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This is the peer reviewed version of the following article:

Lutz, James A.; Furniss, Tucker J.; Johnson, Daniel J.; Davies, Stuart J.; Allen, David; Alonso, Alfonso; Anderson-Teixeira, Kristina J.; Andrade, Ana; Baltzer, Jennifer; Becker, Kendall M.L.; Blomdahl, Erika M.; Bourg, Norman A.; Bunyavejchewin, Sarayudh; Burslem, David F.R.P.; Cansler, C. Alina; Cao, Ke; Cao, Min; Cárdenas, Dairon; Chang, Li-Wan; Chao, Kuo-Jung; Chao, Wei-Chun; Chiang, Jyh-Min; Chu, Chengjin; Chuyong, George B.; Clay, Keith; Condit, Richard; Cordell, Susan; Dattaraja, Handanakere S.; Duque, Alvaro; Ewango, Corneille E.N.; Fischer, Gunter A.; Fletcher, Christine; Freund, James A.; Giardina, Christian; Germain, Sara J.; Gilbert, Gregory S.; Hao, Zhanqing; Hart, Terese; Hau, Billy C.H.; He, Fangliang; Hector, Andrew; Howe, Robert W.: Hsieh, Chang-Fu: Hu, Yue-Hua: Hubbell, Stephen P.: Inman-Narahari, Faith M.; Itoh, Akira; Janík, David; Kassim, Abdul Rahman; Kenfack, David; Korte, Lisa; Král, Kamil; Larson, Andrew J.; Li, YiDe; Lin, Yiching; Liu, Shirong; Lum, Shawn; Ma, Keping; Makana, Jean-Remy; Malhi, Yadvinder; McMahon, Sean M.; McShea, William J.; Memiaghe, Hervé R.; Mi, Xiangcheng; Morecroft, Michael; Musili, Paul M.; Myers, Jonathan A.; Novotny, Vojtech; de Oliveira, Alexandre; Ong, Perry; Orwig, David A.; Ostertag, Rebecca; Parker, Geoffrey G.; Patankar, Rajit; Phillips, Richard P.; Reynolds, Glen; Sack, Lawren; Song, Guo-Zhang M.; Su, Sheng-Hsin; Sukumar, Raman; Sun, I-Fang; Suresh, Hebbalalu S.; Swanson, Mark E.; Tan, Sylvester; Thomas, Duncan W.; Thompson, Jill; Uriarte, Maria; Valencia, Renato; Vicentini, Alberto; Vrška, Tomáš; Wang, Xugao; Weiblen, George D.; Wolf, Amy; Wu, Shu-Hui; Xu, Han; Yamakura, Takuo; Yap, Sandra; Zimmerman, Jess K. 2018. Global importance of large-diameter trees. Global Ecology and Biogeography, 27 (7). 849-864, which has been published in final form at https://doi.org/10.1111/geb.12747

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Research Paper 1

2 **Title: Global importance of large-diameter trees**

3 **Running head: Large-diameter trees**

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- Xiangcheng Mi¹⁵¹⁵, Michael Morecroft⁵³, Paul M. Musili⁵⁴, Jonathan A. Myers⁵⁵, Vojtech 17
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25 Abstract

Aim: To examine the contribution of large-diameter trees to biomass, stand structure, and
 species richness across forest biomes.

28 **Location:** Global

29 Methods: We examined the contribution of large trees to forest density, richness, and biomass

30 using a global network of 48 large (from 2 ha to 60 ha) forest plots representing 5,601,473 stems

31 across 9,298 species and 210 plant families. This contribution was assessed using three metrics:

32 the largest 1% of trees \geq 1 cm diameter at breast height (DBH), all trees \geq 60 cm DBH, and those

33 rank-ordered largest trees that cumulatively comprise 50% of forest biomass.

34 **Results:** Averaged across these 48 forest plots, the largest 1% of trees ≥ 1 cm DBH comprised

35 50% of above ground live biomass, with hectare-scale standard deviation of 26%. Trees \geq 60 cm

36 DBH comprised 41% of aboveground live tree biomass. The size of the largest trees correlated

37 with total forest biomass ($r^2 = 0.62$, *P*<0.001). Large-diameter trees in high biomass forests

represented far fewer species relative to overall forest richness ($r^2 = 0.45$, *P*<0.001). Forests with

39 more diverse large-diameter tree communities were comprised of smaller trees ($r^2 = 0.33$,

40 P<0.001). Lower large-diameter richness was associated with large-diameter trees being

41 individuals of more common species ($r^2 = 0.17$, *P*=0.002). The concentration of biomass in the

42 largest 1% of trees declined with increasing absolute latitude ($r^2 = 0.46$, P<0.001), as did forest

43 density ($r^2 = 0.31$, *P*<0.001). Forest structural complexity increased with increasing absolute 44 latitude ($r^2 = 0.26$, *P*<0.001).

45 **Main conclusions:** Because large-diameter trees constitute roughly half of the mature forest

46 biomass worldwide, their dynamics and sensitivities to environmental change represent

47 potentially large controls on global forest carbon cycling. We recommend managing forests for

- 48 conservation of existing large-diameter trees or those that can soon reach large diameters as a
- 49 simple way to conserve and potentially enhance ecosystem services.
- 50 Keywords: forest biomass, forest structure, large-diameter trees, latitudinal gradient, resource
- 51 inequality, Smithsonian ForestGEO
- 52 Word Count: 4,628
- **Tables:** 2
- 54 **Figures:** 5
- 55 **References:** 68
- 56 Data References: 36

57 Introduction

Concentration of resources within a few individuals in a community is a pervasive property of biotic systems (West et al., 1997), whether marine (Hixon et al., 2014), terrestrial (Enquist et al., 1998), or even anthropogenic (Saez & Zucman, 2016). The concentration of total forest biomass in a few large-diameter trees is no exception (Pan et al., 2013). Large-diameter trees in forests take many decades or even centuries to develop, but human or natural disturbances can decrease their abundance, rapidly changing forest structure (Lutz et al. 2009, van Mantgem et al., 2009, Allen et al. 2010, Lindenmayer et al. 2012).

65 Despite the recognised ecological significance of large-diameter trees within individual 66 forest types, relatively little is known about the distribution and abundance of large-diameter 67 trees at the global scale. Previous studies have showed that large-diameter trees comprise a large 68 fraction of the biomass of many forests (Brown et al., 1995; Clark & Clark, 1996; Lutz et al., 69 2012; Bastin et al., 2015) and that they modulate stand-level leaf area, microclimate, and water 70 use (Martin et al., 2001, Rambo & North, 2009). Large-diameter trees contribute 71 disproportionately to reproduction (van Wagtendonk & Moore, 2010), influence the rates and 72 patterns of regeneration and succession (Keeton & Franklin, 2005), limit light and water 73 available to smaller trees (Binkley et al. 2010), and contribute to rates and causes of mortality of 74 smaller individuals by crushing or injuring sub-canopy trees when their bole or branches fall to 75 the ground (Chao et al., 2009; Das et al., 2016). Large-diameter trees (and large-diameter snags 76 and large-diameter fallen woody debris) make the structure of primary forests and mature 77 secondary forests unique (Spies & Franklin, 1991). Large-diameter trees occur at low stem 78 densities, yet influence spatial patterns over long inter-tree distances (Enquist et al., 2009; Lutz 79 et al., 2014; Das et al., 2018). Consequently, to elucidate the patterns, mechanisms, and

80 consequences of large-diameter tree ecology requires sample plots ≥1 ha (Das et al., 2011;
81 Réjou-Méchain et al., 2014; Lutz, 2015).

82 Changes in climate, disturbance regimes, and logging are accelerating the decline of 83 large-diameter trees (e.g., Lindenmayer et al., 2012; Bennett et al., 2015; Lindenmayer & 84 Laurence, 2016). The dynamics of large-diameter trees is dependent on at least two factors: 1) 85 presence of species capable of attaining a large size, and 2) conditions, including disturbance 86 regimes, that permit the development of large-diameter individuals. If the species richness of the 87 large-diameter assemblage is high, a forest may be better able to respond to perturbations 88 (Musavi et al., 2017) and maintain its structure and ecological function. However, if the large-89 diameter species richness is low, then a forest could be susceptible to any change that affected 90 those few species.

