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Comont, Richard F.; Purse, Bethan V.; Phillips, William; Kunin, William E.; Hanson, Matthew; Lewis, Owen T.; Harrington, Richard; Shortall, Christopher R.; Rondoni, Gabriele; Roy, Helen E. 2014. **Escape from parasitism by the invasive alien ladybird, *Harmonia axyridis***. *Insect Conservation and Diversity*, 7 (4). 334-342. [10.1111/icad.12060](https://doi.org/10.1111/icad.12060)

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# 1 **Escape from parasitism by the invasive alien ladybird, *Harmonia*** 2 ***axyridis***

3  
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## 19 **Abstract:**

- 20  
21 1. Alien species are often reported to perform better than functionally-similar species native to  
22 the invaded range, resulting in high population densities, and a tendency to become invasive.  
23 The Enemy Release Hypothesis (ERH), explains the success of invasive alien species (IAS)  
24 as a consequence of reduced mortality from natural enemies (predators, parasites and  
25 pathogens) compared to native species. The harlequin ladybird, *Harmonia axyridis*, a species  
26 alien to Britain, provides a model system for testing the ERH.  
27  
28 2. Pupae of *H. axyridis* and the native ladybird *Coccinella septempunctata* were monitored for  
29 parasitism between 2008 and 2011, from populations across southern England in areas first  
30 invaded by *H. axyridis* between 2004 and 2009. Additionally, a semi-field experiment was  
31 established to investigate the incidence of parasitism of adult *H. axyridis* and *C.*  
32 *septempunctata* by *Dinocampus coccinellae*.  
33  
34 3. *Harmonia axyridis* pupae were parasitised at a much lower rate than conspecifics in the  
35 native range, and both pupae and adults were parasitised at a considerably lower rate than *C.*  
36 *septempunctata* populations from the same place and time (*H. axyridis*: 1.67%; *C.*  
37 *septempunctata*: 18.02%) or in previous studies on Asian *H. axyridis* (2-67%). We found no  
38 evidence that the presence of *H. axyridis* affected the parasitism rate of *C. septempunctata* by  
39 *D. coccinellae*.  
40  
41 4. Our results are consistent with the general prediction that the prevalence of natural enemies  
42 is lower for introduced species than for native species at early stages of invasion. This may  
43 partly explain why *H. axyridis* is such a successful IAS.  
44

45 **Keywords:** enemy release hypothesis, natural enemies, native species, invasive alien species,  
46 *Harmonia axyridis*, *Coccinella septempunctata*

## 47 **Introduction:**

48  
49 Biological invasions are a major threat to native ecosystems (Millennium Ecosystem Assessment,  
50 2005, UK National Ecosystem Assessment, 2011), and the rate of establishment of alien species is  
51 accelerating (Roy, Bacon, Beckmann *et al.*, 2012b, Roy, Roy & Roques, 2011). Not all the species  
52 introduced to an area establish: 10% is commonly acknowledged as the proportion of new arrivals  
53 that establish successfully, but this is based on an assessment of plant species in Britain (Williamson  
54 & Fitter, 1996) and there is evidence that this rate may not hold true for other taxa (Jeschke, Aparicio,  
55 Haider *et al.*, 2012, Jeschke, 2008).  
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58 Some introduced species have become spectacularly successful in the new regions they occupy, and  
59 are termed invasive alien species (IAS) in view of their rapid spread and high impact on native  
60 diversity. Many new arrivals do less well, failing to establish or surviving only as small, isolated  
61 populations (Lodge, 1993). One commonly-cited potential explanation for this discrepancy is the  
62 Enemy Release Hypothesis (ERH) (Elton, 1958, Torchin, Lafferty, Dobson *et al.*, 2003), also known  
63 as enemy-escape (Brown, Abrahamson, Packer *et al.*, 1995) or the escape-from-enemy hypothesis  
64 (Wolfe, 2002). This predicts that an alien species introduced to a new region will experience reduced  
65 mortality from specialised natural enemies (e.g., predators, parasites and pathogens) compared to  
66 native species. The co-evolved natural enemy species remain in the native range of the IAS, while  
67 natural enemies native to the introduced range of the IAS will not be co-evolved with it, and will often  
68 take time to adapt to a novel prey or host species. This gives the IAS a competitive advantage,  
69 assuming that natural enemies are important in regulating populations (Roy & Lawson Handley,  
70 2012c), and allows a rapid increase in the abundance and distribution of the alien species (Elton,  
71 1958, Torchin *et al.*, 2003, Colautti, Ricciardi, Grigorovich *et al.*, 2004).

