

## Article (refereed) - postprint

---

Newell, Mark; Wanless, Sarah; Harris, Michael P.; Daunt, Francis. 2015.  
**Effects of an extreme weather event on seabird breeding success at a North Sea colony.** *Marine Ecology Progress Series*, 532. 257-268.  
[10.3354/meps11329](https://doi.org/10.3354/meps11329)

© 2015 Inter-Research

This version available <http://nora.nerc.ac.uk/511490/>

NERC has developed NORA to enable users to access research outputs wholly or partially funded by NERC. Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at <http://nora.nerc.ac.uk/policies.html#access>

**This document is the author's final manuscript version of the journal article, incorporating any revisions agreed during the peer review process. There may be differences between this and the publisher's version. You are advised to consult the publisher's version if you wish to cite from this article.**

The definitive version is available at <http://www.int-res.com/home/>

Contact CEH NORA team at  
[noraceh@ceh.ac.uk](mailto:noraceh@ceh.ac.uk)

1 **Effects of an extreme weather event on seabird breeding success at a North Sea colony**

2

3 Mark Newell<sup>1\*</sup>, Sarah Wanless<sup>1</sup>, Michael P. Harris<sup>1</sup> and Francis Daunt<sup>1</sup>

4

5 <sup>1</sup>Centre for Ecology & Hydrology, Bush Estate, Penicuik EH26 0QB, UK

6 \*Correspondence: [manew@ceh.ac.uk](mailto:manew@ceh.ac.uk)

7

8

9

## 10 **Abstract**

11

12 Quantifying the effects of extreme weather is a critical question in population ecology since  
13 climate models predict increased climate variability. Effects will vary among and within  
14 species due to exposure or susceptibility, yet few studies have considered these sources of  
15 variation simultaneously. We investigated the effects of a summer storm on breeding  
16 success of four seabird species at a North Sea colony in relation to aspect, height above sea-  
17 level, distance to cliff edge and laying date. The storm lasted 8 hours with gusts of  $>60 \text{ ms}^{-1}$ .  
18 In exposed plots, razorbills *Alca torda* had higher failure rates (28.5%) than European shags  
19 *Phalacrocorax aristotelis* (15.1%), black-legged kittiwakes *Rissa tridactyla* (15.6%) or  
20 common guillemots *Uria aalge* (10.4%). Conversely, failure rates in sheltered plots were  
21 negligible (shags 0.0%; kittiwakes 1.9%; no guillemot or razorbill plots in sheltered  
22 locations). Guillemots breeding closer to sea-level were more likely to fail, but cliff edge  
23 proximity did not affect failure rate. In razorbills, pairs that laid early were more likely to  
24 survive the storm. In all species, some failed pairs relaid, and success of relays was lower  
25 than pairs that survived. Thus, relaying only provided partial compensation and, overall, the  
26 storm caused a net reduction in annual population production of 4.6%, 10.7%, 8.9% and  
27 22.8% for shags, kittiwakes, guillemots and razorbills, respectively. Increased storm  
28 frequency may therefore have important consequences on seabird populations, but orientation  
29 of storms relative to colonies and timing in relation to the breeding season are likely to be  
30 critical in determining the overall effect.

31

32 **Keywords:** climate change; summer storm; rainfall; IPCC; European shag; black-legged  
33 kittiwake; common guillemot; razorbill

## 34 **Introduction**

35

36 Climate change is having a dramatic effect on the population dynamics of many animal  
37 species, and much research has focussed on the effects of mean temperature, typically at  
38 annual or decadal scales (Walther et al. 2002, Thomas et al. 2004). However, there is  
39 increasing evidence that populations are also affected by climate variability (Parmesan et al.  
40 2000, Moreno & Moller 2011). These effects warrant further investigation since climate  
41 models predict that mean wind speeds and the frequency of severe weather is going to  
42 increase in some regions in the future, in particular at higher latitudes (McInnes et al. 2011,  
43 Young et al. 2011). Such events can take different forms, including extremes of temperature,  
44 high rainfall or strong winds, and usually operate at much shorter time scales than changes in  
45 mean climate, typically hours or days rather than years or decades. The effects of extreme  
46 weather are likely to vary amongst species due to differences in their ecology and life history.  
47 Furthermore, differences are likely among individuals within species due to variation in  
48 exposure or susceptibility. Quantifying variation among and within species is therefore  
49 critical to understanding the impacts of extreme weather events on animal populations.

50 A number of studies have shown that the survival and productivity of seabirds can be  
51 affected by extreme weather (Schreiber 2001, Jenouvrier 2013). Extreme weather events are  
52 likely to be important outside the breeding season since this is typically when most adult  
53 mortality occurs, and the population dynamics of seabirds are generally more sensitive to  
54 changes in adult survival rates than changes in breeding success (Weimerskirch 2001).  
55 However, it has proved challenging to attribute variation in adult survival rates to extreme  
56 winter weather because comprehensive data on the timing and location of deaths are not  
57 typically available (but see Frederiksen et al. 2008). Extreme weather may also be important  
58 during the breeding season since individuals are constrained to remain at or close to the  
59 breeding colony. At this time, adults and chicks can be affected by heat stress (Gaston et al.

60 2002, Oswald et al. 2008, Oswald & Arnold 2012), and high winds and rainfall can result in  
61 breeding failure (White et al. 1976, King et al. 1992, Aebischer 1993, Hennicke &  
62 Flachsbarth 2009, Mallory et al. 2009, Sherley et al. 2012, Wolfaardt et al. 2012, Boersma &  
63 Rebstock 2014, Bonter et al. 2014).

64 While severe weather is widely understood to reduce seabird breeding success, studies  
65 that quantify variation amongst and within species are needed to investigate the consequences  
66 of extreme weather on seabird communities (Wolfaardt et al. 2012). Such heterogeneity may  
67 arise from differences in exposure or susceptibility to wind, waves or rainfall. Thus, species  
68 may vary in susceptibility due to physical size or attributes of the nest site. Within species,  
69 nests that are oriented in the direction of the storm, closer to sea level and nearer to the cliff  
70 edge are likely to be more vulnerable. Intrinsic factors may also be important. For example,  
71 pairs that lay early in the season typically have higher average breeding success than those  
72 that lay late. Early breeders may therefore have greater ability or willingness to withstand  
73 bad weather, may occupy higher quality nest sites that are less exposed and be more likely to  
74 relay if they do fail (Potts et al. 1980, Hipfner et al. 1999). On 23<sup>rd</sup> May 2011, a severe  
75 storm was forecast to hit eastern Scotland. This gave us the opportunity to compare the short  
76 term impact of prolonged gale force winds and rough seas on the breeding success of four  
77 cliff-nesting species, the European Shag *Phalacrocorax aristotelis* (hereafter ‘shag’), black-  
78 legged kittiwake *Rissa tridactyla* (hereafter ‘kittiwake’), common guillemot *Uria aalge*  
79 (hereafter ‘guillemot’) and razorbill (*Alca torda*) at a major breeding colony in the region.  
80 Within species, we tested whether the effect of the storm had a disproportionate effect on  
81 nests located on the exposed (south-westerly) side of the island, closer to sea level and nearer  
82 the cliff edge. We also tested whether a pair’s laying date was important in determining  
83 failure rate in the storm. Finally, we quantified the capacity of species to compensate for

84 losses sustained in the storm by relaying, allowing us to estimate the likely net effect of the  
85 storm on annual population production.