91 Surprisingly, the specific roles of large-diameter trees are not well anchored in two 92 widely referenced theories of global vegetation. Both the Unified Neutral Theory of Biodiversity 93 (Hubbell, 2001) and Metabolic Scaling Theory (West et al., 2009) propose that plants have a 94 degree of functional equivalency. The Unified Neutral Theory makes predictions about the rank-95 order abundance of species in a forest, but it makes no specific predictions about the rank order 96 of large-diameter species or even if large-diameter individuals are members of common or rare 97 species. Metabolic Scaling Theory does predict the abundance of large-diameter trees, and 98 empirical tests of the theory for more abundant, smaller-diameter individuals are generally good. 99 However, Metabolic Scaling Theory often tends to under-predict the abundance of large-100 diameter trees in temperate forests (Anderson-Teixeira et al. 2015b; their Fig. 2) and rather over-101 predict the abundance of large-diameter trees in tropical forests (Muller-Landau et al. 2006; their 102 Table 2) and in some temperate forests (Lutz et al. 2012; their Fig. 2). Metabolic Scaling Theory

also advances its predictions as continuous functions, and the departure from theory (i.e., the
spatial variation) at discrete grain sizes remains unquantified. Accordingly, those theories alone
cannot fully explain global patterns of forest species diversity or the larger portion of the size
distribution (Coomes et al., 2003; Muller-Landau et al., 2006; Lutz et al., 2012; LaManna et al.,
2017).

108 However, studies do suggest that a greater generalization of forest structure in the 109 tropical, subtropical, temperate, and boreal forests of the world may indeed be possible (i.e., 110 Gilbert et al., 2010; Slik et al., 2013; Ostertag et al., 2014). To the extent that forests share 111 structural attributes either globally or regionally, our ability to model forest change may be 112 improved by focusing on global patterns in structure rather than individual species life-history 113 traits. We expected that latitudinal trends in the concentration of biomass in the largest trees 114 would follow trends in forest density (with more stems in the largest diameter classes, relative 115 biomass should be higher). We also expected that relative richness of the large-diameter cohort 116 would be lower in forests with high stem density because the large trees would be a smaller 117 fraction of stems and thus a smaller fraction of species. Our principal hypothesis was that only a 118 small proportion of the largest trees are responsible for the preponderance of forest biomass, and 119 that the abundance and variation of these large-diameter trees reflect latitudinal gradients of 120 forest structure. Specifically we set out to ask four interrelated questions:

121

122

 Are there global relationships between large-diameter trees (defined various ways) and forest biomass?

123 2) Does the richness of the large-diameter cohort depend on the richness or biomass of124 the forest?

- 3) Are there latitudinal gradients in forest density, biomass, concentration of biomass, orstructural complexity?
- 127

4) Are large-diameter trees members of common or rare species in forests?

128 Materials and Methods

129 We used data from the Forest Global Earth Observatory (ForestGEO; Anderson-Teixeira et al., 130 2015a) network of forest dynamics plots coordinated by the Smithsonian Institution, which 131 includes major forest types in the Köppen climate zones of cold, temperate, and tropical forests 132 (Fig. 1, Table S3.1). Forests included in the ForestGEO network include undisturbed primary 133 forests or older secondary forests meeting the United Nations Food and Agricultural 134 Organization definition of forest (trees >5 m tall and canopy cover >10% occurring in patches 135 >0.5 ha; Forest Resource Assessment 2015). The ForestGEO plots feature consistent field 136 methods (Condit, 1998) and data representation (Condit et al., 2014). Importantly, these plots 137 include all standing woody stems ≥ 1 cm diameter at breast height (1.3 m along the main stem; 138 DBH). A representativeness analysis showed that the ForestGEO includes most major forest 139 types of the world, albeit with some exceptions (see Anderson-Teixeira et al., 2015a for details). 140 We analyzed 48 plots in primary or older secondary forest spanning 86.4° of latitude (Fig. 1), 141 covering 1,278 ha (median size 24 ha), and including 5,601,473 stems representing 9,298 species 142 and 210 plant families (Fig. 1, Tables 1, S3.1).

There is no universal definition for what constitutes a large-diameter tree. Generally, a large-diameter tree is of reproductive stature, is tall enough to reach the upper canopy layer of the forest, and that is larger than the majority of woody stems in the forest. In any forest, the largest trees *relative to the rest of the stand* contribute disproportionately to ecological function and represent some of the longest-lived and most fecund components of their respective forests.

148	The definition of large-diameter inherently depends on species and forest type. In cold,
149	continental forests, a large-diameter tree may only be 20 cm DBH (Baltzer et al., 2014). In
150	productive temperate or tropical forests, a large-diameter tree may be >100 cm DBH (Lutz et al.,
151	2012; 2013). To compare dissimilar ecosystems, we used three metrics for defining large
152	diameter trees:
153	1) 99 th percentile diameter (the largest 1% of trees ≥ 1 cm DBH in the forest).
154	2) Fixed diameter. We used a fixed threshold for large-diameter trees of 60 cm DBH, a
155	diameter reached by at least some trees in almost all plots.
156	3) The large-diameter threshold. We defined the large-diameter threshold to be that diameter
157	such that trees greater than or equal to that diameter constituted half of the aboveground
158	live biomass of the plot.
159	We calculated the density, basal area, and biomass of stems ≥ 1 cm DBH and tabulated
160	them within each square hectare (100 m \times 100 m) of the 48 plots. Because the distribution of
161	large-diameter trees within forests is often not homogeneous (e.g., Lutz et al. 2013), we used the
162	one-hectare scale to capture variation in structure across the plots without introducing the
163	spurious high or low values of biomass that could be associated with small extents (Réjou-
164	Méchain et al., 2014). We calculated biomass for tropical forests (absolute latitude $\leq 23.5^{\circ}$) by
165	the methods of Chave et al. (2014), which uses a generic equation to predict biomass based on
166	diameter, climate, and wood density. The Chave et al. (2014) equations are of the form:

167
$$AGB = exp[-1.803 - 0.976E + 0.976\ln(\rho) + 2.676\ln(DBH) - 0.0299\ln(DBH)^2]$$
 Eq. 1

168 where ρ is wood density and *E* is the environmental parameter. Wood specific gravity was taken

169 from Zanne et al. (2009), and we used the values hierarchically, taking species-specific values

where defined, then genus-specific values, then family-specific values. If there was no wood specific gravity data for the plant family, or if the stem was unidentified, we used the global average of 0.615 g cm⁻³. Values for the environmental parameter *E* are listed in Table S3.1.

173 We calculated biomass for cold and temperate plots (absolute latitude >23.5°) using the 174 composite taxa-specific equations of Chojnacky *et al.* (2014). Those equations are of the form

175
$$\ln(biomass) = \beta_0 + \beta_1 \times \ln(DBH)$$
 Eq. 2

176 where β_0 and β_1 are listed in Chojnacky *et al.* (2014; their Table 5).

Species not represented by specific biomass equations were defaulted to an equation or
wood density value for the genus or the family. We used site-specific allometric equations for
Palamanui (Ostertag et al., 2014), Laupahoehoe (Ostertag et al., 2014), Lanjenchi (Aiba &
Nakashizuka, 2009), and Changbaishan (Wang, 2006).

181 We further analyzed the diameter-abundance relationships of each plot based on six tree 182 diameter classes (1 cm \leq DBH < 5 cm, 5 cm \leq DBH < 10 cm, 10 cm \leq DBH < 30 cm, 30 cm \leq 183 $DBH < 60 \text{ cm}, 60 \text{ cm} \le DBH < 90 \text{ cm}, \text{ and } DBH \ge 90 \text{ cm}$). Diameter classes were selected to 184 include recognised differences in tree life-history traits (Memiaghe et al., 2016). We performed 185 nonmetric multidimensional scaling (NMDS; Kenkel & Orloci, 1986) analyses on the density of 186 each diameter class of each 100 m \times 100 m area. We used the Bray-Curtis dissimilarity index 187 and performed the NMDS ordinations in three dimensions using the version 2.4-4 of the vegan 188 package (Oksanen et al. 2016) in R version 3.3.1 (R Development Core Team, 2016). We used 189 the three-dimensional coordinates of each 1-ha in NMDS space to create a metric for structural 190 complexity. For the 1-ha structural ordination values for each plot, we fit a one standard 191 deviation ellipsoid using the orglellipse function from the vegan3d package (Oksanen, 2017). We

then calculated the volume of that ellipsoid as a metric of structural difference (i.e., complexity) to compare the relative differences between $100 \text{ m} \times 100 \text{ m}$ areas within the plot.

