## Report

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Severn-Thames Transfer. A Review of Biological Data.
Volume 2 -Appendices.

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## 1

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tables of references cited in Chapter 2: Fish of the main report

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APPENDIX 2.1 The title page, executive summary and/or conclusions and key tables of references cited in Chapter 2: Fish of the main report.

RIVER THAMES ADULT FISH SURVEX<br>FINAL REPORT<br>JULY - OCTOBER 1991



The issues addressed in this report are dealt with in greater detail in later documents and are not summarised here.

UNIVERSITY OF LONDON

## ROYAL HOLLOWAY AND BEDFORD NEW COLLEGE

1992
RIVER THAMES JUVENILE FISH SURVEY Royal Holloway and Bedford New College University of London
by:

Dr.Nan Duncan
Dr.Jan Kubecka
Dr. Nuha Hanna
Steve Kett
Jackie Skeldon
Simon Hickenbotham
Katie Baker
$\&$
Brian Quilliam.
(1) This reports on a fry survey undertaken in July/August 1992 by Royal Holloway \& Bedford New College (RHBNC) in that part of the River Thames between Oxford and Days Weir where the outlet of a new reservoir proposed by Thames Water Utilities is likely to be sited. The location of the thirteen sites sampled were specified and overlapped with a first fry survey conducted in 1991 by the Environmental Advisory Unit Liverpool (EAU). It is noteworthy that 1992 and 1991 represented the fourth and third years of drought and low river flows.
(2) The RHBNC survey sampled each site with three seine hauls. The three seine hauls per site were used to study the influence of the nature of three kinds of microhabitats on the fry community: (1) shallow sites with water lilies, (2) shallow sites without aquatic plants and (3) deep sites without aquatic plants. The net used was 25 m long made of micromesh with 3 mmesh size.
(3) The baseline fry stocks of the River Thames in July/August 1992 compared with July 1991 from the EAU Report were as follows:

|  | 1992 <br> RHBNC | 1991 <br> EAU | 1991 <br> EAU |
| :--- | :---: | :---: | :---: |
| Number of sites |  |  |  |
| Number of seinings | 38 | $13 *$ | 36 |
| Fry density per min | 3.72 | 4.07 | 3.07 |
| Catch per unit effort | 706 | 607 | 490 |
| (= fish per net haul) |  |  |  |

This shows very similar fry densities and catch per unit effort in the two years.
(4) It is also important to sample different microhabitats because the densities of fry inhabiting them varies greatly:

Nature of microhabitat

> Deep Water-lily Shallow
$\begin{array}{llll}\text { Mean } 0+\text { fry density } & 0.082 & 10.07 & 5.38\end{array}$ (fish/m)

This shows that the two very common fry microhabitats in the River Thames (deep and lilies) vary in $0+f r y$ densities by two orders of magnitude and differ in species diversity and body size. There are probably about seven important microhabitat categories in the River Thames in addition to these three which need to be studied and characterised.
(5) Our recommendations for improving baseline estimates of fish fry stocks are to improve our knowledge of the fish fauna of the River Thames along the lines listed in Recommendations. We offer some comments on impacts of the proposed reservoir and on least damaging designs for reservoir operation.
4. The species diversity and composition of fry fish was different in the three microhabitat categories. There were fewer species in the 'deep' category and fry from rheophilous species of fish favoured the 'shallow' category.
5. There is some evidence that the sizes of some species of fry differ in the different microhabitats but this needs further analysis.

## 7. CONCLUSIONS

1. The baseline fry stocks of the River Thames in July/August 1992 compared with those recorded for July 1991 in the EAU Report were as follows:

|  | 1992 | 1991 | 1991 |
| :--- | ---: | :---: | :---: |
| Number of sites | 13 | $13 *$ | $36 * *$ |
| Number of seinings | 38 | 13 | 36 |
| Fry CPUE (fish per haul) | 706 | 607 | 490 |
| Fry density (fish $/ \mathrm{m}^{2}$ ) | 3.72 | 4.07 | 3.07 |

* comparable sites to the 1992 survey
** includes the intermediate sites and sites above Oxford.

This shows very similar levels of catch-per-unit-effort and mean fry densities in the two survey years. The mean netted area was larger in the 1992 survey ( $190 \mathrm{~m}^{2}$ ) compared with 149 $m^{2}$ in 1991 which affects the CPUE figures. About $50 \%$ of the 36 sampled sites in 1991 were 'shallow with no macrophytes' compared with $33 \%$ in 1992 which affects the mean densities for combined microhabitats.
2. It is important to sample different microhabitats because the densities of fry inhabiting varies greatly:

$$
\begin{array}{lccc}
\text { Nature of microhabitat } & \text { 'deep' } & \text { water-lily' 'shallow } \\
0+\text { fry density } / \mathrm{m}^{2} & 0.082 & 10.07 & 5.38
\end{array}
$$

This shows that the two commonest category of microhabitats in the River Thames ('deep' and 'water-lilies') vary in $0+f r y$ densities by two orders of magnitude. There is also a difference in species diversity, species composition and body size. There are probably about seven additional fry microhabitats in the River Thames which need a similar study.
3. Our main recommendation for improving baseline estimates of fish fry stocks are to improve our knowledge of the fish fauna of the River Thames along the lines listed in section 4.3 Recommendations. A very limited analysis of individual angler catches suggests that angling is a selective mode of sampling the species composition of the river fish fauna.
4. We offer some comments on impacts of the proposed reservoir and on leasi damaging designs for reservoir operation.

## Table 5

Summery of fish densities in different microhabitats of the Sutton Pools compared with the 1991/1992 results for the River Thames


## Table 6

Percentage frequency of fish fry species $(0+\& 1+)$ at sites sampled with a fry seine in the River Thames during the 1391 EAU and 1992 RHBNC surveys

|  | $\begin{array}{r} 1992 \\ 13 \text { sites } \end{array}$ | $\begin{array}{r} 1992 \\ 11 \text { sites } \end{array}$ | 1992 <br> 13 sites | $\begin{array}{r} 1992 \\ 13 \text { sites } \end{array}$ | $\frac{1991}{36 \text { EAU sites }}$ | $\begin{array}{r} 1991 \\ 13 \text { sites } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Microflabitats |  |  |  |  |  |  |
| species | Deep | Lily | Shallow | Combined |  |  |
| pike | 0.05 | ${ }^{3} 0.03$ | 0.04 | 0.04 | 0.03 | 0 |
| gudgeon | 3.33 | 8.03 | 14.86 | 10.82 | 22.43 | 11 |
| silver bream | 0 | 0 | 0.01 | 0.003 | 0 | - 0 |
| bream | 1.22 | 0.29 | 0.77 | 0.46 | 0.63 | 1.01 |
| bleak | 6.39 | 7.27 | 17.84 | 11.69 | 4.76 | 6.7 |
| minnow | 0 | 0.76 | 0.07 | 0.42 | 0 | - |
| roach | 80.18 | 77.2 | 53.23 | 67.13 | 44.9 | 72.4 |
| chub | 2.58 | 1.45 | 7.36 | 4.03 | 18.89 | 3.5 |
| dace | 0.33 | 0.97 | 3.26 | 1.9 | 6.03 | 3.93 |
| stoneioach | 0 | 0 | 0.01 | 0.003 | 0 | 0 |
| stickleback | 0.33 | 0.14 | 0.1 | 0.13 | 0.34 | 0.41 |
| perch | 5.35 | 2.27 | 2.37 | 2.51 | 1.9 | 1.1 |
| ruffe | 0.19 | 0.03 | 0 | 0.03 | 0.012 | 0.03 |
| bullhead | 0 | 0.01 | 0.01 | 0.01 | 0.094 | 0.1 |
| $\mathrm{Ro} / \mathrm{Br}$. Hyb | 0.05 | 1.56 | 0.07 | 0.83 | 0 | 0 |
| barbel | 0 | 0 | 0 | 0 | 0 | 0.013 |

Percentage frequency was calculated from the sum of fish of all sampled sites.

- This was calculated from onty those 1991 sites which were also sampled in 1992
'THE DENSXTIES OF FRY BELONGING TO DIFFERENT AGE GROUPS CAUGHT' BY SEINING IN DIFFERENT MICROHABITAT CATEGORIES IN SUTTON POOLS


MICROHABITAT CATEGORY
i Scirpus
II Scirpus+Typha+lilies
III Water-lilies only
IV Deep bank with Nostoc
V Sparse Sparganiua emersum
VI. Shallow; fringing border
4.70
0.60
0.30

VII Shallow with strong flow
0.00
0.00
0.17
$X$ Shallow, sandy, no plants
6.20
0.10
0.50

DENSITXES OF FRY BELONGING TO DIFPERENT AGE GROUPS CAUGHT BY ELECTROFISHING IN DIFFERENT MICROHABITAT CATEGORIES IN SUTTON POOLS

MICROHABITAT CATEGORY

MICROHABITAT ABUNDANCE*
$\%$
11.1
11.9
12.4
18.5
14.4
8.2
6.6
7.0
4.1
3.3
1.7

|  | $\%$ |
| :--- | ---: |
|  | 11.1 |
| Scirpus | 11.9 |
| Water-lilies | 12.4 |
| Lilies+S.emersum | 18.5 |
| Mixed monocots. | 14.4 |
| Overhanging trees | 8.2 |
| Stony area | 6.6 |
| Weirs | 7.0 |
| Shallow, fringing | 4.1 |
| Deep, fringing | 3.3 |
| Glyceria maxima | 1.7 |

FISH DENSITY PER METRE SQUARE BELONGING TO THE AGE GROUPS

| $0+$ | $1+$ | $>1+$ |
| ---: | :---: | :---: |
| 5.20 | 0.90 | 1.60 |
| 14.60 | 0.10 | 0.10 |
| 3.10 | 0.40 | 0.60 |
| 5.80 | 0.80 | 0.60 |
| 3.90 | 0.20 | 0.00 |
| 6.00 | 0.50 | 0.30 |
| 3.75 | 1.04 | 2.90 |
| 24.50 | 0.00 | 0.00 |
| 0.30 | 0.00 | 0.00 |
| 0.00 | 0.00 | 1.30 |
| 0.00 | 0.00 | 1.70 |

* Microhabitat abundance was calculated as the number of point samples per microhabitat category as a percentage of the total number of point samples taken. This is possible as the total point samples were spread uniformly throughout the littoral circumference of the Sutton pools at approximately 1 point per 5 m of shoreline.

1992
SUTTON POOLS FISHERY SURVEY
Royal Holloway and Bedford New College University of London
by:
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Dr.Nuha Hanna
Steve Kett
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## EXECUTIVE SUMMARY

1. Nine sites in Sutton Pools were sampled quantitatively by a $10 x 2 \mathrm{~m}$ fry seine net and seven other sites by a $25 \times 3 \mathrm{~m}$ fry seine net during July/August 1992. In addition, 243 point samples were taken by electrofishing in all four Sutton pools, using a battery-operated portable electrofisher. The total number of fry caught was 11,245 .
2. The Sutton Pools Report emphasises that it is essential to have a fry microhabitat classification for reliable estimates of the baseline of juvenile fish populations.
3. The most important nursery microhabitats were: (a) those with Scirpus beds ( $65.1 \mathrm{fish} / \mathrm{m}^{2}$ ); (b) shallow sites with fringing vegetation ( $24.5 \mathrm{fish} / \mathrm{m}^{2}$ ) and (c) mixed Monocotyledons Scirpus + Typha with some water lilies (22.6 fish/m²). The roach appears to be ubiquitous as it was always present as over $90 \%$ of the fish catch from the above habitats and between $30-70 \%$ of the other categories.
4. Other types of microhabitats in Sutton Pools, such as shallow and deep unvegetated sites and sites with waterlilies, support a similar species composition and similar or lower level of fry density to comparable microhabitats in the River Thames.
5. The special feature of Sutton Pools is the presence of microhabitats like the three weir pools. These are sites with strong water currents and a much greater relative proportion of shallow bare shores than in the River Thames. None of these microhabitats seems to provide important nursery areas.
6. Apart from the weir pools and sites with strong water currents, all other microhabitats are present in both the River Thames and in the Sutton pools. In the River Thames, the areas of each microhabitat category are far more extensive.
7. From this brief study during July/August 1992, there is no strong evidence for the Sutton Pools to be considered as an outstanding nursery area for fry living near shore. There is still the possibility that they constitute a good spawning area because of absence of disturbance by navigation which is a major influence upon the ecology of the River Thames and its fish stocks.
portable electrofisher. The total number of fry caught was 11,245.
8. Different kinds of fry microhabitats with and without aquatic vegetation were sampled by seining and electrofishing. The highest catches were found in vegetated sites with Scirpus beds ( $65.1 \mathrm{fish} / \mathrm{m}^{2}$ ) and in weed beds of mixed aquatic monotylendons with some water-lilies ( $22.6 \mathrm{fish} / \mathrm{m}^{2}$ ) but also in non-vegetated sites with a terrestrial fringing border (24.5 fish $/ \mathrm{m}^{q}$ ). The roach was ubiquitous both in the above sites (90\% of the $f i s h$ ) and in other sites (30-70\%).
9. Other microhabitats sampled were shallow and deep nonvegetated sites and sites with water-lilies which were similar to those sampled in the River Thames, both in species composition and density levels.
10. The presence of microhabitats such as the weirs and weir pools were a special feature of the Sutton Pools, with strong water currents and hard bottoms. There was also a greater proportion of the littoral with shallow bare shores than in the River Thames.

