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1 **Which persistent organic pollutants in the rivers of the Bohai Region of China**
2 **represent the greatest risk to the local ecosystem?**

3

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16

17 **Abstract**

18 Freshwater aquatic organisms can be exposed to hundreds of persistent organic
19 pollutants (POPs) discharged by natural and anthropogenic activities. Given our limited
20 resources it is necessary to identify, from the existing evidence, which is the greatest
21 threat so that control measures can be targeted wisely. The focus of this study was to
22 rank POPs according to the relative risk they represent for aquatic organisms in rivers
23 in the Bohai Region, China. A list of 14 POPs was compiled based on the available data
24 on their presence in these rivers and ecotoxicological data. Those that were widely
25 detected were benzo[a]pyrene, *p,p'*-DDE, *p,p'*-DDT, endrin, fluoranthene, heptachlor ,
26 hexabromocyclododecane, hexachlorobenzene, α -hexachlorocyclohexane, γ -
27 hexachlorocyclohexane, naphthalene, perfluorooctanoic acid, perfluorooctane
28 sulfonate and phenanthrene. Effect concentrations were compiled for Chinese relevant
29 and standard test species and compared with river aqueous concentrations. Only bed-
30 sediment concentrations were available so water levels were calculated based on the
31 known local sediment organic carbon concentration and the K_{oc} . The POPs were ranked
32 on the ratio between the median river and median effect concentrations. Of the POPs
33 studied, fluoranthene was ranked as the highest threat, followed by phenanthrene,
34 naphthalene and *p,p'*-DDE. The risk from *p,p'*-DDE may be magnified due to being
35 highly bioaccumulative. However, the greatest overlap between river concentrations
36 and effect levels was for lindane. Overall, fish was the most sensitive species group to
37 the risks from POPs. Hotspots with the highest concentrations and hence risk were

38 mainly associated with watercourses draining in Tianjin, the biggest city in the Bohai
39 Region.

40

41 **Key words:**

42 Ecological risk; POPs; Fluoranthene; Risk ranking; Bohai Region

43

44 1. Introduction

45 Persistent organic pollutants (POPs) are of concern globally due to their
46 persistence, long-range transportation, bioaccumulation and toxicity to wildlife.
47 Perhaps the best example of the potentially devastating impact of POPs was that of
48 DDT and the associated DDE on birds of prey (Ratcliffe, 1967). Consequently, many
49 POPs are now subject to a great deal of monitoring to assess the exposure and risks
50 from such chemicals (Wong et al., 2005; Doney, 2010; Letcher et al., 2010; Covaci et
51 al., 2011; Elliott and Elliott, 2013). Many POPs which are extremely persistent, such
52 as PCBs and lindane, have been banned or restricted by international conventions, so
53 some decrease in environmental exposure is starting to occur (Lohmann et al., 2007).
54 However, human society still needs stable organic molecules with properties such as
55 fire resistance (eg HBCDs), non-reactivity (eg PFOS) and plasticising properties (short
56 chain chlorinated paraffins), so the environment will continue to be exposed to such
57 chemicals, but just how much of a risk do they represent?

58 The Bohai coastal region, located to the east of Beijing is one of the China's most
59 important manufacturing areas. It includes the provinces of Shandong, Tianjin, Hebei
60 and Liaoning which have benefitted from rapid industrialisation since the late 1970's.
61 Apart from its industrial base, the region has a combined population of 231 million
62 people (National Bureau of Statistics, 2014). There are more than 40 rivers flowing into
63 the Bohai Sea, a semi-enclosed sea, and they convey many chemical pollutants to the
64 Bohai Sea (Wang et al., 2014; Wang et al., 2015b). With the Chinese rush for growth
65 there have been concerns about a resultant chemical pollution of the environment, such

66 as pesticides (Zhang et al., 2009), polycyclic aromatic hydrocarbons (Wang et al.,
67 2015a), polychlorinated biphenyls (Zhao et al., 2005), perfluoroalkyl and
68 polyfluoroalkyl substances (Wang et al., 2014) and hexabromocyclododecanes (Zhang
69 et al., 2016) in the Bohai region. China is now taking steps to ban many of the most
70 persistent organic pollutants (POPs) as indicated by the Stockholm Convention. Whilst
71 some pollutants may no longer be discharged and could be considered a legacy of the
72 past, others may still be generated, for example from combustion processes.

73 There is now an increasing appreciation for the need to better protect the natural
74 environment in China, such as the Water Pollution Control Action Plan in 2015 and the
75 Soil Pollution Control Action Plan in 2016 issued by the State Council. However, with
76 so many kinds of chemical contaminants being discovered and monitored, it is
77 important to find some ways for identifying which represent the greatest risk. This is a
78 problem for the whole world and not just China. In Europe, as part of the Water
79 Framework Directive chemicals were identified as being of special concern (priority
80 and hazardous substances) on the basis of several properties including persistence and
81 different toxic properties. However, a recent approach has been proposed which argues
82 that only two factors are critical, toxicity and exposure, and that relative risk can be
83 assessed from the proximity of the median exposure and toxicity concentrations
84 (Donnachie et al., 2014; Donnachie et al., 2015). In this study the environmental
85 concentrations of POPs which have been well monitored in the freshwater Bohai coastal
86 region were compared with the available information on toxicity concentrations. The

87 objective was to identify which currently well studied POPs should be considered of
88 greatest threat to wildlife in the region?