194 To examine commonness of species that can reach large diameters, we ranked all species 195 according to their abundance within each plot. We then identified large-diameter species as 196 species which had ≥ 1 individual with a DBH greater than or equal to the large-diameter 197 threshold, and determined the species rank for each of these large-diameter species (i.e., if the 198 third most abundant species was a 'large-diameter species', it would receive rank = 3). We then 199 used the median rank for all large-diameter species ranks within each plot, and normalised this 200 value across plots by dividing rank by the total number of species (i.e., in a plot with 60 species, 201 a median rank of 18 becomes 0.3).

202 To validate our results, we calculated structural accumulation curves for each plot, 203 calculating the area required to estimate forest density and aboveground live biomass to within 204 5% of the entire plot value. Within each plot, for each of density and biomass, we used random sampling of 400 m² quadrats with replacement (from the available quadrats), beginning with a 205 206 random sample of n = 1 quadrat and ending with a random sample of n = total number of 207 quadrats in each plot. This process was repeated based on the number of quadrats in each plot 208 which allowed us to calculate a mean and standard deviation for each value of n. A percent 209 deviation metric was calculated as:

210
$$Percent difference = (abs(mean_n - mean_{plot}) + sd_n)/mean_{plot}$$
 Eq. 3

where mean_n is the mean of a random sampling of n quadrats, mean_{plot} is the mean for the entire plot, and sd_n is the standard deviation for the random sample of n quadrats.

213 **Results**

214	Average stem density in the plots ranged from 608 stems ha ⁻¹ (Mudumalai, India) to 12,075
215	stems ha ⁻¹ (Lanjenchi, Taiwan) with most high-density plots occurring in the tropics (Tables 1,
216	2). Aboveground live tree biomass ranged from 13 Mg ha ⁻¹ (Mpala, Kenya) to 559 Mg ha ⁻¹
217	(Yosemite, USA). The biomass of trees ≥ 60 cm DBH ranged from 0 Mg ha ⁻¹ (Mpala, Kenya,
218	Palamanui, USA, and Scotty Creek, Canada) to 447 Mg ha ⁻¹ (Yosemite, USA). The large-
219	diameter tree threshold (separating the plot aboveground forest biomass into two equal parts)
220	varied from 2.5 cm (Palamanui, USA) to 106.5 cm (Yosemite, USA). Variation in the abundance
221	of trees of different diameter classes at the 1-ha scale was high globally (Tables S3.2, S3.3), and
222	CV of the 1-ha stem densities was highest in the cold temperate / boreal plots and lowest in the
223	tropics (Table 2).

224 There was a strong positive relationship between the large-diameter threshold and overall forest biomass ($r^2 = 0.62$, P<0.001; Fig. 2A). This relationship held for all three of our 225 226 definitions for large diameter trees (Fig. 2A-C). The relationship for large-diameter threshold 227 was strongest, but the biomass of the largest 1% of trees also predicted total biomass ($r^2 = 0.35$, P < 0.001; Fig. 2B) as did the density of stems ≥ 60 cm DBH ($r^2 = 0.49$, P < 0.001; Fig. 2C). 228 229 Results based on basal area were similar to those for biomass (Fig. S1.1). There was a negative 230 relationship between large-diameter species richness and total biomass ($r^2 = 0.45$, P<0.001; Fig. 231 2D) which was consistent with the negative relationship between large-diameter threshold and large-diameter richness ($r^2 = 0.33$, P<0.001; Fig. 2E) and the negative relationship between 232 large-diameter richness and the biomass of the largest 1% of trees ($r^2 = 0.61$, P<0.001; Fig. 2F). 233 234 In other words, plots with high biomass had high large-diameter thresholds and relatively low 235 species richness within this large-diameter structural class.

236 The amount of aboveground forest biomass contained within the largest 1% of trees 237 averaged among the 48 plots was 50% (weighted by the forest biomass of each plot, 45% as an 238 unweighted average of the 48 plots), representing an average of 23% of the total species richness 239 (Table 1). The average large-diameter threshold was 47.7 cm DBH (half of the biomass of the 48 240 plots was contained within trees \geq 47.7 cm DBH). The average portion of biomass contained 241 within trees ≥ 60 cm DBH in the 48 plots was 41%. Forest density gradually decreased with increasing absolute latitude ($r^2 = 0.31$, P<0.001; Fig. 3A), as did the proportion of tree biomass 242 accounted for by the largest 1% of trees ($r^2 = 0.46$, P < 0.001; Fig. 3C), following our 243 expectations and partially a reflection of the higher stem densities in the tropics (Fig. 3A, Tables 244 245 1, S3.2). However, latitudinal gradients were not present for biomass (Fig. 3B) or the large-246 diameter threshold (Fig. 3D).

The three metrics for large-diameter trees were not perfectly correlated (Fig. S1.2). The large-diameter threshold and the density of stems ≥ 60 cm DBH had a linear relationship (r² = 0.80, P<0.001), even though some forests did not have trees ≥ 60 cm DBH. The relationship between the biomass of the 1% of largest diameter trees and both the density of stems ≥ 60 cm DBH and the large-diameter threshold was significant for tropical plots but not for temperate plots.

NMDS ordinations of the abundance of trees in the six diameter classes in each 100 m × 100 m area showed that tropical forests have a higher degree of structural similarity based on their positions in the ordination (Fig. 4A, B). The 1-ha scale variation for tropical plots also showed a high degree of similarity both globally (clustering and high overlap of red ellipses in Fig. 4C, D) and locally (smaller size of individual red ellipses). The volumes occupied by the 1ha NMDS points of temperate plots, conversely, covered a wide range in ordination space,

259 indicating greater structural variability both among and within the plots (greater size and 260 dispersion of green ellipses in Fig. 4C, D, three-dimensional animation in S2). This phenomenon 261 was also mirrored by coefficients of variation of density and biomass of 1-ha quadrats, which 262 differed among regions and were higher in temperate and boreal forests than in tropical plots 263 (Table 2). The grouping of plots with no trees ≥ 60 cm DBH (left of Figs. 4A, B; Table S3.2) 264 shows a structural equivalency of forests growing in stressful environments. Those forests 265 include Scotty Creek, Canada (temperature, nitrogen, and hydrologically limited), Mpala, Kenya (water and herbivory limited) and Palamanui, USA (water limited, limited soil development, and 266 267 with limited species complement). The structural complexity of forests (variation in abundance 268 of the six diameter classes) at 1-ha scale increased with increasing absolute latitude (Fig. 5A).

269 Large-diameter trees consisted primarily of common species (rank <0.5; Fig. 5B), and rarer species reached large diameter in plots with higher large-diameter richness ($r^2 = 0.17$; P =270 271 0.002). The absolute numbers of species that reached the local large-diameter threshold ranged 272 from two in tropical Laupahoehoe, USA to 343 in Yasuni, Ecuador (Table 1). Tropical plots 273 generally had >25 species reaching the large-diameter threshold (minimum nine species in 274 Cocoli, Panama). Temperate plots generally had <10 species that reached the large-diameter 275 threshold (maximum 25 species in SERC, USA). On a percentage basis, large-diameter richness 276 ranged from 5% (Cocoli, Panama and Bukit Timah, Singapore) to 69% (Palamanui, USA). The 277 relative richness of the large-diameter assemblage was highest in plots with low biomass, while 278 plots with high biomass had a lower proportion of richness represented by the large-diameter 279 trees (Fig. 2D, Table 1). In general, forests with lower total richness had a higher proportion of 280 that richness retained in the large-diameter class. Unsurprisingly, plots with lower large-diameter 281 thresholds (<60 cm DBH) had a higher proportion of species represented in the large-diameter

assemblage (mean 34%), whereas plots with large-diameter thresholds ≥60 cm DBH had a lower
proportion of species represented in the large-diameter guild (mean 18%).