86

## 87 **Methods**

88

89 Fieldwork took place during the 2011 breeding season on the Isle of May National Nature  
90 Reserve, south-east Scotland (56° 11' N, 02° 33' W). The island is oriented on a north-  
91 west/south-east axis with high cliffs facing predominantly to the south-west and gently  
92 sloping rocky terrain facing predominantly to the north-east (Fig 1). Breeding phenology and  
93 success of a sample of shag, kittiwake, guillemot and razorbill nest sites were collected at  
94 long established monitoring plots using standardized methods (Walsh et al. 1995). For shags,  
95 104 nest sites at 11 plots (79 nest sites in 8 plots facing south-west, 25 nest sites in 3 plots  
96 facing north-east; Fig 1) were checked every 7 days from before laying to fledging, and the  
97 laying date (within 7 days, taken to be half-way between the first date incubation is observed  
98 and the previous date), number of chicks fledged (range 0-4) and, for unsuccessful pairs, date  
99 of failure (minimum accuracy 7 days) were recorded. For kittiwakes, 166 nests in 6 plots  
100 were checked every 5 days from pre-laying to fledging, and as with shags, the laying date  
101 (minimum accuracy 5 days) the number of chicks fledged (range 0-3) and, for unsuccessful  
102 pairs, date of failure (minimum accuracy 5 days) were recorded. At 9 additional kittiwake  
103 plots, 283 nests were checked when most pairs had finished laying and again from the day  
104 after the first fledged chick was seen in the colony, and the number of chicks fledged was  
105 recorded (Harris 1987). Thus, the total sample size for kittiwakes was 449 nest sites in 15  
106 plots of which 397 were in 12 south-west facing plots and 52 in 3 north-east facing plots (Fig.  
107 1). For guillemots, 828 nest sites at 6 plots (all on south-west facing cliffs; Fig 1) were  
108 checked daily from before laying to fledging, and laying date, breeding success (i.e whether

109 the single chick fledged, since guillemots only lay one egg) and, where applicable, date of  
110 failure was recorded. The protocol for razorbills, which also lay one egg, was similar to that  
111 of guillemots (n = 173 nest sites at 5 of the six plots followed for guillemots; Fig 1). For  
112 guillemot and razorbill nest sites, height above sea level was measured using a marked rope  
113 (range: guillemots, 3-27m; razorbills, 4-26m; Harris et al. 1997). Height above sea level was  
114 not known for shag or kittiwake nests. In one guillemot plot the majority of breeding sites  
115 were located on a series of broad flat ledges and, for these sites, straight line distance from  
116 the cliff edge was measured (n = 250; horizontal distance range: 0.2-3.8m; height above sea  
117 level: 5m; Harris et al. 1997).

118 The storm occurred on 23<sup>rd</sup> May 2011. To assess its severity relative to summer weather  
119 conditions over the last 40 years, hourly wind speeds (mean speed and maximum gust speed)  
120 were extracted from the weather station at Leuchars (56° 23' N, 02° 52' W; 28km from the  
121 Isle of May; source: [www.badc.ac.uk](http://www.badc.ac.uk)) for dates between 1<sup>st</sup> April and 15<sup>th</sup> July each year  
122 from 1969, when hourly records began. These dates covered the core breeding periods of the  
123 four study species. The storm was forecast in advance and its strength was predicted to be of  
124 such magnitude that we considered it important to quantify its effects, so a full check of  
125 breeding status at each study nest was carried out on the day before and again on the day after  
126 the storm. This constituted a departure from standard monitoring frequency for shags and  
127 kittiwakes to ensure that any breeding failures over that period could be unequivocally  
128 attributed to the storm. Since guillemots and razorbills were being monitored on a daily basis,  
129 the effects of the storm could be estimated without the need to depart from standard  
130 protocols. Monitored nest sites were categorised as follows: a) failed before the storm; b)  
131 failed during the storm; c) survived the storm; d) eggs laid for the first time after the storm.  
132 Some pairs that failed before or during the storm (categories a) and b), respectively) relaid  
133 after the storm. These relays were monitored in the same way as other breeding attempts so

134 that final breeding outcome was known for all study pairs. In most cases, we were unable to  
135 confirm that the second laid egg was relayed by the same pair, as opposed to a different pair  
136 occupying the site after the failure of the first pair. However, these species are aggressively  
137 territorial and our long-term studies of these populations has shown that pairs rarely change  
138 sites when relaying. In this study, the same colour-ringed individual was in the pair  
139 associated with the second egg in 5/5 cases for guillemots, 1/1 for razorbill and 7/8 for shags.  
140 The extent to which relaying compensated for failure during the storm (“percentage  
141 compensation”) was estimated as the number of chicks raised as a percentage of the number  
142 that could have been raised had all failed nests relaid and been as successful as those that  
143 survived the storm. Thus, full compensation would have a value of 100%.

144 Our principal variables of interest in analyses of within-species variation in effects of the  
145 storm were aspect, height above sea level, distance from cliff edge and laying date. Where  
146 possible, we also tested whether the effects of exposure (aspect, height and distance) were  
147 dependent on laying date since early breeders may have greater ability or willingness to  
148 withstand bad weather, and therefore we might expect any difference between early and late  
149 layers to become more marked as exposure increased. We could not test the interaction  
150 between aspect and laying date in shags since the former could only be estimated with a  
151 randomisation test (see below). We were also not able to test the interaction between aspect  
152 and laying date in kittiwakes, since the plots where laying date was recorded were all on the  
153 exposed side of the island. For guillemots, we tested the effects of height above sea level,  
154 laying date and the interaction between them (correlation between height above sea level and  
155 laying date:  $r = 0.08$ ). In a separate analysis, we tested the effects of distance from cliff edge,  
156 laying date and the interaction between them (correlation between distance from cliff edge  
157 and laying date:  $r = 0.11$ ). The former model was based on a much larger sample size, so was  
158 a more comprehensive test of the effect of laying date. However, the effects of laying date in

159 the latter model were qualitatively similar. For razorbills, we tested the effects of height  
160 above sea level, laying date and the interaction between them (correlation between height  
161 above sea level and laying date:  $r = -0.01$ ).

162 To test whether nests in south-westerly plots were more affected by the storm than those  
163 in north-easterly plots for kittiwakes, we carried out a binomial Generalized Linear Mixed  
164 Model (GLMM) with logit link function on all breeding attempts active on the day of the  
165 storm with failure in the storm (0 or 1) as the response variable, aspect (north-east or south-  
166 west) as a fixed effect and plot as a random effect. We could not take this approach with  
167 shags since the north-east orientation contained no failed nests, which makes this parameter  
168 inestimable in a model-based test. Instead we carried out a randomization test (Fisher Exact  
169 Test) on the number of nests active on the day of the storm that failed or survived in south-  
170 western and north-eastern plots. To test the effect of height above sea level, laying date and  
171 the interaction between them on nest survival from the storm in guillemots and razorbills, we  
172 carried out a binomial GLMM on each species on all breeding attempts active on the day of  
173 the storm with failure in the storm (0 or 1) as the response variable, height above sea level (in  
174 metres), laying date and a height above sea level by laying date interaction as fixed effects  
175 and plot as a random effect. To test whether guillemots nesting closer to the cliff edge were  
176 more affected, and whether there was an interaction between distance to cliff edge and laying  
177 date, we used a binomial GLM on all active nests with failure in the storm (0 or 1) as the  
178 response variable and distance from the cliff edge (in metres), laying date and a distance to  
179 cliff edge by laying date interaction as fixed effects (nests in this analysis were from a single  
180 plot so a GLMM was not required). To test whether early or late breeding birds were more  
181 likely to survive the storm in kittiwakes and shags, we carried out a binomial GLMM on all  
182 active nests with failure in the storm (0 or 1) as the response variable, laying date as a fixed  
183 effect and plot as a random effect. For shags, we repeated the analysis, substituting breeding

184 stage (incubation or chick-rearing) for laying date, to test whether the effect of the storm was  
185 dependent on whether pairs were incubating eggs or brooding chicks. This breeding stage  
186 test was not undertaken for the other species since all pairs were incubating.

187 To test whether early or late breeding birds were more likely to relay after loss in the  
188 storm, we carried out a binomial GLMM on all nests that failed in the storm with relay  
189 incidence (0 or 1) as the response variable, laying date as a fixed effect and plot as a random  
190 effect. We substituted laying date for breeding stage in shags to test whether individuals that  
191 were incubating eggs at the time of the storm were more likely to relay than those that were  
192 rearing young. Laying date was not available in a small number of cases which is reflected in  
193 the slightly smaller sample sizes in these analyses. All statistical analyses were carried out in  
194 Genstat 16.