89 **2. Method**

90 **2.1. Approach to risk ranking**

91 The risk ranking approach, which compares levels of chemicals in the
92 environments and effect concentrations in ecotoxicological tests, has been applied in
93 the UK for metals and pharmaceuticals (Donnachie et al., 2014; Donnachie et al., 2015).
94 To obtain measured environmental data for the Bohai region, literature from both
95 English and Chinese sources were reviewed. For toxicity information, the US EPA
96 ECOTOX Database, as well as a wider literature review was used. With environmental
97 data and effect data collected, the final risk ranking compared the proximity of the
98 medians of both datasets. In this study the ecotox dataset typically comprised 8 to 90
99 entries (see SI). So the median was considered a robust (or fair) comparator of relative
100 risk between chemicals. It is important to note that this approach is different from
101 traditional risk assessment, where something like the 5% percentile toxicity
102 concentration, or lowest observable ecotoxicity concentration (LOEC) or predicted no
103 effect concentration (PNEC) is used as a comparator. These methods often put great
104 weight on only a few data points and the danger is that some of these studies may be
105 weak and unrepeatably (Harris et al., 2014). It should be acknowledged that where the
106 median ecotox value is quite similar between chemicals, the ranking should not be seen
107 as absolute, and that the output is a relative ranking rather than an absolute risk
108 probability.

109 **2.2. Chemicals selected for this study**

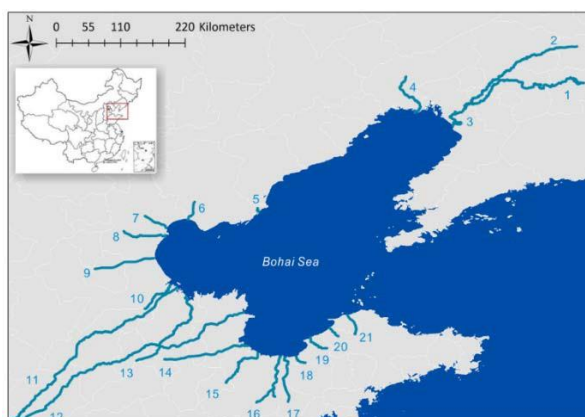
110 The selection of chemicals was determined both by their presence in the rivers of Bohai
111 region (the availability and quality of measured data) and by the degree of concern
112 expressed in the literature over their toxicity, persistence or potential to accumulate.
113 The persistent organic pollutants considered in this research included industrial
114 chemicals, pesticides and by-products of human activity. Fourteen chemicals were
115 selected from more than 20 groups of chemicals on the basis monitoring data
116 availability. The criteria used included having recent monitored data (2010-2015),
117 abundant freshwater sampling sites and a sufficient geographic spread across the Bohai
118 Region (Tab. 1 and Fig. 1). Lakes and reservoirs were not considered due to lack of
119 sufficient measurements.

120

121 **Table 1.** Chemicals assessed in this study

	Chemical name	Usage	Production status
1	α -Hexachlorocyclohexane (α -HCH)	Insecticide by-product	Banned
2	γ -hexachlorocyclohexane (γ -HCH)	Insecticide	Banned
3	Endrin	Insecticide	Banned
4	Heptachlor	Insecticide	Banned

5	<i>p,p'</i> -Dichlorodiphenyltrichloroethane (<i>p,p'</i> -DDT)	Insecticide	Restricted
6	<i>p,p'</i> -Dichlorodiphenyldichloroethylene (<i>p,p'</i> -DDE)	Degradation product of <i>p,p'</i> -DDT	
7	Hexachlorobenzene (HCB)	Industrial use chemical	Banned
8	Hexabromocyclododecane (HBCD)	Flame Retardant	Still produced
9	Perfluorooctanoic acid (PFOA)	Insulators for electric wires, planar etching of fused silica, fire fighting foam, and outdoor clothing	Restricted
10	Perfluorooctane sulfonate (PFOS)	Electric and electronic parts, fire fighting foam, photo imaging, hydraulic fluids and textiles	Restricted
11	Benzo[a]pyrene (B[a]P)	Unintentional production chemical	
12	Fluoranthene (Flu)	Unintentional production chemical	
13	Phenanthrene (Phe)	Unintentional production chemical	
14	Naphthalene (Nap)	Unintentional production chemical	



123

124 **Figure 1.** River segments with POPs concentrations reported in Bohai Region. 1.

125 Taizi River; 2. Hunhe River; 3. Liaohe River; 4. Daling River; 5. Luanhe River; 6.

126 Duohe River; 7. Yongding New River; 8. Dagu Drainage River; 9. Ziyaxihe River;

127 10. Zhangwei New River; 11. Majiahe River; 12. Tuhaihe River; 13. Yellow River;

128 14. Xiaoqinghe River; 15. Mihe River; 16. Weihe River; 17. Jiaolaihe River; 18.

129 Shahe River; 19. Wanghe River; 20. Jiehe River; 21. Huangshuihe River.

130

131 **2.3. Estimation of POPs concentration in water**

132 Whilst it may not be entirely appropriate for POPs, most available ecotoxicity

133 information for these chemicals is based on exposure through the water column.

134 However, most POPs, being moderately to highly hydrophobic, partition strongly to

135 river sediment. Due to the virtual absence of water column measurements or sediment-

136 based toxicity data, predictions for the aqueous concentrations had to be made from

137 Bohai region river sediment values. Measured concentrations in the Bohai Region were

138 searched from the literature in the Web of Science™ database for English publications

139 and CNKI database for Chinese publications. The partition theory can be used to

140 estimate water concentrations from measured sediment concentration. K_{oc} , the organic
 141 carbon-water partition coefficient is defined as