72  
73 Two mechanisms have been proposed to account for the increase in population growth of the alien  
74 species in comparison to the native species: regulatory or compensatory release (Colautti *et al.*  
75 2004). For hosts that are strongly regulated by enemies in their native range a reduction in enemies in  
76 the introduced range may lead to direct changes in survivorship, fecundity, biomass or other  
77 parameters (regulatory release). Alternatively, for hosts that are well-defended and, consequently,  
78 lack natural enemies within their native range, a reduction in enemies may be of minimal  
79 consequence for hosts. Indeed for well-defended hosts, fewer enemies may lead to a reallocation of  
80 resources from defence to population growth over ecological time (Roy *et al.*, 2012c), so-called  
81 compensatory release or as the Evolution of Increased Competitive Ability (EICA). Empirical evidence  
82 for the role of the ERH or EICA in invasion success is lacking (Roy *et al.* 2011), particularly for  
83 invertebrates,

84  
85 The harlequin ladybird, *Harmonia axyridis* (Pallas) (Coleoptera: Coccinellidae), is considered to be a  
86 highly successful IAS (Roy, Brown & Majerus, 2006, Brown, Adriaens, Bathon *et al.*, 2008a, Roy &  
87 Wajnberg, 2008b). It is native to temperate Asia, but is now established across Europe, Africa, and  
88 both North and South America (Brown, Thomas, Lombaert *et al.*, 2011b). There has often been a time  
89 lag between introduction of *H. axyridis*, as a biological control agent for aphids and coccids, and  
90 establishment (Koch, 2003). This species is also known to colonise through natural spread but also  
91 anthropogenically on produce and along transport networks (Brown, Adriaens, Bathon *et al.*, 2008b).  
92 *Harmonia axyridis* established in Britain in 2004 (Majerus, Mabbott, Rowland *et al.*, 2006), and spread  
93 at more than 100 km per year (Brown, Roy, Rothery *et al.*, 2008), contributing to declines in several  
94 native species (Roy, Adriaens, Isaac *et al.*, 2012a, Ware & Majerus, 2008, Brown, Frost, Doberski *et al.*,  
95 2011a) most likely through competition and predation (Ware & Majerus, 2008).

96  
97 The invasive nature of *H. axyridis* is thought to result from several factors. The species has good  
98 dispersal capabilities, occupies a broad range of habitats, has high reproductive potential, broad  
99 climatic tolerance, and a wide dietary range, including a propensity to act as an intraguild predator  
100 within the aphidophagous guild (Majerus *et al.*, 2006, Ware, Majerus, Roy *et al.*, 2005, Roy &  
101 Wajnberg, 2008a, Berkvens, Baverstock, De Clercq *et al.*, 2008, Brown *et al.*, 2011b, Soares, Borges,  
102 Borges *et al.*, 2008). As a well-defended species, with strong chemical defences and large larval  
103 spines (Sloggett, Magro, Verheggen *et al.*, 2011, Ware *et al.*, 2008) introduced to a new continent  
104 many thousands of kilometres from its native range, there is also a strong possibility that enemy  
105 release plays a role in the success of *H. axyridis*.

106  
107 In Britain, several species of endoparasitic Hymenoptera and Diptera attack ladybirds. Some of these  
108 have a Holarctic distribution, and are known to attack *H. axyridis* in its native range: *Dinocampus*  
109 *coccinellae* (Schrank) (Hymenoptera: Braconidae) (Kuznetsov, 1997, Ware, Michie, Otani *et al.*,  
110 2010); *Homalotylus flaminus* Dalman (Hymenoptera: Encyrtidae) (Kuznetsov, 1997); *Oomyzus*  
111 *scaposus* (= *Tetrastichus coccinellae*) Thomson (Hymenoptera: Eulophidae) (Kuznetsov, 1997) and  
112 *Medina separata* (Meigen) (Diptera: Tachinidae) (Kuznetsov, 1997) (often erroneously referred to as  
113 *M. luctuosa* (Hodek, van Emden & Honek, 2012)), while others (*Phalacrotophora fasciata* (Fallén) and  
114 *Phalacrotophora berolinensis* Schmitz (Diptera: Phoridae) (Disney & Beuk, 1997, Disney *et al.*, 1994))  
115 are European in distribution but are closely related to *Phalacrotophora philaxyridis* Disney (Diptera:  
116 Phoridae) which attacks *H. axyridis* in Japan (Disney, 1997).