195 To quantify the net effect of the storm on the annual population production of each species,  
196 we first estimated the predicted breeding success of nests that failed in the storm had the  
197 storm not occurred. The simple approach to this estimation assumed that, but for the storm,  
198 pairs that failed would have been as successful as those in the same plot that were active at  
199 the time of the storm but survived. However, it is possible that the storm affected nests  
200 whose breeding success was higher or lower than average e.g. the latter might occur if young  
201 breeders tend to occupy more exposed sites, but also have lower foraging efficiency thus  
202 increasing the probability of chick mortality from starvation (Daunt et al. 2007). To examine  
203 this possibility, we compared the past breeding success of nest sites that survived the storm  
204 with those that did not in shags (data from 1996-2010), guillemots (1981-2010) and razorbills  
205 (1982-2010). In guillemots, long-term breeding success of nest sites that failed during the  
206 storm was significantly lower (by 4.7%) than breeding success of nest sites that survived the  
207 storm (Appendix A). We therefore reduced the predicted breeding success of pairs that failed  
208 in the storm by 4.7%. In shags and guillemots, we found no significant difference (Appendix

209 A), so we used the simple approach outlined above. For kittiwakes, individual nest identity  
210 was not retained across years, so we used the simple approach.

211 In a second step, we estimated the predicted mean breeding success in exposed  
212 monitoring plots (facing south-west) and sheltered monitoring plots (facing north-east)  
213 separately. For both groups, predicted mean breeding success was estimated as the average  
214 across all nests based on the observed breeding success of pairs unaffected by the storm (i.e.  
215 those that failed before the storm, survived the storm or laid after the storm) and predicted  
216 breeding success of those pairs that failed in the storm as outlined above. We then  
217 extrapolated the mean predicted and observed breeding success in exposed monitoring plots  
218 to the proportion of the whole population in the south-western part of the island on the  
219 assumption that exposure was similar across all nests with this orientation, inside and outside  
220 the monitoring plots. We carried out an identical extrapolation from sheltered monitoring  
221 plots to the proportion of the population in the north-eastern part of the island. Since there  
222 were no guillemot or razorbill monitoring plots in the north-east, we assumed that no nests  
223 with this orientation failed in the storm based on failure rates of north-eastern facing shag and  
224 kittiwake monitoring plots (see results). Relative proportions of the population in south-  
225 western and north-eastern parts of the colony were estimated from whole-island population  
226 counts (Pickett & Squire 2011). Equivalent extrapolations from monitoring plots to the  
227 population as a whole based on height above sea level or distance to cliff edge were not  
228 feasible since these measures were not available for the majority of nests.

229 For each species, we combined the results for the two parts of the colony into estimates of  
230 predicted and observed breeding success for the whole population as follows:

231

232 Predicted breeding success = (predicted breeding success in SW plots \* propn. of population  
233 in SW) + (predicted breeding success in NE plots \* propn. of population in NE)

234

235 Observed breeding success = (observed breeding success in SW plots \* propn. of population  
236 in SW) + (observed breeding success in NE plots \* propn. of population in NE)

237

238 Finally, we combined these two estimates to calculate the net effect of the storm on  
239 annual population production as follows:

240

241 Net effect = (predicted breeding success – observed breeding success) / predicted breeding  
242 success %

243

## 244 **Results**

245

246 The storm lasted approximately eight hours, with mean wind speeds of  $36.3\text{ms}^{-1}$  and  
247 maximum hourly gust speeds  $>50\text{ms}^{-1}$  recorded throughout the period from 12.00h to 20.00h.

248 The storm came from a westerly direction and coincided with high tide. A comparison with  
249 historical weather data showed that it was the most severe summer storm since hourly records  
250 began in 1969, such that the four highest, and seven of the 10 highest hourly maximum gust  
251 speeds during April-mid July 1969-2011 occurred on this day (Fig. 2).

252 Since the storm came from a westerly direction, the majority of monitoring plots were  
253 exposed since they were positioned on the south-west side of the island (shag: 8 out of 11  
254 plots; kittiwake: 12 out of 15 plots; guillemot: 6 out of 6 plots; razorbill: 5 out of 5 plots).

255 The storm occurred during early to mid-incubation for kittiwake, late incubation for  
256 guillemot and razorbill, and late incubation to early chick-rearing for shag (median lay dates:  
257 shag: 11<sup>th</sup> April; kittiwake: 10<sup>th</sup> May; guillemot: 23<sup>rd</sup> April; razorbill: 26<sup>th</sup> April), with laying  
258 almost complete in all species (percentage pairs that laid before the storm: shag 96.2%;

259 kittiwake 98.4%; guillemot 99.5%; razorbill 95.4%; Table 1). A small proportion of nests  
260 failed before the storm, so the percentages of nests that were active when the storm occurred  
261 for shag, kittiwake, guillemot and razorbill were 91.3%, 97.3%, 96.4% and 91.3%  
262 respectively.

263 The storm had a similar impact on shag, kittiwake and guillemot, with 11.5%, 14.0% and  
264 10.4% of active nests failing, respectively. However, a higher percentage of razorbill nests  
265 was affected (28.5%; Table 1). There was a tendency for south-westerly facing shag nests to  
266 be more vulnerable to the storm than north-easterly nests, with 11/73 (15.1%) and 0/22  
267 (0.0%) failing in the storm, respectively (Fisher Exact Test:  $p = 0.06$ ). South-westerly facing  
268 kittiwake nests were significantly more vulnerable to the storm than north-easterly nests, with  
269 60/385 (15.6%) nests in south-westerly plots failing in the storm and 1/52 (1.9%) nests failing  
270 in north-easterly facing plots (GLMM:  $W = 4.05$ ;  $p < 0.05$ ). In guillemots, a lower failure  
271 rate was apparent with increasing height above sea level, but there was no effect of laying  
272 date and no interaction between height above sea level and laying date ( $n=768$ ; GLMM:  
273 height above sea level:  $W = 10.63$ ,  $p < 0.01$ , Fig 3; laying date:  $W = 0.70$ ,  $p = 0.40$ , Fig 4c;  
274 interaction term:  $W = 0.00$ ,  $p = 0.97$ ). In contrast, there was no effect of height above sea  
275 level on failure rate in razorbills, but those nests with an earlier laying date were more likely  
276 to survive; the interaction term was not significant ( $n=153$ ; GLMM: height above sea level:  
277  $W = 0.48$ ,  $p = 0.49$ , Fig 3; laying date:  $W = 5.40$ ,  $p < 0.05$ , Fig 4d; interaction term:  $W = 0.00$ ,  
278  $p = 0.96$ ). Failure rate was not related to distance from the cliff edge in guillemots, and there  
279 was no effect of laying date in this subset of nests (in line with findings from the larger  
280 sample), nor was there a significant interaction between these two variables ( $n = 250$ ; GLM:  
281 distance from cliff edge:  $W = 1.71$ ,  $p = 0.19$ ; laying date:  $W = 1.31$ ,  $p = 0.25$ ; interaction  
282 term:  $W = 3.08$ ;  $p = 0.08$ ). There was no effect of laying date on failure rate in shags ( $n = 95$ ;

283  $W = 0.09$ ,  $p = 0.76$ ; Fig 4a) or kittiwakes ( $n = 148$ ;  $W = 0.33$ ,  $p = 0.57$ ; Fig 4b), nor was  
284 breeding stage related to failure rate in shags ( $n = 95$ ;  $W = 0.06$ ;  $p = 0.80$ ).

285 Not all pairs that failed during the storm relaid (pairs relaying: shag: 45.5%; kittiwake:  
286 39.3%; guillemot: 25.3%; razorbill: 17.8%), and such pairs were ultimately less successful  
287 than those that survived the storm (Table 1). Thus, failure during the storm was only partially  
288 compensated for through re-laying (percentage compensation: shags: 19.6%; kittiwakes:  
289 16.3%; guillemots: 1.4%; razorbills: 0.0%). There was a tendency for early laying pairs to be  
290 more likely to relay among razorbills ( $n = 44$ ;  $W = 3.68$ ,  $p = 0.06$ ; Fig 4h). However, there  
291 was no relationship in shags ( $n = 11$ ;  $W = 0.05$ ,  $p = 0.82$ ; Fig 4e), kittiwakes ( $n=20$ ,  $W =$   
292  $0.00$ ,  $p = 0.98$ ; Fig 4f) or guillemots ( $n = 82$ ;  $W = 0.48$ ;  $p = 0.49$ ; Fig 4g), nor was breeding  
293 stage related to relay probability in shags ( $n = 11$ ;  $W = 0.13$ ;  $p = 0.71$ ).

