group of simple models (i) for which only  $\log_{10}(U)$  was  
614 used (**SSG[U]** in Table 8, and e.g., Figure 11 A); a second group of models (ii) for which a sub composition  
615 including three elements (K, Th, U in  $\text{mg.kg}^{-1}$ ) is used together with the associated residual part (R)  
616 (**SSG[UThK]** in Table 8); and a third group of models (iii) for which the available composition of 41 elements  
617 (in  $\text{mg.kg}^{-1}$ ) is used together with the residual part (R) (**[SSG]** in Table 8, e.g., Figure 11 B). For models with  
618 (ii) or (iii), the isometric logratio of the four part sub-composition or the 42 part composition are used  
619 respectively.

620 For SW England, the **InRn** variation *between* **KM1BSs** explained by **[SSG]** is 32.5% (Table 8, Figure 11 B),  
621 which is clearly above the value given by **SSG[U]** (17.8%, Figure 11 A) or by **SSG[UThK]** (20%), and at the  
622 same level of the explanation given by **[AGR]+G** (32.6%, data (4) in Table 7). Explained variance is enhanced  
623 +4.2% by adding **[SSG]** to the model with **[AGR]+G** ( $32.6\% + 4.2\% = 36.8\%$ , Figure 11 C), and only +0.6% by  
624 adding **SSG[U]** or **SSG[UThK]** to the same **[AGR]+G**.

625 These results show that the stream sediment geochemistry as a whole and using a CODA approach (**[SSG]**)  
626 clearly enhance results provided by uranium alone or U, Th, K in stream sediments, and provide a similar  
627 level of explanation given by the **[AGR]+G** model. The results also suggest that, concerning indoor radon, a  
628 (small) part of the information provided by **[SSG]** is complementary to that from **[AGR]+G**.

629

630 Table 8 – Proportion of the variation (adjusted R<sup>2</sup> in %) of natural log transformed indoor radon  
 631 concentrations (**InRn**) explained by stream sediment geochemistry (**SSG**), or **SSG** combined with **AGR**  
 632 and/or Geology.

SSG elements used	Adjusted R <sup>2</sup>			
	SSG[U]	SSG[U]+G	SSG[U]+[AGR]	SSG[U]+[AGR]+G
(i) Log <sub>10</sub> (U)	17.8	29.1	28.3	33.2
(ii) U, Th, K (CODA)	SSG[UThK]	SSG[UThK]+G	SSG[UThK]+[AGR]	SSG[UThK]+[AGR]+G
	20.0	29.0	28.2	33.2
(iii) ALL 41 SSG elements (CODA)	[SSG]	[SSG]+G	[SSG]+AGR	[SSG]+[AGR]+G
	32.5	34.7	34.7	36.8

633 **SSG[U]**: log<sub>10</sub> of uranium concentrations in stream sediments; **SSG[UThK]**: the isometric logratio of the sub composition U, Th, K  
 634 (including the residual R (Equation 3)) concentrations in stream sediments geochemistry; **[SSG]**: the isometric logratios of the  
 635 available composition with 10 major+31 trace elements (including the residual R (Equation 3)) concentrations in stream  
 636 sediments geochemistry. Data used for all models: **1182** records of data (**4**) in Table 5, for which 41 **BS** classes are available;  
 637 **[AGR]**: airborne gamma-ray; **G**: geology (**BS** classes). The significance for the adjusted R<sup>2</sup> is p < 0.001.

638

639 FIGURE 11 HERE

640

### 641 3.4. Indoor radon explained by topsoil geochemistry

642 Table 9 summarises the log<sub>e</sub> indoor radon (**InRn**) variance explained (adjusted R<sup>2</sup>) by topsoil geochemistry  
 643 after regression analysis with 3 models shown in Figure 12. The **InRn** models have been developed using  
 644 topsoil geochemical data from data (5) (referred in Table 5, Table 6 and Table 7), and are separated in three  
 645 groups. A first group of simple models (i) for which only log<sub>10</sub>(U) was used (**TSG[U]** in Table 9, e.g., Figure 12  
 646 A); a second group of models (ii) for which a sub composition including three elements (K, Th, U in mg.kg<sup>-1</sup>)  
 647 is used together with the associated residual part (R) (**TSG[UThK]** in Table 9); and a third group of models  
 648 (iii) for which the available composition of 41 elements (in mg.kg<sup>-1</sup>) are used together with the residual part

649 (R) ([TSG] in Table 9, e.g., Figure 12 B). For models with (ii) or (iii), the isometric logratio of the four part sub-  
 650 composition or the 42 part composition are used respectively.

651 For SW England, the *lnRn* variation *between* **KM1BSs** explained by [TSG] is 39.7% (Table 9, Figure 12 B),  
 652 which is clearly above the value given by TSG[U] (14.7%, Figure 12 A) or by TSG[UThK] (16.3%), and at the  
 653 same level of the variation explained by [AGR]+G (38.8%, data (5) in Table 7). Explained variance is  
 654 enhanced +7.8% by adding [TSG] to the model with [AGR]+G (38.8% + 7.8% = 46.6%, Figure 12 C), and only  
 655 +1.2% or +1.5% by adding TSG[U] or TSG[UThK] respectively.

656 These results show that the topsoil geochemistry as a whole and using a CODA approach ([TSG]) clearly  
 657 enhances results provided by uranium (TSG[U]) or U, Th, K (TSG[UThK]) in topsoils. The topsoil “whole”  
 658 composition model [TSG] also show a better performance than the stream sediment one [SSG], which does  
 659 not happen if only uranium or the sub-composition with U, Th, K are used. Also, concerning to indoor radon,  
 660 a part of the information provided by [TSG] is complementary to that from [AGR]+G. The [TSG]+[AGR]+G  
 661 seems to provide the best model of all those considered in the present work, with more than **46%** of  
 662 variance explained (Figure 12 C).

663

664 Table 9 – Proportion of the variation (adjusted R<sup>2</sup> in %) of natural log transformed indoor radon  
 665 concentrations (*lnRn*) explained by topsoil geochemistry (TSG), or TSG combined with AGR and/or geology  
 666 (G).

TSG elements used	Adjusted R <sup>2</sup>			
	TSG[U]	TSG[U]+G	TSG[U]+[AGR]	TSG[U]+[AGR]+G
(i) Log <sub>10</sub> (U)	14.7	37.0	33.1	40.0
	TSG[UThK]	TSG[UThK]+G	TSG[UThK]+[AGR]	TSG[UThK]+[AGR]+G
(ii) U, Th, K (CODA)	16.3	37.2	33.2	40.3
	[TSG]	[TSG]+G	TSG+[AGR]	TSG+[AGR]+G
(iii) ALL 41 TSG elements (CODA)	39.7	44.5	42.8	46.6

667 TSG[U]: log<sub>10</sub> of uranium concentrations in topsoils; TSG[UThK]: the isometric logratio of the sub composition U, Th, K (including the  
 668 residual R (Equation 3)) concentrations in topsoils geochemistry; [TSG]: the isometric logratios of the available composition with 10

669 major+31 trace elements (including the residual R (Equation 3)) concentrations in topsoils geochemistry. Data used for all models:  
670 600 records of data (5) in Table 5, for which 29 BSs are available; [AGR]: airborne gamma-ray; G: geology (BS classes). The  
671 significance for the adjusted  $R^2$  is  $p < 0.001$ .

672

673 FIGURE 12 HERE

674

#### 675 4. DISCUSSION

676 For SW England, the indoor radon variability is higher *within* **KM1BS** polygons than *between* them. Thus, a  
677 wide difference can be observed between neighbouring houses built over the same geology, which reflects  
678 the different construction (namely the effectiveness of radon isolation from the ground) and ventilation  
679 (poorly ventilated usually induces higher radon concentrations) of the buildings, apart from any possible  
680 geological difference at *local* level. The **InRn** variation observed *between* **KM1BS** polygons (32.3% of the  
681 *total* variation, last column of Table 6) can be seen as an estimation of the maximum amount of indoor  
682 radon variability which can potentially be explained by the geological (s.l.) datasets here tested.

683 For the SW England, the indoor radon model with uranium (U) in stream sediments (**SSG[U]**, Table 8 and  
684 Figure 11 A) performs (slightly) better than uranium (U) in topsoils (**TSG[U]**, Table 9 and Figure 12 A). If Th  
685 and K are added to U, for both stream sediments (**SSG[UThK]**, Table 8) and soils (**TSG[UThK]**, Table 9) the  
686 model only slightly improves, while the model with stream sediments (20.0%) keeps performing better than  
687 that with soils (16.3%). However, geology (ca. 25% to 32%, Table 6, Figure 9) or [AGR] (ca. 27% to 32.5%,  
688 Table 7, Figure 10) alone or combined ([AGR]+G, ca. 32.5% to 39%, Table 7) provide a better explanation  
689 than any of the models above. Moreover, the U, Th, K sub-composition whether in stream sediments (Table  
690 8) or in soils (Table 9) do not provide further explanation (<1.5%) to models including airborne and geology  
691 ([AGR]+G, Table 7). There is an enhanced **InRn** prediction of the [AGR] model (32.5%, Table 7, Figure 10 B)  
692 relative to the **TSG[UThK]** (16.3%, Table 9), and that of **AGR[U]** (22.9% to 30.5%, Table 7, Figure 10 A)  
693 relative to **SSG[U]** (17.8%, Table 8, Figure 11 A) or **TSG[U]** (14.7%, Table 9, Figure 12 A). In short, these  
694 results indicate that the airborne gamma-ray data generally provide better models than the U, Th, K

695 concentrations in topsoil or stream sediment geochemistry, suggesting that the **AGR** signal may capture  
696 other characteristics than just the uranium (and thorium and potassium) concentration in soils,  
697 characteristics (e.g. permeability, water content, organic matter?) which are favourable for indoor radon  
698 explanation. In fact, the **AGR** data is a geophysical signal corresponding to the radiation emitted from the  
699 ground, thus depending on the “radiation permeability”. A typical case is the ground covered by peat which  
700 returns a depressed **AGR** signal (due to high water content) whatever the bedrock underneath. Also, this  
701 good performance of **AGR** also complies with Appleton et al. (2011) which, in an exercise carried out for the  
702 Northern Ireland, concluded that, among a set of soil geochemistry and airborne gamma-ray variables, the  
703 airborne eU (with 24% of variance explained) was the most significant parameter in modelling the radon  
704 potential.

705 For models with only one dataset, **[TSG]** shows the best performance, with an explained variance of about  
706 40% (Table 9, Figure 12 B), thus clearly above the other datasets, namely **[SSG]** (32.5%, Table 8, Figure 11 B),  
707 **[AGR]** (ca. 29.5% ± 2.7%, e.g., Figure 10 B) and **G** (ca. 29.0% ± 2.5%, e.g., Figure 9 B). Similar or improved  
708 results are obtained from the **[TSG]** model than that using airborne and geology together (**[AGR]+G**, 36.4% ±  
709 3.4%, Table 7, e.g., Figure 10 C) or that using **[SSG]+[AGR]+G** (36.8%, Table 8, Figure 11 C).