117  
118 There is evidence that some natural enemies of ladybirds, particularly the holarctic species known to  
119 attack *H. axyridis* in Asia, are beginning to attack it in Britain too (Ware *et al.*, 2010, Hall, Ware &  
120 Michie, 2009). It is unclear what effect this will have on native ladybird species. As parasitoids are  
121 shared across ladybird species, there is the potential for apparent competition (Holt, 1977, Bonsall &  
122 Hassell, 1997), where high abundance of *H. axyridis* elevates rates of parasitism in susceptible native  
123 species. Alternatively, *H. axyridis* may act as a parasitoid sink, whereby the parasitoid may oviposit in  
124 the IAS but the eggs do not produce an adult, potentially reducing the population density of the  
125 parasitoid and thus the parasite burden on native populations. This has been suggested for the  
126 parasitoid wasp *D. coccinellae*, which in laboratory studies oviposited approximately equally into *H.*  
127 *axyridis* and *Coccinella septempunctata* L., the primary native host in Britain, but which successfully  
128 eclosed from a significantly greater proportion of *C. septempunctata* (Koyama & Majerus, 2008).  
129 Additionally a study comparing parasitism of *Coleomegilla maculata* De Geer with that of *H. axyridis*  
130 by *D. coccinellae* also concluded that *H. axyridis* was an unsuitable host (Hoogendoorn & Heimpel,  
131 2002). Teratocyte cells, produced by *D. coccinellae*, are involved in immunosuppression of the host  
132 and nutrition of the parasitoid, interestingly follow an abnormal pattern of growth within *H. axyridis*  
133 which could explain the impeded development of *D. coccinellae* within this marginal host (Firlej,  
134 2012). Therefore, it is unlikely that *D. coccinellae* will limit the population growth of *H. axyridis* within  
135 invaded ranges (Berkvens, Moensa, Berkvens *et al.*, 2010, Hoogendoorn *et al.*, 2002). Intriguingly a  
136 recent study demonstrated that *H. axyridis* individuals contain high numbers of obligate parasitic  
137 microsporidia which while not seemingly harmful to *H. axyridis* are lethal when artificially injected into the  
138 native ladybird *C. septempunctata* (Vilcinskis, Stoecker, Schmidtberg *et al.*, 2013).  
139

140 In this paper, we examine the following hypotheses:

- 141
- 142 • Time since establishment will affect parasitism of the IAS *H. axyridis* resulting in low rates of  
143 parasitism at sites which have been colonised relatively recently by *H. axyridis*.  
144
- 145 • The parasitism rate of the functionally-similar native ladybird *C. septempunctata* will be higher  
146 than for *H. axyridis*, concomitant with the predictions of the ERH.  
147

148

149 We examined these hypotheses through a field survey of pupal parasitism (monitoring pupae of the  
150 two ladybird species for parasitism by the native parasitoids *P. fasciata*, *P. berolinensis* and *O.*  
151 *scaposus*, at a large spatial and temporal scale) and a semi-field experiment of adult parasitism  
152 (monitoring parasitism by the parasitoid wasp *D. coccinellae* in overwintering aggregations of adults of  
153 the two ladybird species within mesocosms).  
154

## 155 **Methods:**

### 156 **Field Survey - pupal parasitism:**

157 Mature pupae of *C. septempunctata* and *H. axyridis* were collected between May and September  
158 from Loughborough, Leicestershire (during 2008-2011), Oxfordshire (2010-11), London (2010-11),  
159 Cambridge (2011) and Plymouth, Devon (2011) (Table 1). Pupae were collected by visually searching  
160 vegetation, primarily sycamore (*Acer pseudoplatanus* L.), lime (*Tilia* spp.) and nettle (*Urtica dioica* L.),  
161 between ground level and 2.5 metres and removing the leaf on which the pupae were attached.  
162 Collected pupae were kept in individual containers at a constant temperature and light regime (18°C,  
163 16:8 L:D), and checked for emergence or parasitism on a daily basis. Parasitoids which emerged  
164 were allowed to reach adulthood and then identified after death. Owing to time constraints the few  
165 pupae from which nothing emerged were not dissected. Individual Phoridae from 66 *C.*  
166 *septempunctata* pupae and 32 *H. axyridis* pupae (representing half of the Phoridae broods) were  
167 dissected to determine species.  
168

### 169 **Semi-field survey - adult parasitism:**

170 Both *H. axyridis* and *C. septempunctata* were collected from field sites in Crowmarsh Gifford, South  
171 Oxfordshire, in late September 2011. Native ladybirds at this time of year in Britain are still active but  
172 begin to move to overwintering sites. *Harmonia axyridis* has been observed to be active much later  
173 than native species in Britain and larval stages have been recorded in November (and exceptionally  
174 early December). As these individuals had been exposed to the possibility of parasitism in the field,  
175

176 they were monitored for parasitoid eclosion for a week in the laboratory, within a controlled  
177 environment (18°C, 16:8 L:D), before being placed outside in overwintering conditions. During the  
178 time spent within the laboratory the ladybirds were fed artificial diet, to ensure hydration, every other  
179 day (see Roy, Brown, Comont *et al.*, 2013 for details) and supplied with pea aphids (*Acyrtosiphon*  
180 *pisum* Harris) *ad libitum*.