284 **Discussion**

285 The relationship between the large-diameter threshold and overall biomass (Fig. 2A) suggests 286 that forests cannot sequester large amounts of aboveground carbon without large trees, 287 irrespective of the richness or density of large-diameter trees. Species capable of attaining large 288 diameters are relatively few (Fig. 2) but individuals of these species are relatively abundant (Fig. 289 5B). The relationships among biomass and richness across plots held over a range of stem densities (608 stems ha⁻¹ to 12,075 stems ha⁻¹) and among trees of varying wood densities (0.10 g 290 cm⁻³ to 1.08 g cm⁻³). A linear relation of biomass to large-diameter threshold (Fig. 2A) best 291 292 explained the correlation among the 48 plots, although we would expect an upper limit based on 293 maximum tree heights (Koch et al., 2004) or biomass (Sillett et al., 2015; Van Pelt et al., 2016). 294 The generally high proportion of biomass represented by the largest 1% of trees reinforces the 295 importance of these individuals to carbon sequestration and productivity (e.g., Stephenson et al., 296 2014). Larger numbers of small and medium-diameter trees cannot provide equivalent biomass 297 to a few large-diameter trees, although small and medium sized trees can contribute significantly 298 to carbon cycling (Fauset et al., 2015; Meakem et al., 2017). The implication from scaling theory 299 (West et al., 2009) is that large-diameter trees are taller and have heavier crowns, and occupy 300 growing space not available to smaller trees (i.e., at the top of the canopy; West et al., 2009; Van 301 Pelt et al., 2016).

302 Temperate forests featured a higher density of trees ≥60 cm DBH (Table 1), consistent
303 with the presence of the very largest species of trees in cool, temperate forests (Sillett et al.,
304 2015; Van Pelt et al., 2016). Temperate forests also exhibited considerably lower densities of

305 small trees (e.g., $1 \text{ cm} \le \text{DBH} \le 5 \text{ cm}$; Table S3.2) and lower total stem density, which is mainly 306 comprised of small trees than tropical forests (Tables 2, S3.2). Metabolic Scaling Theory does 307 predict the diameter-abundance relationship throughout much of the middle of the diameter 308 range in many forest types (Muller-Landau et al., 2006; Lutz et al., 2012; Anderson-Teixeira et 309 al. 2015b). However, the dichotomy between temperate forests and tropical forests, where 310 temperate forests have lower densities of small trees and higher densities of large trees (and 311 tropical forests the reverse), reinforces the need to examine departures from the theory's 312 predictions. In tropical forests, the lower proportional richness of large-diameter trees likely has 313 at least two explanations. First, tropical forests contain many more stems per ha (Table S3.2) 314 with much higher small-diameter understory diversity (LaFrankie et al., 2006). Secondly, not all 315 of the species capable of reaching large-diameters in that region may be present even in the large 316 ForestGEO plots, and thus even the extensive ForestGEO network may have sampling 317 limitations.

318 The grouping of plots with only small-diameter trees (Fig. 4A) shows that forests in 319 markedly different environments can exhibit convergent structure based on different limiting 320 factors. Large-diameter trees can be abundant in any region (Table S3.1), but different factors 321 may limit the ability of an ecosystem to support a high level of aboveground live biomass. In 322 addition to environmental limits, ecosystems that are environmentally quite productive in terms 323 of annual growth may be limited by frequent, severe disturbance (e.g., typhoons in Fushan, 324 Taiwan and hurricanes in Luquillo). Finally, the regional species pool may not contain species 325 that can attain large diameters in the local combination of climate and resource availability (e.g., 326 Palamanui, USA). The higher levels of structural complexity at 1-ha scales in temperate forests 327 may be due to higher proportions of the forests where small trees predominate and large-

diameter trees are generally excluded (i.e., swamps, rocky outcrops), supported by the higher
coefficient of variation of density in temperate and cold forests (Table 2). The trend of increasing
structural complexity (i.e., 1-ha heterogeneity) with increasing absolute latitude (Fig. 5A) may in
fact be hump-shaped, with decreasing complexity at higher latitudes than the 61.3°N of the
Scotty Creek, Canada plot.

333 There is still considerable uncertainty as to what will happen to large-diameter trees in 334 the Anthropocene when so much forest is being felled for timber and farming, or is being 335 affected by climate change. Bennett et al. (2015) suggested that the current large-diameter trees 336 are more susceptible to drought mortality. Larger trees, because of their height, are susceptible to 337 sapwood cavitation and are also exposed to high radiation loads (Allen et al., 2010; Allen et al., 338 2015), but vigorous large-diameter individuals may also still be sequestering more carbon than 339 smaller trees (Stephenson et al., 2014). Both Allen et al. (2015) and Bennett et al. (2015) 340 suggested that larger trees will be more vulnerable to increasing drought than small trees, and 341 Luo & Chen (2013) suggested that although the rate of mortality of larger trees will continue to 342 increase because of global climate change, smaller trees will experience more drought related 343 mortality. These last two conclusions need not be in conflict as the background mortality rates 344 for smaller trees are higher than those of larger trees within mature and old-growth forests 345 (Larson and Franklin 2010). What remains generally unanswered is whether the increasing 346 mortality rates of large-diameter trees will eventually be offset by regrowth of different 347 individuals of those same (or functionally similar) species. Any reduction in temperate zone 348 large-diameter tree abundance may be compounded by the low large-diameter tree diversity in 349 temperate forests (temperate forests had high relative large-diameter richness, but low absolute 350 large-diameter richness). Large-diameter tree richness in tropical forests suggest more resilience

to projected climate warming in two ways. First, absolute large-diameter tree richness was
highest in tropical forests, suggesting that the large-diameter tree guild may have different
adaptations that will allow at least some species to persist (Musavi et al. 2017). Secondly, the
pool of species that can reach large diameters may have been undersampled in the plots used
here, implying an even higher level of richness may exist in some forests than captured in these
analyses.

The finding that large-diameter trees are members of common species groups (Fig. 5B) contradicts the Neutral Theory assumption of functional equivalency (Hubbell, 2001). Similarly the different structural complexity of forests worldwide (Fig. 5A) contradicts the assumptions of universal size-abundance relationships of Metabolic Scaling Theory (Enquist et al., 1998, 2009). The presence of a latitudinal gradient in forest density (Fig. 3A) and the lack of a latitudinal gradient in forest biomass (Fig. 3B) suggests that size-abundance relationships are not universal but depend on region or site conditions (Table 2).

364 Characterizing forest structural variation did require these large plots (Fig. S1.3), a 365 finding consistent with other studies examining forest biomass (Réjou-Méchain et al., 2014). 366 With large plot sizes and global distribution, ForestGEO is uniquely suited to capture structural 367 variation (i.e., the heterogeneity in the abundance of trees of all diameter classes). The relatively 368 large area required (6.5 ha, on average) to estimate biomass to within 5% of the entire plot value 369 reinforces conclusions that the distribution of large-diameter trees is not homogeneous within 370 forests (e.g., Table 2; Lutz et al., 2012; 2013; Furniss et al., 2017. We note that this calculation of 371 the size of the plot required is a measure of spatial variation within the forest, and does not 372 depend on the accuracy of the allometric equations for calculating tree biomass themselves. 373 Allometric equations can be imprecise for large-diameter trees, both because of their structural

variability and the enormous sampling effort, and therefore our estimates of overall biomass could be off by $\pm 15\%$ (Lutz et al., 2016).

376 Although temperate plots had much lower overall species diversity compared to the 377 tropical plots, tropical plots had much more homogeneous structure, both within- and across-378 plots (Fig. 4), potentially suggesting greater structural equivalency among the many species 379 present. We found that the largest 1% of trees constitute 50% of the biomass (and hence, carbon) 380 supporting our hypothesis of their significance, at least in primary forests or older secondary 381 forests. The conservation of large-diameter trees in tropical and temperate forests is therefore 382 imperative to maintain full ecosystem function, as the time necessary for individual trees to 383 develop large sizes could preclude restoration of full ecosystem function for centuries following 384 the loss of the oldest and largest trees (Lindenmayer et al. 2012). Clearly, these large-diameter 385 trees that are now absent from areas that have been recently logged lack the structural 386 heterogeneity of older forests. That the largest individuals belong to relatively few common 387 species in the temperate zone means that the loss of large-diameter trees could alter forest 388 function - if species that can attain large diameters disappear, forests will feature greatly reduced 389 structural heterogeneity (e.g., Needham et al., 2016), biomass, and carbon storage. In the tropical 390 zones, the larger absolute numbers of species reaching large diameters may buffer those forests 391 against structural changes.. Policies to conserve the tree species whose individuals can develop 392 into large, old trees (Lindenmayer et al., 2014) could promote retention of aboveground biomass 393 globally as well as maintenance of other ecosystem functions.