710 The best explanation of all tested models is given by those including the soil geochemistry (**[TSG]**), i.e.  
711 including all the available 41 elements, 10 major – Al, Ca, Fe, K, Mg, Mn, Na, P, Si, Ti - plus 31 trace, in a  
712 CODA approach), together with geology (**[TSG]+G**, 44.5%) or geology and **[AGR]** (**[TSG]+G+[AGR]**, 46.6%),  
713 shown in Figure 12 C. Models including combinations with **[TSG]** and **[SSG]** together were not included as  
714 the number of samples (83) is too low.

715 Adding geology **G** (**BS** geological classes) to the models **[SSG]** or **[TSG]**, only slightly improves the prediction  
716 (+2.2%, +4.8% respectively), suggesting that the “near-complete” geochemistry of stream sediments and of  
717 soils has already captured most of the bedrock-superficial (geology) variation. The same happens for **[AGR]**  
718 (2.2% for **[SSG]** and 3.1% for **[TSG]** models). This suggest that the information given by the **AGR** is mainly  
719 contained within the information given by the soil (or the stream sediment) geochemistry, at least with  
720 regard to the explanation of indoor radon.



721 The model with U (as well as the model with U, Th, K) concentration in stream sediments provides a slightly  
722 improved outcome than that with U (as well as that with U, Th, K) concentration in topsoil. This may reflect  
723 (i) the different processes behind the formation of these two materials, including the mobilization of U from  
724 soils to the stream sediments; (ii) the existence of organic peat soils over some (U-rich) granitic areas, which  
725 may show low levels of uranium, contrary to the stream sediments derived from the bedrock underneath  
726 peat; and (iii) the different size fractions analysed (<2 mm for soils and 150 µm for stream sediments), as the  
727 fine fraction often concentrate trace elements. Figure 13, providing information about the stability (CSR, the  
728 closest to 1 the more stable is the ratio) and of the median (MDN) of the U (or Th) / K<sub>2</sub>O plot, illustrates the  
729 net difference in U (and Th) concentrations between topsoils (<2 mm) and stream sediments (<150 µm), as  
730 well as over granitic rocks or other rock types.

731 However if the “complete” geochemical composition is used, topsoils ([TSG]) perform better than stream  
732 sediments ([SSG]) and doubles, at least, the explanation provided by SSG[U] models. This suggests that,  
733 from the point of view of the indoor radon, the information provided by the soil chemistry (of <2 mm size  
734 fraction) as a whole is more complete than that provided by the stream sediment (<150 µm), perhaps  
735 containing information about the structure of the soil and a more consistent signature of the bedrock.

736 Although, the lower performance of [SSG] relative to [TSG] models may also be an effect of the expected  
737 lower accuracy for parent material classification of stream sediments relative to soil samples. In fact, this  
738 lower accuracy may be the reason for the lower *lnRn lateral* variation explained by geology observed for  
739 data (4) which includes the stream sediment data (25.1%, in Table 6), relative to that for data(5) which  
740 includes the topsoil data (32.1%, in Table 6).

741

742 FIGURE 13 HERE

743

744

745 Despite obtaining better results from geochemistry datasets, namely that on soils, this type of data has  
746 several obvious disadvantages relative to the other tested datasets (geology and airborne data). Airborne

747 data is usually collected in a much denser set of locations, thus constituting a very rich dataset, with high  
748 potential for (very) detailed mapping of the spatial radon variation. However, the land coverage of **AGR**  
749 datasets is generally relatively small compared with stream sediment or topsoil geochemical datasets, as is  
750 the case in the UK. Geology has an obvious advantage as it is composed of polygons, contrary to the other  
751 datasets which are composed by a set of points. This characteristic of the geology allows the point-based  
752 information to be extrapolated to areas where these datasets do not exist. However, it seems that the  
753 indoor radon variation explained by geology is sensitive to the level of detail used.

754 The *ilr* transformation assures that all requirements for regression analysis of compositional data are fulfilled  
755 as their coordinates are in an orthogonal system. Yet, expressions for the calculation of *ilr* coordinates  
756 (isometric logratios) are complex and various rules to generate them exist (Egozcue et al., 2003). Using an  
757 adequate compositional regression approach, with *ilr* coordinates, the relevance of each original chemical  
758 element generally cannot be assessed from the model and their associated tests should therefore be ignored  
759 (van den Boogaart & Tolosana-Delgado, 2013).

760 At this point, it should be remembered that detailing the results of each *ilr* coordinate (i.e., each *ilr* logratio),  
761 namely in the **[SSG]** and **[TSG]** cases, in attempting to detect which are the most important elements (other  
762 than the obvious U, Th and K) for indoor radon explanation, is out of the scope of the present work.

763 A high number of negative values in an airborne dataset, which may occur after the standard calibrations  
764 and data processing (see section 2.3.), may disable the use of a CODA approach, as this is only valid for  
765 positive values. This was not the case for the present case study as only 3 records out of 8508 have been  
766 removed from data (3). Results (not shown) from **InRn** models with **AGR** data (eU, eTh, K) after a simple  
767 log<sub>10</sub>-transformation, are virtually the same as those reported in table 7 (i.e., after the CODA transformation  
768 used). This suggest that the airborne data is not necessarily compositional in its nature, despite of data  
769 being reported with units (as % for K and as ppm for eU and eTh) usually associated to compositional  
770 datasets. For the present case study we conclude that there is no obvious benefit on a CODA transformation  
771 of the **AGR** dataset.

772 The **AGR[U]**, **SSG[U]** and **TSG[U]**, were developed using a simple classical log<sub>10</sub> transformation of the U  
773 concentrations only (no other elements were taken into account). This is not critical for **AGR[U]** as the **AGR**  
774 data may not be compositional in nature. However, this is important for **SSG[U]** and **TSG[U]** as the **SSG** and  
775 the **TSG** datasets are clearly compositional. Thus, it may be concluded that the explained variance obtained  
776 for the models **SSG[U]** (Table 8) and **TSG[U]** (Table 9) is biased, that is, those values are not exclusively  
777 representative of uranium contents in stream sediments and topsoils respectively.

778

## 779 5. CONCLUSIONS

780 Results from an indoor radon modelling exercise carried out in SW England is presented here. Models were  
781 developed using geology, airborne data, topsoil geochemistry and stream sediment geochemistry. The main  
782 purpose was to understand how much indoor radon variation can be explained by the above earth science  
783 datasets, whether alone or combined. This was achieved by measuring the *lateral* variation of indoor radon,  
784 with *lateral* variation referring to the variation observed *between* the **KM1BSs** polygons, which size is equal  
785 or less than 1 km<sup>2</sup>.

786 For SW England, the indoor radon variability is higher *within* **KM1BS** polygons than *between* them.

787 Results show that any of the tested datasets can be useful for radon mapping purposes as each one is able to  
788 explain part of the indoor radon variance observed *between* **KM1BSs**. The stream sediment geochemistry,  
789 provided that they refer to small watershed areas, can be seen as a valid alternative to topsoil geochemistry,  
790 namely for areas where this last type of data is absence.

791 By combining two or more datasets, models are often enhanced but the variance explained never doubles,  
792 meaning that a large portion of the “information” provided by each one of the datasets is already contained  
793 in the “information” provided by the other datasets. The best model indicates that almost half (ca. 47%) of  
794 the **InRn** variation *between* **KM1BSs** in SW England can be explained by combining soil geochemistry (**[TSG]**),  
795 airborne data (**[AGR]**) and geology (**G**).

796 A “complete” topsoil (or stream sediment) geochemistry (in a CODA approach) is preferable to the use of  
797 uranium concentrations only. Not only because uranium concentration in a sample is dependent on the

798 concentrations of the other elements (leading to the CODA approach), but also because the information  
799 provided by the other elements may reflect other soil (stream sediment) characteristics (such as grain size,  
800 element's associations) which are potentially helpful in explaining indoor radon. This also applies to the U,  
801 Th, K sub-composition.

802 Using "complete" airborne data (eU, eTh, K) is preferable to the use of eU only. There is no obvious benefit  
803 on using a CODA transformation of the **AGR** dataset.

804 The present work was a first step to modelling indoor radon in the SW England using a multi-dataset  
805 multivariate CODA approach. The present work was focused on the significance of each dataset's model, as  
806 a whole, rather than evaluating the significance of each *ilr* coordinate or, in the case of stream sediments  
807 and topsoils, searching for elements, other than U, Th, K that may contribute for the explanation of indoor  
808 radon. Further work is planned to address this last point. Mapping these models is another point to be  
809 addressed in the near future.

810

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815

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942

#### 943 **FIGURE CAPTIONS**

944

945 **Figure 1.** Geological Periods in SW England, with main faults (black lines) and U (large red dots) and other  
946 mines (small black dots). The Carboniferous-Permian units (reddish) essentially refer to the U-rich granite  
947 outcrops (the major five are labelled in *brown italic*) referred in the text.

948

949 **Figure 2.** The 42 simplified bedrock (BED) and 8 simplified superficial (SUP) units from which derived the 134  
950 simplified bedrock-superficial geological combinations (**BSs**) of SW England (based on the BGS DiGMapGB-  
951 50k) used in the present study. The major five granite outcrops are labelled in *brown italic*.

952

953 **Figure 3.** The spatial distribution of the 197464 indoor radon measurements throughout 9144 KM1BS  
954 polygons (out of the total 27633) in the SW England. The underlying greyscale map refers to the BS  
955 geological combinations.

956

957 **Figure 4.** The spatial distribution of indoor radon concentrations (in  $\text{Bq}\cdot\text{m}^{-3}$ ) in SW England after allocating  
958 the each measurement to the underlying KM1BS polygon. Each concentration value refers to the geometric  
959 mean (GM) of measurements allocated to a KM1BS polygon. The histogram (*bars*) refers to the 197,464  
960 indoor radon measurements, while the histogram density *curve* refers to the 9,144 KM1BS GM values. The  
961 underlying greyscale map refers to the BS geological combinations.

962

963 **Figure 5.** Spatial distribution of equivalent U (eU) concentrations ( $\text{mg}\cdot\text{kg}^{-1}$ ) from airborne gamma-ray  
964 spectroscopy (AGR). The total number of data points in the map is 684,384.

965

966 **Figure 6.** Spatial distribution of equivalent U (eU) concentrations ( $\text{mg}\cdot\text{kg}^{-1}$ ). Values are the arithmetic mean  
967 of eU data points (Figure 5) in each KM1BS polygon. The total number of KM1BS polygons with AGR data in  
968 the map is 23,573.

969

970 **Figure 7.** U content ( $\text{mg}\cdot\text{kg}^{-1}$ ) in 3382 stream sediment samples from the SW England. Data from the G-BASE  
971 project. The underlying greyscale map refers to the BS geological combinations.

972

973 **Figure 8.** U content ( $\text{mg}\cdot\text{kg}^{-1}$ ) in 987 topsoil samples from the SW England. Data from the G-BASE project and  
974 the NSI inventory. The underlying greyscale map refers to the BS geological combinations.

975

976 **Figure 9:**  $\ln\text{Rn}$  (Y axis) modelled by Geology (G the BS geological classes derived from the BGS DiGMapGB-  
977 50k), using **A) 197484** log indoor radon measurements used (data (1) of Table 5), 10.1% of the *total  $\ln\text{Rn}$*   
978 variance explained; **B) 8905** records of log indoor radon averages at KM1BS (data (3) of Table 5), 29.8% of  
979 the *lateral  $\ln\text{Rn}$*  variance explained; **C) 600** records of log indoor radon averages at KM1BS (data (5) of Table  
980 5), 32.1% of the *lateral  $\ln\text{Rn}$*  variance explained. The X axis units are not relevant, corresponding to BS  
981 geological classes recoded to numbers.