142
$$K_{oc} = \frac{C_{oc}}{C_w} \quad \text{(Equation 1)}$$

143 And the mass of organic carbon in Equation 1 can be expressed as

144
$$C_{oc} = f_{oc} \cdot C_{oc, total} \quad \text{(Equation 2)}$$

145 So the chemical concentration in water can be expressed as

146
$$C_w = \frac{C_s}{K_{oc} \cdot f_{oc}} \quad \text{(Equation 3)}$$

147 In this study, prediction of water concentration was conducted for all chemicals
 148 except PFOS and PFOA, which had sufficient water measurements. The partition
 149 theory assumes equilibrium status on the surface of sediment. So the estimated water
 150 concentration was close to the pore water concentration, which was likely to be several
 151 times higher than the surface water concentration. The data collected in the previous
 152 studies (Zhou et al., 2006; Tan et al., 2009) was used to test the deviation between the
 153 predicted values and measured values. Compared with the pore water concentrations,
 154 the relative deviations were 0.53 for α -HCH, 0.56 for γ -HCH, 0.48 for p,p'-DDE, 0.02
 155 for p,p'-DDT, 0.02 for heptachlor and 4.1 for endrin. In comparison with the surface
 156 concentrations, the relative deviations were 7.7 for α -HCH, 8.0 for γ -HCH, 4.4 for p,p'-
 157 DDE, -0.1 for p,p'-DDT, 1.2 for heptachlor and 3.9 for endrin. Thus, predicted
 158 concentrations may over-estimate water column levels, however, it could be considered
 159 better to err on the precautionary side.

160 **2.4. Effect data collection and selection**

161 Literature giving effect data for the selected POPs was largely obtained from the
162 US EPA ECOTOX database, and when the dataset was not sufficient more literature
163 was obtained using the Web of Science™ database and searched for via a series of key
164 words (Donnachie et al., 2014; Donnachie et al., 2015). Ecotoxicity data for Chinese
165 local freshwater species and standard test species were selected for each chemical. A
166 range of effect measurements were present in the literature including LOEC, EC50,
167 LC50, acute and chronic toxicity and all of these were collected. The effect data of
168 LC50 and EC50, was preferred for each species in each study. The widest range of
169 species and end-points were considered, to ensure that as representative a picture of
170 species and possible effects as possible was obtained. Where several studies reported
171 effect concentrations using the same end-point for one species then in this case only the
172 lowest effect concentration for a single species was used. Thus, the final ecotoxicity
173 dataset allocated a single value for a single species for a particular end-point.
174 Alternative approaches might have been to use the median of the ecotoxicity dataset
175 points for a single species, or simply used all the data, regardless of whether several of
176 the points are for the same species/end-point. When a comparison was made to look at
177 the impact on the overall median ecotoxicity value for a chemical it was found that
178 these choices made very little difference. The value of only plotting one data point per
179 end-point and species is it reveals clearly to the viewer the number of different species
180 available for analysis and does not give undue weight to commonly studied species.

181 **2.5. Risk analysis**

182 Once the datasets for ecotoxicology and river measurements were considered
183 sufficient, the information included in them could be plotted and the medians noted.
184 The difference between these medians can be described as a risk ratio, which can be
185 used to rank concern; the larger the value, the greater the concern (Equation 4).

186
$$\frac{mW}{mT} \quad (\text{Equation 4})$$

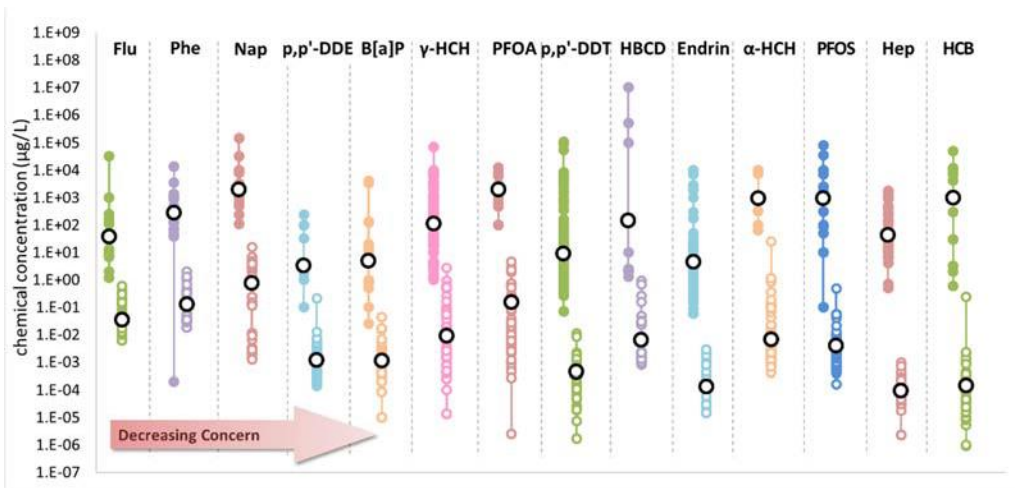
187 Where mW is the median river water concentration ($\mu\text{g/L}$) and mT is the median
188 effect concentration ($\mu\text{g/L}$).

189 3. Results and discussion

190 3.1. Risk ranking: which chemicals posed the greatest threat to wildlife?

191 The approach used was able to rank the 14 chemicals considered on the basis of
192 risk (Fig. 2 and Fig. 3). The risk ratios ranged from 1×10^{-3} to 1×10^{-7} , so this method
193 suggests most wildlife would not be suffering unacceptable direct toxic effects via
194 water exposure in rivers in the Bohai Region. Based on the median risk ratio, the PAHs
195 group tended to be the POPs of greatest concern for the Bohai Region, with Flu, Phe,
196 Nap and B[a]P ranking 1st, 2nd, 3rd and 5th. These were followed in terms of risk by
197 traditional pesticides including p,p' -DDE and γ -HCH ranking 4th and 6th. The novel
198 POPs including PFOA, HBCD and PFOS were further down ranking 7th, 9th and 12th.
199 The other selected POPs including endrin, α -HCH, heptachlor and HCB had the lowest
200 relative risk.