294 The observed mean and predicted mean breeding successes had the storm not occurred of  
295 pairs in exposed and sheltered parts of the island are given in Table 2. The majority of  
296 kittiwakes, guillemots and razorbills were located in exposed areas (85-93%) whilst the  
297 majority of the shag population was breeding in the sheltered part of the island (69%, Table  
298 2). Combining mean predicted and observed breeding success in exposed and sheltered nests  
299 with these proportions, we estimated that the net reduction in annual population production of  
300 shags was 4.6%, reflecting the high proportion of the population located in sheltered areas. In  
301 contrast, estimates of net reduction in annual population production for the other three species  
302 matched closely those in the monitoring plots, demonstrating that the high impact of the  
303 storm on razorbills was apparent in the population as a whole (kittiwake: 10.7%; guillemot:  
304 8.9%; razorbill: 22.8%; Table 2). The greater impact of the storm on razorbills can be seen  
305 when comparing the overall breeding success in 2011 with the long-term mean for each  
306 species (shag: 1.54 in 2011 vs  $1.01 \pm 0.57$  mean  $\pm$  sd chicks/pair, 1985-2010; kittiwake 0.87  
307 vs  $0.55 \pm 0.38$ , 1985-2010; guillemot: 0.73 vs  $0.73 \pm 0.13$ , 1981-2010; razorbill 0.50 vs 0.66

308  $\pm 0.08$ , 1982-2010). Razorbill breeding success in 2011 was the worst on record (previous  
309 range 0.52 – 0.86 chicks/pair).

310

## 311 **Discussion**

312

313 Few studies have investigated among and within-species variation in the effects of extreme  
314 weather on seabird breeding success. We quantified these effects in four species of cliff  
315 nesting seabirds during the most severe summer storm recorded in the region in 40 years. We  
316 demonstrated important, and in some cases unexpected, effects of the extreme weather event  
317 on the breeding success of this seabird community. Our results supported our prediction that  
318 nests on the exposed side of the island and low down the cliff would be more severely  
319 affected. However, we did not foresee that razorbills would be much more vulnerable than  
320 the other species. Razorbills typically breed on more sheltered sites than the closely related  
321 guillemot (Olsthoorn & Nelson 1990). However, on the Isle of May the two species breed in  
322 close proximity and exposure to wind, waves and spray seemed likely to have been similar.  
323 Due to the ferocity of the wind it was not possible to directly observe losses in the storm.  
324 However, razorbills may have been more susceptible to being physically lifted off their sites,  
325 since they are 30% lighter than guillemots. This effect may have been particularly important  
326 higher up the cliff, where wind speeds are typically highest. If both species are susceptible to  
327 spray, whilst guillemots are more able to withstand high wind speeds, this may explain why  
328 we detected an effect of height above sea level on nest survival from the storm in guillemots  
329 but not in razorbills. Further, it may explain why vulnerability to the storm was unrelated to  
330 horizontal distance in guillemots, since level of spray is likely to be determined more by  
331 height than distance to the cliff edge. Whatever the mechanisms driving the variation among  
332 the two species in overall loss and the effect of height, the high impact of the storm on

333 razorbills resulted in the lowest breeding success at this colony since records began in 1982.  
334 In contrast, breeding success for the other three species was at or above the long term  
335 average, despite losses in the storm.

336 The disproportionate effect of the storm on exposed compared to sheltered locations had a  
337 strong influence on the population-level effects for each species in line with their breeding  
338 distribution on the island. Although shag nests in exposed plots were affected by the storm to  
339 a similar extent as kittiwakes and guillemots, the majority of nests of this species are located  
340 on the north-east side of the island, so the overall impact on the population was  
341 comparatively small. In contrast, the other three species are concentrated on the south-west  
342 side of the island, so were more exposed to the storm. The majority of severe summer storms  
343 recorded on the Isle of May over the last 40 years have been from a westerly direction (Fig  
344 2). This would suggest that these interspecific differences have occurred repeatedly in recent  
345 decades. However, in contrast to the other species, the distribution of shag nests on the island  
346 has changed substantially over this period. Thus, a westerly storm in May 1982 had a  
347 dramatic effect on shag breeding success because, at that time, the majority of the population  
348 bred on the south-west side of the island (Aebischer 1993). Our results therefore suggest that  
349 the effects of summer storms on breeding seabirds are likely to be strongly dependent on the  
350 direction of the storm relative to breeding sites. However, predicting future effects of  
351 extreme weather is challenging since both storm direction and changes in breeding  
352 distribution, such as we have observed in the shag population on the Isle of May over the last  
353 three decades, would have to be considered. The percentage of shag nests in exposed  
354 locations that failed was much lower in 2011 than 1982 (15% vs 49%). One possibility for  
355 this difference is that nests were on average closer to sea level when densities were higher in  
356 the south-west. Storm duration may also have been a contributory factor since the 1982

357 storm was longer, with gale force winds experienced for most of the day (Aebischer 1993;  
358 Fig 2).

359 A proportion of breeding pairs that failed in the storm relaid, and success of these pairs  
360 was significantly lower than those that survived the storm. In combination, these two effects  
361 resulted in only partial compensation for the storm, ranging from 0.0-19.6% across the four  
362 species. We do not know how typical this level of compensation is, but breeding success of  
363 those pairs unaffected by the storm was at or above the long-term average in all species,  
364 suggesting that, aside from the storm, environmental conditions were favourable. Thus, it is  
365 possible that compensation would be lower in years when overall conditions are poorer, since  
366 breeding individuals would likely be in poorer condition and therefore less likely to relay.  
367 These estimates only relate to breeding success, yet compensation may have been even lower  
368 in terms of longer term fitness since there is a decline in post-fledging survival with fledging  
369 date in shags and guillemots at this colony (Harris et al. 1994, Lahoz-Monfort et al. 2014).  
370 Post-fledging survival may also be related to the nutritional quality of the egg, which is likely  
371 to be lower on average in relaid eggs (Nager et al. 2000, Krist 2011). However, these  
372 potential effects require formal testing since a study of Brünnich's guillemots *Uria lomvia* in  
373 the Canadian Arctic found no difference in recruitment rates of individuals from first laid and  
374 replacement eggs (Hipfner 2001). Long term fitness effects of relaying may not be limited to  
375 impacts on chicks. The costs associated with relaying may also have reduced the survival  
376 probability of adults (Nager et al. 2001), which could have consequences for population size.

377 Our prediction that early laying pairs would be more likely to survive the storm was  
378 upheld in razorbills. Late breeders may be less able to withstand bad weather, or show  
379 reproductive restraint which may increase likelihood of abandonment in poor conditions  
380 (Williams 1966). Alternatively, they may occupy lower quality nest sites that are more  
381 exposed to wind and spray. It is not clear why laying date was not an important determinant

382 of nest survival from the storm in the other species, but one possibility is that the losses  
383 resulted from catastrophic events (e.g. a nest or clutch being washed away or dislodged by  
384 gusting wind) where intrinsic effects are less likely to play a role. We found a tendency that  
385 early laying pairs that failed in the storm were more likely to relay than late laying pairs in  
386 razorbills. As with the effect of laying date on nest survival probability, this may reflect  
387 intrinsic differences in ability or effort.