982

983 **Figure 10:**  $\ln\text{Rn}$  (Y axis) modelled by airborne gamma ray data (AGR) using **600** records (data (5) of Table 5)  
984 predicted by **A) the  $\log_{10}(\text{eU})$**  (with eU concentrations in  $\text{mg}\cdot\text{kg}^{-1}$ ), 30.5% of the *lateral  $\ln\text{Rn}$*  variance  
985 explained; **B) the [AGR]**, i.e., the 3 isometric logratios of the 4-part composition (eU, eTh, K and the residual  
986 R (Equation 3) with elements concentrations in  $\text{mg}\cdot\text{kg}^{-1}$ ), 32.5% of the *lateral  $\ln\text{Rn}$*  variance explained; **C) the**

987 **[AGR]+G**, with [AGR] as in graph B) and G as the BS geological classes derived from the BGS DiGMapGB-50k,  
988 32.5% of the *lateral InRn* variance explained. The X axis units for B) and C) graphs are dimensionless (the  
989 logarithm of the ratio between two or more elements) and not relevant in compositional data analysis.

990

991 **Figure 11:** *InRn* (Y axis) modelled by stream sediment geochemistry (**SSG**) using **1182** records (data (4) of  
992 Table 5) predicted by **A)** the  $\log_{10}(\mathbf{U})$  (with U concentrations in  $\text{mg.kg}^{-1}$ ), 17.8% of the *lateral InRn* variance  
993 explained; **B)** **[SSG]**, i.e., the 41 isometric logratios of the 42-part **SSG** composition (41 elements and residual  
994 R (Equation 3) with elements concentrations in  $\text{mg.kg}^{-1}$ ), 32.5% of the *lateral InRn* variance explained; **C)**  
995 **[SSG]+[AGR]+G**, with **[SSG]** as in graph B), [AGR] as the 3 isometric logratios of the 4-part composition (eU,  
996 eTh, K and the residual R (Equation 3) with elements concentrations in  $\text{mg.kg}^{-1}$ ) and G as the BS geological  
997 classes derived from the BGS DiGMapGB-50k, 36.8% of the *lateral InRn* variance explained. The X axis units  
998 for B) and C) graphs are dimensionless (the logarithm of the ratio between two or more elements) and not  
999 relevant in compositional data analysis.

1000

1001 **Figure 12:** *InRn* (Y axis) modelled by topsoil geochemistry (**TSG**) using **600** records (data (5) of Table 5),  
1002 predicted by **A)** the  $\log_{10}(\mathbf{U})$  (with U concentrations in  $\text{mg.kg}^{-1}$ ), 14.7% of the *lateral InRn* variance explained;  
1003 **B)** **[TSG]**, i.e., the 41 isometric logratios of the 42-part **TSG** composition (41 elements and the residual R  
1004 (Equation 3) with elements concentrations in  $\text{mg.kg}^{-1}$ ), 39.7% of the *lateral InRn* variance explained; **C)**  
1005 **[TSG]+[AGR]+G**, with **[TSG]** as in graph B), [AGR] as the 3 isometric logratios of the 4-part composition (eU,  
1006 eTh, K and the residual R (Equation 3) with elements concentrations in  $\text{mg.kg}^{-1}$ ) and G as the BS geological  
1007 classes derived from the BGS DiGMapGB-50k, 46.6% of the *lateral InRn* variance explained. The X axis units  
1008 for B) and C) graphs are dimensionless (the logarithm of the ratio between two or more elements) and not  
1009 relevant in compositional data analysis.

1010

1011 **Figure 13:** scatterplots of  $\text{K}_2\text{O}$  and Th (A, B) and  $\text{K}_2\text{O}$  and U (C, D) in soils (A, C) and stream sediments (B, D)  
1012 from SW England. Red dots represent samples over Carboniferous-Permian granitic rocks (CG); black dots  
1013 represent samples over the remaining bedrock units (OB). CSR: coefficient of stability of the ratio (Filzmoser  
1014 et al., 2010); MDN: median; N: number of samples. The dashed lines indicate a constant ratio (corresponding  
1015 to the median of the log-ratio) among the pair, while the solid lines indicate a ratio 2 times higher (upper  
1016 line) or one half (lower line) the constant ratio for samples over OB (in grey) and CG (in red).  
1017  
1018

1019

1020 **SUPPLEMENTARY MATERIAL**

1021

1022 **Extra Table A** – Summary statistics for 41 elements from **stream sediment geochemistry** of the SW England  
 1023 (N= 3,382 data points) and after taking the average of the data points in each KM1BS polygon (N = 3,027  
 1024 data polygons)

	N	min	Q25	mdn	Q75	Q95	Q99	max	IQR	mean	GM	GSD
Al <sub>2</sub> O <sub>3</sub>	3382	6.20	13.99	16.29	18.59	21.29	23.09	28.19	4.60	16.26	15.93	1.23
Al <sub>2</sub> O <sub>3</sub>	3027	6.20	13.99	16.29	18.59	21.29	23.08	28.19	4.60	16.26	15.94	1.23
CaO	3382	0.25	0.49	0.55	0.69	1.25	2.84	7.82	0.20	0.67	0.61	1.47
CaO	3027	0.28	0.49	0.56	0.70	1.29	2.89	7.82	0.21	0.69	0.62	1.48
Fe <sub>2</sub> O <sub>3</sub>	3382	1.00	4.99	6.17	7.44	9.68	12.34	22.08	2.45	6.28	5.93	1.42
Fe <sub>2</sub> O <sub>3</sub>	3027	1.00	5.01	6.17	7.45	9.77	12.37	22.08	2.44	6.30	5.94	1.42
K <sub>2</sub> O	3382	0.53	2.33	2.75	3.18	3.92	4.53	6.78	0.85	2.78	2.69	1.29
K <sub>2</sub> O	3027	0.53	2.33	2.75	3.19	3.91	4.48	6.17	0.86	2.78	2.70	1.28
MgO	3382	0.70	1.00	1.20	1.50	2.00	3.72	15.10	0.50	1.30	1.23	1.36
MgO	3027	0.70	1.00	1.20	1.50	2.00	3.87	15.10	0.50	1.31	1.24	1.37
MnO	3382	0.02	0.12	0.18	0.27	0.60	1.32	8.90	0.15	0.24	0.18	2.06
MnO	3027	0.02	0.12	0.18	0.27	0.60	1.35	6.56	0.15	0.24	0.18	2.04
Na <sub>2</sub> O	3382	0.20	0.70	0.90	1.00	1.30	1.80	3.00	0.30	0.87	0.83	1.38
Na <sub>2</sub> O	3027	0.20	0.70	0.90	1.00	1.30	1.80	3.00	0.30	0.87	0.82	1.38
P <sub>2</sub> O <sub>5</sub>	3382	0.03	0.16	0.20	0.25	0.39	0.53	0.85	0.09	0.21	0.20	1.45
P <sub>2</sub> O <sub>5</sub>	3027	0.03	0.16	0.20	0.25	0.39	0.53	0.85	0.09	0.22	0.20	1.44
SiO <sub>2</sub>	3382	12.69	54.47	59.34	63.24	69.23	74.73	85.61	8.77	58.57	58.03	1.15
SiO <sub>2</sub>	3027	16.98	54.25	59.24	63.14	68.70	74.66	85.61	8.89	58.41	57.89	1.15
TiO <sub>2</sub>	3382	0.24	0.80	0.86	0.92	1.14	1.73	3.37	0.11	0.87	0.85	1.23
TiO <sub>2</sub>	3027	0.24	0.80	0.86	0.92	1.15	1.77	3.37	0.12	0.87	0.85	1.23
As	3382	1.40	11.10	16.90	37.10	197.8	786.7	11000	26.00	65.05	23.07	2.97
As	3027	1.40	11.20	17.15	39.20	200.5	844.7	11000	28.00	67.27	23.64	3.00
Ba	3382	101.0	377.3	444.0	492.0	569.0	642.2	5000	114.8	432.7	417.4	1.31
Ba	3027	142.0	375.0	444.0	491.0	568.0	636.5	5000	116.0	431.5	416.2	1.31
Br	3382	0.50	9.70	15.90	29.00	67.30	139.5	427.0	19.30	24.21	17.11	2.22
Br	3027	0.50	9.80	16.10	29.30	67.27	137.7	427.0	19.50	24.43	17.37	2.21
Ce	3382	22.0	67.0	75.0	88.0	154.0	261.6	1940	21.0	86.4	80.4	1.39
Ce	3027	22.0	67.0	75.0	88.8	152.7	264.7	1940	21.8	86.7	80.6	1.39
Co	3382	1.8	16.1	21.7	30.3	60.4	106.0	1926	14.2	27.0	21.8	1.85
Co	3027	2.4	16.2	21.8	30.3	60.7	106.3	984.6	14.1	26.9	21.9	1.85
Cr	3382	2.10	73.40	86.80	102.8	127.8	274.7	4097	29.40	93.76	81.20	1.65
Cr	3027	7.90	73.60	86.93	102.8	129.2	280.6	4097	29.20	94.52	81.46	1.65
Cs	3382	1.00	7.00	10.00	18.00	40.00	66.00	153.00	11.00	14.20	10.47	2.16
Cs	3027	1.00	7.00	10.00	18.00	40.00	64.74	153.00	11.00	14.36	10.64	2.15
Cu	3382	3.70	17.80	24.10	34.00	91.77	635.8	12000	16.20	59.50	26.94	2.15
Cu	3027	3.70	17.93	24.50	34.63	95.04	780.6	8000	16.70	60.69	27.54	2.17
Ga	3382	2.60	16.70	20.20	23.90	29.40	35.80	67.90	7.20	20.46	19.65	1.34
Ga	3027	4.60	16.70	20.20	23.90	29.40	35.80	67.90	7.20	20.53	19.74	1.33
Ge	3382	0.10	0.80	1.00	1.40	1.80	2.30	7.70	0.60	1.06	0.94	1.74
Ge	3027	0.10	0.80	1.00	1.33	1.80	2.30	7.70	0.53	1.07	0.95	1.72
Hf	3382	0.50	6.50	8.60	11.80	26.10	95.07	1014	5.30	12.73	9.30	1.84
Hf	3027	0.50	6.50	8.60	11.70	25.08	101.8	548.7	5.20	12.50	9.25	1.83
La	3382	10.00	37.00	42.00	49.75	84.00	130.2	320.0	12.75	47.12	44.37	1.38
La	3027	10.00	37.00	42.00	50.00	84.70	131.0	318.0	13.00	47.34	44.54	1.38
Mo	3382	0.10	0.60	0.90	1.50	3.40	6.00	11.70	0.90	1.29	1.00	1.96
Mo	3027	0.10	0.60	0.90	1.50	3.30	6.00	11.70	0.90	1.28	1.00	1.95
Nb	3382	8.40	16.40	17.40	19.00	32.40	50.88	350.0	2.60	19.73	18.73	1.31