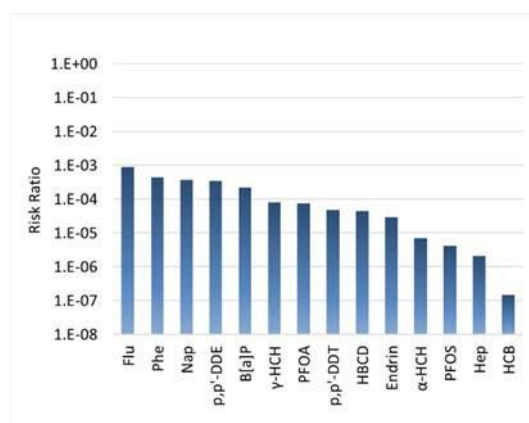
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202

203 **Figure 2.** Risk ranking of 14 chemicals in rivers in Bohai Region. For each chemical
 204 both the: effect concentrations data (solid filled circles) and: water concentrations
 205 predicted from sediment measurements in Bohai Region (unfilled circles) are shown
 206 side by side. The large black circles are the median points.

207



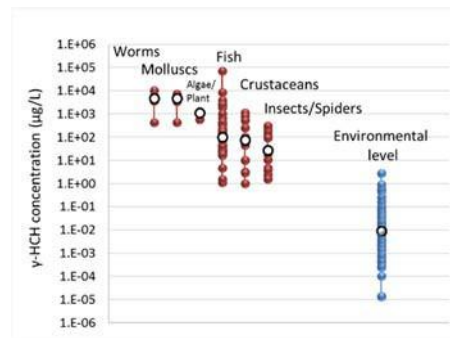
208

209 **Figure 3.** Risk ratio ranking of 14 chemicals based on comparison of the median
 210 ecotoxicity and median river values

211

212 It is noted that for some of these chemicals there is an overlap, with some
213 estimated/measured water values exceeding some of the levels where effects have been
214 reported (Fig. 2). These include phenanthrene, DDE, benzo[a]pyrene, lindane, HBCD
215 and PFOS. Lindane (γ -HCH) had the largest overlap according to the number of species
216 involved. The insects/spiders were the most sensitive category, as well as crustaceans
217 and fish (Fig. 4). Thus, from the available ecotoxicity information the possibility exists
218 for lethal effects on insects such as mosquito (*Culex sitiens*) (Oh et al., 2013). Some
219 crustaceans, such as ostracod (*Cypris subglobosa*), might receive adverse effects such
220 as immobility (Cheng et al., 2011). Fish, such as walking catfish (*Clarias batrachus*)
221 and pool barb (*Puntius sophore*) may also experience lethal effects.

222



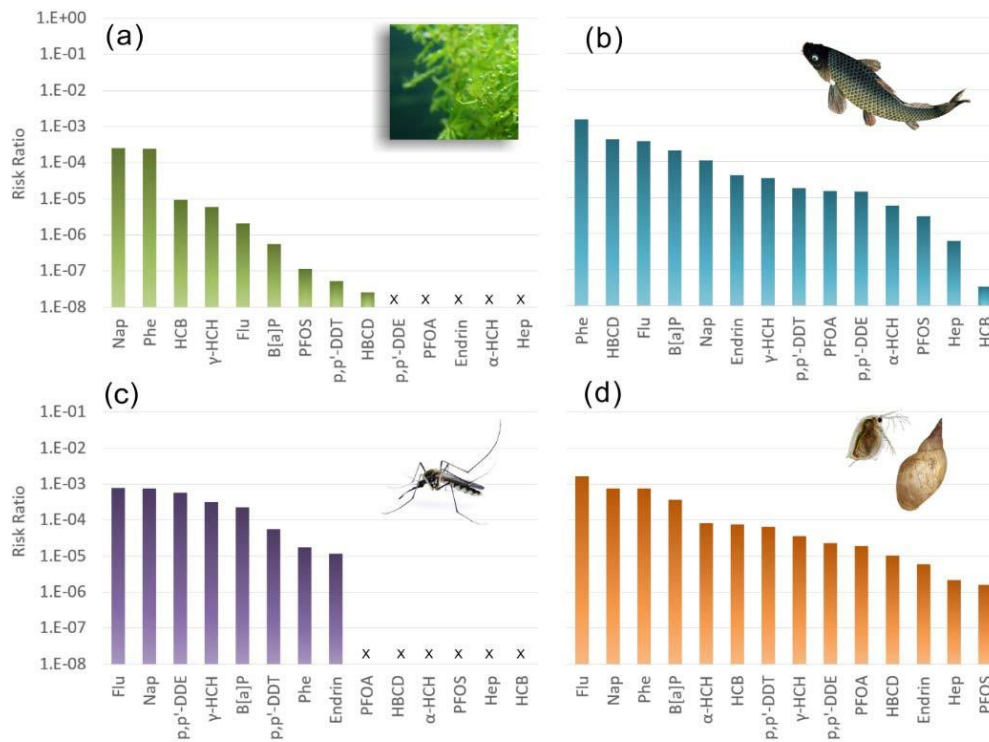
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224 **Figure 4.** Effect concentrations of different species groups (red) and environmental
225 level (blue) of γ -HCH in rivers in Bohai Region.

226

227 The effects data can be disaggregated into algae, fish, insects/spiders and
228 invertebrates/molluscs/crustaceans to examine their different sensitivities to these POPs.

229 Generally, fish were the most sensitive group of species to this group of POPs.
 230 Insects/spiders and molluscs/crustaceans/invertebrates were less sensitive to POPs with
 231 algae being the least sensitive to these POPs (Fig. 5).