394 Acknowledgements

Funding for workshops during which these ideas were developed was provided by NSF grants
1545761 and 1354741 to SD Davies. This research was supported by the Utah Agricultural

- 397 Experiment Station, Utah State University, and approved as journal paper number 8998.
- 398 Acknowledgements for the global support of the thousands of people needed to establish and
- 399 maintain these 48 plots can be found in Supplementary Material S4. References to locations refer
- 400 to geographical features and not to the boundaries of any country or territory.

401 **References**

- 402 Aiba, M. & Nakashizuka, T. (2009) Architectural differences associated with adult stature and
 403 wood density in 30 temperate tree species. Functional Ecology 23(2), 265-273.
- Allen, C.D., et al. (2010) A global overview of drought and heat-induced tree mortality reveals
 emerging climate change risks for forests. Forest Ecology and Management 259, 660-684.
- Allen, C.D., Breshears, D.D. & McDowell, N.G. (2015) on underestimation of global
 vulnerability to tree mortality and forest die-off from hotter drought in the Anthropocene.
 Ecosphere 6(8): 1-55.
- Anderson-Teixeira, K.J., et al. (2015a) CTFS-ForestGEO: A worldwide network monitoring
 forests in an era of global change. Global Change Biology 21(2), 528-549.
- Anderson-Teixeira, K.J., McGarvey, J.C., Muller-Landau, H.C., Park, J.Y., Gonzalez-Akre,
 E.B., Herrmann, V., et al. (2015b) Size-related scaling of tree form and function in a mixedage forest. Functional Ecology, 29(12), 1587-1602.
- Baltzer, J.L., Venes, T., Chasmer, L.E., Sniderhan, A.E. & Quinton, W.L. (2014) Forests on
 thawing permafrost: fragmentation, edge effects, and net forest loss. Global Change Biology
 20, 824-834.
- 417 Bastin, J-F., et al. (2015) Seeing central African forests through their largest trees. Scientific
 418 Reports 5, 13156.
- Bennett, A.C., McDowell, N.G., Allen, C.D. & Anderson-Teixeira, K.J. (2015) Larger trees
 suffer most during drought in forests worldwide. Nature Plants 1(10), 15139.
- Binkley, D., Stape, J.L., Bauerle, W.L. & Ryan, M.G. (2010) Explaining growth of individual
 trees: Light interception and efficiency of light use by Eucalyptus at four sites in Brazil.
 Forest Ecology and Management 259, 1704-1713.
- Brown, I.F., et al. (1995) Uncertainty in the biomass of Amazonian forests: An example from
 Rondônia, Brazil. Forest Ecology and Management **75**, 175–189.
- 426 Chao, K-J., Phillips, O.L., Monteagudo, A., Torres-Lezama, A. & Vásquez Martínez, R. (2009)
- How do trees die? Mode of death in northern Amazonia. Journal of Vegetation Science 20,
 260–268.
- Chave, J., et al. (2014) Improved allometric models to estimate the aboveground biomass of
 tropical trees. Global Change Biology 20(10), 3177-3190.
- Chojnacky, D.C., Heath, L.S. & Jenkins, J.C. (2014) Updated generalized biomass equations for
 North American tree species. Forestry 87, 129-151.
- 433 Clark, D.B. & Clark, D.A. (1996) Abundance, growth and mortality of very large trees in
 434 Neotropical lowland rain forest. Forest Ecology and Management 80, 235–244.
- 435 Condit, R. (1998) Tropical forest census plots. Berlin: Springer-Verlag, Georgetown: R.G.
 436 Landes Company. 211 p.
- 437 Condit, R., Lao, S., Singh, A., Esufali, S. & Dolins, S. (2014) Data and database standards for 438 permanent forest plots in a global network. Forest Ecology and Management **316**, 21–31.

- Coomes, D.A., Duncan, R.P., Allen, R.B. & Truscott, J. (2003) Disturbances prevent stem sizedensity distributions in natural forests from following scaling relationships. Ecology Letters 6,
 980–989.
- 442 Das, A., Battles, J., Stephenson, N.L. & van Mantgem, P.J. (2011) The contribution of
 443 competition to tree mortality in old-growth coniferous forests. Forest Ecology and
 444 Management 261, 1203–1213.
- Das, A.J., Larson, A.J. & Lutz, J.A. (2018) Individual species-area relationships in temperate
 coniferous forests. Journal of Vegetation Science DOI: 10.1111/jvs.12611
- 447 Das, A.J., Stephenson, N.L. & Davis, K.P. (2016) Why do trees die? Characterizing the drivers
 448 of background tree mortality. Ecology 97(10), 2616-2627.
- Enquist, B.J., Brown, J.H. & West, G.B. (1998) Allometric scaling of plant energetics and
 population density. Nature 395, 163-165.
- 451 Enquist, B.J., West, G.B. & Brown, J.H. (2009) Extensions and evaluations of a general
 452 quantitative theory of forest structure and dynamics. Proc Natl Acad Sci USA 106, 7046–
 453 7051.
- 454 Fauset, S., et al. (2015) Hyperdominance in Amazonia forest carbon cycling. Nature
 455 Communications 6, 6857.
- 456 Forest Resource Assessment Working Paper 180. 2015. Terms and Definitions. Food and
 457 Agriculture Organization of the United Nations. Downloaded 12/12/2017 from
 458 http://www.fao.org
- 459 Furniss, T.J., Larson, A.J. & Lutz, J.A. (2017) Reconciling niches and neutrality in a subalpine
 460 temperate forest. Ecosphere 8(6), Article01847.
- 461 Gilbert, G.S., et al. (2010) Beyond the tropics: forest structure in a temperate forest mapped plot.
 462 Journal of Vegetation Science 21(2), 388-405.
- Hixon, M.A., Johnson, D.W. & Sogard, S.M. (2014) BOFFFFs: on the importance of conserving
 old-growth structure in fishery populations. ICES Journal of Marine Science 71(8), 21712185.
- Hubbell, S.P. (2001) The Unified Neutral Theory of Biodiversity and Biogeography. Princeton
 University Press. 375 pp.
- Keeton, W.S. & Franklin, J.F. (2005) Do remnant old-growth trees accelerate rates of succession
 in mature Douglas-fir forests? Ecological Monographs **75**, 103–118.
- Kenkel, N.C. & Orloci, L. (1986) Applying metric and nonmetric multidimensional scaling to
 ecological studies: some new results. Ecology 67(4), 919-928.
- 472 Koch, G.W., Sillett, S.C., Jennings, G.M. & Davis, S.D. (2004) The limits to tree height. Nature
 473 428, 851-854.
- 474 LaFrankie, J.V., et al. (2006) Contrasting structure and composition of the understory in species475 rich tropical rain forests. Ecology 87(9), 2298-2305.
- 476 LaManna, J.A., et al. (2017) Negative density dependence contributes to global patterns of plant
 477 biodiversity. Science 356, 1389-1392.
- 478 Larson, A.J. & Franklin, J.F. (2010). The tree mortality regime in temperate old-growth
 479 coniferous forests: the role of physical damage. Canadian Journal of Forest Research 40(11):
 480 2091–2103.
- 481 Lindenmayer, D.B., Laurence, W.F. & Franklin, J.F. (2012) Global decline in large old trees.
 482 Science 338, 1305–1306.
- 483 Lindenmayer, D.B. & Laurence, W.F. (2016) The ecology, distribution, conservation and
 484 management of large old trees. Biological Reviews doi: 10.1111/brv.12290