Nb	3027	8.40	16.50	17.40	19.10	32.70	50.77	350.0	2.60	19.84	18.80	1.32
Nd	3382	3.00	31.40	35.70	41.60	71.60	118.7	313.2	10.20	39.91	37.44	1.39
Nd	3027	3.00	31.50	35.90	41.80	72.67	122.4	313.2	10.30	40.12	37.60	1.40
Ni	3382	3.80	32.73	44.10	57.50	96.09	173.3	1689	24.78	49.91	40.23	1.94
Ni	3027	4.10	32.90	44.40	57.80	97.05	178.7	1041	24.90	50.13	40.43	1.95
Pb	3382	5.00	26.43	34.10	50.58	121.0	481.1	10000	24.15	68.65	39.87	1.95
Pb	3027	5.00	26.80	34.60	51.45	125.5	508.4	10000	24.65	70.04	40.54	1.96
Rb	3382	27.20	99.70	126.9	164.6	308.1	477.0	916.4	64.88	147.07	131.26	1.58
Rb	3027	27.20	100.2	127.8	167.4	307.0	475.3	916.4	67.20	147.84	132.12	1.58
Sb	3382	0.10	0.70	1.10	2.00	6.60	16.34	128.3	1.30	2.13	1.19	2.62
Sb	3027	0.10	0.70	1.10	2.00	6.70	16.64	128.3	1.30	2.16	1.21	2.59
Sc	3382	0.80	10.40	13.30	16.10	19.80	23.00	34.50	5.70	13.25	12.51	1.44
Sc	3027	0.80	10.40	13.35	16.00	19.70	23.00	34.50	5.60	13.28	12.53	1.43
Se	3382	0.10	0.50	0.80	1.40	2.81	4.83	11.63	0.90	1.09	0.79	2.28
Se	3027	0.10	0.50	0.90	1.40	2.80	4.87	11.60	0.90	1.11	0.81	2.27
Sm	3382	0.10	4.60	6.00	8.00	14.10	22.21	72.60	3.40	6.86	5.96	1.74
Sm	3027	0.10	4.70	6.00	7.90	14.24	22.85	72.60	3.20	6.89	6.00	1.73
Sn	3382	1.00	3.40	5.50	62.00	1571	2000	2212	58.60	190.66	16.05	8.08
Sn	3027	1.00	3.50	5.90	75.85	1627	2000	2212	72.35	199.10	17.17	8.24
Sr	3382	18.80	66.40	80.70	97.28	129.1	164.6	334.2	30.88	83.06	78.60	1.41
Sr	3027	18.80	66.65	80.70	97.10	129.5	165.2	334.2	30.45	83.27	78.93	1.40
Th	3382	2.60	9.80	11.30	12.90	43.38	77.24	321.7	3.10	15.01	12.53	1.64
Th	3027	2.60	9.88	11.30	13.00	44.20	76.41	321.7	3.13	15.07	12.58	1.64
U	3382	1.00	2.50	2.90	3.50	24.48	44.44	206.9	1.00	5.70	3.64	2.10
U	3027	1.00	2.50	2.90	3.50	23.91	43.87	206.9	1.00	5.70	3.66	2.11
V	3382	10.30	81.60	105.1	131.0	160.0	191.7	289.1	49.38	104.7	96.77	1.53
V	3027	20.40	82.05	105.0	131.5	160.0	193.1	289.1	49.45	104.9	96.98	1.53
W	3382	0.10	2.30	3.10	5.40	53.59	153.0	2000	3.10	14.15	4.61	3.18
W	3027	0.10	2.30	3.10	6.00	55.21	157.7	1314	3.70	14.35	4.74	3.22
Y	3382	12.60	28.30	30.30	34.20	67.79	113.2	531.8	5.90	35.43	33.06	1.37
Y	3027	12.60	28.30	30.30	34.30	67.94	114.3	531.8	6.00	35.36	33.06	1.37
Zn	3382	20.00	98.33	127.7	175.2	427.7	1300	12000	76.85	192.4	138.7	1.88
Zn	3027	20.00	99.30	128.7	176.9	445.3	1408	7594	77.60	194.7	140.5	1.88
Zr	3382	30.00	240.0	329.8	438.1	871.1	2000	2000	198.1	402.1	343.8	1.66
Zr	3027	68.30	239.9	327.9	433.1	852.1	2000	2000	193.2	398.3	341.9	1.65

1025 N: number of samples; Min: minimum value; Max: maximum value; Q##: ##th quantile; Mdn: median; IQR: interquartile range; AM:

1026 arithmetic mean; GM: geometric mean; GSD: geometric standard deviation. Units (in brackets) are valid for all parameters except

1027 GSD (dimensionless).

1028

1029 **Extra Table B** – Summary statistics for 41 elements from **top soil geochemistry** of the SW England (N= 987

1030 data points) and after taking the average of the data points in each KM1BS polygon (N = 947 data polygons)

	N	min	Q25	mdn	Q75	Q95	Q99	max	IQR	mean	GM	GSD
Al <sub>2</sub> O <sub>3</sub>	987	0.30	12.50	15.00	17.29	20.30	21.90	23.00	4.79	14.45	13.41	1.62
Al <sub>2</sub> O <sub>3</sub>	947	0.30	12.50	15.00	17.29	20.30	21.80	23.00	4.80	14.44	13.40	1.62
CaO	987	0.04	0.40	0.53	0.68	1.60	4.34	26.73	0.28	0.74	0.53	1.96
CaO	947	0.04	0.40	0.53	0.68	1.67	4.36	26.73	0.29	0.74	0.53	1.98
Fe <sub>2</sub> O <sub>3</sub>	987	0.29	3.86	5.49	6.73	8.15	10.34	12.06	2.87	5.19	4.44	1.95
Fe <sub>2</sub> O <sub>3</sub>	947	0.29	3.87	5.49	6.73	8.15	10.32	11.80	2.86	5.20	4.46	1.94
K <sub>2</sub> O	987	0.01	1.96	2.47	2.96	3.43	3.96	4.94	1.01	2.41	2.20	1.84
K <sub>2</sub> O	947	0.01	1.95	2.47	2.96	3.43	3.92	4.94	1.01	2.41	2.19	1.86
MgO	987	0.20	0.80	1.00	1.30	1.80	3.21	6.90	0.50	1.10	1.00	1.56
MgO	947	0.20	0.80	1.00	1.30	1.80	3.25	6.90	0.50	1.10	1.00	1.57

MnO	987	0.00	0.05	0.12	0.18	0.30	0.58	1.75	0.13	0.14	0.09	2.83
MnO	947	0.00	0.06	0.12	0.18	0.30	0.58	1.75	0.13	0.14	0.09	2.83
Na <sub>2</sub> O	987	0.10	0.50	0.70	0.90	1.20	1.50	2.70	0.40	0.73	0.66	1.58
Na <sub>2</sub> O	947	0.10	0.50	0.70	0.90	1.20	1.50	2.70	0.40	0.72	0.66	1.59
P <sub>2</sub> O <sub>5</sub>	987	0.02	0.21	0.28	0.35	0.50	0.66	0.82	0.14	0.29	0.26	1.61
P <sub>2</sub> O <sub>5</sub>	947	0.02	0.21	0.28	0.35	0.50	0.67	0.82	0.14	0.29	0.26	1.61
SiO <sub>2</sub>	987	1.70	52.30	57.80	62.50	69.57	73.03	79.70	10.20	55.96	53.13	1.54
SiO <sub>2</sub>	947	1.70	52.28	57.80	62.57	69.52	73.11	79.70	10.30	55.94	53.07	1.55
TiO <sub>2</sub>	987	0.00	0.70	0.83	0.90	1.14	1.90	2.64	0.20	0.80	0.72	1.75
TiO <sub>2</sub>	947	0.00	0.70	0.83	0.90	1.12	1.87	2.64	0.20	0.80	0.72	1.76
As	987	4.70	15.70	23.50	39.70	134.4	356.6	1949.2	24.00	46.50	28.09	2.29
As	947	4.70	15.81	23.60	39.80	134.4	351.8	1949.2	24.00	46.00	28.18	2.27
Ba	987	13.20	284.0	374.6	435.5	514.0	556.0	641.0	151.5	354.7	326.5	1.63
Ba	947	13.17	284.0	374.2	435.9	513.7	556.0	610.0	152.0	354.1	325.5	1.64
Br	987	4.00	24.20	32.60	44.95	131.2	254.0	459.1	20.75	44.95	35.10	1.88
Br	947	4.00	24.66	32.90	44.80	130.5	256.1	459.1	20.15	45.15	35.26	1.88
Ce	987	6.40	54.90	66.90	75.00	84.79	92.91	134.10	20.10	63.07	59.57	1.47
Ce	947	6.35	54.78	66.90	75.05	84.77	92.96	134.09	20.27	63.07	59.56	1.47
Co	987	0.10	5.70	13.10	18.00	27.70	40.70	89.30	12.30	13.12	9.23	2.74
Co	947	0.10	5.71	13.00	18.05	27.71	40.98	89.30	12.35	13.14	9.26	2.74
Cr	987	7.60	60.75	80.90	95.90	112.8	267.3	1577.8	35.15	83.17	67.33	1.90
Cr	947	7.60	61.04	80.93	96.00	113.9	298.5	1577.8	34.97	83.84	67.75	1.90
Cs	987	1.80	7.00	10.00	17.00	36.87	64.21	240.0	10.00	14.67	11.16	2.00
Cs	947	1.79	7.00	10.00	17.00	36.00	63.08	184.5	10.00	14.42	11.12	1.98
Cu	987	0.30	16.70	25.70	37.20	85.86	189.0	690.5	20.50	34.27	24.58	2.21
Cu	947	0.30	16.95	25.88	37.30	85.12	189.4	690.5	20.35	34.42	24.76	2.21
Ga	987	0.20	15.40	19.10	23.60	28.17	30.10	34.60	8.20	19.06	17.39	1.73
Ga	947	0.20	15.39	19.10	23.50	27.94	29.95	32.30	8.11	19.01	17.34	1.73
Ge	987	0.10	0.70	1.00	1.30	1.80	2.20	7.10	0.60	1.04	0.90	1.82
Ge	947	0.10	0.70	1.00	1.30	1.80	2.20	7.10	0.60	1.04	0.91	1.80
Hf	987	0.50	5.00	6.20	7.40	9.60	11.10	14.60	2.40	6.26	5.87	1.48
Hf	947	0.46	5.00	6.20	7.40	9.51	11.08	14.58	2.40	6.26	5.87	1.49
La	987	0.20	29.00	35.00	41.00	47.00	51.84	79.40	12.00	33.62	30.46	1.85
La	947	0.24	29.00	35.00	41.00	46.95	51.96	79.35	12.00	33.59	30.39	1.86
Mo	987	0.10	0.80	1.10	1.40	2.47	4.31	17.50	0.60	1.28	1.10	1.70
Mo	947	0.10	0.85	1.11	1.40	2.47	4.30	17.50	0.56	1.28	1.11	1.70
Nb	987	0.60	15.10	16.80	18.00	22.91	39.18	107.0	2.90	16.93	16.00	1.46
Nb	947	0.64	15.10	16.72	18.00	22.69	36.96	107.0	2.90	16.87	15.93	1.47
Nd	987	0.20	21.60	28.30	32.30	37.40	44.71	86.90	10.70	26.38	23.83	1.74
Nd	947	0.20	21.60	28.20	32.30	37.40	44.42	72.83	10.70	26.33	23.82	1.74
Ni	987	1.70	13.15	26.00	38.95	58.57	148.9	564.0	25.80	30.64	21.40	2.39
Ni	947	1.65	13.32	26.00	38.95	59.04	154.3	564.0	25.63	30.89	21.56	2.39
Pb	987	10.80	35.00	45.10	66.75	125.50	206.6	521.2	31.75	57.60	49.59	1.66
Pb	947	10.80	35.40	45.95	66.88	125.95	207.0	521.2	31.48	58.09	49.99	1.66
Rb	987	3.40	104.2	140.1	179.9	286.6	389.8	551.3	75.65	150.4	132.1	1.79
Rb	947	3.40	104.4	140.1	179.3	283.1	383.7	551.3	74.90	149.6	131.4	1.80
Sb	987	0.10	0.70	1.20	2.30	6.80	19.71	79.70	1.60	2.29	1.32	2.57
Sb	947	0.10	0.70	1.20	2.30	6.86	20.59	79.70	1.60	2.32	1.33	2.58
Sc	987	0.50	8.30	12.20	14.95	18.20	21.40	24.20	6.65	11.48	10.08	1.80
Sc	947	0.50	8.40	12.20	15.00	18.20	21.35	24.15	6.60	11.51	10.12	1.80
Se	987	0.10	0.60	0.80	1.10	1.90	3.31	6.80	0.50	0.95	0.84	1.67
Se	947	0.10	0.60	0.80	1.10	1.90	3.35	4.80	0.50	0.95	0.84	1.66
Sm	987	0.10	2.90	4.20	5.30	7.17	8.73	20.40	2.40	4.19	3.63	1.90
Sm	947	0.10	2.98	4.20	5.30	7.16	8.69	15.72	2.32	4.19	3.63	1.90
Sn	987	2.10	5.85	12.50	41.10	230.2	494.5	2000	35.25	51.28	17.64	3.74
Sn	947	2.10	5.90	12.47	40.15	231.2	495.8	2000	34.26	51.33	17.60	3.72
Sr	987	12.50	54.40	70.00	90.20	144.0	321.5	1305	35.80	81.06	69.51	1.67