388 The relative timing of an extreme weather event is likely to be critical to the overall effect  
389 on breeding success. Losses to extreme weather may be higher during chick-rearing than  
390 incubation since chicks are sensitive to exposure to extreme rain, wind and temperature  
391 (White et al. 1976, Demongin et al. 2010, Boersma & Rebstock 2014). Furthermore, chick  
392 mortality during extreme weather may occur indirectly through a reduction in adult foraging  
393 success. A study of guillemots at this colony showed that during stormy weather, chick-  
394 rearing adults increased their foraging effort, caught smaller fish and showed reduced nest  
395 attendance (Finney et al. 1999). A recent study on southern rockhopper penguins (*Eudyptes*  
396 *chrysocome*) provides further evidence that wind affects foraging success (Dehnhard et al.  
397 2013). Such indirect effects of weather on breeding performance are likely to be more  
398 profound during chick-rearing than incubation, where there is greater capacity for the non-  
399 attending bird to extend the time away from the nest since no provisioning is required.  
400 However, it is possible that the storm was not of sufficient duration for these indirect effects  
401 to occur, since shags did not do worse than the other species, despite breeding being more  
402 advanced with some pairs rearing chicks at the time of the storm. The extent to which  
403 individuals compensate through replacing lost eggs may also be linked to the timing of  
404 extreme weather relative to the breeding season. Studies of shags, kittiwakes and guillemots  
405 (both common and Brünnich's) have shown that the proportion relaying and success of relays  
406 declines with date (Gaston & Nettleship 1981, Harris & Birkhead 1985, Aebischer 1993,

407 Wanless & Harris 1997, Daunt 2000, Coulson 2011). As shown in razorbills in this study,  
408 breeders that lay earlier in the season are more likely to relay than later breeders (Hipfner et  
409 al. 1999). Furthermore, at the individual level, relaying is more likely if failure occurs sooner  
410 after laying, perhaps linked to body condition which is on average higher at that time than  
411 later in the breeding season (unpublished data on guillemots on the Isle of May from 1982-  
412 present: correlation between days incubated prior to loss and probability of relaying:  $r = -$   
413 0.95). Thus, if the 2011 storm had occurred earlier in the breeding season, relaying might  
414 have compensated more fully for clutches that were lost. Conversely, reduced compensation  
415 from relaying is likely had the storm occurred later in the season. However, species  
416 differences are clearly apparent since shags had a comparatively high relay rate despite their  
417 breeding season being more advanced.

418 Extreme weather events are an understudied but potentially important driver of seabird  
419 breeding success (Schreiber 2001, Jenouvrier 2013). A number of studies have shown that  
420 extremes of temperature, high rainfall, strong wind and rough seas can all result in major  
421 offspring mortality (King et al. 1992, Aebischer 1993, Gaston et al. 2002, Oswald et al. 2008,  
422 Hennicke & Flachsbarth 2009, Mallory et al. 2009, Oswald & Arnold 2012, Sherley et al.  
423 2012, Wolfaardt et al. 2012, Boersma & Rebstock 2014, Bonter et al. 2014). Here, we have  
424 shown that the overall effect of extreme weather varies both among and within species, likely  
425 due to variation in exposure and susceptibility. Our study highlights the value in recording  
426 immediate impacts and compensation from relaying to enable estimation of the net effects on  
427 annual breeding output of such events. Quantifying the impacts of extreme weather on  
428 breeding success is likely to become increasingly important, since many models predict that  
429 their frequency is going to increase in some regions, in particular at higher latitudes (McInnes  
430 et al. 2011, Young et al. 2011). An isolated event such as this is unlikely to have a dramatic  
431 effect on population size, especially in seasons such as the study year where (razorbills

432 excepting) the breeding season was moderately good, despite the storm. However, a greater  
433 frequency of summer storms of this severity could result in discernible impacts on population  
434 size (Descamps et al. 2015), although orientation of storms relative to breeding sites, storm  
435 duration and timing in relation to the breeding season are likely to be critical in determining  
436 the overall effect.

437

## 438 **Acknowledgements**

439

440 We thank Bethany Nelson, Hanna Granroth-Wilding and Katherine Herborn for help with  
441 fieldwork, Adam Butler for statistical advice, Sarah Burthe for useful discussions and  
442 comments on the manuscript and three anonymous reviewers for very helpful comments. We  
443 thank the Joint Nature Conservation Committee for funding and Scottish Natural Heritage for  
444 access to the Isle of May National Nature Reserve. The work was conducted under research  
445 licences from Scottish Natural Heritage.

446

## 447 **References**

448

- 449 Aebischer NJ (1993) Immediate and delayed effects of a gale in late spring on the breeding of  
450 the shag *Phalacrocorax aristotelis*. *Ibis* 135:225-232
- 451 Boersma PD, Rebstock GA (2014) Climate change increases reproductive failure in  
452 Magellanic penguins. *Plos One* 9:e85602
- 453 Bonter DN, MacLean SA, Shah SS, Moglia MC (2014) Storm-induced shifts in optimal  
454 nesting sites: a potential effect of climate change. *J Ornithol* 155:631-638
- 455 Coulson JC (2011) The kittiwake. T & AD Poyser, London
- 456 Daunt F (2000) Age-specific breeding performance in the European shag *Phalacrocorax*  
457 *aristotelis*. Unpublished PhD thesis, University of Glasgow
- 458 Daunt F, Reed TE, Newell M, Burthe S, Phillips RA, Lewis S, Wanless S (2014)  
459 Longitudinal bio-logging reveals interplay between extrinsic and intrinsic carry-over  
460 effects in a long-lived vertebrate. *Ecology* 95:2077-2083
- 461 Daunt F, Wanless S, Harris MP, Money L, Monaghan P (2007) Older and wiser:  
462 improvements in breeding success are linked to better foraging performance in  
463 European shags. *Funct Ecol* 21:561-567

464 Dehnhard N, Ludynia K, Poisbleau M, Demongin L, Quillfeldt P (2013) Good days, bad  
465 days: wind as a driver of foraging success in a flightless seabird, the southern  
466 rockhopper penguin. *Plos One* 8:e79487

467 Demongin L, Poisbleau M, Strange IJ, Quillfeldt P (2010) Effects of severe rains on the  
468 mortality of southern rockhopper penguin (*Eudyptes chrysocome*) chicks and its  
469 impact on breeding success. *Ornitol Neotrop* 21:439-443

470 Descamps S, Tarroux A, Varpe O, Yoccoz NG, Tveraa T, Lorentsen SH (2015) Demographic  
471 effects of extreme weather events: snow storms, breeding success, and population  
472 growth rate in a long-lived Antarctic seabird. *Ecol Evol* 5: 314-325

473 Finney SK, Wanless S, Harris MP (1999) The effect of weather conditions on the feeding  
474 behaviour of a diving bird, the common guillemot *Uria aalge*. *J Avian Biol* 30:23-30

475 Frederiksen M, Daunt F, Harris MP, Wanless S (2008) The demographic impact of extreme  
476 events: stochastic weather drives survival and population dynamics in a long-lived  
477 seabird. *J Anim Ecol* 77:1020-1029

478 Gaston AJ, Hipfner JM, Campbell D (2002) Heat and mosquitoes cause breeding failures and  
479 adult mortality in an Arctic-nesting seabird. *Ibis* 144:185-191

480 Gaston AJ, Nettleship DN (1981) The thick-billed murre of Prince Leopold Island.  
481 Canadian Wildlife Service Monograph Series, Book 6

482 Harris MP (1987) A low-input method of monitoring kittiwake *Rissa tridactyla* breeding  
483 success. *Biol Conserv* 41:1-10

484 Harris MP, Birkhead TR (1985) Breeding ecology of the Atlantic Alcidae. In: Nettleship DN,  
485 Birkhead TR (eds) *The Atlantic Alcidae*. Academic Press, London

486 Harris MP, Buckland ST, Russell SM, Wanless S (1994) Post fledging survival to breeding  
487 age of shags *Phalacrocorax aristotelis* in relation to year, date of fledging and brood  
488 size. *J Avian Biol* 25:268-274

489 Harris MP, Wanless S, Barton TR, Elston DA (1997) Nest site characteristics, duration of use  
490 and breeding success in the guillemot *Uria aalge*. *Ibis* 139:468-476

491 Hennicke JC, Flachsbarth K (2009) Effects of cyclone Rosie on breeding red-tailed  
492 tropicbirds *Phaethon rubricauda* on Christmas Island, Indian Ocean. *Marine*  
493 *Ornithology* 37:175-178

494 Hipfner JM (2001) Fitness-related consequences of relaying in an arctic seabird: survival of  
495 offspring to recruitment age. *Auk* 118:1076-1080