- 485 Lindenmayer, D.B., et al. (2014) New policies for old trees: averting a global crisis in a keystone
 486 ecological structure. Conservation Letters 7(1), 61-69.
- 487 Luo, Y. & Chen, H.Y.H. (2013) Observations from old forests underestimate climate change
 488 effects on tree mortality. Nature Communications 4, 1655.
- 489 Lutz, J.A. (2015) The evolution of long-term data for forestry: large temperate research plots in
 490 an era of global change. Northwest Science 89(3), 255-269.
- Lutz, J.A., Larson, A.J., Freund, J.A., Swanson, M.E. & Bible, K.J. (2013) The importance of
 large-diameter trees to forest structural heterogeneity. PLoS ONE 8(12), e82784.
- 493 Lutz, J.A., Larson, A.J., Swanson, M.E. & Freund, J.A. (2012) Ecological importance of large494 diameter trees in a temperate mixed-conifer forest. PLoS ONE 7(5), e36131.
- Lutz, J.A., et al. (2014) Spatially non-random tree mortality and ingrowth maintain equilibrium
 pattern in an old-growth *Pseudotsuga-Tsuga* forest. Ecology **95(8)**, 2047-2054.
- Lutz, J.A., Matchett, J.R., Tarnay, L.W., Smith, D.F., Becker, K.M.L., Furniss, T.J. & Brooks,
 M.L. 2017. Fire and the distribution and uncertainty of carbon sequestered as aboveground
 tree biomass in Yosemite and Sequoia & Kings Canyon National Parks. Land 6(10), 1-24.
- Lutz, J.A., van Wagtendonk, J.W. & Franklin, J.F. (2009) Twentieth-century decline of large diameter trees in Yosemite National Park, California, USA. Forest Ecology and Management
 257(11), 2296-2307.
- Martin, T.A., et al. (2001) Control of transpiration in a 220-year old *Abies amabilis* forest. Forest
 Ecology and Management 152, 211–224.
- Meakem, V., et al. (2017) Role of tree size in moist tropical forest carbon cycling and water
 deficit response. New Phytologist, doi: 10.1111/nph.14633.
- Memiaghe, H.M., Lutz, J.A., Korte, L., Alonson, A. & Kenfack, D. (2016) Ecological
 importance of small-diameter trees to the structure, diversity, and biomass of a tropical
 evergreen forest at Rabi, Gabon. PLoS ONE 11(5), e0154988.
- 510 Muller-Landau, H.C., et al. (2006) Comparing tropical forest tree size distributions with the 511 predictions of metabolic ecology and equilibrium models. Ecology Letters **9**, 589–602.
- Musavi, T., et al. (2017) Stand age and species richness dampen interannual variation of
 ecosystem-level photosynthetic capacity. Nature Ecology and Evolution 0048.
- Needham, J., Merow, C., Butt, N., Malhi, Y., Marthews, T.R., Morecroft, M. et al. (2016) Forest
 community response to invasive pathogens: the case of ash dieback in a British woodland.
 Journal of Ecology, 104(2), 315-330.
- 517 Oksanen. J, et al. (2017) vegan: Community Ecology Package. R package version 2.44.
- 518 Oksanen, J., Kindt, R. & Simpson, G.L. (2016). vegan3d: Static and Dynamic 3D Plots for the 519 'vegan' Package. R package version 1.0-1.
- Ostertag, R., Inman-Narahari, F., Cordell, S., Giardina, C.P., Sack, L. (2014) Forest structure in
 low-diversity tropical forests: A study of Hawaiian wet and dry forests. PLoS ONE 9(8),
 e103268.
- Pan, Y., Birdsley, R.A., Phillips, O.L. & Jackson, R.B. (2013) The structure, distribution, and
 biomass of the world's forests. Ann Rev Ecol Evol Sys 44, 593-622.
- R Development Core Team (2016) R: A language and environment for statistical computing. R
 Foundation for Statistical Computing, Vienna, Austria. <u>http://www.R-project.org/</u>.
- Rambo, T. & North, M. (2009) Canopy microclimate response to pattern and density of thinning
 in a Sierra Nevada forest. Forest Ecology and Management 257, 435–442.
- 529 Réjou-Méchain, M., et al. (2014) Local spatial structure of forest biomass and its consequences
- for remote sensing of carbon stocks. Biogeosciences **11**, 6827-6840.

- 531 Saez, E. & Zucman, G. (2016) Wealth inequality in the United States since 1913: evidence from 532 capitalized income tax data. The Quarterly Journal of Economics **131(2)**, 519-578.
- Sillett, S.C., Van Pelt, R., Kramer, R.D., Carroll, A.L. & Koch, G.W. (2015) Biomass and
 growth potential of *Eucalyptus regnans* up to 100 m tall. Forest Ecology and Management
 348, 78-91.
- Slik, J.W., et al. (2013) Large trees drive forest aboveground biomass variation in moist lowland
 forests across the tropics. Global Ecology and Biogeography 22(12), 1261-1271.
- Spies, T.A. & Franklin, J.F. (1991) The structure of natural young, mature and old-growth
 Douglas-fir forests in Oregon and Washington. Pages 91-109 in Wildlife and Management of
 Unmanaged Douglas-fir Forests. USDA Forest Service PNW-GTR-285, Portland, Oregon.
- 541 Stephenson, N.L, et al. (2014) Rate of tree carbon accumulation increases continually with tree 542 size. Nature 507:90-93.
- van Mantgem, P.J., et al. (2009) Widespread increase of tree mortality rates in the western
 United States. Science 323, 521-524.
- Van Pelt, R., Sillett, S.C., Kruse, W.A., Freund, J.A. & Kramer, R.D. (2016) Emergent crowns
 and light-use complementarity lead to global maximum biomass and leaf area in *Sequoia sempervirens* forests. Forest Ecology and Management **375**, 279-308.
- van Wagtendonk, J.W. & Moore, P.E. (2010) Fuel deposition rates of montane and subalpine
 conifers in the central Sierra Nevada, California, USA. Forest Ecology and Management 259,
 2122–2132.
- Wang, C. (2006) Biomass allometric equations for 10 co-occurring tree species in Chinese
 temperate forests. Forest Ecology and Management 222, 9-16.
- West, G.B., Brown, J.H. & Enquist, B.J. (1997) A general model for the origin of allometric
 scaling laws in biology. Science 276, 122-126.
- West, G.B., Enquist, B.J. & Brown, J.H. (2009) A general quantitative theory of forest structure
 and dynamics. Proc Natl Acad Sci USA 106(17), 7040-7045.
- Zanne, A.E., Lopez-Gonzalez, G, Coomes, D.A., Ilic, J., Jansen, S., Lewis, S.L., Miller, R.B.,
 Swenson, N.G., Wiemann, M.C. & Chave, J. (2009) Global wood density database.
 https://doi.org/10.5061/dryad.234 Accessed 12/31/2017.

560 Data Accessibility

- 561 Data for plots in the ForestGEO network is available through the online portal at:
- 562 <u>http://www.forestgeo.si.edu</u>

563 Biosketch

- 564 **James A. Lutz** is an Assistant Professor of Forest Ecology at Utah State University. He studies
- 565 forest ecosystems to contribute to science-based conservation and management with particular
- 566 emphasis on demography and spatial patterns of tree mortality and the effects of fire on old-
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 and demographic processes.
- 570 **The ForestGEO Network** includes the senior investigators who collaborated on this research.
- 571 The Smithsonian ForestGEO network conducts long-term, large-scale research on forests around
- 572 the world. This collaborative effort seeks to increase scientific understanding of forest
- 573 ecosystems, guide sustainable forest management and natural-resource policies, monitor the
- 574 impacts of global change, and build capacity in forest science.

575 Data References

- Allen, D., Vandermeer, J. & Perfecto, I. (2009) When are habitat patches really islands? Forest
 Ecology and Management 258, 2033–2036.
- Arias Garcia, J.C., Duque, A. & Cárdenas, D. (2009) Crecimiento Diamétrico de un bosque del nor occidente Amazónico. Revista Colombia Amazónica 2, 57–64.
- Bourg, N.A., McShea, W.J., Thompson, J.R., McGarvey, J.C. & Shen, X. (2013) Initial census,
 woody seedling, seed rain, and stand structure data for the SCBI SIGEO Large Forest
- 582 Dynamics Plot. Ecology **94(9)**, 2111-2112.
- Bunyavejchewin, S., Baker, P.J., LaFrankie, J.V. & Ashton, P.S. (2001) Stand structure of a
 seasonal dry evergreen forest at Huai Kha Khaeng Wildlife Sanctuary, western Thailand.
 Natural History Bulletin of the Siam Society 49, 89-106.
- Butt, N., et al. (2009) Initial results from establishment of a long-term broadleaf monitoring plot
 at Wytham Woods, Oxford, UK. University of Oxford Report.
- 588 Cao, M., et al. (2008) Xishuangbanna tropical seasonal rainforest dynamics plot: Tree
 589 distribution maps, diameter tables and species documentation. Yunnan Science and
 590 Technology Press, Kunming. 266 pp.
- 591 Chao, W-C., et al. (2010) Lowland rainforests in southern Taiwan and Lanyu, at the northern
 592 border of paleotropics and under the influence of monsoon wind. Plant Ecology 210, 1–17.
- 593 Chen, L., et al. (2010) Community-level consequences of density dependence and habitat
 594 association in a subtropical broad-leaved forest. Ecology Letters 13, 695-704.
- 595 Co, L., et al. (2004) Palanan Forest Dynamics Plot, Philippines. In: Tropical forest diversity and
 596 dynamism: Findings from a large-scale plot network (eds Losos E, Leigh E), pp. 574–584.
 597 University of Chicago Press, Chicago.
- 598 Condit, R., et al. (2004) Tropical forest dynamics across a rainfall gradient and the impact of an
 599 El Nino dry season. Journal of Tropical Ecology 20, 51–72.
- Georgiadis, N.J. (2011) Conserving Wildlife in African Landscapes: Kenya's Ewaso Ecosystem.
 Smithsonian Contributions to Zoology 1–123.
- Gomes, A.C.S., et al. (2013) Local plant species delimitation in a highly diverse Amazonian
 forest: do we all see the same species? Journal of Vegetation Science 24, 70-79