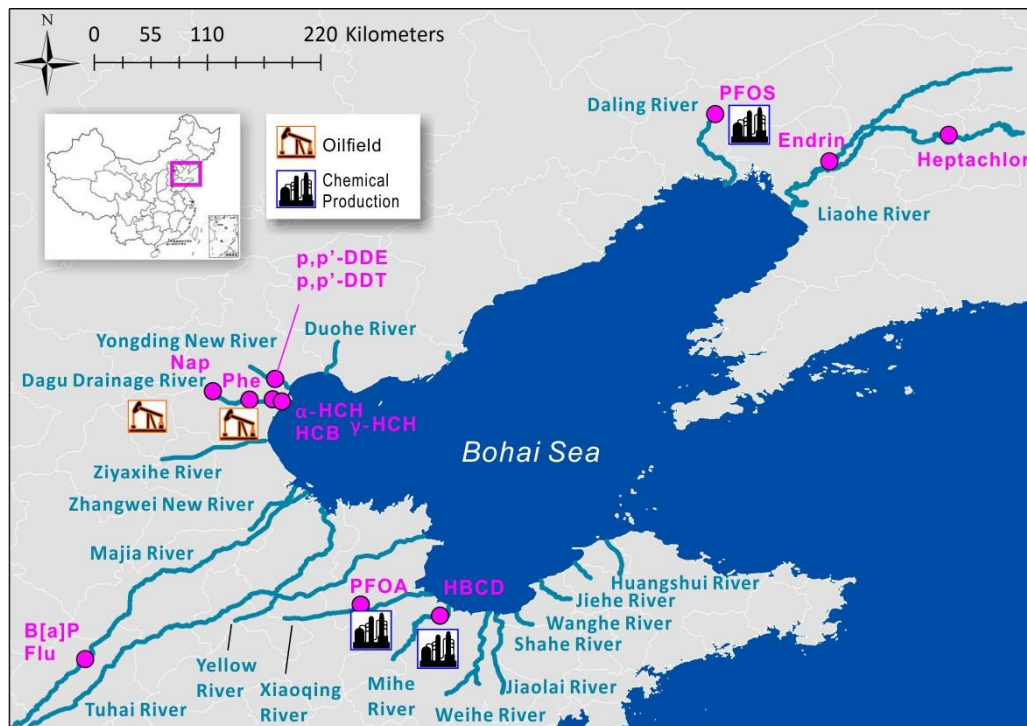


232

233 **Figure 5.** Risk ranking of chemicals for different groups of species. (a) algae, (b) fish,
 234 (c) insects/spiders, (d) molluscs/crustaceans/invertebrates. X means not enough
 235 ecotoxic data available for this chemical.

236

237 **3.2. Hot-spots: in which areas of the Bohai Region might POPs have the**
 238 **greatest impacts?**



239

240 **Figure 6.** Locations where the maximum concentrations were recorded.

241

242 The Dagu Drainage River featured as a hot-spot where 5 chemicals were found at
 243 their highest concentrations (Fig. 6). It is 68 km long and is the primary drainage canal
 244 for Tianjin City, which is one of the four municipalities directly under the National
 245 Central Government and is an important industrial centre, with a population of 11
 246 million. The Dagu Drainage River, receives 0.8 million m³/day effluent from municipal
 247 wastewater treatment plants and industrial and agricultural wastewaters along its way
 248 to Bohai Sea (Li et al., 2011). In order to improve water quality, local government
 249 dredged the contaminated sediment in this waterbody in 2008 and 2009, but clearly the
 250 POPs have not been eliminated. The Yongding New River, had the highest
 251 concentrations of *p,p'*-DDE and *p,p'*-DDT, is a similar artificial river located in

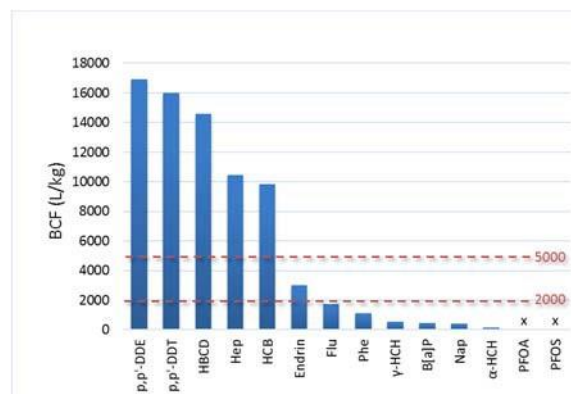
252 Tianjin, which receives the river flows from Haihe River Basin and municipal and
253 industrial wastewater along its way to Bohai Bay as well.

254 3.3. Bioconcentration Factor (BCF) ranking of POPs

255 Whether a chemical is bioaccumulative has been a traditional concern of chemicals
256 in risk assessment. A judgement on whether a chemical could be considered
257 bioaccumulative has been linked to the bioconcentration factor, which is the
258 partitioning of a chemical between the water phase and an aquatic organism. According
259 to the European standard, a BCF value above 2000 is considered to be bioaccumulative
260 and 5000 is considered very bioaccumulative (EC, 2006).

261 The median BCF value of each POP was examined, using data from the US EPA
262 Ecotox Database and additional literature (Fig. 7). Of the top ranked POPs in this study
263 (Fig. 2 and 3) only *p,p'*-DDE would be considered bioaccumulative. This could be an
264 argument for raising our concern over this chemical within the top five ranked POPs in
265 the rivers of the Bohai region.

266



267

268 **Figure 7.** Ranking of POPs based on median value of Bioconcentration Factor (BCF).
269 A BCF value above 2000 is considered to be bioaccumulative and 5000 is considered
270 very bioaccumulative.