496 Hipfner JM, Gaston AJ, Martin DL, Jones IL (1999) Seasonal declines in replacement egg-  
497 layings in a long-lived, Arctic seabird: costs of late breeding or variation in female  
498 quality? *J Anim Ecol* 68:988-998

499 Jenouvrier S (2013) Impacts of climate change on avian populations. *Global Change Biol*  
500 19:2036-2057

501 King BR, Hicks JT, Cornelius J (1992) Population changes, breeding cycles and breeding  
502 success over 6 years in a seabird colony at Michaelmas Bay, Queensland. *Emu* 92:1-  
503 10

504 Krist M (2011) Egg size and offspring quality: a meta-analysis in birds. *Biol Rev* 86:692-716

505 Lahoz-Monfort JJ, Harris MP, Morgan BJT, Freeman SN, Wanless S (2014) Exploring the  
506 consequences of reducing survey effort for detecting individual and temporal  
507 variability in survival. *J Appl Ecol* 51:534-543

508 Mallory ML, Gaston AJ, Forbes MR, Gilchrist HG (2009) Influence of weather on  
509 reproductive success of northern fulmars in the Canadian high Arctic. *Polar Biol*  
510 32:529-538

511 McInnes KL, Erwin TA, Bathols JM (2011) Global Climate Model projected changes in 10 m  
512 wind speed and direction due to anthropogenic climate change. *Atmos Sci Lett*  
513 12:325-333

514 Moreno J, Moller AP (2011) Extreme climatic events in relation to global change and their  
515 impact on life histories. *Curr Zool* 57:375-389

516 Nager RG, Monaghan P, Houston DC (2000) Within-clutch trade-offs between the number  
517 and quality of eggs: experimental manipulations in gulls. *Ecology* 81:1339-1350

518 Nager RG, Monaghan P, Houston DC (2001) The cost of egg production: increased egg  
519 production reduces future fitness in gulls. *J Avian Biol* 32:159-166

520 Olsthoorn JCM, Nelson JB (1990) The availability of breeding sites for some British  
521 seabirds. *Bird Study* 37:145-164

522 Oswald SA, Arnold JM (2012) Direct impacts of climatic warming on heat stress in  
523 endothermic species: seabirds as bioindicators of changing thermoregulatory  
524 constraints. *Integr Zool* 7:121-136

525 Oswald SA, Bearhop S, Furness RW, Huntley B, Hamer KC (2008) Heat stress in a high-  
526 latitude seabird: effects of temperature and food supply on bathing and nest  
527 attendance of great skuas *Catharacta skua*. *J Avian Biol* 39:163-169

528 Parmesan C, Root TL, Willig MR (2000) Impacts of extreme weather and climate on  
529 terrestrial biota. *B Am Meteorol Soc* 81:443-450

530 Pickett DA, Squire J (2011) Isle of May National Nature Reserve Annual Report 2011.  
531 Scottish Natural Heritage, Cupar.

532 Potts GR, Coulson JC, Deans IR (1980) Population dynamics and breeding success of the  
533 shag, *Phalacrocorax aristotelis*, on the Farne Islands, Northumberland. *J Anim Ecol*  
534 49:465-484

535 Schreiber EA (2001) Climate and weather effects on seabirds. In: Schreiber EA, Burger J  
536 (eds) *Biology of marine birds*. CRC Press, Boca Raton, FL

537 Sherley RB, Ludynia K, Underhill LG, Jones R, Kemper J (2012) Storms and heat limit the  
538 nest success of bank cormorants: implications of future climate change for a surface-  
539 nesting seabird in southern Africa. *J Ornithol* 153:441-455

540 Thomas CD, Cameron A, Green RE, Bakkenes M, Beaumont LJ, Collingham YC, Erasmus  
541 BFN, de Siqueira MF, Grainger A, Hannah L, Hughes L, Huntley B, van Jaarsveld  
542 AS, Midgley GF, Miles L, Ortega-Huerta MA, Peterson AT, Phillips OL, Williams  
543 SE (2004) Extinction risk from climate change. *Nature* 427:145-148

544 Walsh PM, Halley DJ, Harris MP, del Nevo A, Sim LMW, Tasker ML (1995) Seabird  
545 monitoring handbook for Britain and Ireland: a compilation of methods for survey and  
546 monitoring of breeding seabirds. Peterborough: Joint Nature Conservation  
547 Committee.

548 Walther GR, Post E, Convey P, Menzel A, Parmesan C, Beebee TJC, Fromentin JM, Hoegh-  
549 Guldborg O, Bairlein F (2002) Ecological responses to recent climate change. *Nature*  
550 416:389-395

551 Wanless S, Harris MP (1997) *Phalacrocorax aristotelis* shag. *Birds of the Western Palearctic*  
552 Update 1:3-13

553 Weimerskirch H (2001) Seabird demography and its relationship with the marine  
554 environment. In: Schreiber EA, Burger J (eds) *Biology of marine birds*. CRC Press,  
555 Boca Raton, FL

556 White SC, Robertson WB, Ricklefs RE (1976) The effect of Hurricane Agnes on growth and  
557 survival of tern chicks in Florida. *Bird Banding* 47:54-71

558 Williams GC (1966) *Adaptation and natural selection: a critique of some current evolutionary*  
559 *thought*. Princeton University Press, Princeton

560 Wolfaardt AC, Crofts S, Baylis AMM (2012) Effects of a storm on colonies of seabirds  
561 breeding at the Falkland Islands. *Marine Ornithology* 40:129-133

562 Young IR, Zieger S, Babanin AV (2011) Global trends in wind speed and wave height.  
563 *Science* 332:451-455

564

565

566

567 Table 1: Immediate effects of the storm on 23<sup>rd</sup> May, percentage that relaid and final breeding  
 568 success (mean chicks fledged per pair) of four species on the Isle of May in 2011.  
 569

	Shag	Kittiwake	Guillemot	Razorbill
Number of monitored nests	104	449	828	173
Number failed before storm	5	5	26	7
Number active when storm occurred	95	437	798	158
Number laid after storm	4	7	4	8
% active in storm and failed	11.6	14.0	10.4	28.5
% relaid after failing in storm	45.5	39.3	25.3	17.8
Breeding success, failed before storm	0.00	0.00	0.00	0.00
Breeding success, survived storm	1.86	1.01	0.84	0.74
Breeding success, failed in storm and relaid	0.80	0.42	0.05	0.00
Breeding success, failed in storm, all nests	0.36	0.16	0.01	0.00
Breeding success, laid after storm	0.00	0.00	0.75	0.38
Mean breeding success of study nests	1.54	0.87	0.73	0.50

570

571

572 Table 2: Input values for the estimate of net effect of the storm on annual population  
573 production (observed mean breeding success and predicted mean breeding success in  
574 exposed, sheltered and all sites, and proportion of the population in exposed and sheltered  
575 sites) and net effect as a percentage reduction of predicted breeding success. The observed  
576 breeding success for guillemots and razorbills in sheltered sites was estimated based on  
577 failure rates during the storm for shags and kittiwakes. Breeding success is mean chicks  
578 fledged per pair.  
579  
580

Species	Exposed			Sheltered			Whole population		
	Observed breeding success	Predicted breeding success	Propn. of population	Observed breeding success	Predicted breeding success	Propn. of population	Observed breeding success	Predicted breeding success	Net effect (%)
Shag	1.58	1.81	0.31	1.40	1.40	0.69	1.46	1.53	-4.6
Kittiwake	0.92	1.04	0.85	0.50	0.49	0.15	0.86	0.96	-10.7
Guillemot	0.73	0.81	0.93	0.81	0.81	0.07	0.73	0.81	-8.9
Razorbill	0.51	0.69	0.85	0.69	0.69	0.15	0.54	0.69	-22.8

581

582 **Figure legends**

583

584 Fig 1: Location of monitoring plots on the Isle of May National Nature Reserve.

585

586 Fig 2: Mean hourly wind speed on 23<sup>rd</sup> May 2011 and the seven other storms between April  
587 and mid-July 1969-2011 where maximum gust speed exceeded  $45 \text{ ms}^{-1}$  in five hours or more.