181  
182 Mesocosms were placed outdoors underneath Lombardy poplar trees (*Populus nigra* L. var. 'Italica')  
183 from the beginning of October 2011 (the usual time to begin overwintering) until mid-January 2012  
184 (before the end of the winter dormancy period). Each individual mesocosm consisted of a black 14-  
185 litre bucket, with three holes drilled in the base to allow rainwater to drain. Mesocosms were filled with  
186 2.5l of compost (John Innes No. 10), pressed down firmly and covered by 6.5l of fresh uncompressed  
187 leaf litter, collected from the area surrounding the mesocosms and frozen at -20°C for five days  
188 between collection and use in the mesocosms to kill any animals present. A 300x18x18mm L-shaped  
189 section of wood was placed on the leaf litter leaning against the side of the bucket. The top of the  
190 bucket was covered with dark green netting, with a mesh size (approximately 2 mm) small enough to  
191 prevent the ladybirds escaping, but large enough to allow *D. coccinellae* to enter.

192  
193 The field-collected ladybirds were placed in the mesocosms so that each contained 40 adult  
194 ladybirds, either all *H. axyridis*, all *C. septempunctata*, or 20 of each of the two species. Twenty-four  
195 mesocosms (eight of each treatment) were located in Crowmarsh Gifford, Oxfordshire (10 km square  
196 SU68 in the UK Ordnance Survey grid system where the first *H. axyridis* was recorded in 2007), and a  
197 further 15 (5 of each treatment) were located at the Leeds University Farm, Tadcaster, West  
198 Yorkshire (10 km square SE44 in the UK Ordnance Survey grid system where the first *H. axyridis* was  
199 recorded in 2009). An additional fifty individuals of each species were kept individually in the  
200 laboratory and monitored for the duration of the experiment as a control, to test for pre-trial parasitism  
201 rates.

202  
203 The mesocosms were collected in from the field in mid-January and the individual ladybirds were  
204 recovered. The number of individual ladybirds which had been parasitised by *D. coccinellae* was  
205 recorded at time of collection. All remaining ladybirds were monitored for *D. coccinellae* emergence  
206 over a period of six weeks. Owing to time constraints the individuals from which nothing emerged  
207 were not dissected.

208  
209

#### 210 **Analysis:**

211 Analysis was restricted to pupae from which either a ladybird or a parasitoid emerged. Parasitism rate  
212 was the proportion of these pupae which were parasitised.

213

214 For both datasets, the relationship between parasitism rate, host species, and years with *H. axyridis*  
215 presence was analysed using binomial General Linear Models (McCullagh & Nelder, 1989) in R  
216 version 2.15.2 (R Development Core Team, 2011), with full subsets model selection using Akaike's  
217 Information Criterion (AIC) values (Burnham & Anderson, 2002, Akaike, 1974). The model with the  
218 lowest AIC was chosen unless a model with fewer parameters was within 2 AIC units of this model, in  
219 which case the simpler model was preferred. The discriminatory ability of the models was evaluated  
220 using the Area Under the Receiver Operating Curve (AUC) (Fawcett, 2006).

221

222 For the pupal parasitism dataset, Generalized Linear Mixed Models (GLMMs) were constructed using  
223 the R package LME4 (Bates, Maechler & Bolker, 2011) and used to evaluate the relationship between  
224 parasitism rate and host species (*C. septempunctata* or *H. axyridis*), years of *H. axyridis* presence (1-  
225 6 years, mean 4.17 ± 1.19) and the interaction between host species and years of *H. axyridis*  
226 presence. The 10 km sq of collection was included as a random effect to take account of site-level  
227 effects not explained by the years of *H. axyridis* presence.

228

229 For the adult parasitism dataset, Generalized Linear Models (GLMs) were used to evaluate the per-  
230 mesocosm relationship between parasitism rate and host species (*C. septempunctata* or *H. axyridis*),  
231 treatment type (40 *C. septempunctata*, 40 *H. axyridis*, or a mixture of 20 *C. septempunctata* and 20 *H.*  
232 *axyridis*), and site (Oxfordshire or Yorkshire). Site was included as a fixed effect in this analysis, to  
233 test whether *H. axyridis* is parasitised at a lower rate in Yorkshire, where it had invaded more recently,  
234 than in Oxfordshire, having standardised the mesocosm environment as far as possible.

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## Results:

### Field survey: pupal parasitism:

Overall, *H. axyridis* was parasitised at a much lower rate than *C. septempunctata*. Of the 4,595 live pupae collected (Table 1), 219 were parasitised (67/3868 *H. axyridis* (1.73%), 152/727 *C. septempunctata* (20.91%); see Table 2). Gregarious wasps (*O. scaposus*) emerged from one *H. axyridis* pupa and 21 of the *C. septempunctata* pupae, resulting in 423 individual wasps: 18 from the *H. axyridis* pupa and a mean  $19.29 \pm 7.79$  (range 10-42) per *C. septempunctata* pupa.

Gregarious flies of the family Phoridae ("phorids") parasitised both *H. axyridis* (66 pupae parasitised, mean  $3.44 \pm 2.72$  parasitoids per brood (range 1-15)) and *C. septempunctata* (131 pupae parasitised, mean  $5.38 \pm 3.15$  parasitoids per brood (range 1-15)), producing 932 individual parasitoids. All 241 individuals (66 broods) examined from *C. septempunctata* were *P. fasciata*, as were 114 individuals (28 broods) from *H. axyridis*, while a further 10 individuals (4 broods) from *H. axyridis* were *P. berolinensis*. There was no evidence of hyperparasitism or cross-species multiparasitism in either host species.