Sr	947	12.50	54.45	70.08	90.62	142.5	351.3	1305	36.17	81.49	69.73	1.68
Th	987	0.30	8.80	10.30	11.60	13.40	17.21	22.20	2.80	9.99	9.43	1.51
Th	947	0.30	8.78	10.25	11.60	13.40	17.25	22.20	2.82	9.99	9.42	1.51
U	987	0.70	2.50	2.90	3.40	6.10	11.71	45.70	0.90	3.34	3.01	1.49
U	947	0.66	2.50	2.90	3.43	5.87	11.70	35.35	0.93	3.31	3.00	1.48
V	987	11.60	79.95	108.9	135.1	170.2	204.3	250.8	55.10	103.9	89.42	1.89
V	947	11.60	80.40	109.4	135.1	170.3	203.5	241.7	54.66	104.1	89.71	1.88
W	987	0.10	2.50	3.30	6.35	16.17	42.70	279.60	3.85	6.44	4.06	2.25
W	947	0.10	2.50	3.30	6.30	15.88	38.79	270.35	3.80	6.14	4.02	2.22
Y	987	1.30	20.50	25.40	27.70	30.40	40.51	82.20	7.20	23.57	22.01	1.54
Y	947	1.30	20.50	25.30	27.61	30.42	39.55	76.98	7.11	23.51	21.96	1.54
Zn	987	14.60	56.75	84.20	120.4	204.6	389.4	720.9	63.60	97.48	81.70	1.80
Zn	947	17.25	57.23	84.74	120.4	204.6	389.6	720.9	63.21	97.85	82.23	1.79
Zr	987	0.50	180.0	223.6	268.0	334.5	385.3	551.2	88.00	223.2	196.9	2.18
Zr	947	0.50	180.5	222.9	267.7	334.4	383.8	551.2	87.15	222.7	195.9	2.20

1031 **N**: number of samples; **Min**: minimum value; **Max**: maximum value; **Q##**: ##th quantile; **Mdn**: median; **IQR**: interquartile range; **AM**:

1032 arithmetic mean; **GM**: geometric mean; **GSD**: geometric standard deviation. Units (in brackets) are valid for all parameters except

1033 **GSD** (dimensionless).

1034

Figure01  
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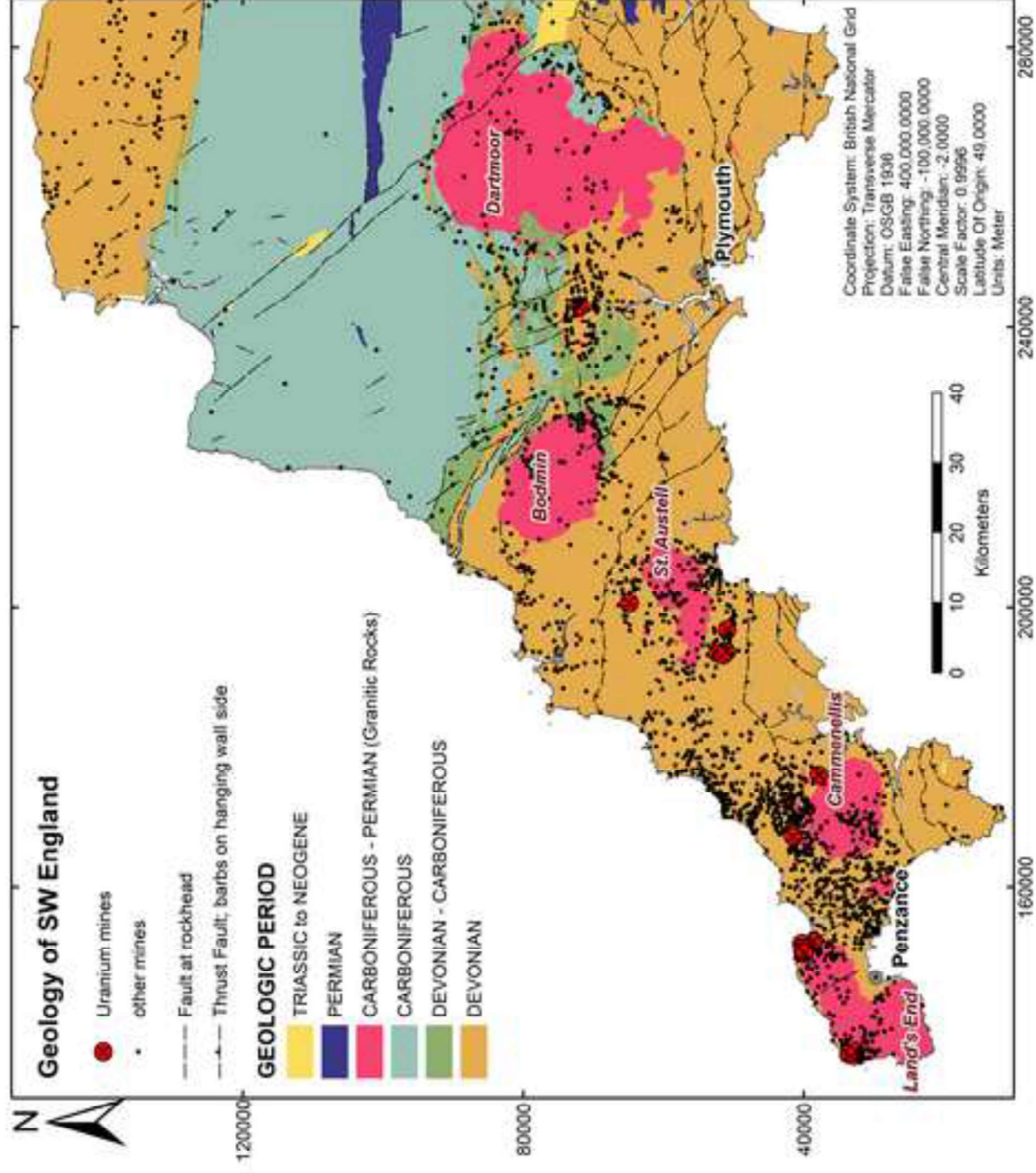




Figure02

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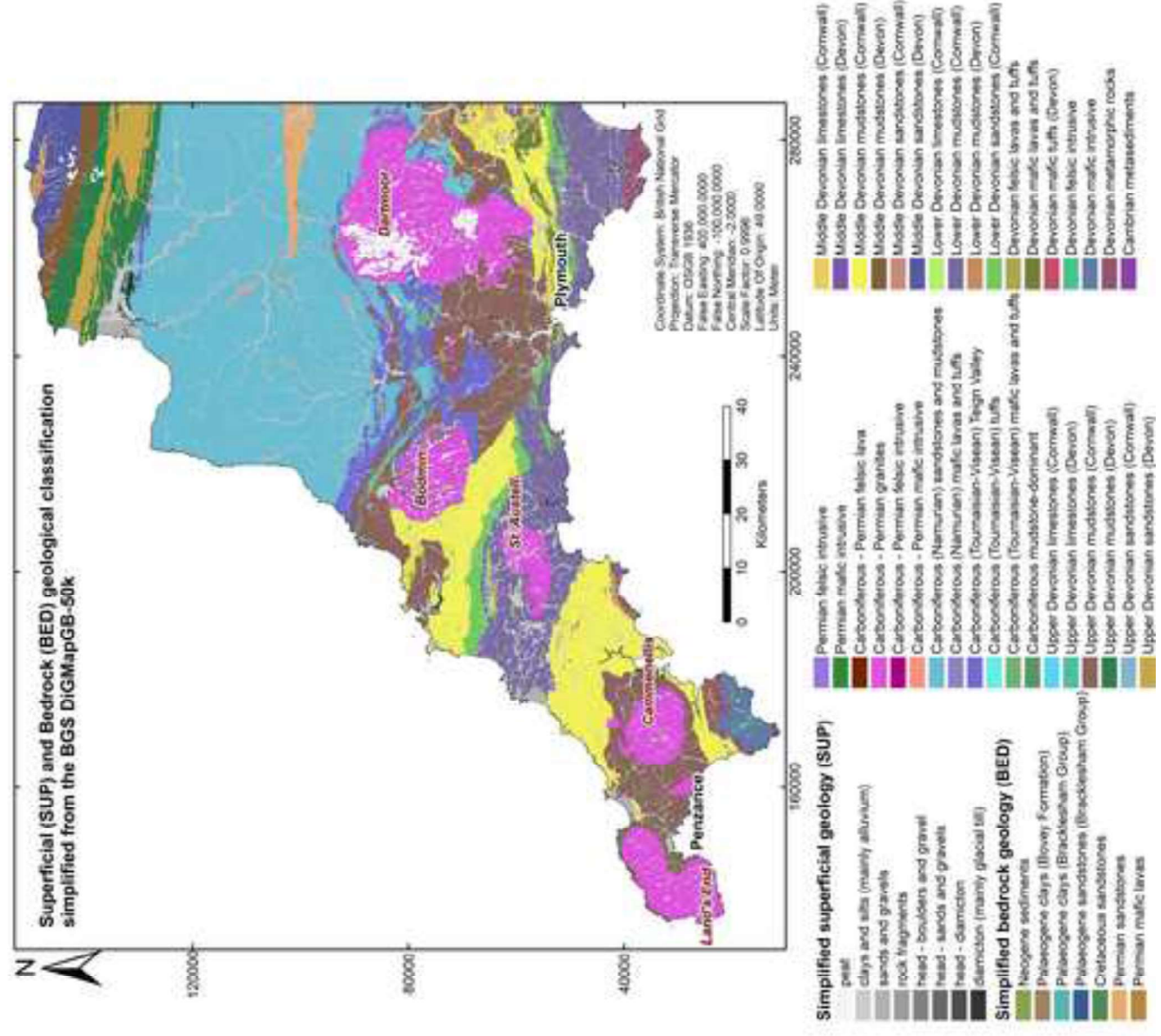