588 Seven of these storms were in a westerly direction and one in an easterly direction (3<sup>rd</sup> April  
589 1998).

590

591 Fig 3: Fitted lines ( $\pm 95\%$  C.I.) for nest survival from the storm in relation to height above sea  
592 level from the GLMMs for guillemots (solid lines;  $n = 768$  active nests when the storm  
593 occurred) and razorbills (dashed lines;  $n = 153$ ).

594

595 Fig 4: Fitted lines ( $\pm 95\%$  C.I.) for survival from the storm in relation to laying date for the  
596 GLMMs for a) shags ( $n=95$ ); b) kittiwakes ( $n=148$ ); c) guillemots ( $n=768$ ) and d) razorbills  
597 ( $n=153$ ); fitted lines ( $\pm 95\%$  C.I.) from the GLMMs of post-storm relaying in relation to  
598 laying date for e) shags ( $n=11$ ); f) kittiwakes ( $n=20$ ); g) guillemots ( $n=82$ ) and h) razorbills  
599 ( $n=44$ ).

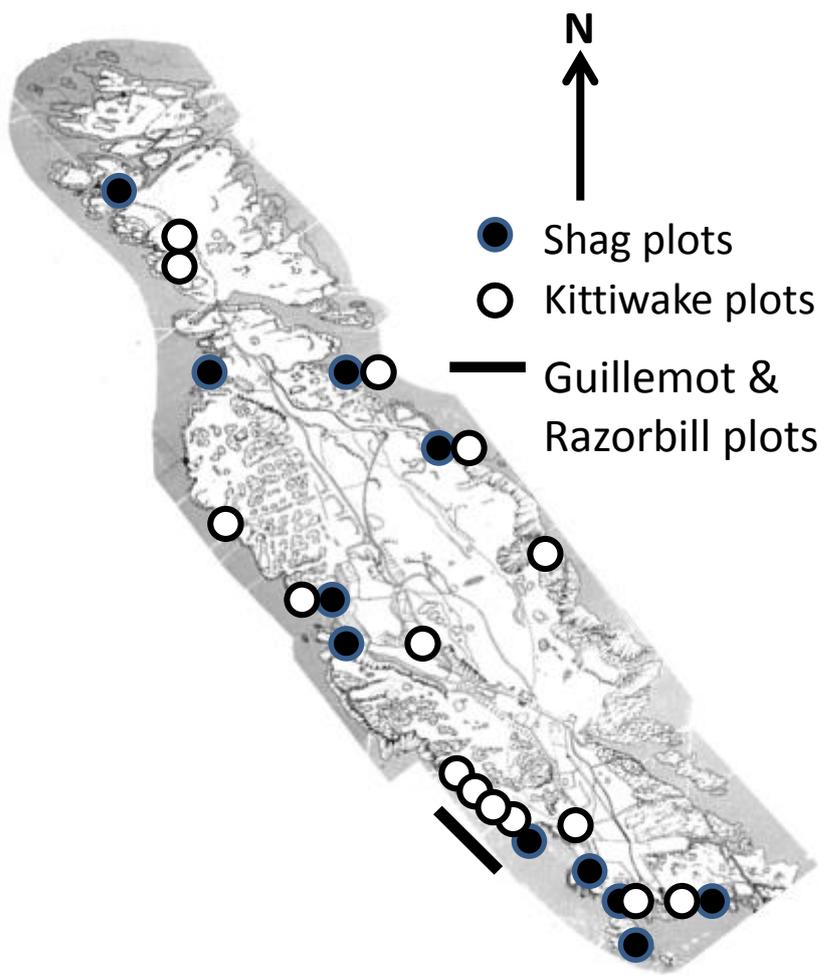
600

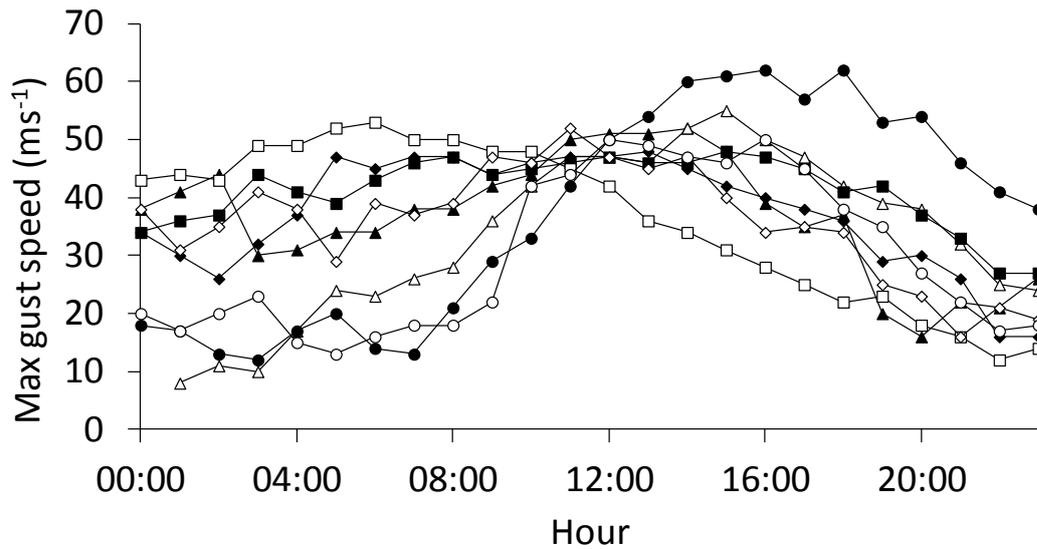
601

602 Fig 1

603

604

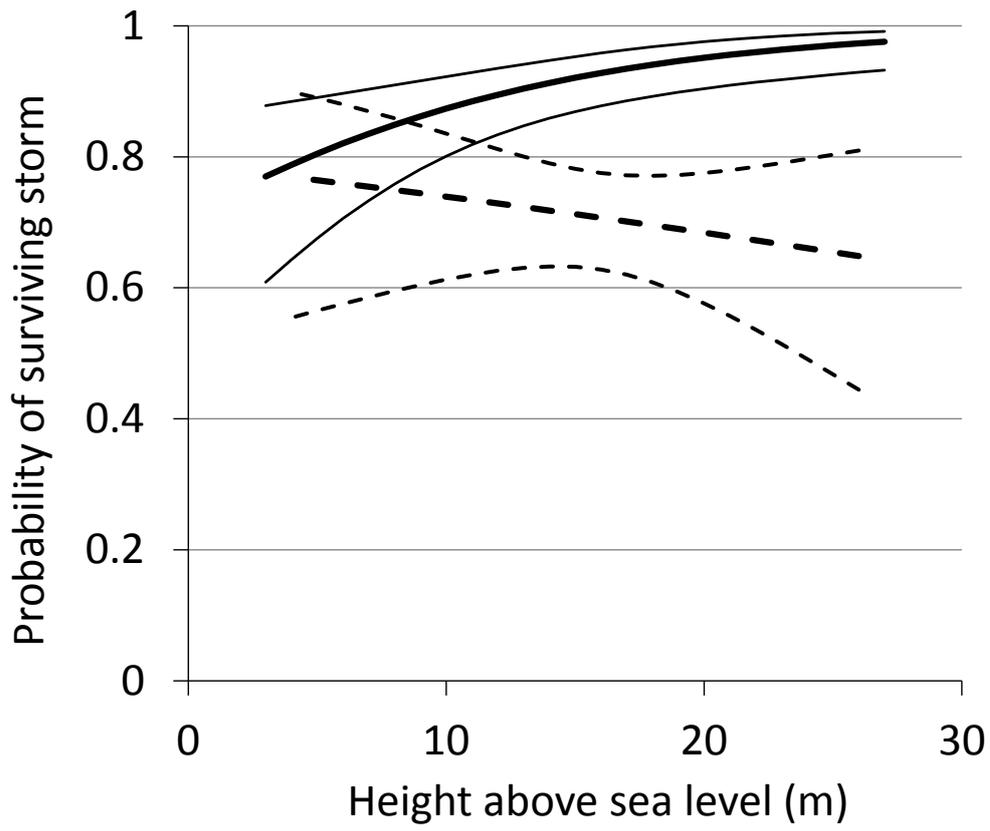




605

606

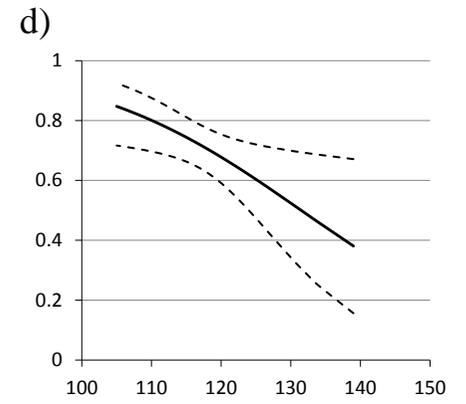
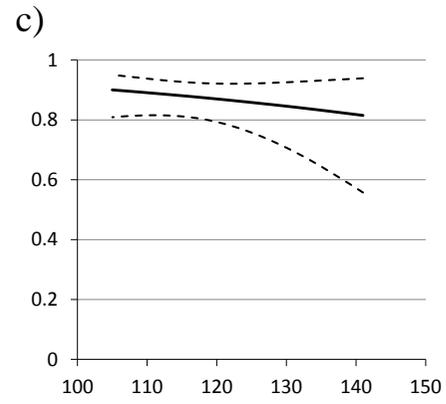
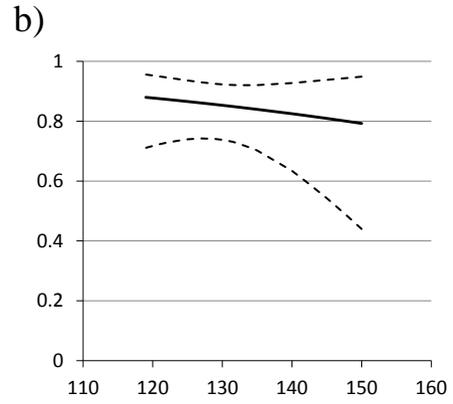
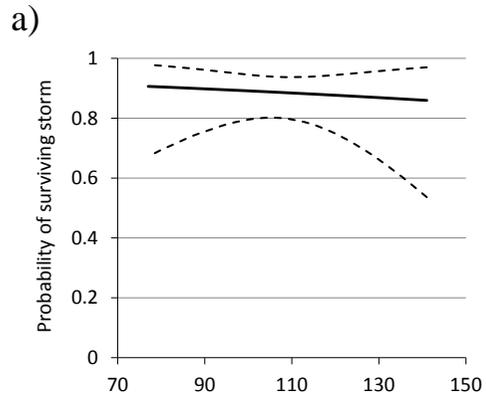
607 Fig 2



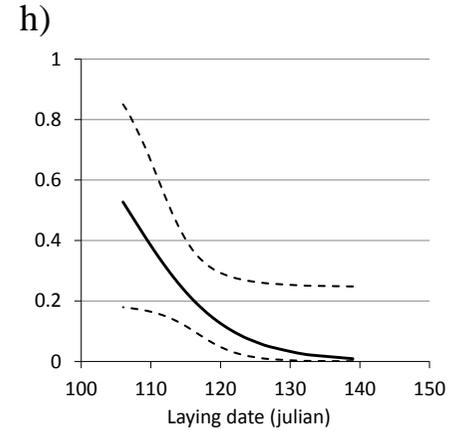
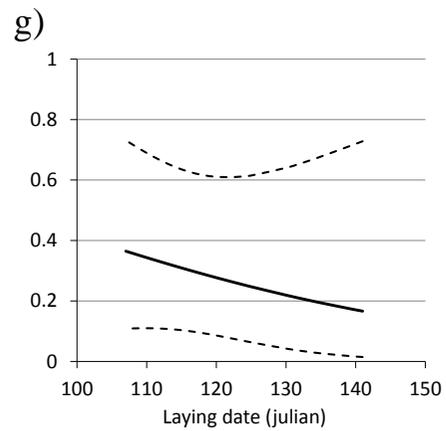
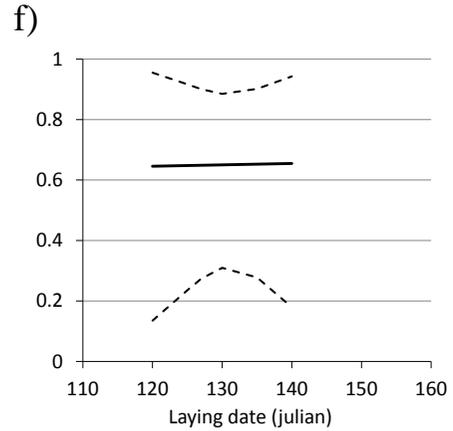
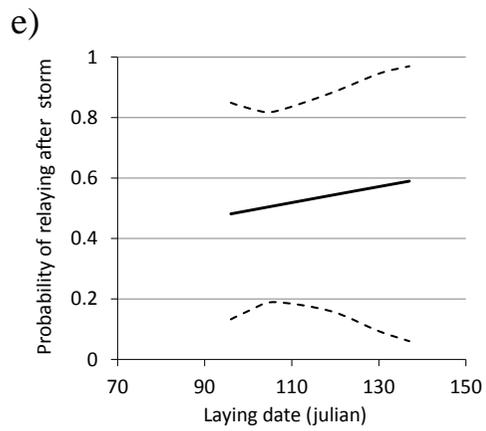
608

609 Fig 3

610



611



612

613 Fig 4

## 614 **Appendix A: Analysis of historical breeding data**

615

### 616 **Methods**

617 To examine whether the storm affected nests where failure rate is higher or lower on  
618 average, we compared the past breeding success of nest sites that survived the storm  
619 with those that did not. Data on past breeding success of study nests in monitoring  
620 plots where individual nest identity was retained across years were available from  
621 1996 for shags (n = 1,497 breeding records; breeding success was first recorded in  
622 1985 but individual nest identity was only retained across years from 1996 onwards),  
623 1981 for guillemots (n = 16,773 breeding records) and 1982 for razorbills (n = 3,800  