271

272 **Local source of the PAHs**

273 PAHs can be introduced into the environment from both natural and anthropogenic
274 processes including biomass burning, fossil fuel combustion, transportation emissions,
275 and petroleum industries (Yunker et al., 2002). As for the individual research in the
276 Bohai Region, the sources of PAHs in the environment were usually attributed by their
277 characteristic isomer ratios, such as InP/(InP+BghiP), Flu/(Flu+Pyr), An/(An+Phe) and
278 BaA/(BaA+Chr) (InP for indeno[1,2,3-cd]pyrene, BghiP for benzo[ghi]perylene, Flu
279 for fluoranthene, Pyr for pyrene, An for anthracene, Phe for phenanthrene, BaA for
280 benz[a]anthracene and Chr for chrysene). The ratios, as well as their sources
281 represented, varied in locations in these investigations, and even varied in sampling
282 sites in individual campaign. For the rivers involved, biomass and coal combustion was
283 the main source of PAHs in the rivers in the north of the Bohai Region such as the Dagu
284 Drainage River (He et al., 2011), Yongdingxinhe River (Wang et al., 2015a) and Daliao
285 River (Zheng et al., 2016). But in the south of the Bohai Region, the main source of
286 PAHs was petroleum and its combustion in the watersheds such as the Yellow River
287 (Wang et al., 2015a) and the Tuhai-Majia River (Liu et al., 2012). Previous studies on
288 PAHs in the Bohai Rim indicated that their origin was a mix of combustion and

289 petroleum production (Zhang et al., 2009; Jiao, 2012; Jiao et al., 2013). Rapid
290 industrialization and urbanization in the Bohai Region increased the fossil fuel
291 consumption due to power generation, heating supply, industrial and commercial
292 activities and residents. In 2014, energy consumption in these four provinces was 932-
293 million-ton standard coal equivalent including coal, petroleum and natural gas (Hebei
294 Government, 2015; Liaoning Statistical Bureau, 2015; Shandong Statistical Bureau,
295 2015; Tianjin Statistical Bureau, 2015), which amounted to 22% of the total energy
296 consumption of China. In addition to biomass and coal combustion, the petroleum
297 industry may also be a direct source of PAHs from oilfield operations such as in Shengli
298 Oilfield, Jinzhou Oilfield and other oilfield drilling platforms in the Bohai Region.

299 **Local source of pesticides**

300 The compound *p,p'*-DDE is a degradation product of DDT, a pesticide which had
301 been widely used globally. DDTs had been produced in China since the 1950s, and
302 despite the the official ban in 1983 the use of DDTs in agriculture had not been
303 stopped until 2000 due to the use of pesticide dicofol with high impurity of DDTs
304 compounds (Tao et al., 2007; Liu et al., 2008). The ratios such as (DDE+DDD)/DDTs,
305 *o,p'*-DDT/*p,p'*-DDT were usually used to distinguish the sources of DDTs. These ratios
306 indicated that the DDTs in the rivers were the legacy from historical production and use
307 (Li et al., 2013; Gao et al., 2015), especially the use of the technical DDT before 1987
308 and the use of dicofol after 1987 in the Daling River (Wang et al., 2013)

309 Two types of HCHs had been used in China as pesticides, technical HCHs and
310 lindane. Technical HCHs (18% of γ -HCH) was used from the 1950s to 1983, while
311 lindane (99.9% of γ -HCH) in the 1990s. The ratio α -HCH/ γ -HCH indicated the
312 historical use of both technical HCHs and lindane in the Haihe River and the Daling
313 River (Li et al., 2013; Wang et al., 2013).

314 **Local sources of flame retardants and per-fluorinated compounds**

315 HBCDs are used as flame retardant in extruded/expanded polystyrene insulation
316 boards, textile, and electric/electronic products. Due to the limited effect data for the
317 individual isomers for α -, β - and γ -HBCD, they were considered as a whole technical
318 mixture. PFOS and PFOA, known as PFASs are widely used in polymer, surfactants,
319 lubricants for their surface activity and heat/acid resistance. The biggest HBCD and
320 PFASs manufacturers in China are located in the Bohai Rim and support the whole
321 industrial chain in this region. Spatial analysis of PFAAs levels in the samples taken
322 from the rivers and producers indicated that the fluoropolymer industries along the
323 Xiaoqing River and the Daling River were the major sources of PFOA and PFOS in the
324 rivers in the Bohai Region (Wang et al., 2014; Meng et al., 2015; Zhu et al., 2015). The
325 spatial analysis and isomer ratio of γ -HBCD/ α -HBCD indicated that the manufacturing
326 was the major source of HBCD in the environment (Li et al., 2012; Zhang et al., 2016).
327 Extremely high levels of HBCD and PFASs in the world had been found in the
328 environment in the Bohai Region due to their high production and use.

329 Of the other chemicals examined, endrin, heptachlor and HCB presented much
330 lower risks. Endrin has not been produced in China and the production of heptachlor
331 and HCB and their application in agriculture were banned in 1983.

332

333 **3.4. Uncertainties and limitations of the study**

334 This study can only be as strong as the existing monitoring and ecotoxicity data
335 allows it to be. It may be that some other POPs, so far not measured, may have a much
336 higher risk ranking to the compounds studied here. Similarly, a wider, more systematic
337 monitoring programme may reveal higher concentrations for some chemicals in the
338 Bohai Region than reported so far. The ecotoxicity database is driven by water exposure
339 studies, yet most POPs measurements are not from the water column but from the river
340 bed-sediments. It is necessary to predict the water concentration and the method used
341 is most likely to several times overestimate, rather than underestimate, this
342 concentration.

343 Finally, for some POPs the ecotoxicity database is still not as wide or complete as
344 would be desirable. For traditional POPs such as γ -HCH and *p,p'*-DDE, abundant
345 effects data could be found in EPA ECOTOX database. But for novel POPs such as
346 HBCD and PFOS/PFOA, very limited ecotoxicity data was available, especially for
347 individual HBCD isomers.

348 **Conclusions**

349 From this group of POPs, the PAHs congeners posed the greatest risk for aquatic
350 wildlife in rivers around the Bohai Sea, followed by *p,p'*-DDE and γ -HCH. However,
351 there was still more than 3-orders of magnitude distance between the median ecotox
352 and median environmental concentrations suggesting the risks to wildlife through water
353 exposure were not large, although the potential for these chemicals to bioconcentrate
354 must be acknowledged. It was observed that there were locations where some water
355 concentrations, for example for lindane, exceeded effect levels for some of the aquatic
356 wildlife. The greatest impacts of POPs on wildlife would be expected in the Dagu
357 Drainage River in Tianjin. The results suggest that regarding threats from POPs to the
358 environment in the Bohai Region the greatest efforts should be in reducing fossil fuel
359 combustion (to lower PAHs). The highly bioaccumulative metabolite of DDT, *p,p'*-
360 DDE, was also flagged up as high risk but has now been controlled by the Government,
361 so its environmental concentrations are expected to reduce in future. It was somewhat
362 encouraging that some of the emerging POPs such as PFOS, PFOA and HBCD were
363 not posing the highest risks.