- Hubbell, S.P., et al. (1999) Light gap disturbances, recruitment limitation, and tree diversity in a
 neotropical forest. Science 283, 554-557.
- Kenfack, D., Thomas, D.W., Chuyong, G.B. & Condit, R. (2007). Rarity and abundance in a
 diverse African forest. Biodiversity Conservation 16, 2045-2074.
- Janík, D., et al. (2016) Tree spatial patterns of *Fagus sylvatica* expansion over 37 years. Forest
 Ecology and Management **375**, 134–145.
- Johnson, D.J., et al. (2014) Conspecific negative density-dependent mortality and the structure of
 temperate forests. Ecology 95, 2493-2503.
- LaFrankie, J.V., Davies, S.J., Wang, L.K., Lee, S.K. & Lum, S.K.Y. (2005) Forest trees of Bukit
 Timah: Population ecology in a tropical forest fragment. Simply Green, Singapore. 178 pp.
- LaManna, J.A., Walton, M.L., Turner, B.L. & Myers, J.A. (2016) Negative density dependence
 is stronger in resource-rich environments and diversifies communities when stronger for
 common but not rare species. Ecology Letters 19(6), 657-667.
- 617 Lee, H.S., et al. (2005) The 52-hectare Forest Research Plot at Lambir Hills, Sarawak, Malaysia:
- Tree distribution maps, diameter tables and species documentation. Forest Department
 Sarawak, The Arnold Arboretum-CTFS Asia Program, Smithsonian Tropical Research
 Institute, Kuching, Sarawak, Malaysia.
- Lin, Y-C., Chang, L-W., Yang, K-C., Wang, H-H. & Sun, I-F. (2011) Point patterns of tree
 distribution determined by habitat heterogeneity and dispersal limitation. Oecologia 165, 175–
- 623 184.
- Makana, J., et al. (2004) Ituri Forest Dynamics Plot, Democratic Republic of Congo. In: Tropical
 forest diversity and dynamism: Findings from a large-scale plot network (eds Losos E, Leigh
 E), pp. 492–505. University of Chicago Press, Chicago.
- Manokaran, N., et al. (2004) Pasoh Forest Dynamics Plot, Malaysia. In: Tropical forest diversity
 and dynamism: Findings from a large-scale plot network (eds Losos E, Leigh E), pp. 585–
 598. University of Chicago Press, Chicago.
- McMahon, S.M. & Parker, G.G. (2014) A general model of intra-annual tree growth using
 dendrometer bands. Ecology and Evolution 5(2), 243-254.
- Oliveira, A.A., et al. (2014) Habitat specialization and phylogenetic structure of tree species in a
 coastal Brazilian white-sand forest. Journal of Plant Ecology 7, 134–144.
- 634 Orwig, D.A., Foster, D.R. & Ellison, A.M. (2015) Harvard Forest CTFS-ForestGEO Mapped
 635 Forest Plot since 2014. Harvard Forest Data Archive: HF253. Available
- 636 online:http://harvardforest.fas.harvard.edu:8080/exist/apps/datasets/showData.html?id=hf253
- Su, S.H., Hsieh, C.F., Chang-Yang, C.H., Lu, C.L. & Guan, B.T. (2010) Micro-topographic
 differentiation of the tree species composition in a subtropical submontane rainforest in
 northeastern Taiwan. Taiwan Journal of Forest Science 25(1), 63-80.
- Sukumar, R., Sathyanarayana, S., Dattaraja, H., John, R. & Joshi, N. (2004) Mudumalai Forest
 Dynamics Plot, India. In: Tropical forest diversity and dynamism: Findings from a large-scale
- 642 plot network (eds Losos E, Leigh E), pp. 551–563. University of Chicago Press, Chicago.
- Valencia, R., et al. (2004) Yasuni Forest Dynamics Plot, Ecuador. In: Tropical forest diversity
 and dynamism: Findings from a large-scale plot network (eds Losos E, Leigh E), pp. 609–
 645 620. University of Chicago Press, Chicago.
- 646 Vallejo, M., Samper, C., Mendoza, H. & Otero, J. (2004) La Planada Forest Dynamics Plot,
- 647 Colombia. In: Tropical forest diversity and dynamism: Findings from a large---scale plot 648 network (eds Losos E, Leigh E), pp. 517–526. University of Chicago Press, Chicago.

- Vincent, J.B., Henning, B., Saulei, S., Sosanika, G. & Weiblen, G.D. (2014) Forest carbon in
 lowland Papua New Guinea: local variation and the importance of small trees. Austral
 Ecology 40, 151-159.
- Wang, X., et al. (2011) Spatial patterns of tree species richness in two temperate forests. Journal
 of Ecology 99, 1382-1393.
- Wu, S-H., et al. (2011) Kenting Karst Forest Dynamics Plot: tree species characteristics and
 distribution patterns. Taipei: Taiwan Forestry Research Institute. 306 p.
- Ku, H., et al. (2015) Community characteristics of a 60 ha dynamics plot in the tropical montane
 rain forest in Jianfengling, Hainan Island. Biodiversity Science 23, 192–201.
- Yin, D. & He, F. (2014) A simple method for estimating species abundance from occurrence
 maps. Methods in Ecology and Evolution 5, 336–343.
- Yuan, Z. et al. (2016) Multiple metrics of diversity have different effects on temperate forest
 functioning over succession. Oecologia 182, 1175-1185.
- 2662 Zimmerman, J.K., Comita, L.S., Thompson, J., Uriarte, M. & Brokaw, N. (2010) Patch dynamics
- and community metastability of a subtropical forest: compound effects of natural disturbance
- and human land use. Landscape Ecology **25**, 1099-1111.

665 Tables

Table 1. Structural characteristics of global forests. Values for density and biomass include trees ≥ 1 cm DBH within each square hectare (100 m × 100 m) of the plots, with the mean and ≤ 100 m ≥ 10

standard deviation (SD) calculated for each full hectare. The large-diameter threshold represents

- the diameter where half the biomass is contained within trees above that threshold. The biomass (70)
- 670 of the 1% indicates the proportion of total live aboveground tree biomass contributed by the
- 671 largest 1% of trees \geq 1 cm DBH. Plots are listed by declining large-diameter threshold.