624 breeding records). The analysis was not possible for kittiwakes, since individual nest  
625 identity is not retained across years in this species. For shags, a linear mixed model  
626 (LMM) was fitted to historical breeding success by Restricted Maximum Likelihood  
627 estimation (REML), with number of chicks fledged per pair as the response variable  
628 (range 0-4), year, plot id and nest site (with nest site nested in plot) as random effects  
629 and storm effect (survived vs not survived) as a fixed effect. Shag breeding success  
630 could be treated as a Poisson variable, given it can only take on integer values.  
631 However, we found that model residuals were approximately normally distributed. To  
632 ensure that model outcomes did not arise from the choice of error structure, we  
633 repeated the analysis in a GLMM with Poisson errors and fixed effects inferences  
634 were very similar (following Daunt et al. (2014)). We therefore only present results  
635 based on the LMM. For guillemots and razorbills, where number of chicks fledged is  
636 binomial since they only lay one egg, we carried out a GLMM of number of chicks

637 fledged (0 or 1) with year, plot and nest site (with nest site nested in plot) as random  
638 effects and storm effect (survived vs not survived) as a fixed effect, with binomial  
639 errors and a logit link function.

640

## 641 **Results**

642 There was no difference in past breeding success of nest sites that survived the storm  
643 and those that did not for shag (LMM: storm survival:  $W = 0.44$ ,  $p = 0.51$ ) or razorbill  
644 (GLMM: storm survival:  $W = 2.52$ ,  $p = 0.12$ ). However, there was a significant  
645 relationship between past breeding success of nest sites and survival in the storm in  
646 guillemots (GLMM: storm survival:  $W = 6.95$ ,  $p < 0.01$ ), with nest sites that failed in  
647 the storm fledging 4.7% fewer chicks on average over the period 1981-2010 than  
648 those that survived the storm. Therefore, for guillemots we reduced the predicted  
649 breeding success of pairs that failed in the storm by 4.7%.