## 5. CONCLUSIONS

1. From this brief study during July/Augsut 1992, there is no strong evidence for the Sutton Pools to be considered as an outstanding nursery area for fry living near the shore.
2. The Sutton Pools study shows that (a) it is essential to develop a fry microhabitat classification for reliable estimates of the baseline stocks of juvenile fish and (b) that this categorization is applied in the River Thames itself.
3. There is still the possibility that the Sutton Pools constitute a good spawning area because of the absence of disturbance by navigation which is a major influence upon the
```
ecology of the River Thames and its fish stocks.
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ADULT FISH COMMUNITIES OF THE
RIVER THAMES
BETWEEN SANDFORD AND BENSON LOCKS 1993


National Rivers Authority Thames Region

# SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY <br> A REPORT FOR <br> NRA THAMES REGION <br> AND <br> THAMES WATER UTILITIES LTD. 

ADULT FISH COMMUNITIES
OF THE
RIVER THAMES
BETWEEN SANDFORD AND BENSON LOCKS 1993

Compiled by Simon Hughes
Hydroacoustic data analysis by J Kubecka and A Duncan, RHUL.
Fieldwork by:
A Killingbeck
A Irvine
J Perkins
S Symonds
D Willis
A Butterworth
J Sutton
L Richardson
J Ruck
H Stone

Table 1. Total Values of Density and Biomass by Reach and Phase.

|  | Density | Echo <br> Counting | $\begin{aligned} & \text { (Wish } \\ & \left.100 \mathrm{~m}^{3}\right) \end{aligned}$ | Density | Echo Integ. | (Fish $100 \mathrm{~m}^{-3}$ ) | Biomass | $\begin{aligned} & \left(\mathrm{Kg}_{\mathrm{m}}\right. \\ & \left.10 \mathrm{~m}^{\mathrm{s}}\right) \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Phase | I | II | Mean | I | II | Mean | I | U | Mean |
| Reach |  |  |  |  |  |  |  |  |  |
| 1 | 2.198 | 3.800 | 2.999 | 2.052 | 2.272 | 2.162 | 0.159 | 0.256 | 0.208 |
| 2 | 5.709 | 3.961 | 4.835 | 6.561 | 1.746 | 4.154 | 0.430 | 0.388 | 0.409 |
| 3 | 7.790 | 6.138 | 6.964 | 11.908 | 3.365 | 7.637 | 0.407 | 0.505 | 0.456 |
| 4 | 2.371 | 6.989 | 4.680 | 1.232 | 5.120 | 3.176 | 0.206 | 0.698 | 0.452 |
| 5 | 2.115 | 4.550 | 3.333 | 1.553 | 2.653 | 2.103 | 0.263 | 0.286 | 0.275 |

### 1.3 Potential Impacts.

Potential impacts from the reservoir scheme may be divided into two categories:
a) Direct impacts likely to occur in the short term and involving large changes in water or habitat quality, possibly resulting in significant and rapid changes in fish populations. This type of impact may well be mitigated by reservoir management and operating agreements.
b) Indirect impacts involving subtle changes to the ecosystem and evinced as a change in selection pressure on fish populations, possibly resulting in changes of fish population abundance and species composition. There is insufficient understanding of the critical pathways involved to model the resulting changes in fish populations.

### 1.4 Recommendations For Future Work

Continued monitoring of adult and juvenile fish populations comprising similar components as this study, for at least two further consecutive years to begin establishing levels of variation in fish population abundance, including an intensive study of one reach to measure diumal and seasonal variability in hydroacoustic results.

Identification of critical physical and chemical parameter limits together with food and habitat requirements at different life stages for key fish species identified in this survey.

Further comparisons of hydroacoustic measures of fish population abundance with results from catch depletion methods applied to the same site will add weight to the results of future surveys.

Electric fishing equipment should be modified in time for future surveys to allow effort to be quantified.

Options for gauging the extent and significance of fish migration should be explored.

### 4.0 RESULTS

4.10 Introduction.

Results from the studies described in this report are presented in this section; the data upon which they are based are appended where appropriate.

### 4.20 Electric fishing Results.

Electric fishing operations were carried out between $21: 00$ and $07: 30$ on $20,21,22$ and 23 September 1993, from downstream of Sandford Weir to upstream of Benson Lock; a total river length of over 30 km . Tweive principal coarse fish species (Table 4) and 2109 individuals were sampled in the five reaches. One roach/bream hybrid was caught. Table 5 shows the number of fish sampled by electric fishing in each reach.

Table 4. Fish Species Sampled (All Reaches Combined).

| Common Name | Generic Name |
| :--- | :--- |
| Barbel | Barbus barbus |
| Bleak | Alburnus alburnus |
| Bream (Common) | Abramis brama |
| Chub | Leuciscus cephalus |
| Dace | Leuciscus leuciscus |
| Gudgeon | Gobio gobio |
| Perch | Perca Fluvitalis |
| Pike | Esox lucius |
| Roach | Rutilus rutilus |
| Ruffe | Gymnocephalus cernua |
| Silver Bream | Blicca bjoerkna |
| Tench | Tinca tinca |

the difference in mean $\mathrm{F} L$, which is significant at $p=0.02$ (Reach 4) and $p=0.00 \mathrm{I}$ (Reach 5).
Roach length frequency for margin zones in Reach 4 (Fig 28) shows a marked bias towards smaller fish (Mean $\mathrm{FL}=98.87 \mathrm{~mm}$ ) than for the centre channel zones (Mean $\mathrm{FL}=149.51 \mathrm{~mm}$ ). The difference between these means is significant at $p=0.001$. This is reflected, although to a lesser extent, by the length frequency for margin zones in Reach 5 (Fig 29), in which mean lengths $($ centre $=130.13 \mathrm{~mm}$, margin $=110.58 \mathrm{~mm})$ are significantly different at $\mathrm{p}=0.002$.

Length frequencies for perch between Reaches 4 \& 5 (Figs 30 and 31) appear to be almost identical, with marked modes of 120 mm and very similar mean lengths.

Population length frequency also provides a tool for comparison between hydroacoustic and electrofished samples, which is presented elsewhere in this report.

### 4.22 Species Percentage Abundance in Samples.

Species percentage abundance for all centre channel zone samples in each reach are presented in Figures 32 to 36 and for all margin zone samples in Figures 37 and 38. Table 6 below provides a summary of this information.

Table 6. Species Percent Abundance by Reach and Sample Zone.

| Species | Reach 1 | Reach 2 | Reach 3 | Reach | 4 | Reach | 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zone | $\begin{gathered} \text { Centre } \\ \Phi \\ \hline \end{gathered}$ | $\begin{gathered} \text { Centre } \\ \text { a } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Centre } \\ \$ \end{gathered}$ | Centre $\%$ | $\begin{gathered} \text { Margin } \\ \% \end{gathered}$ | $\begin{gathered} \text { Centre } \\ \% \end{gathered}$ | $\begin{aligned} & \text { Margin } \\ & \mathbf{x} \end{aligned}$ |
| Barbel | 0 | 0.66 | 0 | 0 | 0 | 0 | 0 |
| Bleak | 32.76 | 51.50 | 59.58 | 26.82 | 13.26 | 67.97 | 18.60 |
| Bream (Common) | 2.69 | 3.99 | 3.75 | 2.35 | 0.55 | 0.78 | 0.41 |
| Chub | 1.96 | 1.66 | 10.83 | 4.47 | 0 | 4.30 | 1.24 |
| Dace | 1.22 | 0.66 | 1.25 | 4.71 | 0.55 | 0 | 1.24 |
| Gudgeon | 0.73 | 0.33 | 0 | 3.76 | 2.21 | 1.17 | 0.83 |
| Perch | 4.89 | 0.66 | 1.67 | 4.00 | 23.20 | 3.91 | 19.42 |
| Pike | 3.67 | 4.65 | 3.33 | 1.18 | 2.21 | 0.78 | 1.65 |
| Roach | 50.86 | 35.55 | 19.58 | 52.71 | 56.91 | 21.09 | 54.55 |
| Ruffe | 0 | 0 | 0 | 0 | 0.55 | 0 | 1.24 |
| Silver Bream | 0.98 | 0.33 | 0 | 0 | 0.55 | 0 | 0.41 |
| Tench | 0.24 | 0 | 0 | 0 | 0 | 0 | 0.41 |

Table 35. Total Values of Density and Biomass by Reach and Phase.

|  | Density | Echo Couating | (Fish $100 \mathrm{ma}^{-3}$ ) | Density | Echo Integ. | $\begin{aligned} & \text { (Fish } \\ & \left.100 m^{3}\right) \end{aligned}$ | Biomas | $\begin{aligned} & \left(\mathrm{K}_{\mathrm{g}}\right. \\ & \left.100 \mathrm{~m}^{3}\right) \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pbase | I | II | Mean | I | It | Mean | I | II | Mean |
| Reach |  |  |  |  |  |  |  |  |  |
| 1 | 2.198 | 3.800 | 2.999 | 2.052 | 2.272 | 2.162 | 0.159 | 0.256 | 0.208 |
| 2 | 5.709 | 3.961 | 4.835 | 6.561 | 1.746 | 4.154 | 0.430 | 0.388 | 0.409 |
| 3 | 7.790 | 6.138 | 6.964 | 11.908 | 3.365 | 7.637 | 0.407 | 0.505 | 0.456 |
| 4 | 2.371 | 6.989 | 4.680 | 1.232 | 5.120 | 3.176 | 0.206 | 0.698 | 0.452 |
| 5 | 2.115 | 4.550 | 3.333 | 1.553 | 2.653 | 2.103 | 0.263 | 0.286 | 0.275 |

These three figures give a good summary of quantitative results collected, and they all show a generally increasing trend in biomass and density towards Reach 3 in phase I and Reach 4 in Phase II, with a decline in values in Reach 5 evident from the results for both phases.

These densities are similar to results for hydroacoustic surveys carried out on the River Thames at Cherstey ( 3.9 fish $100 \mathrm{~m}^{-3}$ ), The River Wey ( 6.5 fish $100 \mathrm{~m}^{-3}$ ), and the River Vltava at Prague ( 6.6 fish $100 \mathrm{~m}^{-3}$ ) (Duncan and Kubecka, 1993).

### 6.0 CONCLUSIONS

- Populations of similar length distribution were sampled in all five reaches.
- Bleak and roach populations sampled in margin zones from Reaches 4 \& 5 have a significantly smaller mean size than populations sampled in centre channel zones for the same reach.
- Reaches 1 \& 2 samples contained a greater number of species than Reaches 3 to 5 . Margin zone samples contained more species than centre channel zones.
- Populations in Reaches 1, 2 and 4 show a similar relationship between length and weight.
- Pooled results for roach from all reaches shows an age length relationship similar to a nationally derived standard.
- Fish densities show scattered zones of high fish density in all reaches.
- Mean values of fish density and biomass are similarly low in Reaches 1 and 5 and similarly high in reaches 2,3 and 4 , although variation about the mean of results from both phases is relatively high.
- Populations sampled by both methods have similar length frequency distribution.
- Abingdon Marina provides an important habitat for smaller fish.

Figure 2. The Five Study Reaches.


Figure 64. Total Fish Density (Echo Counting) by Reach
Phase I, Phase II and Mean of Both.


Figure 65. Total Fish Biomass by Reach
Phase I, Phase II and Mean of Both.

-- Phase I
*. Phase II Mean

Figure 66. Total EI Fish Density by Reach
Phases I, Phase II and Mean of Both.


# ABINGDON RESERVOIR: FISH STOCK ASSESSMENT 

FRY SURVEY
FINAL REPORT
JUNE - SEPNEMBER 1991


## ENVIRONMENTAL ADVISORY UNIT LTD.

It therefore appears that there may be important differences in the distribution of species which are apparent when $0+$ and $1+$ age groups of fish are considered.

It was noted during sampling that more fry and older groups seemed to be captured on sunny days and were not present on the shallows in the evening. This possibly reflects a movement on to shallow water habitats to take advantage of the different spatial thermal regime available (O'Hara et al. in prep). Abundance data and field records were of insufficient rigour to test this hypothesis.

## 5. AGE AND GROWTH

Mean lengths of fish at age 1 were calculated from the June sample on the assumption that their birthdate was ist June. The analysis of growth in the subsequent samples from July and September concentrated on the size of the $0+f i s h e s$ since the objective of the study was to investigate the characteristics and vulnerability to capture of fry. Older fishes were generally aged as $1+$ and older because the length distributions overlapped. The results are presented in tables 6, 7 and 8. There is perhaps some evidence of variation in the mean length of fish between sites but without detailed statistical analysis and confirmation of ages by scale reading particularly for $1+$ and older fish this cannot be confirmed as significant. It should also be noted that variation in measurement by different operatives has not been accounted for, to overcome this potential problem the measuring team should be kept consistent and should be experienced in the identification of juvenile stages.

The mean length of fish achieved at the end of their first year of life was toward the mid-range of results from other rivers in Britain (see Williams 1967, Mann 1976 \& 1982, Weatherley 1986). However it should be noted that there is some discussion over the accuracy of some studies because small fishes may not have been representatively sampled or aged accurately. In contrast to other methods of capture, we consider that micromesh seine netting does provide an accurate estimate of fish growth rates for young of the year.

## 6. CONCLUSIONS AND RECOMMENDATIONS

1. On the basis of this year's work we consider it is possible to capture most of the recreationally important species in their year of hatching by late July. The exception to this appears to be bleak and this may reflect a naturally late spawning time for this species.
2. Fish hatched in the previous year remained vulnerable to capture by micromesh seine netting until late July.
3. For most species of fish, and assuming normal climatic conditions, it is possible to obtain relative abundance data and thereby information on year class strengths for fish from 2 years by sampling in late July. This would enable not only an estimation of spawning success but also overwinter survival and mortality.
4. Ideally a thorough appraisal of the causes of variability in catches between months, between sites and within and between days should be undertaken. However, in the absence of such a rigorous strategy, the use of more samples taken over a restricted single time period, possibly at a limited number of sites, may provide a better database of relative abundance than sampling over three time periods. This would give a more accurate estimate of the relative abundance but would not give such good information on distributions along the length of the river. This may be important, particularly if the observed differing distributions of age classes is a real phenomenon that is repeated in future years.

It was suggested that there is an indication of reduced abundances in the more "channelised" areas of river but this may reflect difficulties in sampling rather than species abundance. Further work would be needed to validate this observation but it may be advantageous to establish a sampling strategy to include areas along the river where the banks shelve gently. It would prove possible with this sampling regime to retain sites above and below the projected location of the reservoir.
5. Any sampling programme should be conducted in the most appropriate month to sample and from the present study this appears to be late July possibly extending into August. This period would combine the likely time of warm weather with that when most species are of a sufficient length in their first year to be vulnerable to the capture method.
6. Ageing of $1+$ fish from scales should be undertaken to validate the age structure of the population. For younger fish, length frequency analysis is a satisfactory method of ageing.
7. It is important to maintain a strict quality assurance when measuring and identifying small fish to prevent errors developing. Personnel experienced in the identification of juvenile stages should undertake sample processing.







Numbers of fish (<10cms) caught by site and month

# ANGLER CATCHES IN THE RIVER THAMES BETWEEN SANDFORD LOCK AND DAY'S LOCK <br> 1993-1994 <br> S N Hughes. 



## National Rivers Authority Thames Region

# SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY A REPORT FOR <br> NRA THAMES REGION <br> AND <br> THAMES WATER UTILITIES LTD 

ANGLER CATCHES IN THE RIVER THAMES
BETWEEN SANDFORD LOCK AND DAY'S LOCK
1993-1994

S N Hughes.

### 1.0 EXECUTIVE SUMMARY

The aims of this study were to identify and measure catches of match and non-match anglers during the coarse fishing season within the SWORP study area and to quantify use of the fishery by anglers and thereby gauge the importance of the study reach of the River Thames as an angling resource.

### 1.1 Methods

The study reach of the River Thames was divided into three subsections, each forming a standard walk of between 3 and 4 km and assigned a colour code:

Blue - Sandford Mill to Radley, Right bank<br>Red - Abingdon Bridge to Culham, Left bank<br>White - Clifton Hampden Bridge to Day's Lock, Right bank

Each reach was walked by a member of NRA Fisheries staff at monthly intervals on Sundays between 20/6/93 and 13/3/94 inclusive. Anglers encountered on the walk were divided into two categories:

Match anglers - Involved in an organised angling match.
Non-match anglers - not involved in an angling match.
Non-match anglers were interviewed and the following information recorded:
a) Distance (m) to the nearest access point.
b) Distance \& time travelled to site.
c) Length of time spent fishing prior to interview.
d) Number of fishing rods in use.
e) Method used (ie float, leger, lure etc).
f) Bait used.
g) Species targeted.
if) Number of days in the week and which days normally fisted.
i) Number of times fished the reach during the current season.
j) Number of years fishing experience.
k) Rating of own experience (Novice/Intermediate/Very experienced).

Total weight of any catch was estimated and species composition noted by NRA staff on examination of catches retained in keepnets. A tally of numbers of principal species in 50 mm fork length (FL) size classes were recorded. If an organised match was encountered, anglers were not interviewed but match results were obtained at a later date. Total catch (kg), effort (rod hours) and catch per unit effort (CPUE) in grammes per rod hour were calculated.

Attempts to collect match angler catch information from large matches were unsuccessful due to poor weather and river conditions.

Match CPUE results were classified using the NRA National Fisheries Classification Scheme.

### 1.2 Results.

Species occurrence in non-match angler catches differ between reaches, bleak are dominant in the Blue reach catch, roach are dominant in catches from the other two reaches.

Maggots are the most common bait, the most common fisting method is the swimfeeder, followed by
float and leger in most reaches.
The majority of non-match anglers do not target specific fish species.
Anglers in the Red and White reaches tend to be within 250 m of the nearest access point.
The majority of anglers interviewed had between 15 and 20 years angling experience.
The Red reach is principally used by local anglers, the Blue reach is equally used by local and non local anglers and the White reach is principally used by non local anglers.

Each reach achieves Class A rating (in the upper quartile of national results) for the match fishery.
The greatest total angling effort (Match and Non-match anglers combined) is concentrated on the White reach, followed by the Blue and Red reaches.

### 1.3 Recommendations for Future Work.

Further studies of this kind are essential to fully describe this important and valuable fishery resource. Future studies should make collection of match angler catch information a priority, and should investigate alternative methods of data collection if necessary.
5.10 Methods.

### 5.11 Angler Categories.

The main reasons for the division of anglers into two groups was principally undertaken to reflect the different needs for data collection in the field. There is also some evidence to suggest that match anglers actively target particular species (Steele, pers.comm.) which may influence some of the results.

### 5.12 Non-match Angler Sampling.

The standard walk creel census is widely used in the USA and is a very successful means of collecting information on both the biological and recreational aspects of the fishery (Bayley, pers.comm.) Building upon the experience of North American exercises, and drawing from work being carried out in the UK for the NRA as part of the research and development programme, a questionnaire was designed to fit the requirements of this investigation. Sunday was selected as the most suitable day on which the maximum number of anglers would be fishing, based upon the experience of the NRA TR Water Bailiffs. As such, all the results of this study relate to data collected on Sundays, and may not accurately reflect use of the resource at other times.

The majority of anglers interviewed were very cooperative and weicomed the opportunity to actively contribute to the study.

Quality assurance checks on fieldwork were carried out by Fisheries Officers, and no problems were identified. A planned exercise to compare estimated to actual catch weights was inadvertently not completed, however this is planned to be carried out during the 1994/5 survey, and the findings retrospectively applied to this study.

### 5.13 Match Angler Catch Sampling.

A method was developed to collect this information with as little disturbance to anglers as possible. In order for the method to succeed, a large number of Fisheries personnel were required to assemble at a weekend to collect catch data. This inevitably required a long time to organise and resulted in a fairly inflexible sampling strategy. In order to justify this mobilisation of resources, it was planned that only large matches would be targeted in order to collect as large a sample as possible. On each occasion that the numbers of staff were available and sufficientiy large matches were planned, the weather conditions deteriorated and very high river flows resulted in extremely low catches.
5.20 Non-match Angler Results.

### 5.21 Species Composition.

The results presented for the Blue reach are based upon catches from only 10 angler's catches, compared with around 50 in both the other reaches. There is therefore little value in comparing these species occurrence results with those from the other two reaches.

Species composition of catches in any fishery is a function of many different factors, including the method and bait used, species targeted and species composition of the fish population. It is possible to account for some of these factors in this survey.

Results from previous electric fishing surveys (Hughes, 1994) indicate that the species composition of
the fish populations in each reach are broadly similar. Anglers were using similar methods and baits in both reaches, and the majority of anglers were not targeting any particular species.

In general terms, the species occurrence in catches from the Red and White reaches show a similar range of species, with roach as the most abundant and perch and bream contributing a significant proportion. The difference in contribution of other species to the catch make up is difficult to explain and is probably due to a combination of many subtle selection parameters, which may include gear and bait selectivity, time of day fished and river flow, temperature and turbidity conditions.

### 5.22 Length Distribution.

Length distribution data were collected to provide a guide to the length composition of catches, rather than identify individual year classes, which are likely to be masked by the relatively large class width employed.

The results show fish over a wide range of lengths are caught in both the Red and White reaches. (results for the Blue reach are not presented for the reasons outlined above). The length ranges and clear modes for roach and perch appear to be similar to those described for fish sampled by electric fishing in the same reach (Hughes 1994). This evidence suggests that fish from across the whole population are being caught by anglers.

This is further confirmed by the length selection ogives which demonstrate the similar length (and age) structure of roach perch and bream captures in both reaches. It is apparent that roach and bream do not become vulnerable to angling before reaching an approximate age of $1+$.

The apparent difference in length structure of chub catches between the two reaches may well be anomalous due to the small proportion of chub in sampled in the Red reach. If not, it indicates a greater selection for smaller fish in this reach, the reasons for which are unclear.

### 5.23 Fishing Methods.

Fishing methods were divided into a number of categories to reflect the range of methods used by anglers. Conventional methods suited to the reiatively deep, slow flowing River Thames were by far the most popular within each reach, with legers, swimfeeders and float being the three most common. These methods are selective for certain species over others, and are suitable for the most common small fish species found in the river. A similar range and number of each method in each reach reflects the similar nature of the river and fish populations within it, and must be a large factor in determining the catch species content.

Maggots formed the principal bait item, worms and caster were secondary, reflecting the wide availability of these baits and the general acceptance by anglers of their efficacy.

### 5.24 Reach Use and Angler Experience.

In the Red and White reaches, the number of days fished by the large majority of anglers was two or less and this is probably a reflection of a need to restrict fishing to weekends. The relatively low sample size for the Blue reach may have biassed the results. Anglers who fished most often ( 5 days a week) were found in this reach.

The distance of anglers from the nearest access point is an indicator of thow evenly spread anglers are along a reach, and highlight areas where effort is concentrated. The equipment-conscious modern angler may be less willing to transport a large amount of fishing gear any significant distance along
a river bank, especially where access paths are not good. The red reach results show the greatest amount of clumping around access points, and this may be a reflection of the reach passing through the town of Abingdon and the presence of Abingdon Town Council fishery - a reach of the Thames that is covered by a general licence and free to residents of Abingdon. The White reach shows a similar but less marked concentration of anglers, despite having a excellent bankside path along the whole reach.

The results show that anglers in these reaches tend to remain relatively close to the access points, and similar results have been demonstrated for rivers in the NRA North West Region (Steele, pers. comm.).

In the Blue reach, anglers are much more widely distributed along the whole length. Again this reach has an excellent footpath and access to a large car park at the upstream extent of the reach. There is, however, no evidence to suggest why the pattern of the other two reaches should not be repeated. It is possible that with much more frequent matches on this reach, non-match anglers were forced to travel further before finding a suitable spot.

Information on angler experience is important as it provides an insight into the commitment and ability of anglers utilising a reach. Both of these factors may have some influence on catching ability and therefore on CPUE values. The results show that there is not a great difference in the experience profile (in terms of years involved in the sport) of anglers between reaches, and reflect the vast resource of experience that exists on the river bank. It is interesting to compare this information with a self assessment of experience, which showed a great deal of variation between reaches, and is probably not an objective measure. One angler who had been fishing for over 50 years rated his own experience as "Novice". For the purpose of this type of study, the number of years an angler has been involved in the sport is sufficient, and a self assessment of experience should not be collected in future surveys.

The distance travelled by anglers to a fishery is one means of assessing its financial worth, assuming that there is a relationship between distance travelled and cost to the angler. In other words, a greater distance travelled to the fishery will, in general, reflect a greater cost incurred by the angler to reach that fishery. There are a number of reasons why anglers may wish to travel a greater distance; the lack of suitable venues near to home for example. These results highlight the relative importance of a fishery to local and outside anglers. Information on time spent travelling provides a useful check against these results; the distribution of each should be similar.

Results for the Red reach demonstrate that the fishery is principaliy used by local anglers, with half of the anglers travelling less than 5 km and for less than 10 minutes to the fishery. This is likely in part to be due to the Abingdon Town fishery described above.

The Blue reach appears to be used principally by anglers living within 15 km of the reach and traveling for less than 30 minutes, and these anglers probably come from Oxford and the surrounding towns. There was, however a fairly large proportion of anglers (about $30 \%$ ) who had travelled for 80 km or more.

The white reach is a further distance from towns with Oxford, Abingdon and Didcot between 1-15 km distant. This is partly reflected in the distance travelled by anglers to the reach, with a greater proportion travelling further than to the other reaches.

### 5.25 Catch, Effort and CPUE.

On their own, catch and effort results provide an insight into the frequency and scale of angling in a
reach, and therefore a means to gauge its importance as a recreational resource. CPUE gives an indication of anglers' rate of success in a reach, and therefore a means to gauge the quality and value of the biological resource.

Total catch and effort values for each reach place the Red reach highest on terms of effort, meaning that this reach is the most frequently used by non match anglers during the sampling period. The White reach was the second and the Blue reach the least frequently used by this group of anglers.

Peaks in effort were seen in the summer months, between June and September and are probably the result of the generally more comfortable weather conditions at this time of year.

Catch per unit effort results rank the Red reach highest, followed by the Blue reach and the White reach. This does not reflect the different population abundance figures measured in each reach (Hughes 1994), and there is still some question over the validity of angler catches as indicators of fish population abundance in UK rivers (Steele, O'Hara and Aprahamian, 1994).

If these CPUE values were for match anglers, they would place the reaches in class A, B and C of the National Fisheries Classification for match fisheries. This scheme was not developed for non-match angler catches, and there is some evidence to suggest in this study that they are not comparable. In this case, these classifications should therefore not be treated as definitive.

### 5.30 Match Angler Results.

Match angling represents an important component of the total fishing effort applied and is greater than non-match angling effort in each reach. The distribution of match effort throughout the season varies between reaches, the Blue reach experienced fairly constant and relatively low levels of effort to the White reach which had a few very high levels of effort in the season. Results for the total effort (by both groups of angler combined) applied to each reach place the White reach as the most significant, followed by the Blue and Red reaches.

CPUE results for match anglers consistently place all reaches in the bighest category for match fisheries nationally, confirming their excellent quality. CPUE values follow the ranking for total effort above; Red reach with the highest value, followed by Blue and White reaches.

### 5.31 Historical Match Data.

The CPUE data again confirm the consistently excellent quality of match fisheries in the White and Blue reaches of the Thames since 1989. The lower values noted in 1993 reflect feedback from angling clubs whose members have noted a reduction in match catches over the $1993 / 4$ season, which they largely attributed to the perceived higher flows experienced this year making match angling more challenging.

### 6.0 CONCLUSIONS

- Species occurrence in non-match angler catches differ between reaches.
- Bleak are dominant in the Blue reach catch, roach are dominant in the other two.
- More young chub are caught in the Red reach than the White reach.
- $50 \%$ of roach caught in each reach were aged $3+$ or less.
- Maggots are the most common bait.
- The most common fishing method is the swimfeeder, followed by float and leger in most reaches.
-The majority of non-match anglers do not target specific fish species.
- Anglers in the Red and White reaches tend to be within 250 m of the nearest access point.
- The majority of anglers in each reach have between 15 and 20 years experience.
- The Red reach is principaliy used by local anglers.
- The Blue reach is equally used by local and non local anglers.