Figure03  
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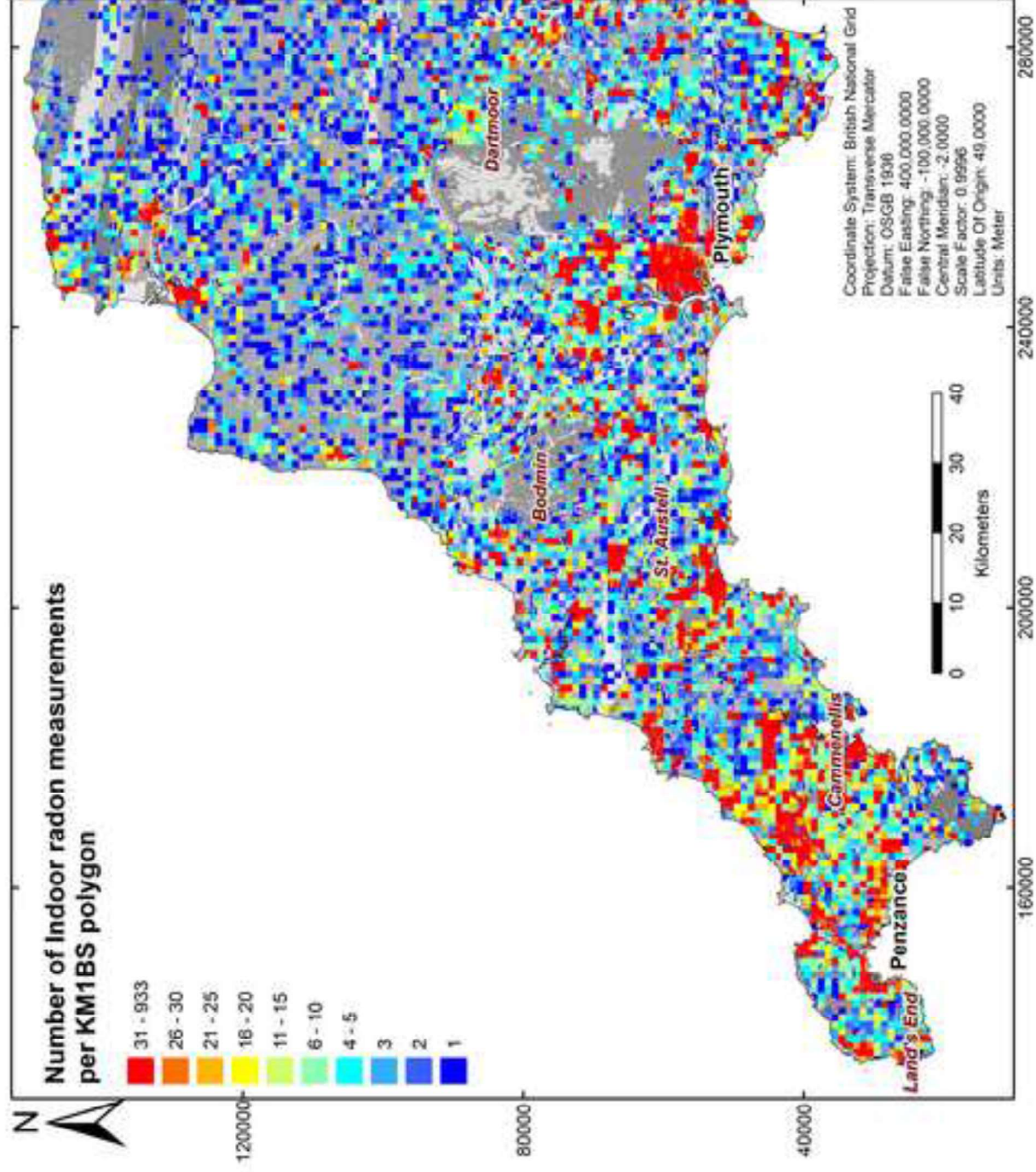


Figure04  
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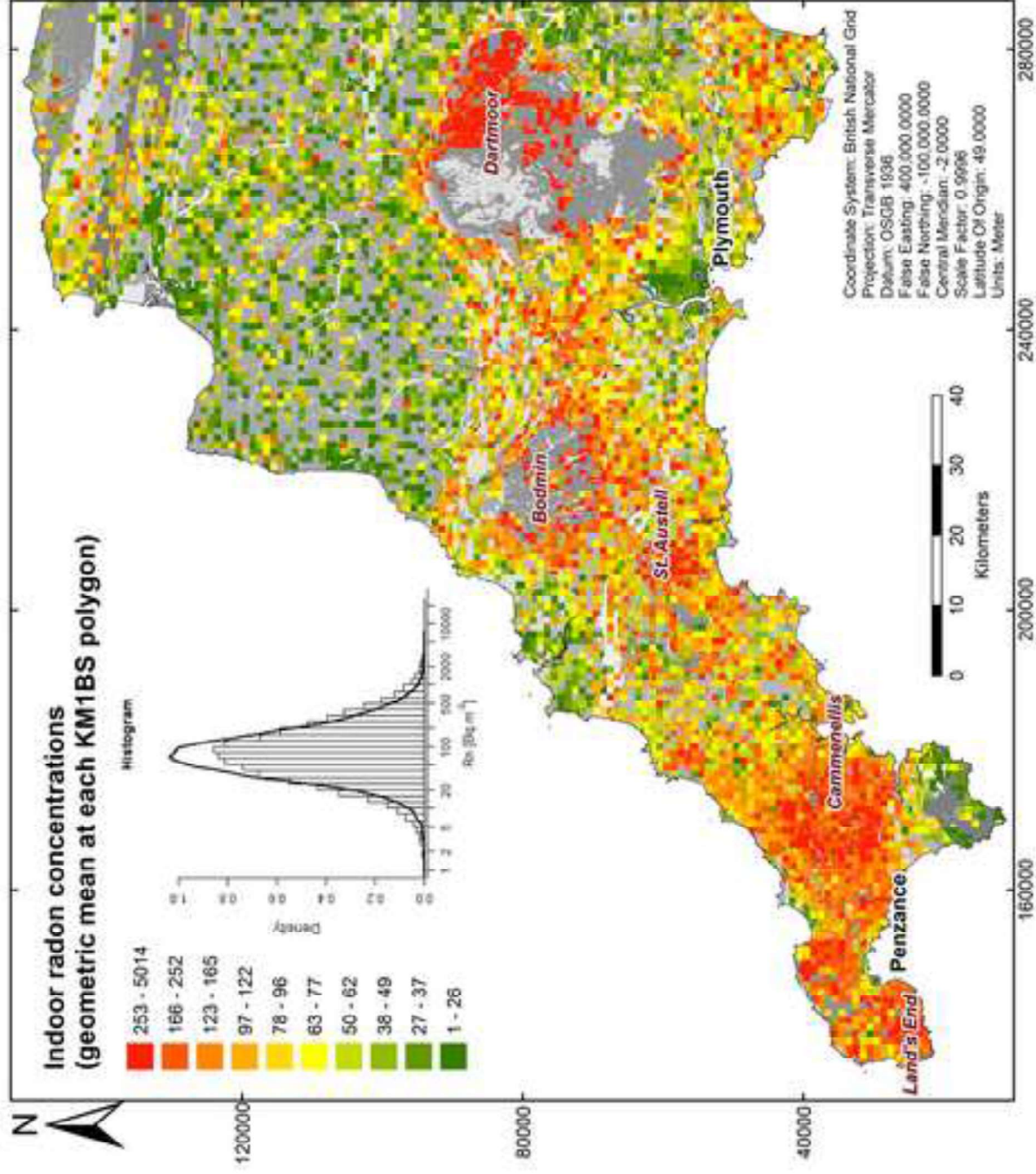


Figure05  
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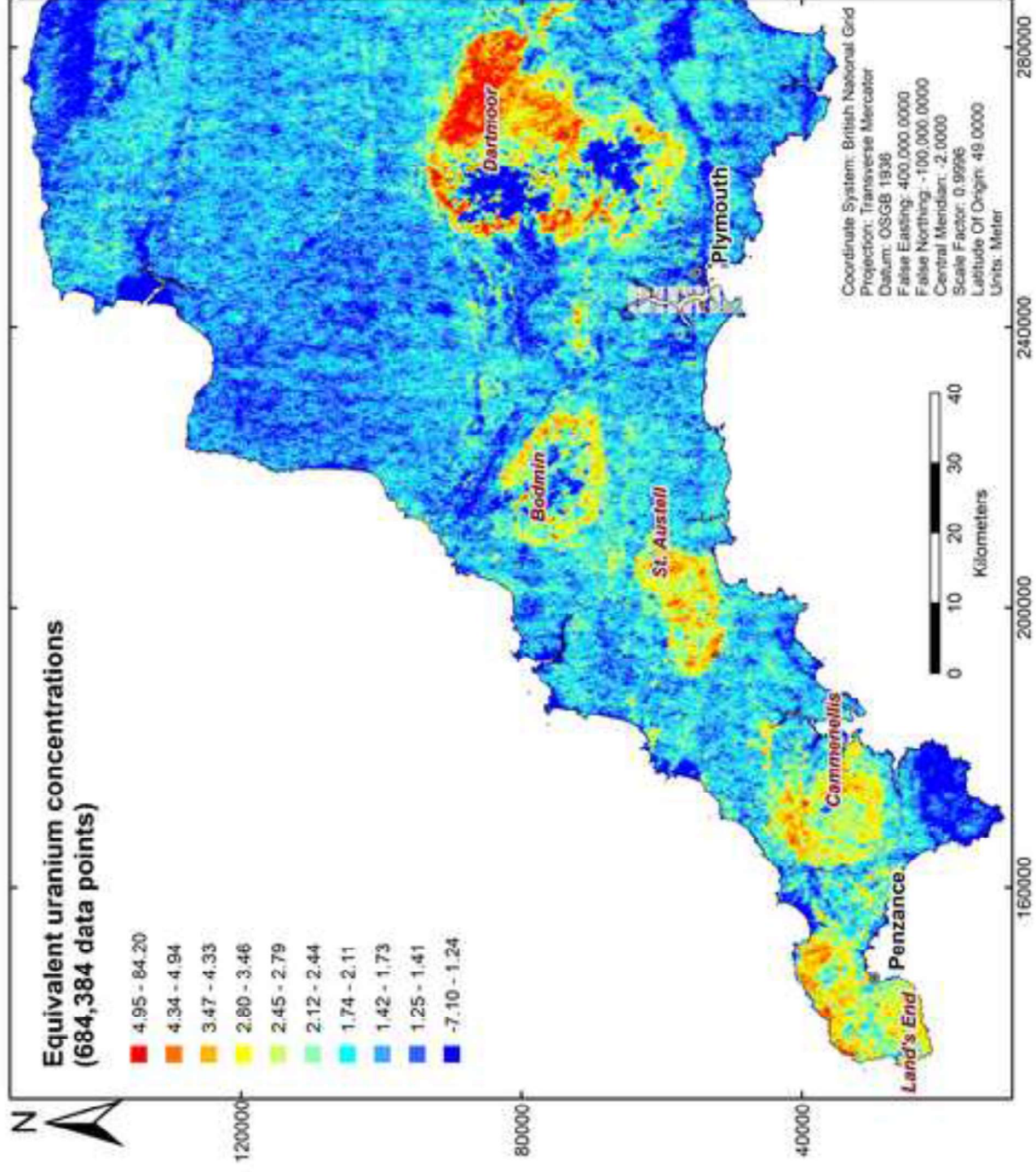