Overall pupal parasitism rate was significantly lower in the IAS *H. axyridis* than in the native *C. septempunctata* (slope from *C. septempunctata* to *H. axyridis*  $-2.67 \pm 0.18$ ,  $z = -15.23$ ,  $p < 0.001$ , AUC = 0.86). There was no significant effect on parasitism rate of the years of *H. axyridis* presence, or of the interaction between host species and years of *H. axyridis* presence (Table 3).

### Semi-field survey: adult parasitism

Of the 1560 ladybirds originally placed in the mesocosms (780 of each species), 1475 were recovered (Table 4: 759 *C. septempunctata*, 716 *H. axyridis*) and the remaining 85 individuals were missing (presumably dead and decomposing, indeed elytra were found within the leaf litter). No successful parasitism was found in *H. axyridis*, but 43 (5.67%) *C. septempunctata* were successfully parasitised by *D. coccinellae*. There were no deaths from parasitism in the control samples, so the pre-experiment background parasitism rate was taken to be zero.

The best model to explain parasitism by *D. coccinellae* contained species and region (Table 5). Ladybirds were significantly more likely to be parasitised in Yorkshire than in Oxfordshire (Table 6:  $-2.63 \pm 0.45$ ,  $z = -5.859$ ,  $p < 0.001$ ), and the native *C. septempunctata* was more likely to be parasitised than the IAS *H. axyridis*. Although not significant in the model (Table 6), inclusion of species as an explanatory variable improved overall model performance considerably (Table 5), and the model containing only 'species' was the best of the univariate models (Table 5), and was significant under the Kruskal-Wallis test ( $X^2_1 = 17.11$ ,  $p = < 0.001$ ). Parasitism rate did not differ significantly between the single- and mixed-species treatments for either ladybird species (Table 5).

## Discussion:

*Harmonia axyridis* in its introduced range is parasitised at a considerably lower rate than either *H. axyridis* in its native range, or populations of *C. septempunctata* native to the introduced range of *H. axyridis*. Pupae of the invasive alien ladybird *H. axyridis* were parasitised at an exceptionally low level across Britain (1.73%) and adults were not found to be parasitized at all in our study. This is in contrast to the co-occurring native species *C. septempunctata*, which experienced reasonably high parasitism (20.91% pupae, 5.67% adults). This is consistent with the predictions of the ERH that enhanced performance of an IAS in the introduced range may result from a reduction or absence of natural enemies (Elton, 1958, Torchin, Lafferty & Kuris, 2001). The aposematic colouration and chemical defences of ladybirds causes them to suffer low rates of attack by non-specialist natural enemies (Roy *et al.*, 2013), and consequently this discrepancy in the rate of parasitism by specialist parasitoids is likely to represent an advantage for *H. axyridis*. However, further research is required to explore the population-level effects of this difference in parasitism rates.

Less than two percent of the *H. axyridis* pupae succumbed to parasitism by phorids in our study whereas *C. septempunctata* experienced high phorid parasitism rates (18 percent) at the same sites. It is also notable that parasitism of *H. axyridis* by phorids in Japan is much higher than in Britain; indeed nearly 15 percent of *H. axyridis* were parasitized by the native phorid *P. philaxyridis* found in Japan by Osawa (1992). However, phorid parasitism rates are known to vary considerably, for example phorid parasitism of *H. axyridis* across the native range was between 2% and 67% (Disney,

294 1997, Osawa, 1992, Park, Park, Hong *et al.*, 1996, Maeta, 1969). The absence of *H. axyridis*  
295 parasitism by *D. coccinellae* was notable in the studied overwintering aggregations particularly when  
296 considering that nearly six percent of the *C. septempunctata* were parasitized. It should be noted that  
297 the mesocosms did not represent ideal overwintering conditions particularly for *H. axyridis* which often  
298 overwinters in buildings or at elevated positions (Roy *et al.*, 2013, Nalepa, Kidd & Ahlstrom, 1996).  
299 However the high proportion of *H. axyridis* and *C. septempunctata* retrieved from the mesocosm is  
300 encouraging.

301  
302 Pupal *H. axyridis* were found to be successfully parasitised by three species of parasitoid, one more  
303 than was recovered from *C. septempunctata*. Although the additional species, *P. berolinensis*, is  
304 known to parasitise *C. septempunctata* (Disney, Majerus & Walpole, 1994, Hodek *et al.*, 2012), in the  
305 study area it was only recovered from *H. axyridis* and two conifer-specialist ladybird species,  
306 *Aphidecta oblitterata* L. and *Anatis ocellata* L. (R. Comont, unpublished data). This suggests that *H.*  
307 *axyridis* is susceptible to the suite of parasitoids in the system. The lack of evidence for increased  
308 rates of parasitism on *H. axyridis* over time from colonisation suggests that there is little or no  
309 adaptation of parasitoids to the arrival of *H. axyridis* so far. Despite the potential resource presented  
310 by the presence of high numbers of this large ladybird species, it is perhaps unsurprising that the  
311 parasitism rate is low given the short period of time since arrival of this IAS.