- The White reach is principally used by non local anglers.
- The order of recreational value for the three reaches to non match anglers is (best first) Red, White, Blue.
- The order of fishery quality for non-match anglers is Red, Blue, White.
-The order of recreational value to match anglers and both groups combined is White, Blue, Red.
- The order of match fishery quality is Red, Blue, White.
- Each reach achieves Class A rating for the match fishery.



# ADULT FISH COMMUNITIES <br> OF THE RIVER THAMES <br> BETWEEN SANDFORD AND BENSON LOCKS <br> 1994 <br> VOLUME 1 - EXECUTIVE SUMMARY 

Simon Hughes
Fisheries Officer


National Rivers Authority
Thames Region

A REPORT FOR

# ADULT FISH COMMUNITIES <br> OF THE RIVER THAMES <br> BETWEEN SANDFORD AND BENSON LOCKS 1994 <br> VOLUME 1 - EXECUTIVE SUMMARY 

Simon Hughes<br>Fisheries Officer

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Hydroacoustic data analysis by Simon Hughes.
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A Killingbeck
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### 1.0 EXECUTIVE SUMMARY

### 1.1 Study Area and Methods.

Adult fish populations in five contiguous reaches of the River Thames separated by locks between Sandford and Benson Locks were surveyed as a repeat of the SWORP 1993 Adult Fish Communities investigation. Hydroacoustic methods were used to provide a quantitative assessment of fish populations in two phases of field work during mid July and late November, electric fishing (July only) provided a sample of 3943 fish from all five reaches which provided comparative species specific information. Electric fishing samples were taken separately from the centre channel zone and the margin zone; both electric fishing and hydroacoustic work was carried out at night.

The methods and equipment used were the same as employed in the 1993 study, and catches by each method were validated by comparison of length frequency results; they seem to be sampling statistically similar populations.

Results from Phase II hydroacoustic work were hampered by gear failure and a reduction in the vulnerability of fish to sampling by the equipment through a seasonal reduction in activity observed by other workers in mid autumn.

## 1. 2 Results and Conclusions.

A total of fourteen coarse fish species and roach/bream hybrids were sampled - two more species (carp and eel) were found this year. Bleak and roach were the most abundant species in both centre and margin zone samples. Species composition was similar to samples taken in 1993 with small changes in some reaches, and shows parity with a sample of fish taken from the River Thames at Reading in 1958/59.

Age frequency results provide a comparative index of year class abundance, although for bleak and bream there is evidence to suggest incorrect ageing of some length classes. Results for other species appear to be of good quality and provide key information on their population dynamics.

Recruitment to the adult bleak population appears to be excellent in the past few years, but may be independent of $0+$ densities of a given cohort, although possible errors in assigning age to length classes may have masked a relationship. The lack of older year classes may be due to incorrect ageing, but could also be a demonstration of the sensitivity of this population to environmental change, although growth is normal.

The bream population has missing cohorts, which may be a sampling artefact, a problem with the ageing process, or a reflection of recruitment success. Extremely successful cohorts illustrate the potential for change in this population. Results indicate that factors affecting growth do not limit this population's balanced development.

Chub populations are generally stable but appear to be able to exploit favourable circumstances, resulting in some successful cohorts. Growth does not appear to be a limiting factor.

Dace populations have reduced since 1993, but are not widely distributed in the study area in either year.

Pike populations show a sensitivity to change that may be related to a reduced rate of growth,
although food items (ie roach) are plentiful. There is supporting evidence that pike are poorly represented within the study area, which may be a function of the influence of bioaccumulant pollutants upon the balanced development of this population.

The roach population is susceptible to enormous changes in cohort abundance, although in the last four years, recruitment to the adult population has been very good. A depressed growth rate is evident possibly due to intraspecific competition for food, and may be a factor in population change. The variability of growth with age may be a result of changes in preference or availability of diet items. The limited evidence available does not suggest a relationship between $0+$ abundance and cohort success.

All species have acceptable levels of parasite infestation that would be expected from a river of this type.

There is evidence to suggest that electric fishing catch per unit effort (CPUE) is closely related to acoustic density for each reach. Even if electric fished samples are not fully quantitative, they do provide an index of relative change between years.

A summary of hydroacoustic density by reach is shown in Table I.
Table I. Fish Densities and CPUE by Reach and Year.

| Reach | Density (n <br> $\left.100 \mathrm{~m}^{-3}\right)$ |  |  | CPUE <br> (n min |
| :---: | :---: | :---: | :---: | :---: |
|  | 1994 <br> (July) | 1993 PI <br> (July) | 1993 PII <br> (September) | 1994 |
| 1 | 3.87 | 2.20 | 3.80 | 2.77 |
| 2 | 3.38 | 5.71 | 3.96 | 3.30 |
| 3 | 10.36 | 7.79 | 6.14 | 6.78 |
| 4 | 8.27 | 2.37 | 6.99 | 3.80 |
| 5 | 4.83 | 2.12 | 4.55 | 4.81 |

The results show a degree of consistency between years with a close relationship to electric fishing CPUE. Increases in density over time appear to be related to a reduction in mean FL, suggesting a recruitment of smaller fish to the population. It is difficult to suggest reasons for the difference in fish density between reaches, although it is likely to be a function of available habitat, water and habitat quality.

There is considerable spatial variability in fish density, but little evidence for the key factors influencing fish aggregation about a given point.

The fisheries status of the study area appears to be very good in some reaches, but only moderate to good in others. It is comparable to other reaches of the River Thames, and slightly better than parts of the River Ouse.

Most of the conclusions from this investigation relate to aspects of fish population dynamics and development that are likely to be strongly influenced by interactions within and changes to the trophic links of the river ecosystem, possibly initiated by environmental change. Some parts of the population have been shown to be extremely sensitive and capable of conspicuous change, which could be related to change in growth characteristics. Given the potential for the proposal to impinge upon the ecosystem of the river, it is essential that we understand as much as possible, the mechanisms and links within this system.

### 1.3 Recommendations.

Further studies of this type are important in allowing us to compare spawning success and subsequent recruitment to adult populations in the study reach; cohorts that were assessed in 1992/3 are only now being fully sampled by the equipment we have available. The comparison of juvenile abundance to cohort success is important in identifying a number of factors, principally the influence of spawning success, environmental factors, juvenile fish growth and survival on cohort success. A description of these relationships may allow prediction of the impact of the proposal on fish populations. The best sampling methods available can only sample a cohort fully about two years after hatching, so this type of survey must continue for a minimum of two years after the final juvenile fish survey.

The results of future studies will allow us to identify any trends in the changes fish populations are undergoing, and will begin to identify the background or natural variability in these populations. A robust statistical investigation of the principal factors affecting observed variance in results will describe the degree of change to the population that we will be able to detect with the methods available.

A study of the status of bioaccumulant pollutants in pike (as a top predator) in the study reach should be carried out to assess whether this is a factor affecting the balanced development of this population.

An assessment of the preferred food items of $0+$ and $1+$ dace (and potential competitor species) should be made to identify whether food availability is a limiting factor to the success of this species in the study reach.

All field work in studies to be compared with this one should be carried out in July to allow more robust comparisons of length and growth data. (A study of the influence of sampling time season on hydroacoustic measures of fish abundance should be made to determine whether July is the optimum sampling period.)

A quality audit of scale age data provided by NRA Anglian Region should be carried out.

Figure II. Species Percent Occurrence.
Reach 1


$\xrightarrow[\text { Reach Total }]{\text { Combined }}$


# ADULT FISH COMMUNITIES <br> OF THE RIVER THAMES <br> BETWEEN SANDFORD AND BENSON LOCKS 1994 

VOLUME 2 - MAIN REPORT

Simon Hughes<br>Fisheries Officer



National Rivers Authority
Thames Region

# SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY <br> A REPORT FOR <br> NRA THAMES REGION 

## ADULT FISH COMMUNITIES <br> OF THE RIVER THAMES <br> BETWEEN SANDFORD AND BENSON LOCKS 1994 <br> VOLUME 2 - MAIN REPORT

Simon Hughes<br>Fisheries Officer

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A Butterworth
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R Preston
$S$ Hughes

The results presented and discussed in this report lead us to draw a number of conclusions about the status and sensitivity to change of fish populations in the study reach, as well as about how they appear to have changed in the period of one year. We cannot yet identify trends in these changes, but a number of recommended future studies should allow this.

### 6.1 Methods.

The methods used provide statistically similar samples of the fish population and there is some positive correlation between hydroacoustic densities and electric fishing catch per unit effort.

Electric fishing is more effective in the shallower margin zones than the deeper centre channel zones, but it is important to sample both areas to minimise sampling bias.

### 6.2 Species Occurrence, Population Dynamics and Health.

Species diversity provides a basic assessment of fishery quality and the relatively small variability between reaches and years is probably a product of the sampling methods used. The difference in species composition between the margin and centre channel samples is probably due in part to species habitat preference.

Bleak and roach are the most abundant species in all reaches, and perch are the third most abundant species in many margin samples.

Recruitment to the adult bleak population has been good in the past although possible errors in assigning age to length classes may have given misleading results. The lack of older year classes may be a result of this error, but could also be a demonstration of the sensitivity of this population to change.

Missing cohorts in the bream population may be an artefact, however if this is not the case, age frequency results suggest a population that is relatively sensitive to change, but that does not appear to be limited by growth. It will be possible to relate juvenile abundance to year class success in this species from the results of future surveys.

Dace populations have reduced since 1993, but are not widely distributed in the study area in either year.

Chub populations appear generally stable with some variability in year class abundance. Growth does not appear to be a limiting factor, and the relationship between juvenile abundance and cohort success should be possible to describe with results from future surveys.

Pike populations show a sensitivity to change that may be related to growth although their preferred food items (roach) are abundant. There is evidence that pike are poorly represented in the study area compared to other large lowland rivers, which may be a function of the influence of bioaccumulant pollutants upon the balanced development of this population.

The roach population is susceptible to enormous changes in cohort abundance, although in the last several years recruitment to the adult population has been good. A depressed growth rate is evident, possibly due to intra-specific competition for food, and may be a factor in population
change. There is a marked variability in growth rate with age which may be a result of changes in preference for or availability of diet items. The limited evidence available does not suggest a relationship between juvenile abundance and cohon success.

All species have acceptable levels of parasite infestation that would be expected from a river of this type.

### 6.3 Population Abundance.

There is evidence to suggest that electric fishing catch per unit effort (CPUE) is closely related to acoustic density for each reach. Even if electric fished samples are not fully quantitative, they do provide an index of relative change between years.

The results show a degree of consistency between years with a close relationship to electric fishing CPUE. Increases in density over time appear to be related to a reduction in mean FL, suggesting a recruitment of smaller fish to the population. It is difficult to suggest reasons for the difference in fish density between reaches, although it is likely to be a function of available habitat, water and habitat quality.

There is considerable spatial variability in fish density, but little evidence for the key factors influencing fish aggregation about a given point.

### 6.4 Denouement.

Most of the conclusions above relate to aspects of fish population dynamics and development that are likely to be strongly influenced by interactions within and changes to the trophic links of the river ecosystem, possibly initiated by environmental change. Some parts of the population have been shown to be extremely sensitive and capable of drastic change, which could be related to change in growth characteristics. Given the potential for the proposal to impinge upon the ecosystem of the river, it is essential that we understand as much as possible, the mechanisms and links within this system.

Strategic Water Resources Assessment:
Review of Great Ouse (1988-1993) and River Thames (1958-1973) Fisheries Community Data

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## EXECUTIVE SUMMARY

The report reviews studies of coarse fish communities undertaken on the River Thames (Dreadnought Reach) during 1958-73 and on the River Great Ouse since 1988 . The results are examined together with a wide range of other relevant studies to identify the main environmental influences on the fish communities of rivers like the Thames, and to comment on the sensitivity of fish to these influences. The review is intended for comparison with current studies on the Thames and to highlight requirements for further, focused studies.

Habitat requirements change as fish grow and develop. Changes are particularly rapid during the first year of life and it is important to maintain habitat diversity in the river. Young fish need cover to provide protection from predation and from high water velocities (over $2 \mathrm{~cm} \mathbf{s}^{-1}$ ). Beds of water lilies and other areas of slack water are important habitats.

Lily beds are also important feeding areas because they contain large populations of small invertebrates during the summer months. Small invertebrates may be a limiting resource for the growth of some species, such as roach, dace and small perch.

Older fish have greater ranges of habitat tolerance but, in all species, the requirements for successful spawning are shown to be more precise than those for feeding and refuge.

Water temperature, current velocity, food availability, refugia and spawning habitats are identified as key factors for the fish community.

It is recommended, in order of priority, that:

1) Further studies should be made of the habitat preferences and diel movements of 0 group fish and of the requirements for successful over-wintering.
2) Data collected from the Dreadnought Reach of the Thames in 1970-71 should be analyzed more fully to show the effects of a large reduction in the roach population during the 1960s, and to provide information intermediate between 1958-59 and current studies.
3) Sonar detection and sonic tags should be used to provide information on the spawning (and other) movements of adult fish, and on the effect of locks and weirs on such movements.
4) An assessment should be made of the availability of spawning habitats for selected fish species from the high priority Category A list (barbel, common bream, chub) that appear to have limited recruitment in the Thames.
group fish can feed and find cover from predation and high water current velocities. The young fish need areas with a flow <2 cm $s^{-1}$. Nuphar beds and other areas of slack water are important habitats, which need to be maintained in the river.

The studies of 0 group fish in the Great Ouse confirm the importance of water temperature in determining growth rates, especially in the first two or three months of life. However, they also showed evidence of food limitation in mid/late summer, when most if not all the fish species had spawned and the eggs had hatched. Changes in river temperatures induced by anthropogenic changes could add to the year-to-year variations that occur naturally though climatic influences.

Much of the food taken by young fish is cladocerans and chironomid larvae known to be in the Nuphar beds. There is evidence of scarcity of invertebrate food for several species of fish. Nuphar beds are the richest areas for invertebrates during the summer months and more food would be available during the main growth season of the fish if there was a greater area of Nuphar in the Thames. Although the data from rivers other than the Thames provide strong indications of the optimum environmental conditions for most species, they also show the need for local knowledge concerning the river and its fish community.

The studies on the Thames, based at the University of Reading, continued from 1958 to 1972. However, the study of the growth, survival and population densities of larger fish were based on samples taken in 1958-59, and those of juvenile fish were based on samples taken in 1967 and 1968. Studies on the energetics and production of the fish and of the ecosystem as a whole are based on these two sets of samples. The study of larger fish was repeated in 1970-71 but the data were never fully analyzed because Dr Berrie left the University to take up another post before this had been achieved. These data are still of considerable interest because of the changes that took place in the structure of the fish community between 1958-59 and 1970-71.

### 5.2 Summary of key factors

The key environmental factors influencing critical life stages of coarse fish are summarized below. Note that these factors do not operate in isolation; they are often interrelated and may act on fish in a synergistic manner.

### 5.2.1 Water temperature

Above average temperatures decrease egg incubation periods, increase fish growth rates and improve the swimming ability of 0 group fish. All these elements lead to higher 0 group survival rates and to improved year-class strength. The timing of spawning is partly temperature controlled, and some species require high temperatures ( $>18^{\circ} \mathrm{C}$ ) before spawning occurs (e.g. tench, carp). Sudden decrease in water temperature during the spawning period can inhibit spawning activity and may cause egg resorption.

Velocities $>2 \mathrm{~cm} \mathrm{~s}^{-1}$ can lead to the displacement of 0 group fish, especially during their first two or three weeks of life. Hatching of the different fish species may extend from April to July in many rivers, including the Thames. Consequently, newlyhatched fish are present throughout this period and will require suitable refugia.

Most fish species require some flow of water over their spawning substrata to enable the eggs to be well aerated. The flows observed for the target species are indicated in Table II, ; they range from near zero in backwaters (tench, pike) to c. $50 \mathrm{~cm} \mathrm{~s}^{-1}$ on gravel beds (dace, chub, barbel).

### 5.2.3 Food availability and fish refugia

These two elements are often related, especially with respect to 0 group fish. Marginal vegetation and/or backwater areas provide shelter from high water velocities for 0 group fish, and also constitute feeding areas.

O group fish require small food particles when they start feeding on external food sources (as opposed to feeding on yolk sac supplies). Synchronisation of the phytoplankton and zooplankton cycles with hatching sometimes occurs but is not consistent between species or between years. The phytoplankton cycle is influenced by river discharge rates in the early spring and on the light regime. The timing of this cycle influences that of the zooplankton population (especially the rotifers), which eats the phytoplankton.

There is evidence from the Great Ouse studies that food supplies for 0 group fish can be limited and that this can override the influence of water temperature on growth rates. This may cause some species to switch to other food sources. Thus, roach in the main river channel switch to a detritus diet in July, but continue with a zooplankton diet in marina backwaters. Similarly, roach, dace and young perch in the Thames appeared to be short of small invertebrate food at the time of the studies. This led to a high dependence on organic detritus and to low rates of growth. The effect on survival is not known.

### 5.2.4 Spawning habitats

Older fish have a wider range of general habitat preferences than 0 group fish, but the spawning requirements for each species are defined more narrowly. Gravel beds washed by fast-flowing water are needed by barbel, dace and chub, whereas most other species spawn on various plant substrata in more slow-flowing areas of the river. These include submerged and emergent aquatic plants, and the submerged tree roots of willow and alder trees. Pike, tench and carp generally prefer vegetated areas with water velocities $<5 \mathrm{~cm} \mathrm{~s}{ }^{-1}$.

Access to spawning areas is vital, but little is known concerning the extent of coarse fish spawning migrations and how these may be affected by locks, weirs and sluices (see below).

### 5.3 Recommendations for future research <br> 5.3.1 General

The review highlights the need for more information on the habitat selection and diel movements of different species of 0 group fish, although it is known that habitat preferences vary between species and with the growth and development of individual species. Also, very little is known about the requirements for successful over-wintering, although the Great Ouse studies point to the importance of backwater areas such as marinas. The growth and early survival of 0 group fish are critical aspects in the life stages of coarse fish, and are those most likely to be affected by changes to the river environment and hydrological regime.

Age determinations on the scales and further analyses on the 1970-71 data from the Dreadnought Reach of the River Thames will show whether the large reduction in the roach population during the late 1960 s produced other effects on the fish community. The growth rate and food of roach could have changed and there could be consequential effects on other species. This study will also provide growth and age structure information for the fish community that is intermediate in time between the 1958-59 data and any recent studies by the NRA. As the material has already been collected, the cost of the study will be much smaller than if field work was involved.

Further information is needed on the movements of adult fish, particularly for spawning but also at other times of the year. The role of locks and weirs in relation to fish movements is unknown and may be of considerable importance. The use of sonar detection and sonic tags in these respects should be considered, especially for the major fish species.

Information from the 1970-71 studies indicates that, of the nine Category A species, the roach, bleak, dace, gudgeon and perch all spawn successfully in the Thames (Dreadnought Reach). Progeny of barbel, common bream, chub and pike were less numerous, which may point to restrictions in the spawning habitats. Emphasis on the spawning movements of barbel, common bream and chub would be valuable if it is intended to enhance the stocks of these species.

### 5.3.2

## Priorities for River Thames studies

The following recommendations are placed in order of priority, but it should be noted that many aspects are interrelated.
a) Determine the distribution of suitable marginal, vegetated refuge areas for 0 group fish, especially areas where the flow velocities are less than $2 \mathrm{~cm} \mathrm{~s}^{-1}$ under most summer

Table $I$. Summary of the spawning habitat requirements of 13 species of freshwater fish. Ind $=$ indifferent.

| Species | Depth (cm) | $\begin{gathered} \text { Flow } \\ \left(\mathrm{cm} \mathrm{~s}^{-1}\right) \end{gathered}$ | Substratum <br> diam. (cm) | Vegetation |
| :---: | :---: | :---: | :---: | :---: |
| Category A |  |  |  |  |
| Barbel | 14-22 | 35-49 | 2-5 | Absent |
| Bleak |  | $<20$ |  | Scimpus \& filamentous algae. |
| Bream (common) | 15-120 | $<20$ |  | Glyceria, Saqittaria, Scirpus \& filamentous algae, Salix roots, Veronica, Myosotis, Phragmites, Elodea. |
| Chub | 10-30 | 20-50 | $>0.5$ | Occasional |
| Dace | 25-40 | 20-50 | 3-25 | Occasional |
| Gudgeon | 1-8 | 2-80 | S-30 | Fontinalis, tree roots, potamogeton pectinatus. |
| Perch | Variable |  |  | Submerged Salix \& Alnus roots. |
| Pike |  | $<5$ | Ind. | Myriophyllum, flooded riparian vegetation. |
| Roach | 5-45 | $>20$ | 5-15 | Fontinalis, Elodea, Scirpus, Salix roots. |
| Category B |  |  |  |  |
| Carp | Variable | $<5$ | Ind. | Submerged riparian vegetation, Carex, Glyceria, Phragmites, Rorippa, Scirpus \& filamentous algae. |
| Ruffe |  |  | $<5$ | Plants/moss. |
| Bream <br> (silver) | Variable | < 20 | Ind. | variety of aquatic plants. |
| Tench |  | $<5$ | Ind. | Myriophyllum \& Lemna |

References: Baras 1992, Baras \& Philippart 1993, Bastl 1969, Copp \& Mann 1993, Diamond 1985, Fedorova \& Vetkasov 1971, Grandmottet 1983, Hancock et al., 1976, Holcik \& Hruska 1966, Kaufmann et al., 1991, Kennedy 1969, Kennedy \& Fitzmaurice 1968, 1972, Kovaleva 1967, Lelek 1987, Mann 1978, 1993, Mills 1981 a, b, c, Sych 1955, Vollestad \& L'Abée-Lund 1987.

Table II. Published spawning times and water temperatures for 13 species of frestrwater fish, and including published (1) and unpublished (2) data from the Thames and unpublished data from the Great Ouse (3).

| Species | Months | Temp. <br> ${ }^{\circ} \mathrm{C}$ | References |
| :---: | :---: | :---: | :---: |
| Category A |  |  |  |
| Barbel | VI | 13.5 | Hancock et al. 1976 |
| Bleak (1,3) | V-VII | > 12 | R. Stour (Mann, unpublished) Leeming 1963, Mackay \& Mann 1969 |
| C. Bream (3) | V -VI | 12-18 | Hartley 1947, Leeming 1963, Holcik\& Hruska 1966, Kennedy \& Fitzmaurice 1968 |
| Chub (3) | V-VII | $>12$ | Cragg-Hine 1963, Leeming 1963, Hellawell 1971, Mann 1976b, Krupka 1988 |
| Dace $(2,3)$ | IT-IV | 8-12 | Cragg-Hine 1963, Kennedy 1969, Mann 1974, Hellawell 1974 Mills 1981 a,b |
| Gudgeon (1, 3) | V-VII | 13-17 | Hartley 1947, Mathews 1971, Kennedy \& Fitzmaurice 1972, Penaz \& Prokes 1978, Mann 1980a |
| Perch (1,3) | IV-V | 8-14 | Hartley 1947, Williams 1963 Thorpe 1977, Mann 1978, Zeh et al. 1989 |
| Pike (3) | III-V | 6-14 | Mann 1976a, Raat 1988 |
| Roach (1,3) | V-VI | 7-17 | Hartley 1947, Mackay \& Mann 1969, Hellawell 1972, Mann 1973, Diamond 1985, Vollestad \& L'Abée-Lund 1987 |
| Category B |  |  |  |
| Carp (3) | VI-VII | $>15$ | Crivelli 1981 |
| Ruffe (3) | ITI-VII |  | Leeming 1963, Bastl 1988 |
| S. Bream (3) | V-VII |  | Hartley 1947, Leeming 1963 |
| Tench (3) | VI | > 18 | Mills 1991, Copp \& Mann 1993 |

South West Oxfordshire Reservoir Proposal (SWORP)

## River Thames Juvenile Fish Survey 1993 <br> Volume 1 - Main Report




April 1994

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## Executive Summary

This report presents the results of a juvenile fish survey of the River Thames between Oxford and Days Weir, undertaken in July and August 1993 by King's Environmental Services. It is along this stretch of the Thames that the outlet of the South West Oxfordshire Reservoir proposed by Thames Water Utilities is likely to be sited. The 1993 juvenile fish survey follows on from previous surveys undertaken in 1991 by the Environmental Advisory Unit (EAU), and in 1992 by Royal Holloway and Bedford New College (RHBNC). Fourteen sites were sampled in 1993, thirteen of which corresponded approximately to the 13 sites sampled in 1992.

Each survey site was sampled with three seine net hauls using a $25 \mathrm{~m} \times 3 \mathrm{~m}$ micromesh net with 3 mm mesh size. With the exception of the additional site at Abingdon Marina, where three essentially identical net hauls were taken, three contrasting sub-sites were sampled at each site, representing three distinct habitat types, namely:
i shallow with macrophytes,
ii shallow without macrophytes,
iii deep without macrophytes.
A comparison of the baseline juvenile fish stocks in the River Thames in 1993 with those recorded in 1992 and 1991 is as follows:

|  | 1993 | 1992 | 1991 | 1991 |
| :--- | :--- | :--- | :--- | :--- |
|  | KES | RHBNC | EAU | EAU |
| Number of sites | 14 | 13 | $13^{*}$ | 36 |
| Number of seinings | 42 | 38 | 13 | 36 |
| Juvenile density $\left(\mathrm{nm} \mathrm{m}^{-2)}\right.$ | 3.84 | 5.81 | 4.07 | 3.07 |
| Catch per unit effort | 355 | 706 | 607 | 490 |
|  | *sites corresponding to 1992 | E 1993 surveys |  |  |

Both juvenile density and catch per unit effort were found to be substantially lower in 1993 than for the two previous surveys. This difference in densities is almost entirely due to differences in the numbers of juvenile roach. Although the dominant species in each of the three years, the density of $0^{+}$roach in 1992 was more than twice that of 1993. If roach are subtracted from the overall mean density, the resulting densities for the remaining species are markedly similar: $1.96 \mathrm{~nm}^{-2}$ in 1993, $1.84 \mathrm{~nm}^{-2}$ in 1992.

Mean densities of $0^{+}$fry were compared for the three habitat types and found to be as follows:

| Habitat type | With macrophytes |  | Shallow without <br> macrophytes |
| :--- | :---: | :---: | :---: | | Deep without |
| :---: |
| macrophytes |

Thus the very clear trends in overall fry distribution between the habitat types recorded in 1992 were not observed in 1993. In 1992 and 1993, roach densities were found to be highest in macrophyte sites and lowest in deep sites. Relationships between habitat type and density of other species was not the same between survey years.
3.4.9 Instantaneous rates of mortality ( $Z$ ) and survival (S)
3.4.9.1 Instantaneous rates of mortality and survival for the $0^{+}$to $1^{+}$year classes in 1993 and 1992 are presented in Table 5.
3.4.9.2 The following observations can be made:
i) Instantaneous rates of mortality for bleak, chub, perch and roach were all higher in 1992 than in 1993. Only gudgeon had a higher instantaneous rate of mortality in 1993.
ii) Survival values for bleak, chub, perch, and roach were all higher in 1993 than in 1992. Only gudgeon had a higher survival rate in 1992.
iii) The highest 1993 Z value (and the lowest S value), was calculated for gudgeon, followed by perch, roach, chub, bleak then dace.
iv) The highest 1992 Z value (and the lowest S value), was calculated for roach, followed by chub, perch, gudgeon and bleak.
3.4.10 Length weight conversions
3.4.10.1 Scatter plots (including 95\% confidence limits of the regression line), of $\log$ length ( mm ) vs. $\log$ preserved weight ( g ) are presented in Figure 13.
3.4.10.2 The equations for length weight conversion are presented below:

Bleak $\quad \log \mathrm{w}=-4.8654+2.9296(\log \mathrm{l})$
Chub $\quad \log w=-4.6662+2.8603(\log 1)$
Dace $\quad \log w=-4.4374+2.7062(\log 1)$