Figure06  
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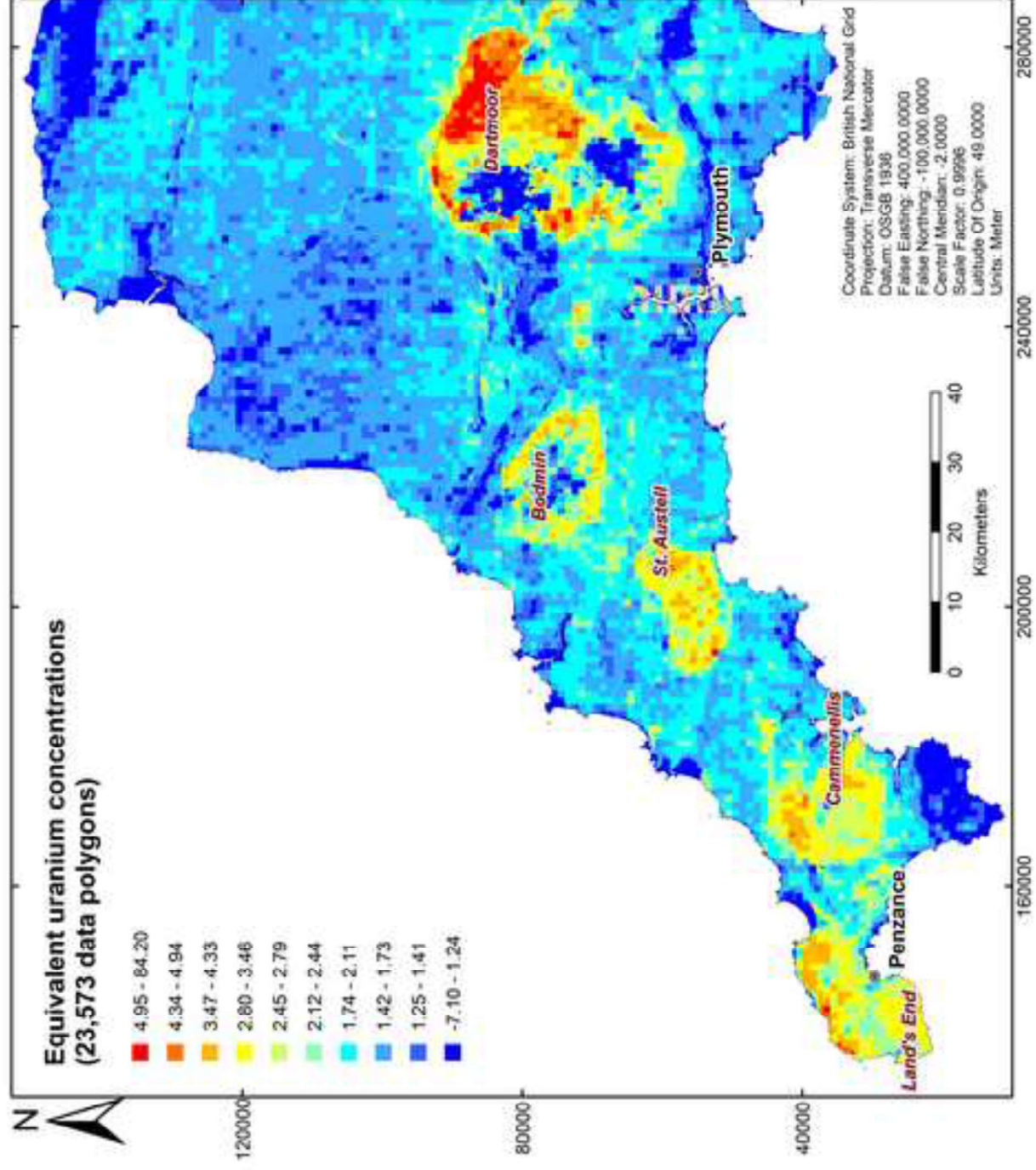


Figure07  
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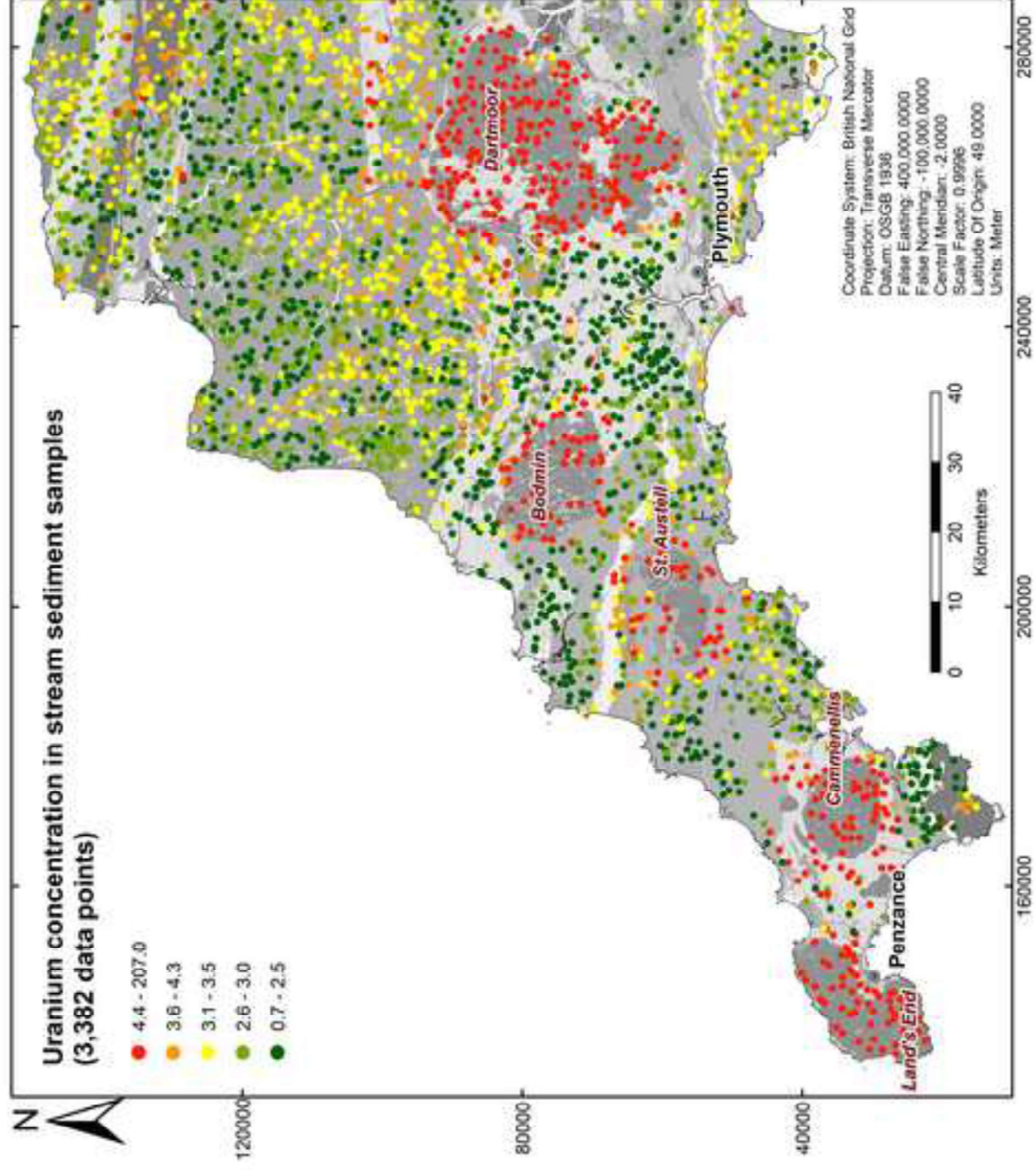


Figure08  
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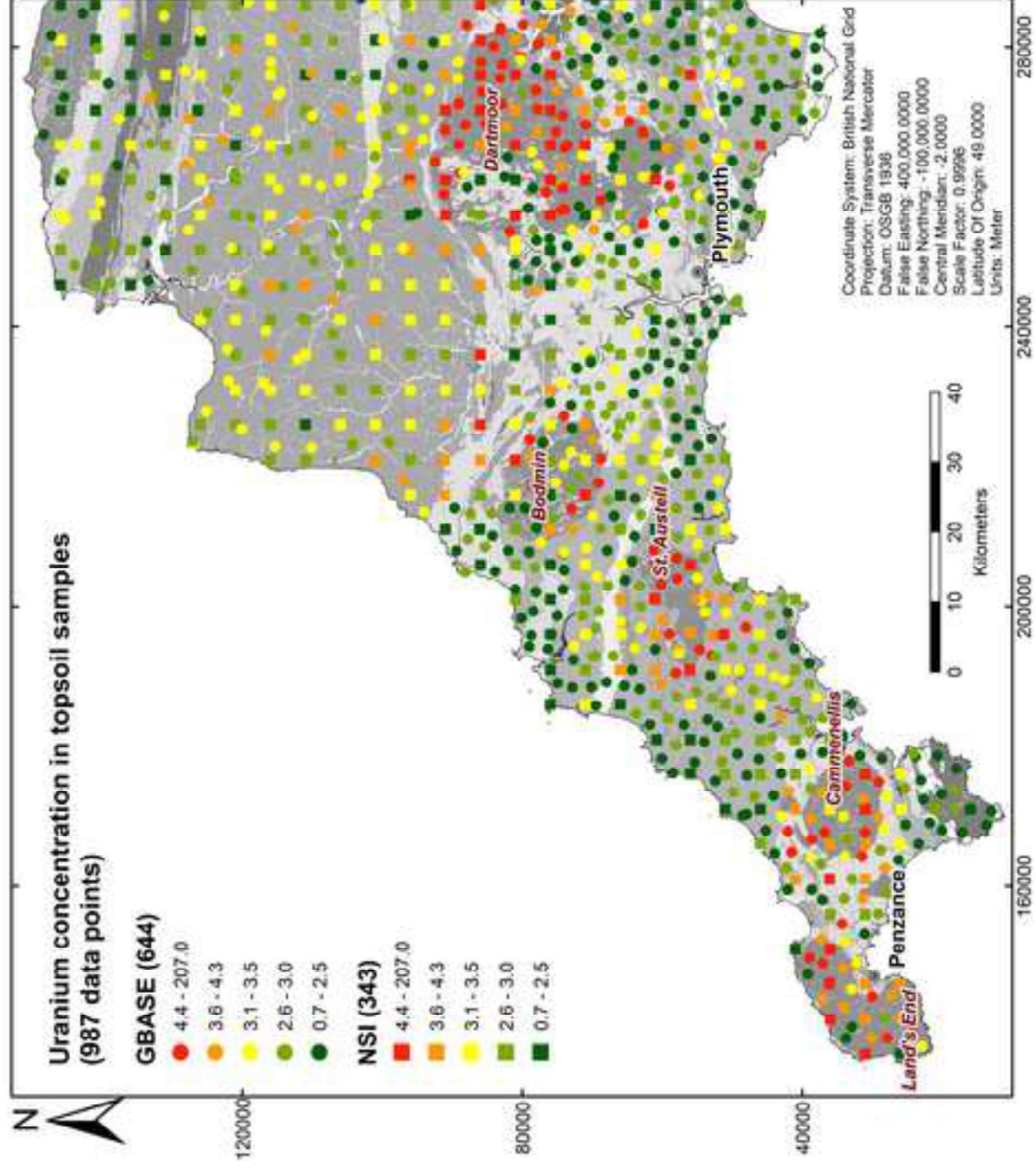