312  
313 This limited adaptation to the presence of an IAS is consistent with results from other range-  
314 expanding or invasive taxa (Girardoz, Kenis & Quicke, 2006, Cornell & Hawkins, 1993, Menéndez,  
315 Gonzalez-Megias, Lewis *et al.*, 2008), including parasitism and inquilinism (in which an animal  
316 characteristically lives commensally in the nest, burrow, or dwelling place of an animal of another  
317 species) in the alien gall-wasp *Andricus quercuscalicis* (Burgsdorf), which has been monitored in  
318 Britain since its arrival in the late 1950s (Schönrogge, Stone & Crawley, 1996). Very low levels of  
319 parasitism/inquilinism (one species, <0.01 inquilines per gall) were recorded up to 30 years after  
320 establishment at sites in southern England, but after an additional five years, another 12 species were  
321 recorded from the gall and the wasp larva, and the mean number of inquilines per gall had risen to  
322 0.26 (Schönrogge *et al.*, 1996). Even after around 55 years in the invaded range, however, parasitism  
323 of the wasp remained low (<10%) (Schönrogge *et al.*, 1996, Schönrogge, Stone & Crawley, 1995).

324  
325 By contrast, rapid responses of natural enemies to invaders of some taxa have been observed. For  
326 instance, parasitism of the moth *Phyllonorycter leucographella* (Zeller) reached similar levels in the  
327 introduced range (Britain) compared to the native range (Turkey) approximately 20 years after  
328 establishment (Gröbler & Lewis, 2008). The ladybird *Olla v-nigrum* (Mulsant) was assimilated even  
329 more quickly into a native ecosystem. A native of North America and Oceania (Gordon, 1985), the  
330 species was introduced to New Caledonia in early 1987, quickly becoming widespread and abundant  
331 (Chazeau, Bouyé & Bonnet de Larbogne, 1991). Parasitism by the native phorid fly *Phalacrotophora*  
332 *quadrimaculata* Schmitz was first recorded in the wild in 1988, and by March 1989, two years after  
333 introduction, parasitism rates of 79% were reported from some populations (15-79%, mean 39%)  
334 (Disney & Chazeau, 1990).

335  
336 Previous studies (Koyama *et al.*, 2008, Hoogendoorn *et al.*, 2002, Berkvens *et al.*, 2010) found that *D.*  
337 *coccinellae* showed no oviposition preferences between *H. axyridis* and *C. septempunctata* but  
338 successfully eclosed significantly less often from the IAS, suggesting the species might act as a sink  
339 for the parasitoid (Berkvens *et al.*, 2010, Hoogendoorn *et al.*, 2002). Other studies have found that *H.*  
340 *axyridis* produces chemicals that attract *D. coccinellae* (Durieux, Fischer, Brostaux *et al.*, 2012,  
341 Richerson & DeLoach, 1972, Al Abassi, Birkett, Pettersson *et al.*, 2001). We found no evidence that  
342 the presence of *H. axyridis* affected the rate of parasitism of *C. septempunctata* by *D. coccinellae*,  
343 and hence no support for a role for apparent competition. We did find a strong site effect for  
344 parasitism of adult ladybirds by *D. coccinellae* and it is possible that high rates of parasitism of native  
345 species may lead to parasitoid spillover into *H. axyridis* populations, if the resulting large numbers of  
346 parasitoids are unable to find many native hosts after depleting the populations. We only compared  
347 parasitism between *H. axyridis* and *C. septempunctata* and did not assess population-level effects on  
348 either host species. However, the low levels of parasitism found to date in *H. axyridis* populations  
349 suggests that the impacts of *H. axyridis* presence found by Roy *et al* (2012a) on common native  
350 ladybird species across Europe are likely to result entirely from direct interactions (predation and  
351 competition), rather than indirect interactions mediated by shared parasitoids.