Gudgeon $\log w=-4.8433+2.9882(\log 1)$
Perch $\quad \log w=-4.7558+2.9407(\log 1)$
Roach $\log w=-4.5888+2.8590(\log 1)$

### 3.4.11 Biomass (standing crop)

3.4.11.1 Overall estimates of biomass ( $\mathrm{gm}^{-2}$ ) for the six major species (preserved weight); bleak, chub, dace, gudgeon, perch and roach are presented in decreasing order in the table below (these results are illustrated graphically in Figure 14):

| Roach | 2.32 |
| :--- | :--- |
| Perch | 1.12 |
| Gudgeon | 0.56 |
| Chub | 0.17 |
| Dace | 0.14 |
| Bleak | 0.07 |
| Total | 4.38 |

### 3.5.9 Instantaneous rates of mortality ( $Z$ ) and survival (S)

3.5.9.1 Values of instantaneous rate of mortality and survival for 1993 and 1992 are presented in Table 5. Observations in both surveys are as follows:
i) Instantaneous rate of mortality of chub and perch was greatest in deep sub-sites.
ii) Instantaneous rate of mortality of gudgeon was greatest in shallow sub-sites.
iii) Instantaneous rate of mortality of perch was lowest in shallow subsites.
iv) Instantaneous rate of mortality of gudgeon was lowest in the macrophyte sub-sites.
3.5.9.2 Survival values reiterate these results showing a direct inverse relationship between survival and mortality.

### 3.5.10 Biomass

3.5.10.1 Values of biomass for the six major fish species in the three different habitat types are ranked in descending order in the following table:

| Deep | Shallow |  | Macrophytes |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Roach | 1.61 | Gudgeon | 1.03 | Roach | 3.60 |
| Perch | 0.82 | Perch | 0.52 | Perch | 1.06 |
| Gudgeon | 0.27 | Reach | 0.50 | Gudgeon | 0.39 |
| Bleak | 0.09 | Dace | 0.27 | Chub | 0.19 |
| Chub | 0.06 | Chub | 0.14 | Bleak | 0.09 |
| Dace | 0.04 | Bleak | 0.03 | Dace | 0.03 |
| Total | 2.89 | Total | 2.49 | Total | 5.36 |

These results are illustrated graphically in Figure 22
3-5.10.2 The above table indicates that although total densities were relatively similar in each of the habitat types, biomass in the macrophyte sub-sites appeared considerably higher. The biomass values of roach and perch were particularly high in the macrophytes relative to the other habitat types. Gudgeon biomass was relatively high in the shallow sites, where it ranked first. The biomass of dace was also higher in the shallow sites relative to the other two habitat types.
3.5.10.3 Statistical analysis was not carried out on the biomass data, as it was beyond the scope of this investigation. The point raised in 3.5.6.4 should be borne in mind when making inferences from these data.



Figure 14

## Combined total biomass



Combined total biomass $=4.3844 \mathrm{~g} / \mathrm{m}^{2}$

## Strategic Water Resource Studies

## River Thames <br> Juvenile Fish Survey 1994

## Volume 1 - Main Report




November 1994

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# River Thames Juvenile Fish Survey 1994 

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## Cover Photograph

Measuring current velocity at Site 13 'shallow without macrophytes' with the Sensa RC2 electromagnetic flow meter. The net has been set and pulled but the net marker buoys and transect posts remain in situ. The buoys and posts define fixed points on the perimeter of the netted arc so that the total netted area can be calculated from the distances and angles of these markers to the centre point on the bank. The posts define the parallel and perpendicular transects along which depth, temperature, and water velocity are measured.

## Executive Summary

This report presents the results of a juvenile fish survey of the River Thames between Oxford and Days Weir, undertaken in July and August 1994 by King's Environmental Services (KES). It is along this stretch of the Thames that the outlet of the South West Oxfordshire Reservoir proposed by Thames Water Utilities is likely to be sited. The 1994 juvenile fish survey follows on from previous surveys undertaken in 1991 by the Environmental Advisory Unit (EAU), in 1992 by Royal Holloway and Bedford New College (RHBNC) and in 1993 by KES. Fourteen sites were sampled in 1994, which corresponded very closely to the 14 sites sampled in 1993.

Each survey site was sampled with three seine net hauls using a $25 \mathrm{~m} \times 3 \mathrm{~m}$ micromesh net with 3mm mesh size. With the exception of the Abingdon Marina site, where three essentially identical net hauls were taken, three contrasting subsites were sampled at each site, representing three distinct habitat types:
i) deep without macrophytes (mean depth $1.37 \mathrm{~m} ; 0-3 \%$ macrophyte cover),
ii) shallow without macrophytes (mean depth $0.52 \mathrm{~m} ; 0-5 \%$ macrophyte cover),
iii) shallow with macrophytes (mean depth $0.88 \mathrm{~m} ; 10-95 \%$ macrophyte cover).

A comparison of the baseline juvenile fish stocks in the River Thames in 1994 with those recorded in 1993, 1992 and 1991 is as follows:

|  | 1994 | 1993 | 1992 | 1991 | 1991 |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  | KES | KES | RHBNC | EAU | EAU |
| Number of sites | 14 | 14 | 13 | $13^{*}$ | 36 |
| Number of seinings | 42 | 42 | 38 | 13 | 36 |
| Juvenile density(nm-2) | 13.03 | 3.84 | 5.81 | 4.07 | 3.07 |
| Catch per unit effort | 1095 | 355 | 706 | 607 | 490 |
|  | *sites corresponding to | 1992 \& 1993 surveys |  |  |  |

Both juvenile density and catch per unit effort in 1994 were substantially higher than for all previous surveys. This difference in densities is largely due to differences in the numbers of juvenile roach. The mean density of $0^{+}$roach was $9.23 \mathrm{~nm}^{-2}$ in 1994, $1.29 \mathrm{~nm}^{-2}$ in 1993 and $2.52 \mathrm{~nm}^{-2}$ in 1992. The densities of other species are less variable between years with combined mean densities for all other species of 3.59 $\mathrm{nm}^{-2}$ in 1994, $1.96 \mathrm{~nm}^{-2}$ in 1993 and $1.84 \mathrm{~nm}^{-2}$ in 1992.

Mean densities of $0^{+}$fry were compared for the three habitat types and found to be as follows:

| Habitat type | Deep without <br> macrophytes | Shallow without <br> macrophytes | Shallow with <br> macrophytes |
| :--- | :---: | :---: | :---: |
| Density $0^{+}$fry $1992\left(\mathrm{~nm}^{-2)}\right.$ | 0.72 |  | 5.36 |

In both 1992 and 1994 there were clear trends in juvenile distribution between habitat types although clear cut trends were fewer in 1993. The preference for macrophyte-rich sub-sites was most marked for roach and was apparent for this species in all three years.

The following relationships between habitat type and juvenile fish species were found in 1994:
i $0+$ dace, chub, gudgeon and perch showed a clear preference for the shallow rather than the deep sites,
ii $0+$ dace and chub showed a preference for the shallow macrophyte-poor sites rather than for the shallow macrophyte-rich sites,
iii $0+$ perch and gudgeon showed a preference for the shallow macrophyte-rich sites rather than the shallow macrophyte-poor sites,
iv $0+$ roach showed a very strong preference for the macrophyte-rich sites but in macrophyte-poor areas appeared to prefer the deep to the shallow sub-sites.
In addition, it was found that:
i perch, pike and roach density was positively correlated with the percentage macrophyte cover,
ii chub, dace and gudgeon densities were negatively correlated with depth,
iii chub and dace densities were positively correlated with the percentage of sand and gravel in the substrate.
There was no apparent correlation between juvenile fish density and water velocity.

This report also discusses the potential impacts to juvenile fish which may arise from construction and operation of the proposed South West Oxfordshire Reservoir and presents proposals for further studies on juvenile cyprinids in the Oxford region of the River Thames.

## 6 Summary and Conclusions

6.1 The current survey recorded the highest mean density of juvenile fish for the three survey years (1992-1994) in which densities could be calculated. Roach comprised $72.4 \%$ of the total 1994 catch of 46,030 fish. Of the remaining species, gudgeon comprised $18.09 \%$ of the 1994 catch with chub, bleak, dace and perch together accounting for $8.13 \%$. Roach was also the most abundant species in 1991, 1992 and 1993, although the relative abundance of the other species varied from year to year.
6.2 The combined mean density of all species was $5.81 \mathrm{~nm}^{-2}$ in $1992,3.84 \mathrm{~nm}^{-2}$ in 1993 and $13.03 \mathrm{~nm}^{-2}$ in 1994. The great majority of the density difference between years was due to the varying year class strength of $0+$ roach, with densities of 2.52, 1.29, and $9.23 \mathrm{~nm}^{-2}$ recorded in 1992, 1993 and 1994 respectively. The mean density of $1+$ roach showed much smaller variation between years with values of $0.15,0.26$ and $0.19 \mathrm{~nm}^{-2}$ for 1992-94 respectively.
6.3 Z and S values have been calculated for the major species by following cohorts from year to year. Calculated survival rates vary greatly between years and between species and are difficult to interpret because factors other than mortality affect the observed density of an individual species' year class. Calculated $S$ values for $0+/ 1+$ roach, dace, chub and perch are within the range of 0.02 to 0.3 and appear relatively realistic. Calculated first year survival rates for bleak and bream are particularly erratic, ranging over the survey years from 0.05 to $>5$. The most likely explanation for these results is that only a part of the $0+$ cohort is being captured because of the relatively late spawning period of these two species.
6.4 Calculated survival values for roach, together with the relative constancy of the $1+$ year class strength over the survey years, provides some indication that density-dependent mortality may operate for this species in the first year of life.
6.5 Although total juvenile density varied enormously in the different sites and sub-sites, there was a general overall trend of increasing densities from the top to the bottom of the survey stretch. This general trend was also apparent in 1992 and 1993, although the reasons for the trend are not clear.
6.6 Analysis of the measured habitat variables for the 1994 survey confirmed that, despite a small degree of overlap, the three habitat (sub-site) types sampled, namely 'deep without macrophytes', 'shallow without macrophytes' and 'shallow with macrophytes' formed statistically distinct sets with regard to macrophyte cover and depth. Although statistically less distinct, the habitat types also differed with respect to water velocity. Mean velocities were $0.023,0.31$ and $0.34 \mathrm{msec}^{-2}$ in the 'deep', 'shallow without'
and 'shallow with macrophyte' sub-sites respectively. Although substrate composition varied greatly between sites and between sub-sites, there was no apparent relationship between substrate nature and sub-site type.
6.7 The 1994 survey indicated clear relationships between habitat type and fish density and in most cases these relationships were statistically significant. Mean total densities of $0+$ fish were $32.48,5.12$ and $4.17 \mathrm{~nm}^{-2}$ for the 'shallow with macrophyte', 'shallow without macrophyte' and 'deep' subsites respectively. These total densities are heavily influenced by the dominant species, roach.
6.8 The $19940+$ cohorts of perch, chub, dace and gudgeon, together with the $1+$ gudgeon cohort, all showed a clear preference for the shallow (with or without macrophytes) rather than the deep sub-sites. Although not statistically significant, bleak appeared to show a similar preference. The $0+$ cohorts of gudgeon and perch showed a statistically significant preference for the 'with macrophyte' sub-sites although this trend was not apparent for the $1+$ gudgeon. Conversely, the $0+$ chub and dace showed a preference for the shallow macrophyte-free sites.
6.9 Both $0+$ and $1+$ roach showed a highly significant association with the macrophyte-rich sites. Unlike the other species however, $0+$ roach in the macrophyte free habitats displayed a preference for the deep rather than the shallow sub-sites.
6.10 When the habitat variables were considered independently rather than by sub-site groups, a number of significant correlations were obtained:
i perch, pike and roach were positively correlated with percentage macrophyte cover,
ii chub, dace and gudgeon were negatively correlated with depth,
iii chub and dace were positively correlated with the percentage of sand and gravel in the substrate.
There was no apparent correlation between fish density and water velocity.
6.11 Stepwise multiple regression using nine potential habitat predictors suggested that $34.8 \%$ of observed variation in total fish density is attributed to \% macrophyte cover ( $22.57 \%$ ) and water velocity ( $34.79 \%$ ). Biomass showed similar trends with $41 \%$ of the variation attributed to $\%$ cover, water velocity and \% sand and gravel. Stepwise multiple regression was not attempted for individual species or year class cohorts.
6.12 The associations demonstrated between habitat variables and fish density in the 1994 survey generally confirm and extend observations made in the earlier surveys. The potential value of stepwise multiple regression has been demonstrated. However, the present system of sub-site selection, which seeks uniformity of habitat variables within one of three defined habitat types, may not be optimum for this type of analysis.
6.13 The possible underlying reasons for the variation in $0+$ year class strength of the different fish species in the Thames have been discussed in this report. Although direct evidence from the Thames is tenuous, published information for other systems supports the hypothesis that spring and summer water velocities, and spring water temperatures, are likely to be the principal controlling factors. A longer term data set for juvenile fish abundance and more comprehensive seasonal temperature records will be required to adequately test this hypothesis.
6.14 Mean lengths and ranges of the various fish species show significant variation over the four survey years, although generally these variations are not great. Growth rates of $0+$ to $2+$ fish are generally similar to national standards and are thus within the range expected.
6.15 Experience gained from the 1991-1994 surveys of the Thames suggests that late July/early August is an appropriate sampling period for juveniles of most species of coarse fish inhabiting the river. However, at least in some years, this date appears to be rather early for the later spawning species such as bleak and bream, leading to under estimation of the $0+$ year class strength for these two species
6.16 Considering the proposed South West Oxfordshire Reservoir, the most significant impacts to juvenile fish are likely to arise from changes to the flow and temperature regime and from changes in food availability. Changes to flow regime may affect juvenile fish directly, if their velocity tolerance is exceeded, or indirectly, by affecting the distribution and abundance of aquatic macrophyte stands. Water quality impacts would probably be of relatively minor significance if the proposed reservoir were fully mixed but would be potentially more severe if the reservoir were allowed to stratify thermally in summer.
6.17 Although of necessity very tentative, the preliminary assessment of potential impacts presented in this report is nevertheless important. It provides a preliminary focus for the development of further studies on both the population dynamics of juvenile cyprinids in the Thames and the vulnerability of these species to the environmental perturbations that may arise from reservoir development.




Species

|  | Deep |  |  |  | Shallow |  |  |  | Macrophytes |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0+$ | $1+$ | >1+ | Total | 0+ | $1+$ | >1+ | Total | 0+ | 1+ | >1+ | Total |
| Bleak | 0.013 | 0.002 | 0.004 | 0.018 | 0.244 | 0.022 | 0.005 | 0.270 | 0.291 | 0.445 | 0.024 | 0.760 |
| Bream | 0.000 | 0.000 | 0.002 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.025 | 0.009 | 0.011 | 0.045 |
| Chub | 0.129 | 0.067 | 0.001 | 0.197 | 0.395 | 0.031 | 0.002 | 0.428 | 0.338 | 0.066 | 0.002 | 0.405 |
| Dace | 0.137 | 0.004 | 0.000 | 0.141 | 0.401 | 0.042 | 0.000 | 0.443 | 0.084 | 0.015 | 0.000 | 0.099 |
| Gudgeon | 0.732 | 0.088 | 0.010 | 0.830 | 2.404 | 0.514 | 0.341 | 3.260 | 2.903 | 0.382 | 0.062 | 3.346 |