Figure09

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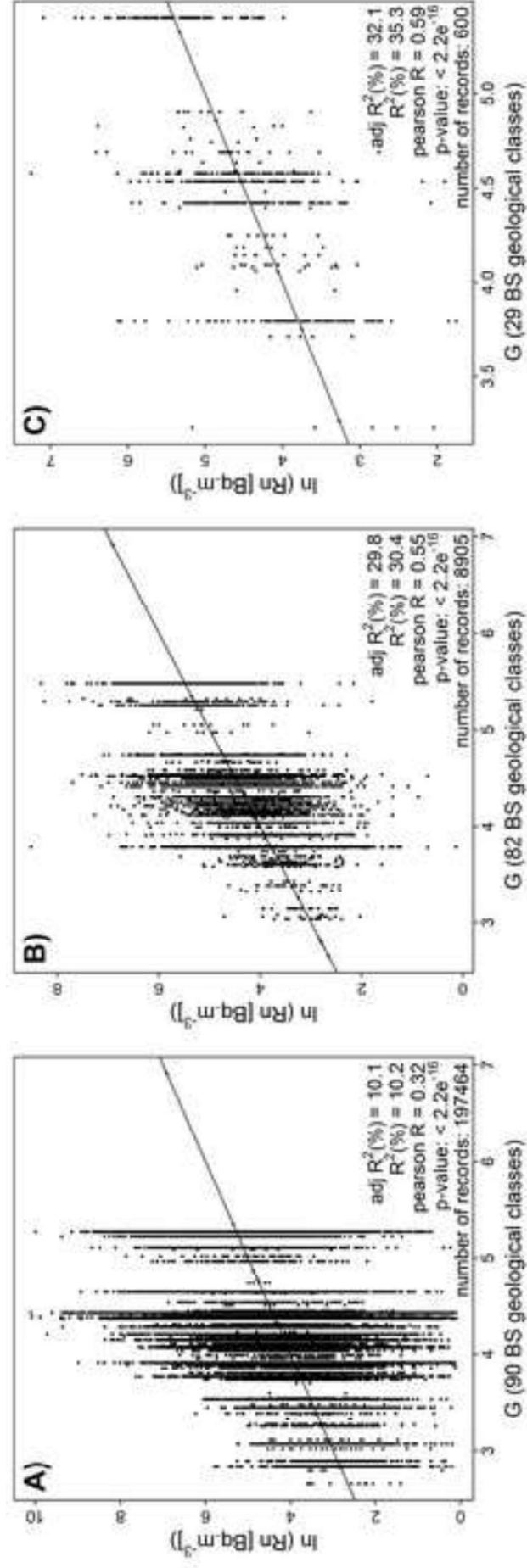




Figure 10

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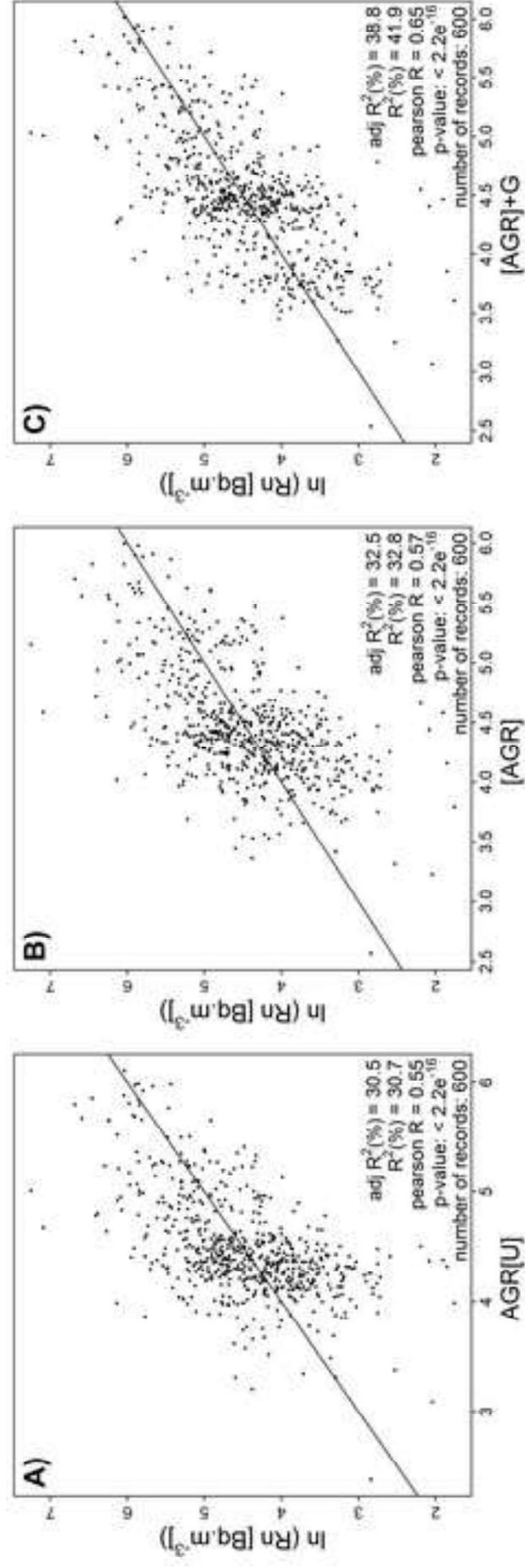


Figure 11

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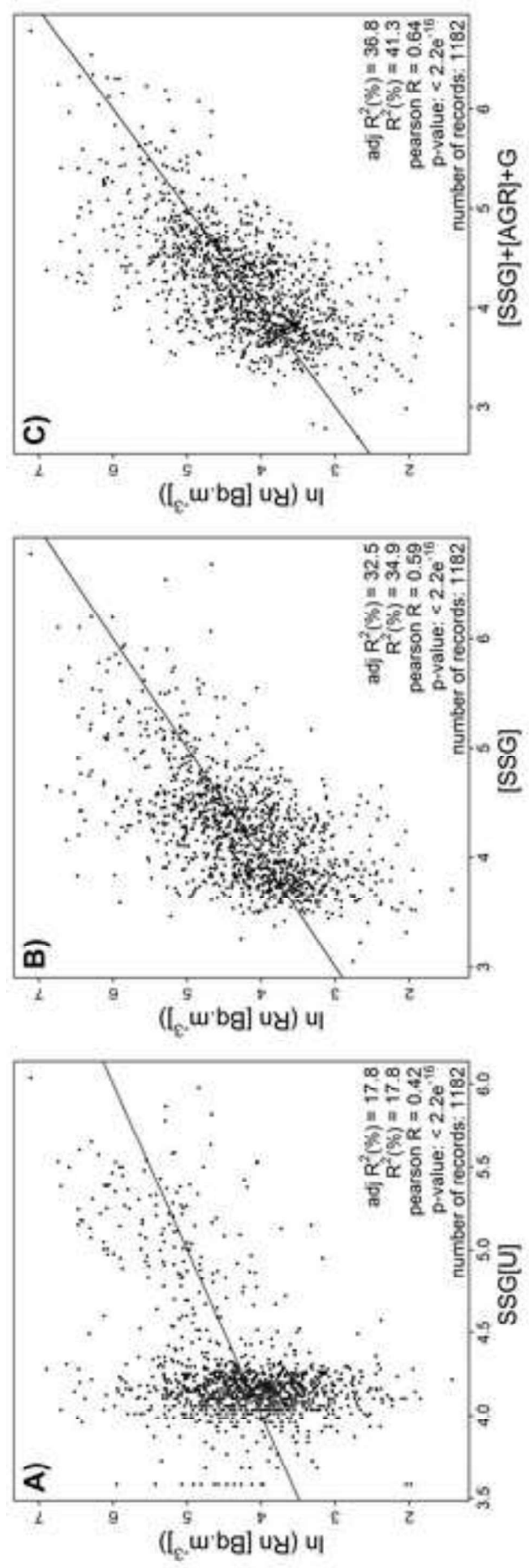


Figure12  
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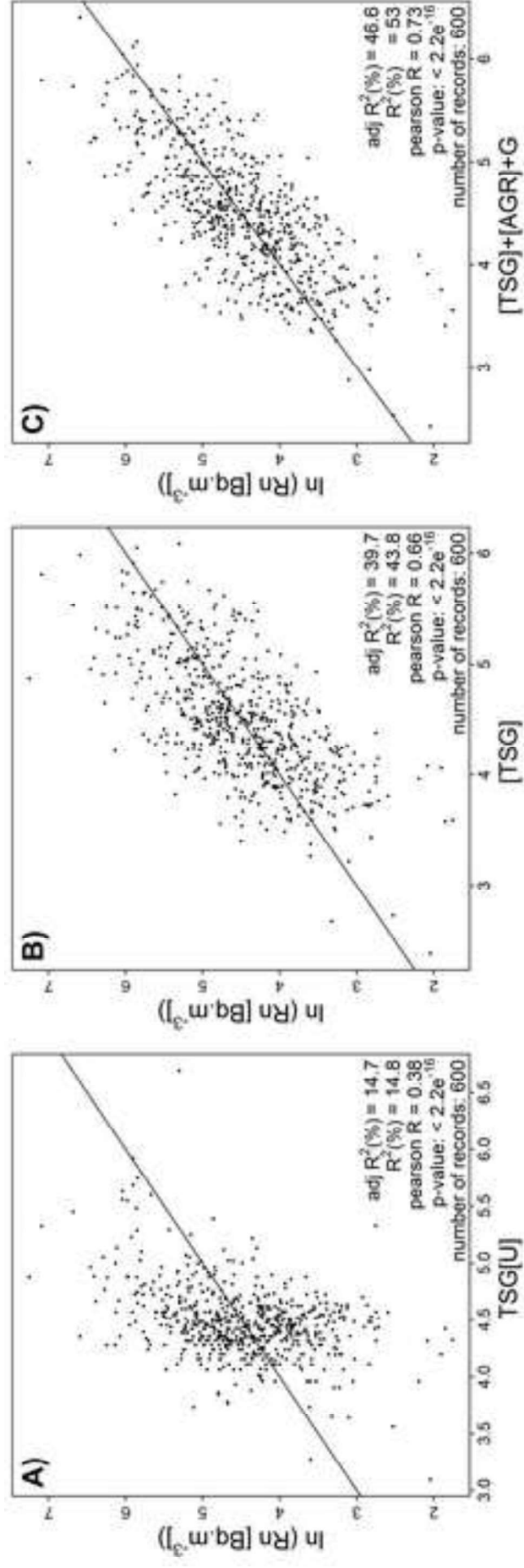


Figure13  
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