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372 **References**

- 373 Cheng, Y., Cui, Y., Chen, H.-m., Xie, W.-p., 2011. Thyroid disruption effects of
374 environmental level perfluorooctane sulfonates (PFOS) in *Xenopus laevis*.
375 *Ecotoxicology* 20, 2069-2078.
- 376 Covaci, A., Harrad, S., Abdallah, M.A.E., Ali, N., Law, R.J., Herzke, D., de Wit, C.A.,
377 2011. Novel brominated flame retardants: A review of their analysis, environmental
378 fate and behaviour. *Environment International* 37, 532-556.
- 379 Doney, S.C., 2010. The Growing Human Footprint on Coastal and Open-Ocean
380 Biogeochemistry. *Science* 328, 1512-1516.
- 381 Donnachie, R.L., Johnson, A.C., Moeckel, C., Pereira, M.G., Sumpter, J.P., 2014.
382 Using risk-ranking of metals to identify which poses the greatest threat to freshwater
383 organisms in the UK. *Environmental Pollution* 194, 17-23.
- 384 Donnachie, R.L., Johnson, A.C., Sumpter, J.P., 2015. A rational approach to selecting
385 and ranking some pharmaceuticals of concern for the aquatic environment and their
386 relative importance compared with other chemicals. *Environmental Toxicology and*
387 *Chemistry*.
- 388 EC, 2006. Regulation No 1907/2006 Registration, Evaluation, Authorisation and
389 Restriction of Chemicals (REACH), Establishing a European Chemicals Agency,
390 Amending Directive 1999/45/EC and Repealing Council Regulation (EEC) No 793/93

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392 and Commission Directives 91/155/EEC, 93/67/EEC, 93/105/EC and 2000/21/EC.

393 Elliott, J.E., Elliott, K.H., 2013. Tracking Marine Pollution. *Science* 340, 556-558.

394 Gao, L.R., Xia, D., Tian, H.Z., Zhang, H.J., Liu, L.D., Wang, Y.W., 2015.
395 Concentrations and distributions of 18 organochlorine pesticides listed in the
396 Stockholm Convention in surface sediments from the Liaohe River basin, China.
397 *Journal of Environmental Science and Health, Part B* 50, 322-330.

398 Harris, C.A., Scott, A.P., Johnson, A.C., Panter, G.H., Sheahan, D., Roberts, M.,
399 Sumpter, J.P., 2014. Principles of Sound Ecotoxicology. *Environmental Science &*
400 *Technology* 48, 3100-3111.

401 He, Z., Hu, P., Yu, Y., 2011. Distribution and Source Analysis of Classic Persistent
402 Organic Pollutants in Sediments from Dagu Drainage Canal, Tianjin, China. *Journal of*
403 *Agro-Environment Science* 10, 031.

404 Hebei Government, 2015. Hebei Economic Yearbook. China Statistics Press, Beijing.

405 Jiao, W., Wang, T., Khim, J.S., Luo, W., Hu, W., Naile, J.E., Giesy, J.P., Lu, Y., 2013.
406 Polycyclic aromatic hydrocarbons in soils along the coastal and estuarine areas of the
407 northern Bohai and Yellow Seas, China. *Environmental Monitoring & Assessment* 185,
408 8185-8195.

409 Jiao, W.T., 2012. PAHs in surface sediments from coastal and estuarine areas of the
410 northern Bohai and Yellow Seas, China. *Environmental Geochemistry and Health* 34,
411 445-456.

412 Letcher, R.J., Bustnes, J.O., Dietz, R., Jenssen, B.M., Jorgensen, E.H., Sonne, C.,
413 Verreault, J., Vijayan, M.M., Gabrielsen, G.W., 2010. Exposure and effects
414 assessment of persistent organohalogen contaminants in arctic wildlife and fish.
415 *Science Of the Total Environment* 408, 2995-3043.

416 Li, C., Zheng, M., Gao, L., Zhang, B., Liu, L., Xiao, K., 2013. Levels and distribution of
417 PCDD/Fs, dl-PCBs, and organochlorine pesticides in sediments from the lower
418 reaches of the Haihe River basin, China. *Environmental monitoring and assessment*
419 185, 1175-1187.

420 Li, F., Sun, H., Hao, Z., He, N., Zhao, L., Zhang, T., Sun, T., 2011. Perfluorinated
421 compounds in Haihe River and Dagu Drainage Canal in Tianjin, China. *Chemosphere*
422 84, 265-271.

423 Li, H., Zhang, Q., Wang, P., Li, Y., Lv, J., Chen, W., Geng, D., Wang, Y., Wang, T.,
424 Jiang, G., 2012. Levels and distribution of hexabromocyclododecane (HBCD) in
425 environmental samples near manufacturing facilities in Laizhou Bay area, East China.
426 *Journal of Environmental Monitoring* 14, 2591-2597.

427 Liaoning Statistical Bureau, 2015. *Liaoning Statistical Yearbook*. China Statistics Press,
428 Beijing.

429 Liu, F., Liu, J., Chen, Q., Wang, B., Cao, Z., 2012. Pollution characteristics, ecological
430 risk and sources of polycyclic aromatic hydrocarbons (PAHs) in surface sediment from
431 Tuhai-Majia River system, China. *Procedia Environmental Sciences* 13, 1301-1314.