352

353 *Harmonia axyridis* has aposematic colouration, strong morphological defences (spines) in the larval  
354 stage, and alkaloid-rich chemical defences at all life stages. Despite occasional observations of  
355 predation by birds, e.g. *Delichon urbicum* L. (R. Comont, personal observation) and mammals, e.g.  
356 *Rhinolophus ferrumequinum* (Shreber) (R. Comont, personal observation) and *Ursus arctos horribilis*  
357 Ord (Wang, Zhang & Zhang, 2007), predation is unlikely to have a regulatory effect on *H. axyridis*  
358 populations. Consequently, parasitism, which can reach 95% in populations of some species (Hodek  
359 *et al.*, 2012), is likely to be the dominant top-down factor regulating populations, and is effectively  
360 missing from *H. axyridis* populations in the introduced range at this early stage of invasion. The  
361 presence of several species of native parasitoid parasitising *H. axyridis* less than a decade after  
362 invasion does, however, suggest that the future recruitment of natural enemies to this abundant, albeit  
363 well-defended species is possible, particularly in the light of the documented time lags for recruitment  
364 of parasitoids to many invasive alien taxa (Cornell *et al.*, 1993, Schönrogge *et al.*, 1996, Schönrogge  
365 *et al.*, 1995, Girardoz *et al.*, 2006). Future research should focus on both the rate of parasitism and  
366 possible effects on the population dynamics of different species of ladybird. Indeed continued  
367 monitoring of *H. axyridis*, native species of ladybird and the parasites associated with them is  
368 essential to unravelling the web of interactions around this invasive alien species.

369

### 370 **Acknowledgements:**

371 RFC, HER, BVP are funded by the Natural Environment Research Council. HER also receives  
372 funding from the Joint Nature Conservation Committee. The Rothamsted Insect Survey is a BBSRC-  
373 supported National Capability. The authors would like to thank Emmeline Favreau, Emma Rhule,  
374 Georgie Knibbs, Carly Hoskin, and Sarah Turner for their assistance collecting pupae. We are  
375 grateful for discussions and advice on *Phalacrotophora* spp. ecology and identification from Remy  
376 Poland and Henry Disney.

377

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547 **Table 1.** Number of pupae of the native ladybird *Coccinella septempunctata* and the invasive alien  
548 species *Harmonia axyridis* collected from sites given as 10km squares (10 x 10 km grid squares of  
549 the Ordnance Survey British National Grid) across England (Cambridge, Cambridgeshire; London,  
550 Middlesex; Loughborough, Leicestershire; various locations in south Oxfordshire; and Plymouth,  
551 Devon) between 2008 and 2011. Numbers in brackets indicate living pupae which produced either an  
552 adult ladybird or a parasitoid. 0 indicates no pupae found and – indicates site not monitored.  
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Site	10km square	H. axyridis arrival	2008		2009		2010		2011	
			C. septempunctata	H. axyridis	C. septempunctata	H. axyridis	C. septempunctata	H. axyridis	C. septempunctata	H. axyridis
Cambridge, Cambridgeshire	TL36	2005	-	-	-	-	-	0	72 (72)	94 (94)
London, Middlesex	TQ27	2004	-	-	-	-	0	227 (212)	-	-
	TQ28	2005	-	-	-	-	0	342 (316)	0	311 (297)
	TQ39	2004	-	-	-	-	0	164 (141)	-	-
Loughborough, Leicestershire	SK51	2007	0	35 (35)	97 (92)	62 (58)	0	28 (28)	58 (58)	272(272)
	SK52	2007	0	63 (61)	31 (30)	191 (191)	0	230 (230)	73 (73)	581 (581)
Various locations, Oxfordshire	SP30	2007	-	-	-	-	19 (18)	21 (21)	-	-
	SP50	2006	-	-	-	-	-	-	0	9 (9)
	SU58	2009	-	-	-	-	14 (14)	279 (270)	-	-
	SU59	2007	-	-	-	-	0	165 (147)	3 (3)	331 (265)
	SU68	2007	-	-	-	-	38 (37)	197 (190)	171 (148)	446 (388)
	SU69	2008	-	-	-	-	189 (182)	3 (2)	-	-
Plymouth, Devon	SX45	2007	-	-	-	-	-	-	0	60 (60)

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558 **Table 2.** Summary of the observations on emergence of *H. axyridis* and *C. septempunctata* pupae.  
 559 Of the 4111 *H. axyridis* and 766 *C. septempunctata* pupae collected 3868 and 727 respectively  
 560 produced either an adult ladybird or parasites. The number and percentage of adult ladybirds  
 561 emerging or pupae successfully parasitized is provided. The parasites were identified as *O.scaposus*  
 562 and Phoridae. The number and percentage of the hosts from which these parasites emerged is  
 563 documented. The data is summarised across all collection sites and dates (2008-2011: Cambridge,  
 564 Cambridgeshire; London, Middlesex; Loughborough, Leicestershire; various locations in south  
 565 Oxfordshire; and Plymouth, Devon).

	Species of ladybird			
	<i>Harmonia axyridis</i>		<i>Coccinella septempunctata</i>	
	Number	%	Number	%
Adult ladybirds emerging	3801	98.27	575	79.09
Pupae successfully parasitized	67	1.73	152	20.91
Pupae parasitized by <i>Oomyzus scaposus</i>	1	0.03	21	2.89
Pupae parasitised by flies of family Phoridae	66	1.71	131	18.02

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**Table 3.** Comparison of the GLMMs for the field survey of pupal parasitism by *Phalacrotophora fasciata*, *P. berlinensis* and *Oomyzus scaposus* on the IAS *H. axyridis* and native *C. septempunctata*.  $\Delta$  AIC is calculated as the AIC of each model in turn minus that of the null model; lower AIC values indicate better-fitting models. The best model is highlighted in bold text.