Plot	Large-	Density	Biomass	Total			Biomass	
	diameter	` /	$(Mg ha^{-1})$	-		diameter		≥60 cm
	threshold	(SD)	(SD)	(n)	-	richness	1% (%)	DBH
	(cm)				(n)	(%)		(stems
	1065	1000 (0(6)	550(100)	1.4			1.0	ha^{-1}
Yosemite		1399 (266)	559(130)	14	3		46	52
Wind River		1207 (273)	532(161)	26	5			72
Žofín	78.0	· · · ·	248 (66)	11	4			
Ituri Lenda	72.0	7553 (829)	467 (62)	396	25			34
Danum Valley	65.7	7573 (526)	486(152)	784	62			27
SERC [†]	65.4	· · · ·	299 (49)	79	25			
Laupahoehoe		3925 (859)	241 (45)	22	2			37
Santa Cruz [†]		1945 (593)	361(102)	31	7	-		34
Cocoli		2164 (248)	281 (37)	170	9			32
Huai KhaKhaeng		2506 (674)	258 (65)	284	80			20
SCBI^\dagger	59.7	1850(1637)	259 (43)	64	22			35
Ituri Edoro	59.3	· · · ·	375 (46)	426	63			
Changbaishan	56.2	1230 (188)	288 (33)	52	15			34
Bukit Timah	55.6	6273 (180)	363(140)	353	18	5	73	19
Rabi	54.7	7988 (926)	323 (74)	346	74	21	73	14
Lambir	51.9	7635(1233)	495 (99)	1387	223	16	69	27
Barro Colorado	51.2	4938 (463)	257 (49)	297	80	27	67	17
Lilly Dickey [†]	51.2	1112 (441)	214 (29)	34	19	56	22	20
Xishuangbanna	49.8	4565 (650)	280 (81)	450	93	21	57	19
Wanang	49.6	5523 (520)	324 (61)	581	170	29	61	14
Palanan	49.4	4981 (489)	414(119)	324	41	13	62	27
Pasoh	48.5	5735 (631)	324 (55)	926	194	21	63	13
Michigan Woods	47.5	1981 (515)	192 (25)	44	16	36	26	14
Tyson [†]	45.4	1601 (751)	176 (16)	45	18	40	24	10
Wytham Woods [†]	44.8	1016 (309)	310 (46)	23	13	57	23	18
Korup	42.9	7283 (920)	345 (88)	485	143	29	67	10
Manaus	42.2	6234 (441)	344 (54)	1529	260	17	59	9
Cedar Breaks	41.9	1542 (961)	168 (53)	17	8	47	34	13
Mudumalai	41.7	608 (210)	205 (33)	72	35	49	18	12
Jianfengling	40.8		392 (37)	290	116	40	48	24
La Planada		4030 (243)	270 (30)	241	74			8
Fushan		4478(1139)	224 (25)	106	33			
Sherman		3662 (550)	275 (41)	224	31	14		13
Amacayacu		4948 (518)	268 (33)	1233	326			7

Kenting	36.1 3760 (410)	255 (38)	92	40	43	36	7
Lienhuachih	35.7 6131(1760)	170 (25)	145	49	34	51	10
Harvard Forest [†]	35.5 3104(2600)	260 (66)	55	17	31	23	7
Luquillo	35.5 2903 (626)	283 (53)	133	47	35	39	12
Heishiding	34.5 5277 (706)	149 (27)	213	59	28	43	12
Wabikon [†]	31.1 1692(1017)	111 (14)	31	15	48	17	1
Gutianshan	31.0 5833(1580)	185 (27)	159	40	25	34	2
Ilha do Cardoso	31.0 4660 (578)	148 (17)	135	43	32	41	7
Yasuni	29.1 5834 (692)	261 (48)	1075	343	32	50	8
Hong Kong [†]	28.6 5860(1056)	142 (20)	172	43	25	39	3
Lanjenchi	17.212075(2795)	113 (7)	128	72	56	29	1
Mpala	10.0 2963(2902)	13 (8)	68	35	51	30	0
Scotty Creek	7.6 4136(1407)	22 (11)	11	7	64	15	0
Palamanui	2.5 8205(1084)	30 (5)	16	11	69	13	0

672 [†]Mature secondary forest

Table 2. The effect of geographical region on tree density and biomass and their variation at 1 ha scale and the abundance of large-

674 diameter trees as measured by the three metrics of proportion of biomass in the largest 1% of trees, density of trees ≥ 60 cm DBH, and

675 large-diameter threshold. The SD of density and the SD of biomass represent the within-region (between-plot) variation. The CV of

676 density and CV of biomass represent the average of the individual plot 1-ha CVs, with each plot weighted equally. SD: standard

677 deviation, CV: coefficient of variation.

Zone	Plots	Density	Density	Density	Biomass	Biomass	Biomass	Biomass	Density	Large-
	n	(trees ha ⁻¹)	SD	CV	(Mg ha ⁻¹)	SD	CV	of the 1%	trees ≥60	diameter
								(%)	cm dbh	threshold
									(trees ha ⁻¹)	(cm)
Cold temperate / boreal	6	2,281	1,114	47	174	98	24	23	11	37
Temperate	16	3,339	2,193	31	266	126	18	38	24	53
All Tropics	26	5,735	1,072	18	278	57	20	61	16	5 44
Tropical Africa	5	6,949	2,317	29	305	172	27	76	16	i 48
Tropical Asia	10	5,767	3,149	16	330	124	21	53	18	47
Tropical Latin America	8	4,339	1,410	12	280	27	15	54	13	42
Tropical Oceania	3	5,884	2,162	15	198	152	18	61	17	38

Figures



- 682 Fig. 1. Location of the 48 plots affiliated with the Smithsonian Forest Global Earth Observatory
- (ForestGEO) used in this study.



686 Fig. 2. Contribution of large-diameter trees to forest structure of 48 large forest plots.

687 Aboveground live tree biomass increases with increasing large-diameter threshold (A). The

688 large-diameter threshold reflects the tree diameter that segments biomass into two equal parts.

689 Below the large-diameter threshold are a large number of small-diameter trees, and above the

- 690 large-diameter threshold are a smaller number of large-diameter trees. Aboveground live
- 691 biomass also increases with the concentration of biomass in the largest 1% of trees (B) and the 692 density of stems \geq 60 cm DBH (C). Large-diameter richness declines with increasing biomass
- 692 density of stems \geq 60 cm DBH (C). Large-diameter richness declines with increasing biomass 693 (D), which is consistent with the declining relationship between large-diameter threshold and
- 694 large-diameter richness (E). The concentration of biomass in the largest 1% of trees has a strong
- 695 negative relationship with large-diameter richness (F). Colours indicate increasing absolute
- 696 latitude from red to green. Grey areas around regression lines indicate 95th percentile confidence
- 697 intervals.

698



699

Fig. 3. Gradients of forest structural attributes by absolute latitude for 48 forest plots in the

701 ForestGEO network. Absolute latitudinal gradients in density (A) and concentration of biomass

in the largest 1% of trees (C) were significant. The relationships for biomass (B; $r^2 = 0.04$, P =

0.106)) and the large-diameter threshold (D; $r^2 = 0.01$, P = 0.551) were not. Colours indicate increasing absolute latitude from red to green. Grey areas around regression lines indicate 95th

increasing absolute latitude from red to green. Grey areas arounpercentile confidence intervals.



Fig. 4. Three-dimensional nonmetric multidimensional scaling (NMDS) results for density of

- trees organized into six diameter classes in 1260, $100 \text{ m} \times 100 \text{ m}$ hectares of 48 forest plots in
- the ForestGEO network (A, B). The structural classes (diameter bins) used in the NMDS
- 710 ordination are superimposed in black text (A, B). The within plot variation of structure for each
- 711 plot is shown by depiction of the standard deviation ellipses of the individual $100 \text{ m} \times 100 \text{ m}$
- hectares within each plot (C, D; where C reflects the variation of NMDS1 vs. NMDS2 (A) and D
- reflects the variation of NMDS1 vs. NMDS3 (B). Ordination stress = 0.047. Colours indicate

- 714 increasing absolute latitude from red to green, with plot centroids numbered (A, B). See
- 715 Supplemental Material for a three-dimensional animation of the structural ordination.





718 function of absolute latitude (A). The metric of structural complexity is the volume of the three-

dimensional ellipsoid generated from the NMDS ordination of abundance in structural classes

(see Fig. 4 for two dimensional projections and the Supplementary Material S2 for a three-

dimensional animation). The rank order of large-diameter species in 48 forest plots (B). Rank
 order is normalized to the range from zero to one to compare plots with differing species

order is normalized to the range from zero to one to compare plots with differing species
 richness. Lower proportions of large-diameter species rank correspond to more abundant species

(median large-diameter species rank <0.5 for all 48 forest plots). Species attaining large-

diameters were the more common species in the forest plots. Colours indicate increasing absolute

726 latitude from red to green.

727 Supplementary Material

- Fig. S1.1. Relationship among basal area, large-diameter threshold, and large-diameter richness
- 729 Fig. S1.2. Relationships among metrics for large-diameter trees
- 730 Fig. S1.3. Area needed to estimate structural characteristics of forests
- 731 Video S2. Animation of the three-dimensional structural NMDS ordination of 48 forest plots
- 732 [Note: For convenience of review, this animation can also be found at:
- 733 http://westernforestinitiative.org/Lutz_et_al_S2_20180109.gif
- 734 Table S3.1. Environmental characteristics of 48 large forest plots
- Table S3.2. Density data by diameter class for 48 large forest plots
- Table S3.3. Biomass data by diameter class for 48 large forest plots
- 737 S4. Acknowledgements for 48 forest plots in the ForestGEO network