| Perch | 0.092 | 0.011 | 0.008 | 0.111 | 0.023 | 0.005 | 0.001 | 0.029 | 0.247 | 0.043 | 0.012 | 0.302 |
| Pike | 0.002 | 0.000 |  | 0.002 | 0.000 | 0.001 |  | 0.001 | 0.006 | 0.004 |  | 0.011 |
| Roach | 3.047 | 0.056 | 0.016 | 3.119 | 1.877 | 0.045 | 0.008 | 1.929 | 28.364 | 0.507 | 0.034 | 28.905 |
| Barbel | 0.001 |  |  | 0.001 | 0.022 |  |  | 0.022 | 0.007 |  |  | 0.007 |
| Bullhead | 0.011 |  |  | 0.011 | 0.011 |  |  | 0.011 | 0.069 |  |  | 0.069 |
| Minnow | 0.003 |  |  | 0.003 | 0.014 |  |  | 0.014 | 0.199 |  |  | 0.199 |
| Ruffe | 0.002 |  |  | 0.002 | 0.009 |  |  | 0.009 | 0.071 |  |  | 0.071 |
| Stickieback | 0.002 |  |  | 0.002 | 0.012 |  |  | 0.012 | 0.011 |  |  | 0.011 |
| Stone loach | 0.001 |  |  | 0.001 | 0.002 |  |  | 0.002 | 0.014 |  |  | 0.014 |
| Total |  |  |  | 4.439 |  |  |  | 6.432 |  |  |  | 34.244 |



Site number

Relative biomass of major fish species in each habitat
Deep without macrophytes


Shallow without macrophytes


Biomass $=14.72 \mathrm{gm}-2$

Shallow with macrophytes




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# River Thames Juvenile Fish Survey 1995 

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## Cover Photograph

Measuring current velocity at Site 10 'shallow with macrophytes' with the Sensa RC2 electromagnetic velocity meter. Following velocity measurement, the enclosed macrophytes are cut with a scythe and removed prior to hauling and landing the net. The buoys and posts define fixed points on the perimeter of the netted are so that the total netted area can be calculated from the distances and angles of these markers to the centre point on the bank. Depth, temperature and water velocity are measured along two transects, one parallel to and the other perpendicular to the bank. The two posts define the parallel transect.

## 6 Summary and Conclusions

6.1 The 1995 survey recorded the highest mean density of juvenile fish for the four survey years (1992-1995) in which densities could be calculated. As in previous years, roach was the dominant species comprising $65.7 \%$ of the total 1995 catch of 59,926 fish. Previous total catches were 31,891 in 1992, 14,936 in 1993 and 46,030 in 1994. Of the remaining species, gudgeon comprised $17.19 \%$ of the 1995 catch, chub $7.76 \%$ and bleak $4.22 \%$, with dace, perch and bream together accounting for $3.58 \%$. Although roach dominated the catch in all survey years, the relative abundance of the other species varied significantly from year to year.
6.2 The combined mean density of all species was $4.38 \mathrm{~nm}^{-2}$ in $1992,3.52 \mathrm{~nm}^{-2}$ in $1993,13.03 \mathrm{~nm}^{-2}$ in 1994 and $16.05 \mathrm{~nm}^{-2}$ in 1995. The great majority of the density difference between years was due to the varying year class strength of $0+$ roach, with densities of $2.52,1.29,9.23$ and $10.55 \mathrm{~nm}^{-2}$ recorded in 1992 to 1995 respectively. The mean density of $1+$ roach showed parallel interyear variation with values of $0.15,0.26$ and 0.19 and $1.10 \mathrm{~nm}^{-2}$ for 1992 to 1995 respectively.
6.3 Z and S values were calculated for the major species by following cohorts from year to year over the period 1992-1995. Calculated survival rates vary greatly between years and between species and in many cases are difficult to interpret because factors other than mortality may affect the observed density of an individual species' year class. Calculated $S$ values for $0+/ 1+$ roach, dace, chub and perch are generally within the range of 0.02 to 0.35 and appear relatively realistic. Calculated first year survival rates for bleak and bream, and to a lesser extent for gudgeon, are particularly erratic, ranging over the survey years from 0.05 to $>5$. The probable explanation for these results is that only a part of the $0+$ cohort was captured. This is due to the relatively late spawning period of bleak and bream and probably because the bottom living habit of gudgeon makes it difficult to net the smallest individuals so that they are under-represented in the catch.
6.4 Calculated one year survival values for the 1992, 1993 and 1994 cohorts of roach and chub are markedly similar (mean $S$ roach $=0.12$, mean $S$ chub $=$ 0.3 ), despite substantial variation in initial cohort strength, suggesting that density-dependent mortality does not operate for these two species. In contrast, the calculated $S$ values for perch suggest that density-dependent mortality may be occurring in this species.
6.5 Although total juvenile density varied enormously in the different sites and sub-sites in all survey years, a general overall trend of increasing densities from the top to the bottom of the survey stretch was clearly apparent in the 1992-1994 survey years. In contrast, fish densities were much more uniformly distributed along the survey stretch in 1995. The reasons for this trend, and its apparent absence in 1995, remain obscure.
6.6 Uniquely amongst the Thames fish species, the mean density of ruffe increased consistently from 1992 to 1995 ( $0.001,0.01,0.05$ and $0.17 \mathrm{~nm}^{-2}$ 1992-1995 respectively) possibly indicating a long term trend.
6.7 As in 1993 and 1994, analysis of the measured habitat variables for the 1995 survey confirmed that, despite a small degree of overlap, the three habitat (sub-site) types sampled, namely 'deep without macrophytes', 'shallow without macrophytes' and 'shallow with macrophytes' formed statistically distinct sets with regard to macrophyte cover and depth.
6.8 Mean water velocity in 1995 for the 39 river sub-sites, at $0.012 \mathrm{~ms}^{-1}$, was the lowest recorded for the three years in which velocity was measured (mean velocity $1993=0.0623 \mathrm{~ms}^{-1} ; 1994=0.030 \mathrm{~ms}^{-1}$ ). There were no significant differences in water velocity or temperature between the sub-site types in 1995. There was a clear longitudinal trend of increasing water temperature from the top to the bottom of the survey stretch but this was simply a reflection of prevailing weather conditions during the survey period. Although substrate composition varied greatly between sites and between sub-sites, the only clear pattern was the positive relationship between depth and the percentage of bare clay substrate.
6.9 As in previous surveys, the 1995 data indicated a clear relationship between habitat type and total fish density. Mean total densities of $0+$ fish were 27.49, 10.52 and $3.02 \mathrm{rm}^{-2}$ for the 'shallow with macrophyte', 'shallow without macrophyte' and 'deep' sub-sites respectively. These total $0+$ densities were heavily influenced by, but not exclusively due to, the dominant species, roach.
6.10 Statistical analysis of the 1995 juvenile fish and environmental data established a number of associations and correlations which largely confirmed the findings of the 1993 and 1994 surveys. The more significant findings from the 1995 data set are as follows:
i $0+$ chub and gudgeon showed a clear preference for the shallow rather than the deep sub-sites;
ii $0+$ perch and roach showed a very strong preference for the macrophyte-rich sub-sites over either the deep or shallow macrophyte-poor sub-sites;
iii $1+$ roach showed a preference for the macrophyte-rich sub-sites rather than for the deep macrophyte-poor sub-sites;
iv $0+$ bleak, perch, roach and ruffe densities were positively correlated with the percentage macrophyte cover;
v 0+ bleak, chub and gudgeon densities were negatively correlated with depth and gradient;
vi $0+$ chub were positively correlated with the percentage of sand and negatively correlated with the percentage of clay in the substrate;
vii 1+ bleak, chub and gudgeon and $2+$ gudgeon were positively correlated with the percentage of gravel substrate.
6.11 When the 1993 to 1995 data sets were pooled to give a larger data set (I17 sub-sites), relationships and correlations generally became more clear cut, especially for the $1+$ and $2+$ age classes. Several additional relationships became apparent that were not evident from the 1995 data set alone, namely:
i $0+$ chub showed a preference for shallow macrophyte-poor sub-sites over both shallow macrophyte-rich and deep macrophyte-poor subsites;
ii $0+$ dace showed similar preferences to 0+ chub;
iii $0+$ perch showed a strong preference for macrophyte rich sub-sites - but in the absence of macrophytes preferred deep to shallow subsites;
iv both $0+$ and $1+$ pike showed a preference for macrophyte-rich sites over the other two habitat types.
6.12 The associations demonstrated between habitat variables and fish density in the 1992-1994 data sets (RHBNC 1992; KES 1993, 1994) and in the 1995 and combined 1993-95 data sets means that it is now possible to define provisionally the habitat preferences of juveniles of the main fish species occurring in the Oxford region of the Thames. As a general rule, it would appear that for a given species, the habitat preferences of $0+$ and $1+$ fish are similar, but the usually weaker statistical relationships for $1+$ fish suggest that they become more catholic in their habitat requirements with increasing age. Habitat preferences appear to be as follows:
i roach show a very strong preference for habitats with macrophyte cover, the preference for cover appearing to eclipse all other habitat requirements;
ii for perch, macrophyte cover is the single most important factor governing juvenile distribution. Perch juveniles display a marked aversion to macrophyte free shallows but can occur in significant numbers in deeper cover-free sites;
iii the preferred habitat for both chub and dace is gently shelving shallows with a sandy or gravelly bed;
iv juvenile pike lurk in weedy shallows, as do the adults;
v juvenile bleak appear to prefer shallow water with a gravelly substrate and good macrophyte cover;
vi bream show a fairly clear preference for sites containing macrophyte cover.
6.13 The possible underlying reasons for the variation in $0+$ year class strength of the different fish species in the Thames are discussed in this report. Published information for other river systems supports the hypothesis that spring and summer water velocities and temperatures, are likely to be major controlling factors. Examination of flow records for the 1992-1995 survey years has suggested that there might be a negative relationship - between early summer river flows and the $0+$ year class strength for roach
and bream. However, a much longer term data set will be required for these or other possible relationships to be confirmed. Evaluation of the potential effects of spring and summer temperature on year class strength has been precluded because of the fragmentary nature of the currently available temperature record.
6.14 Mean lengths and ranges of the various fish species showed some variation over the five survey years, although generally these variations were not great. Growth rates of $0+$ to $2+$ fish were generally similar to national standards and were within the range expected. The $1995 b$ constants for bleak, chub, dace, perch and roach were higher than those determined in 1994 but the gudgeon $b$ constants were lower.
6.15 As shown by the 1992-1995 surveys of the Thames, late July/early August is an appropriate sampling period for juveniles of most species of coarse fish inhabiting the river. However, this date is rather early for the later spawning bleak and bream, leading to under estimation of the $0+$ year class strength for these two species. A greater number of 0+ bleak and bream were captured in 1995 than in 1993 or 1994, probably indicating earlier spawning, but it is likely that even in 1995, $0+$ individuals of both species were under-represented in the catch.
6.16 In the absence of any additional information concerning the design or operation of the proposed South West Oxfordshire Reservoir, no further assessment of potential impacts of the scheme to juvenile fish was undertaken.

Table 4: Comparison of mean densities ( $\mathrm{nm}^{-2}$ ) for all species 1992-1995

| 1995 Survey |  | 1994 Survey |  | 1993 Survey |  | 1992 Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Roach | 10.55 | Roach | 9.44 | Roach | 1.88 | Roach | 3.97 |
| Gudgeon | 2.76 | Gudgeon | 2.36 | Gudgeon | 0.99 | Bleak | 0.68 |
| Chub | 1.25 | Chub | 0.33 | Chub | 0.27 | Gudgeon | 0.63 |
| Bleak | 0.68 | Bleak | 0.31 | Perch | 0.26 | Chub | 0.23 |
| Perch | 0.25 | Dace | 0.24 | Dace | 0.23 | Perch | 0.13 |
| Bream | 0.17 | Perch | 0.19 | Bleak | 0.07 | Dace | 0.08 |
| Dace | 0.15 | Minnow | 0.06 | Minnow | 0.05 | Hybrids | 0.03 |
| Rutte | 0.14 | Ruffe | 0.05 | Buthead | 0.02 | Bream | 0.02 |
| Minnow | 0.08 | Bullhead | 0.03 | Bream | 0.01 | Minnow | 0.01 |
| Stickleback | 0.013 | Bream | 0.01 | Stickleback | 0.01 | Ruffe | 0.001 |
| Buthead | 0.009 | Barbel | 0.01 | Ruffe | 0.001 | Pike | 0.0009 |
| Stone loach | 0.004 | Stickleback | 0.01 | Pike | 0.004 | Stickleback | 0.0009 |
| Barbel | 0.0003 | Stone foach | 0.007 | Stone loach | 0.001 | Bullhead | 0.0004 |
| Hybrids | 0.0003 | Pike | 0.004 | Tench | 0.001 | Tench | 0.00 |
| Pke | 0.00 | Tench | 0.00 | Hybrids | 0.00 | Stone loach | 0.00 |
| Tench | 0.00 | Hybrids | 0.00 | Barbel | 0.00 | Barbel | 0.00 |

Table 5: Comparison of mean densities ( $\mathrm{nm}^{-2}$ ) for all species 1991-1995

| Year | $n$ | Mean <br> Density | C.V. |
| :--- | :---: | :---: | :---: |
| $1991^{*}$ | 36 | 3.07 | - |
| $1991^{\star \wedge}$ | 13 | 4.07 | - |
| $1992^{\wedge \wedge}$ | 37 | 4.38 | 137.18 |
| 1993 | 42 | 3.52 | 107.50 |
| 1994 | 42 | 13.03 | 148.29 |
| 1995 | 42 | 16.05 | 98.59 |

* Densities calculated in 1992 survey
^ Sites corresponding to those in 1992 and 1993 surveys
A^ Recalculated values
$n=$ number of sites sampled

Table 7: Comparison of mean CPUE 1991-1995

| Year | n | Mean CPUE | C.V. |
| :---: | :---: | :---: | :---: |
| $1991^{*}$ | 36 | 470.40 | 114.40 |
| 199 *^^ $^{*}$ | 13 | 607.20 | 97.50 |
| 1992 | 37 | 861.92 | 107.52 |
| 1993 | 42 | 355.57 | 104.60 |
| 1994 | 42 | 1095.95 | 124.86 |
| 1995 | 42 | 1426.81 | 102.21 |

[^0]Species ranked in descending order of \% frequency
Table 6 and relative importance value 1991-1995

| \%Frequency |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | n=42 | 1994 | $\mathrm{n}=42$ | 1993 | $n=42$ | 1992 | $\mathrm{n}=37^{\circ}$ | 1991 | $\mathrm{n}=13^{*}$ |
| Roach | 65.72 | Roach | 72.42 | Roach | 45.96 | Roach | 67.81 | Roach | 72.36 |
| Gudgeon | 17.19 | Gudgeon | 18.09 | Gudgeon | 27.31 | Bleak | 11.79 | Gudgeon | 10.83 |
| Chub | 7.76 | Chub | 2.53 | Perch | 7.75 | Gudgeon | 10.75 | Bleak | 6.70 |
| Bleak | 4.22 | Bleak | 2.36 | Chub | 7.39 | Chub | 4.06 | Dace | 3.93 |
| Perch | 1.56 | Dace | 1.82 | Dace | 6.38 | Perch | 2.58 | Chub | 3.51 |
| Bream | 1.06 | Perch | 1.43 | Bleak | 1.94 | Dace | 1.78 | Perch | 1.08 |
| Dace | 0.95 | Minnow | 0.45 | Mintow | 1.37 | Hybrids | 0.82 | Bream | 1.01 |
| Rufie | 0.85 | Rutie | 0.38 | Bullhead | 0.66 | Bream | 0.19 | Stickleback | 0.44 |
| Minnow | 0.48 | Bullhead | 0.20 | Bream | 0.54 | Minnow | 0.16 | Bullhead | 0.10 |
| Stickleback | 0.08 | Bream | 0.11 | Rutfe | 0.27 | Stickleback | 0.02 | Ruffe | 0.03 |
| Bullhead | 0.06 | 8arbel | 0.08 | Stickleback | 0.25 | Pike | 0.02 | Barbel | 0.01 |
| Stoneloach | 0.03 | Stickleback | 0.06 | Pike | 0.12 | Ruffe | 0.01 | Hybrids | 0.00 |
| Barbel | 0.02 | Stone loach | 0.04 | Tench | 0.03 | Bullhead | 0.01 | Minnow | 0.00 |
| Hybrids | 0.02 | Pike | 0.03 | Stone loach | 0.03 | Stone loach | 0.00 | Pike | 0.00 |
| Tench | 0.00 | Tench | 0.00 | Barbel | 0.00 | Barbel | 0.00 | Stone loach | 0.00 |
| Pike | 0.00 | Hybrids | 0.00 | Hybrids | 0.00 | Tench | 0.00 | Tench | 0.00 |
| Relative Importance Value |  |  |  |  |  |  |  |  |  |
| 1995 | $\mathrm{n}=42$ | 1994 | $\mathrm{n}=42$ | 1993 | $\mathrm{n}=42$ | 1992 | $\mathrm{n}=37^{\circ}$ | 1991 | $\mathrm{n}=13^{*}$ |
| Roach | 165.73 | Roach | 162.90 | Roach | 143.58 | Roach | 157.00 | Roach | 172.36 |
| Gudgeon | 112.43 | Gudgeon | 106.19 | Gudgeon | 113.02 | Bleak | 95.57 | Perch | 101.08 |
| Perch | 101.56 | Perch | 82.38 | Perch | 100.61 | Perch | 94.47 | Dace | 88.54 |
| Chub | 91.10 | Chub | 76.34 | Bleak | 73.37 | Gudgeon | 83.72 | Gudgeon | 87.76 |
| Bleak | 87.56 | Bleak | 69.02 | Chub | 69.30 | Chub | 68.93 | Bleak | 68.24 |
| Bream | 51.07 | Bullhead | 59.73 | Dace | 46.86 | Dace | 39.62 | Chub | 65.05 |
| Minnow | 50.49 | Dace | 49.44 | Buthead | 45.90 | Hybrids | 17.04 | Stickleback | 38.90 |