432 Liu, M., Cheng, S., Ou, D., Yang, Y., Liu, H., Hou, L., Gao, L., Xu, S., 2008.
433 Organochlorine pesticides in surface sediments and suspended particulate matters
434 from the Yangtze estuary, China. *Environ. Pollut.* 156, 168-173.

435 Lohmann, R., Breivik, K., Dachs, J., Muir, D., 2007. Global fate of POPs: Current and
436 future research directions. *Environmental Pollution* 150, 150-165.

437 Meng, J., Wang, T., Wang, P., Zhu, Z., Li, Q., Lu, Y., 2015. Perfluoroalkyl Substances
438 in Daling River Adjacent to Fluorine Industrial Parks: Implication from Industrial
439 Emission. *Bulletin of environmental contamination and toxicology* 94, 34-40.

440 National Bureau of Statistics, 2014. *Statistical Bulletin on Domestic Economics and*
441 *Social Development*.

442 Oh, J.H., Moon, H.-B., Choe, E.S., 2013. Alterations in Differentially Expressed Genes
443 After Repeated Exposure to Perfluorooctanoate and Perfluorooctanesulfonate in Liver
444 of *Oryzias latipes*. *Archives of Environmental Contamination and Toxicology* 64, 475-
445 483.

446 Ratcliffe, D.A., 1967. Decrease in eggshell weight in certain birds of prey. *Nature* 215,
447 208-210.

448 Shandong Statistical Bureau, 2015. *Shandong Statistical yearbook*. China Statistics
449 Press, Beijing.

450 Tan, L., He, M., Men, B., Lin, C., 2009. Distribution and sources of organochlorine
451 pesticides in water and sediments from Daliao River estuary of Liaodong Bay, Bohai
452 Sea (China). *Estuarine Coastal & Shelf Science* 84, 119-127.

453 Tao, S., Li, B.G., He, X.C., Liu, W.X., Shi, Z., 2007. Spatial and temporal variations
454 and possible sources of dichlorodiphenyltrichloroethane (DDT) and its metabolites in
455 rivers in Tianjin, China. *Chemosphere* 68, 10-16.

456 Tianjin Statistical Bureau, 2015. *Tianjin Statistical Yearbook*. China Statistics Press,
457 Beijing.

458 Wang, L., Jia, H., Liu, X., Sun, Y., Yang, M., Hong, W., Qi, H., Li, Y.-F., 2013. Historical
459 contamination and ecological risk of organochlorine pesticides in sediment core in
460 northeastern Chinese river. *Ecotoxicology and environmental safety* 93, 112-120.

461 Wang, M., Wang, C., Hu, X., Zhang, H., He, S., Lv, S., 2015a. Distributions and
462 sources of petroleum, aliphatic hydrocarbons and polycyclic aromatic hydrocarbons
463 (PAHs) in surface sediments from Bohai Bay and its adjacent river, China. *Marine
464 pollution bulletin* 90, 88-94.

465 Wang, P., Lu, Y., Wang, T., Fu, Y., Zhu, Z., Liu, S., Xie, S., Xiao, Y., Giesy, J.P., 2014.
466 Occurrence and transport of 17 perfluoroalkyl acids in 12 coastal rivers in south Bohai
467 coastal region of China with concentrated fluoropolymer facilities. *Environmental
468 Pollution* 190, 115-122.

469 Wang, R.M., Tang, J.H., Xie, Z.Y., Mi, W.Y., Chen, Y.J., Wolschke, H., Tian, C.G., Pan,
470 X.H., Luo, Y.M., Ebinghaus, R., 2015b. Occurrence and spatial distribution of
471 organophosphate ester flame retardants and plasticizers in 40 rivers draining into the
472 Bohai Sea, north China. *Environmental Pollution* 198, 172-178.

473 Wong, M.H., Leung, A.O.W., Chan, J.K.Y., Choi, M.P.K., 2005. A review on the usage
474 of POP pesticides in China, with emphasis on DDT loadings in human milk.
475 Chemosphere 60, 740-752.

476 Yunker, M.B., Macdonald, R.W., Vingarzan, R., Mitchell, R.H., Goyette, D., Sylvestre,
477 S., 2002. PAHs in the Fraser River basin: a critical appraisal of PAH ratios as indicators
478 of PAH source and composition. Organic Geochemistry 33, 489-515.

479 Zhang, P., Song, J., Yuan, H., 2009. Persistent organic pollutant residues in the
480 sediments and mollusks from the Bohai Sea coastal areas, North China: An overview.
481 Environment International 35, 632-646.

482 Zhang, Y., Li, Q., Lu, Y., Jones, K., Sweetman, A.J., 2016. Hexabromocyclododecanes
483 (HBCDDs) in surface soils from coastal cities in North China: Correlation between
484 diastereoisomer profiles and industrial activities. Chemosphere 148, 504-510.

485 Zhao, X., Zheng, M., Liang, L., Zhang, Q., Wang, Y., Jiang, G., 2005. Assessment of
486 PCBs and PCDD/Fs along the Chinese Bohai Sea coastline using mollusks as
487 bioindicators. Archives of Environmental Contamination and Toxicology 49, 178-185.

488 Zheng, B., Wang, L., Lei, K., Nan, B., 2016. Distribution and ecological risk assessment
489 of polycyclic aromatic hydrocarbons in water, suspended particulate matter and
490 sediment from Daliao River estuary and the adjacent area, China. Chemosphere 149,
491 91-100.

492 Zhou, R., Zhu, L., Yang, K., Chen, Y., 2006. Distribution of organochlorine pesticides
493 in surface water and sediments from Qiantang River, East China. Journal of Hazardous
494 Materials 137, 68-75.

495 Zhu, Z., Wang, T., Meng, J., Wang, P., Li, Q., Lu, Y., 2015. Perfluoroalkyl substances
496 in the Daling River with concentrated fluorine industries in China: seasonal variation,
497 mass flow, and risk assessment. *Environmental Science and Pollution Research* 22,
498 10009-10018.

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500