Explanatory variables within the model	AUC	AIC	Deviance	Model comparisons (to best model)		
				$\Delta$ AIC	$\Delta$ Deviance	p-value
<b>Species</b>	<b>0.85</b>	<b>1319.7</b>	<b>1313.67</b>	-	-	<b>&lt;0.001</b>
Species + years <i>H. axyridis</i> present	0.86	1321.2	1313.25	1.5	-0.42	<0.001
Species + years <i>H. axyridis</i> present + species*years <i>H. axyridis</i> present	0.86	1323.2	1313.25	2.5	-0.42	<0.001
Null (intercept-only)	0.78	1577.7	1573.71	258	260.04	1
Years <i>H. axyridis</i> present	0.77	1578.0	1572.03	258.3	258.36	0.20

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577 **Table 4.** Number and percentage of *C. septempunctata* or *H. axyridis* parasitized by *D. coccinellae*  
578 from mesocosms, situated in Crowmarsh Gifford (Oxfordshire) or Leeds (Yorkshire), in which  
579 overwintering aggregations of single species (either 40 *C. septempunctata* or 40 *H. axyridis*) or both  
580 species (20 *C. septempunctata* and 20 *H. axyridis*) had been placed. n = total number of individuals  
581 recovered whereby in Oxfordshire there were a total of 320 individuals per treatment (24 mesocosms)  
582 and in Yorkshire there was a total of 200 individuals per treatment (15 mesocosms). The number of  
583 individuals recovered is lower than the total number of individuals originally placed in the mesocosms  
584 because a small proportion of individuals were not retrieved.  
585

	Single species				Both species			
	<i>C. septempunctata</i> (Oxfordshire n=317; Yorkshire n = 184)		<i>H. axyridis</i> (Oxfordshire n=310; Yorkshire n = 179)		<i>C. septempunctata</i> (Oxfordshire n=160; Yorkshire n = 88)		<i>H. axyridis</i> (Oxfordshire n=145 Yorkshire n = 82)	
	No. of <i>D. coccinellae</i>	% parasitism	No. of <i>D. coccinellae</i>	% parasitism	No. of <i>D. coccinellae</i>	% parasitism	No. of <i>D. coccinellae</i>	% parasitism
<b>Oxfordshire</b>	<b>2</b>	<b>0.63</b>	<b>0</b>	<b>0</b>	<b>4</b>	<b>2.50</b>	<b>0</b>	<b>0</b>
<b>Yorkshire</b>	<b>28</b>	<b>15.22</b>	<b>0</b>	<b>0</b>	<b>9</b>	<b>9.18</b>	<b>0</b>	<b>0</b>

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**Table 5.** Comparison table for GLMs of the semi-field survey of parasitism by *D. coccinellae* on overwintering aggregations of adult ladybirds (the native *C. septempunctata* and the IAS *H. axyridis*).  $\Delta$  AIC is calculated as the AIC of each model in turn minus that of the null model; lower AIC values indicate better-fitting models. The best model is highlighted in bold text

Explanatory variables within the model	AUC	AIC	Deviance	Model comparisons (to best model)		
				$\Delta$ AIC	$\Delta$ deviance	p-value
<b>Species, region</b>	<b>0.52</b>	<b>78.1</b>	<b>37.53</b>	-	-	<b>&lt;0.001</b>
Species, region, treatment	0.52	81.3	36.63	3.2	-0.9	<0.001
Region, treatment	0.52	94.4	51.75	16.3	14.22	<0.001
Species	0.51	128.9	90.30	50.8	52.77	<0.001
Region	0.53	130.0	91.36	51.9	53.83	<0.001
Species, treatment	0.52	132.5	89.91	54.4	52.38	<0.001
Treatment	0.52	143.6	102.99	65.5	65.46	<0.001
Null (intercept-only)	0.53	173.7	137.03	95.6	99.5	1

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595 **Table 6.** Coefficients for each explanatory variable retained within the best model (species (*C.*  
596 *sempunctata* or *H. axyridis*) and region: Leeds (Yorkshire) or Crowmarsh Gifford (Oxfordshire).  
597 Individual significance levels (z-value and p-value) are also given.  
598

<b>Explanatory variable</b>	<b>Coefficient (<math>\pm</math> SE)</b>	<b>z-value</b>	<b>p-value</b>
Intercept	-1.56 ( $\pm$ 0.18)	-8.623	<0.001
Species	-19.77 ( $\pm$ 1856.92)	-0.011	0.992
Region	-2.63 ( $\pm$ 0.45)	-5.859	<0.001

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