| Dace | 43.81 | Minnow | 36.17 | Minnow | 34.70 | Bream | 11.00 | Bream | 24.09 |
| Ruffe | 41.32 | Ruffe | 36.10 | Pike | 28.69 | Pike | 10.83 | Bullhead | 23.18 |
| Bulhead | 31.01 | Pike | 21.46 | Stickleback | 24.08 | Minnow | 8.27 | Ruffe | 7.72 |
| Stickleback | 19.13 | Bream | 19.16 | Bream | 17.21 | Stickleback | 8.13 | Bamel | 7.70 |
| Stoneloach | 16.69 | Barbel | 19.12 | Ruffe | 9.79 | Bullhead | 8.12 | Hybrids | 0.00 |
| Barbel | 2.38 | Stickleback | 16.73 | Tench | 4.79 | Ruffe | 5.42 | Minnow | 0.00 |
| Hybrids | 2.38 | Stone loach | 14.32 | Stone loach | 4.79 | Barbel | 0.00 | Pike | 0.00 |
| Pike | 0.00 | Hybrids | 0.00 | Barbel | 0.00 | Stone loach | 0.00 | Stone loach | 0.00 |
| Tench | 0.00 | Tench | 0.00 | Hybrids | 0.00 | Tench | 0.00 | Tench | 0.00 |

[^1]n=number of sub-sites sampled
Table 9

Table 10


- data not included in previous reports

| Species | $r$ | $p$ |
| :---: | :---: | :---: |
| Fish density v macrophyte cover |  |  |
| Bream $0+$ | 0.197 | < 0.05 |
| Gudgeon 2+ | 0.203 | < 0.05 |
| Perch $0+$ | 0.344 | < 0.001 |
| Perch 1+ | 0.311 | < 0.001 |
| Pike $0+$ | 0.373 | < 0.001 |
| Pike 1+ | 0.317 | < 0.001 |
| Roach 0+ | 0.449 | < 0.001 |
| Roach 1+ | 0.298 | < 0.001 |
| Roach 2+ | 0.372 | $<0.001$ |
| Bullhead | 0.457 | < 0.001 |
| Ruffe | 0.273 | < 0.01 |
| Stone loach | 0.316 | $<0.001$ |
| Fish density v \% clay |  |  |
| Chub 0+ | -0.246 | $<0.01$ |
| Perch 1+ | -0.203 | < 0.05 |
| Fish density y \% silt |  |  |
| Perch $0+$ | 0.185 | $<0.05$ |
| Stone loach | 0.206 | $<0.05$ |
| Fish density $\mathbf{v} \%$ sand |  |  |
| Chub 0+ | 0.273 | $<0.01$ |
| Dace 0+ | 0.267 | $<0.01$ |
| Barbel | 0.232 | $<0.05$ |
| Minnow | 0.363 | $<0.001$ |
| Stickleback | 0.184 | < 0.05 |
| Fish density v \% gravel |  |  |
| Bleak 1+ | 0.237 | < 0.001 |
| Chub 1+ | 0.313 | < 0.001 |
| Fish density v depth |  |  |
| Bleak $0+$ | -0.222 | $<0.05$ |
| Chub 0+ | -0.235 | $<0.05$ |
| Gudgeon 1+ | -0.229 | < 0.05 |
| Perch 0+ | 0.252 | $<0.01$ |
| Perch 2+ | 0.194 | < 0.05 |
| Fish density v gradient |  |  |
| Gudgeon 2+ | 0.194 | < 0.05 |
| Perch $0^{+}$ | 0.249 | $<0.01$ |


| $p>0.05$ | Not significant |
| :--- | :--- |
| $p<0.05$ | Significant |
| $p<0.01$ | Highly significant |
| $p<0.001$ | Very highly significant |

Estimated physical habitat preferences of 6 major fish species Table 19 (adapted from Bullock et al. 1991)

| Modal value | Velocity (m/s) | Depth (m) |
| :---: | :---: | :---: |
| Species |  |  |
| Bream |  |  |
| Spawning | 0.00-0.10 | 0.50-1.00 |
| Fry | 0.00-0.05 | 0.05-0.50 |
| Juveniles | 0.00-0.10 | 0.50-3.00 |
| Adults | 0.00-0.10 | 1.70-3.00 |
| Chub |  |  |
| Spawning | 0.25-0.90 | 0.40-1.70 |
| Fry | 0.05-0.30 | 0.50-0.90 |
| Juveniles | 0.30-0.70 | 0.50-1.60 |
| Adults | 0.20-0.60 | 0.50-1.60 |
| Dace |  |  |
| Spawning | 0.55-1.00 | 0.20-0.80 |
| Fry | 0.05-0.25 | 0.10-0.30 |
| Juveniles | 0.15-0.35 | 0.30-0.70 |
| Adults | 0.20-0.70 | 0.50-1.00 |
| Perch |  |  |
| Spawning | 0.00-0.30 | 0.30-1.50 |
| Fry | 0.00-0.10 | 0.10-0.50 |
| Juveniles | 0.00-0.30 | 0.20-0.80 |
| Aduits | 0.00-0.40 | 0.30-2.50 |
| Pike |  |  |
| Spawning | 0.00-0.10 | 0.20-0.80 |
| Fry | 0.00-0.10 | 0.20-0.90 |
| Juveniles | 0.00-0.20 | 0.10-0.70 |
| Adults | 0.00-0.20 | 0.40-2.90 |
| Roach |  |  |
| Spawning | 0.40-0.80 | 0.30-3.00 |
| Fry | 0.00-0.20 | 0.25 |
| Juveniles | 0.00-0.40 | 1.00-3.00 |
| Adults | 0.00-0.40 | 1.00-3.00 |

Average depth profiles for shallow and deep sites (1995) Figure 2

Distance from bank (m)


Average temperature for sites 1-14 (1995)
Figure 3



1+ Juveniles


0+ Juveniles




Total
國 Dace
biomass $=20.38 \mathrm{gm}-2$
Gudgeon
國 Perch
$\square$ Roach


Bleak
Bream
$\square$ Chub
A Dace Total
图 Gudgeon $\underset{\text { biomass }}{\text { bion }}=\mathbf{2 5 . 6 1} \mathrm{gm}-2$
Ey Pench
R Roach



## River Thames O Group Fish Gut Contents Study 1995

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## EXECUTIVE SUMMARY

This project on the diets of 0 group fish in the River Thames at Abingdon had four main objectives:
a) Assessment of the between-species, between-habitats and between-season changes in fish diets,
b) Comparison of the results from the River Thames with those of similar studies by IFE in the River Great Ouse,
c) Assessment of the likely effects of changes in the flow regime of the River Thames on the food sources of 0 group fish,
d) Recommendation for future studies that would increase the robustness of the results of the River Thames studies.

The diets were analyzed for five fish species: roach, gudgeon, chub, silver bream, perch. Although differences were observed between the diets of single species caught in different plant habitats, these were small compared with those observed between species. In general, the diet of each species changed from small prey (mostly rotifers) to larger items (mostly microcrustacea and insect larvae) as the fish grew in size. However, larger 0 group perch and gudgeon contained many copepods, whereas the other fish species contained more cladoceran taxa. The most striking difference between species was that, in July, the roach switched from an invertebrate diet to one dominated by detritus (aufwuchs) that accumulates on the underwater surfaces of plants.

The results from the River Thames show a close parallel with those of the same fish species in the River Great Ouse. This similarity adds weight to the reliability of the Thames results, even though they are based on small numbers of fish per sample.
However, the Great Ouse data showed marked differences in the numbers and types of prey eaten in different years, which reflected changes in the abundance of these taxa in the river.

The Thames and Great Ouse studies highlight the importance of aquatic plants as feeding and refuge areas for young fish. Any major reduction in their areas, as could occur through large changes in the flow regime, would be detrimental to the growth of the young fish and to the numbers surviving to the adult stage.

Further studies to increase the reliability of the results from the River Thames are:
a) examination of more of the fish caught in the 1995 study,
b) collection of additional samples to determine the extent of year-to-year fluctuations.

For the greatest understanding of fish-prey dynamics, such studies should accompanied by assessments of prey availability (planktonic and non-planktonic invertebrates).

## 6. CONCLUSIONS

a) As they grew in size the 0 group fish progressed from a diet of small prey items to one that included larger prey taxa. In July, the roach switched from an invertebrate diet to one predominantly of detritus (aufwuchs). All the other species continued to feed on invertebrates.
b) Between-habitat differences in the diets of the same fish species could be explained by differences in the sizes of fish caught, or by the presence/absence of planktonic Cladocera (Bosmina). However, the results may have been affected by the ability of the larger 0 group fish to move between habitats.
c) Between-species differences in diets greatly exceeded those observed for the same species in different habitats. These partly reflected where the fish were feeding in the water column, e.g. gudgeon fed on the river bed, whereas roach fed much nearer the water surface. Moreover, as each species spawned and hatched at a different time, their 0 group fish were at different stages of development on any one sampling occasion.
d) The results closely parallel those obtained from IFE's studies in the Great Ouse. Both show the importance of aquatic plants as feeding and refuge areas for 0 group fish.

Table 7. 0 group roach: Chi-square analyses showing the statistical significance of differences in the contributions (numbers of animals) by each prey taxon. Ph $=$ Phragmites, $\mathrm{N}=$ Nuphar, $\mathrm{S}=$ Salix, $\mathrm{A}=$ Acorus, Mix $=$ Mixed reeds. Symbols refer to individual chi-square vales: $=\langle 3.0,+$ and o $3.0-15.0,++$ and $00>15.0 ;+$ values indicate higher than expected contribution, o values indicate lower than expected contribution; e/f = electrofishing.

| Ph | $N$ <br> seine | $N$ <br> $e / f$ | $S$ | $A$ | Mix |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 6 June $($ chi-square $=$ | $18.01,6 \mathrm{df})$ |
| :--- | :---: |
| Rotifers | $=$ |
| Chydoridae | $=$ |
| Other taxa | 0 |
|  |  |
| 20 June (chi-square $=$ | $555.1,30 \mathrm{df})$ |


| Rotifers | 00 | + | $=$ | $=$ | 0 | $=$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Chydoridae | $=$ | 0 | 0 | $=$ | $=$ | ++ |
| Bosmina | $=$ | $=$ | ++ | $=$ | $=$ | 00 |
| Polyphemus | + | 0 | 00 | $=$ | + | 00 |
| Other Cladocera | $=$ | $=$ | $=$ | $=$ | $=$ | $=$ |
| Chiron. l. | $=$ | $=$ | $=$ | + | $=$ | $=$ |
| Other taxa | $=$ | $=$ | $=$ | 0 | + | $=$ |

4 July (chi-square $=330.2,15 \mathrm{df}$ )

| Rotifers | 0 | $=$ | + | $=$ |
| :--- | :--- | :--- | :--- | :--- |
| Chydoridae | ++ | 0 | 00 | + |
| Polyphemus | 0 | + | ++ | 0 |
| Other Cladocera | 0 | $=$ | 0 | ++ |
| Chiron. I. | $=$ | $=$ | 0 | ++ |
| Other taxa | $=$ | $=$ | $=$ | + |

25 July (chi-square $=374.0,12 \mathrm{df}$ )

| Chydoridae | $=$ | 00 | $=$ | $=$ | + |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Other Ciadocera | 00 | + | $=$ | 00 | 00 |
| Other taxa | + | 0 | $=$ | $=$ | ++ |
| Aufwuchs | + | 00 | $=$ | + | $=$ |

$$
22 \text { August (chi-square }=14.99,3 \mathrm{df} \text { ) }
$$

| All taxa | 0 | $=$ | $=$ | + |
| :--- | :--- | :--- | :--- | :--- |
| Aufwuchs | $=$ | $=$ | $=$ | $=$ |

19 September (chi-square $=17.85,4 \mathrm{df}$ )

| All taxa | $=$ | + | $=$ | $=$ |
| :--- | :--- | :--- | :--- | :--- |
| Aufurchs | $=$ | $=$ | $=$ | $=$ |
|  |  | $=$ | $=$ |  |

On 25 July comparisons between the same two sites revealed small but statistically significant differences (chi-square $=18.98,6 \mathrm{df}$ ), with more chydorids and fewer copepods being found in fish from the Salix site, and the reverse at the Nuphar site. Nevertheless, the diets were very similar, with considerable overlap.

### 4.2.4 Chub

Table 9. 0 group chub: Chi-square analyses showing the statistical significance of differences in the contributions (number of animals) by each prey taxon. Key to symbols given in Table 7.

| Habitat | Salix | Mixed | Nuphar |  |
| :---: | :---: | :---: | :---: | :---: |
| 20 June (chi-square $=67.22,6 \mathrm{df}$ ) |  |  |  |  |
| Rotifers | + | 0 | = |  |
| Chydoridae | 00 | + | $+$ |  |
| Other Cladocera | = | = | = |  |
| Other taxa | $=$ | = | $=$ |  |
|  | Salix | Mixed | Nuphar | Acorus |
| 4 July (chi-square $=1229.34,15 \mathrm{dt}$ ) |  |  |  |  |
| Rotifers | + | 00 | oo | + |
| Chydoridae | $=$ | $+$ | $\bigcirc$ | 00 |
| Polyphemus | 00 | 00 | + | $\bigcirc 0$ |
| Other Cladocera | = | + | = | 0 |
| Chiron. 1. | = | + | = | $\bigcirc$ |
| Other taxa | $=$ | + | + | 00 |
|  | Mixed | $\mathrm{d} / \mathrm{s} A$ | s Nuphar |  |
| 25 July (chi-square $=85.02,10 \mathrm{df}$ ) |  |  |  |  |
| Chydoridae | 0 | $=$ | + |  |
| Other Cladocera | $=$ | ++ | 0 |  |
| Copepoda | $=$ | = | $=$ |  |
| Chiron. 1. | ++ | = | o |  |
| Terrestrial prey | = | = | = |  |
| Other taxa | $=$ | = | = |  |

In the 4 July sample, the number of Polyphemus eaten increased with the length of chub (Figure 2), the data giving the following linear relationship:

$$
\text { Number of Polyphemus }=5.113(\mathrm{SLmm})-52.57 \quad \mathrm{r}^{2}=0.68
$$

### 4.3 Comparisons between fish species

The following tables (11A-11Q)show the percentage number of prey items of various taxa that were eaten by different fish species caught at the same sites on the same sampling occasions.

Table 11. Comparison of prey taken by different fish species caught in the same habitat. Values are the percentage numbers of animals in each prey taxon, the highest contributions ( $\$ 20 \%$ ) being indicated in bold type. Aufwuchs values (index scale 0-3) are the means for each sample.
$11(A) 6$ June: Nuphar (combined seine and electro-fished samples)
Roach Gudgeon Perch

| No. of fish | 29 | 10 | 20 |
| :---: | :---: | :---: | :---: |
| No. of prey items | 999 | 132 | 273 |
| Rotifers | 93.50 |  |  |
| Chydoridae | 4.90 | 90.91 | 31.50 |
| Polyphemus |  | 6.06 | 10.62 |
| Other Cladocera | 1.17 |  | 3.66 |
| Copepoda |  | 2.27 | 49.45 |
| Chiron. 1. | 0.43 |  | 4.03 |
| Other taxa |  | 0.76 | 0.73 |

11(B) 20 June: Nuphar

| No. of fish | Roach <br> 40 | Gudgeon $18$ | Chub 13 |
| :---: | :---: | :---: | :---: |
| No. of prey items | 892 | 214 | 235 |
| Rotifers | 56.05 | 3.27 | 73.62 |
| Chydoridae | 21.52 | 76.17 | 17.02 |
| Polyphemus | 5.38 | 0.47 | 1.70 |
| Other Cladocera | 14.01 | 0.47 | 5.11 |
| Copepoda | 1.23 | 13.55 | 0.43 |
| Chiron. I. | 0.67 | 2.34 | 1.70 |
| Other taxa | 1.12 | 3.74 | 0.43 |

11(C) 20 June: Salix


| II(D) 20 June:Mixed reeds <br> Roach | Chub |  |
| :--- | :---: | :---: |
| No. of fish | 20 | 12 |
| No. of prey items | 404 | 121 |
| Rotifers | $\mathbf{4 5 . 5 4}$ | $\mathbf{6 2 . 8 1}$ |
| Chydoridae | $\mathbf{4 7 . 5 2}$ | $\mathbf{3 1 . 4 1}$ |
| Polyphemus | 0.50 |  |
| Other Cladocera | 1.24 | 3.31 |
| Copepoda | 1.24 | 1.65 |
| Chiron. I. | 1.24 |  |
| Other taxa | 2.72 | 0.83 |

11(E) 20 June: Acorus

|  | Roach | Silver Bream |
| :--- | :---: | :---: |
| No. of fish | 20 | 19 |
| No. of prey items | 308 | 226 |


| Rotifers | 36.69 | $\mathbf{9 8 . 2 3}$ |
| :--- | ---: | ---: |
| Chydoridae | 29.87 | 1.33 |
| Polyphemus | 12.66 |  |
| Other Cladocera | 14.94 |  |
| Copepoda. | 2.60 | 0.44 |
| Chiron. l. | 0.32 |  |
| Other taxa | 2.92 |  |
|  |  |  |

11(F) 4 July: Nuphar

|  | Roach | Chub | Perch |
| :---: | :---: | :---: | :---: |
| No. of fish | 20 | 20 | 20 |
| No. of prey items | 564 | 564 | 413 |
| Rotifers | 2.13 | 31.38 |  |
| Chydoridae | 22.52 | 8.16 | 11.38 |
| Polyphemus | 65.60 | 50.18 | 10.41 |
| Other Cladocera | 6.91 | 1.77 | 61.50 |
| Copepoda | 0.18 | 6.56 | 12.59 |
| Chiron. I. | 2.48 | 1.60 | 2.91 |
| Other taxa | 0.18 | 0.35 | 1.21 |

11(G) 4 July: Phragmites
Roach Silver Bream
No. of fish 60
No. of prey items $131 \quad 361$

| Rotifers |  | $\mathbf{5 7 . 8 9}$ |
| :--- | ---: | ---: |
| Chydoridae | $\mathbf{5 7 . 2 5}$ | 29.09 |
| Polyphemus | 35.11 | 10.80 |
| Other Cladocera | 3.05 | 2.21 |
| Chiron. l. | 4.58 |  |

11(H) 4 July: Mixed reeds

|  | Roach | Chub |
| :--- | :---: | :---: |
| No. of fish | 20 | 15 |
| No. of prey items | 235 | 180 |


| Rotifers | 3.40 | 13.33 |  |
| :--- | :---: | :---: | :--- |
| Chydoridae | $\mathbf{4 5 . 5 3}$ | 65.00 |  |
| Polyphemus | 20.85 | 6.67 |  |
| Other Cladocera | 15.74 a | 2.78 | $\mathrm{a}=$ Sida |
| Copepoda | 0.85 | 1.11 |  |
| Chiron. I. | 10.21 | 3.33 |  |
| Other taxa | 3.40 | 7.78 |  |
|  |  |  |  |



|  | Chub | Silver Bream |
| :--- | :---: | :---: |
| No. of fish | 15 | 15 |
| No. of prey items | 612 | 107 |


| Rotifers | 96.73 | 18.69 |
| :--- | :---: | :---: |
| Chydoridae | 3.10 | 50.46 |
| Polyphemus |  | 30.84 |
| Chiron. l. | 0.16 | . |

## 11(J) 4 July: Salix



11(K) 25 July: Nuphar

| No. of fish | Roach 37 | Gudgeon 15 | $\begin{aligned} & \text { Chub } \\ & 15 \end{aligned}$ | Perch 19 |
| :---: | :---: | :---: | :---: | :---: |
| No. of prey items | 322 | 186 | 170 | 681 |
| Chydoridae | 6.83 | 33.33 | 44.12 | 4.11 |
| Sida | 2.80 |  | 1.76 | 16.89 |
| Daphnidae |  | 5.91 | 0.59 | 47.13 |
| Other Cladocera | 86.96a |  | 9.41 | 3.52 |
| Copepoda |  | 34.41 | 11.18 | 20.70 |
| Chiron. I. | 1.86 | 7.52 | 0.59 | 2.06 |
| Large crustacea |  |  |  | 2.64 |
| Other taxa | 1.55 | 18.82b | 32.35 c | 2.94 |

Aufwuchs (index) $\mathbf{2 . 5 4}$
$\mathrm{a}=$ Bosmina $\quad \mathrm{b}=$ Ostracoda $\mathrm{c}=$ Terrestrial insects

11(L) 25 July: Phragmites

| - | Roach | Silver Bream |  |
| :---: | :---: | :---: | :---: |
| No. of fish | 20 | 20 |  |
| No. of prey items | 24 | 303 |  |
| Rotifers | 33.00 |  |  |
| Chydoridae | 20.83 | 18.81 |  |
| Other Cladocera | 33.33a | 24.09a | $\mathrm{a}=$ mostly Bosmina |
| Copepoda | 20.83 | 11.22 |  |
| Chiron. 1. | 12.50 | 11.55 |  |
| Other taxa | 12.50 | 1.32 |  |
| Aufwuchs (index) | 2.65 | 0.30 |  |

11(M) 25 July: Mixed reeds

|  | Roach | Chub |
| :--- | :---: | :---: |
| No. of fish | 20 | 15 |
| No. of prey items | 36 | 161 |


| Chydoridae | 55.56 | 22.36 |  |
| :---: | :---: | :---: | :---: |
| Other Cladocera | 11.11 | 16.15 | $\mathrm{a}=$ mostly Ostracoda |
| Copepoda |  | 19.25 |  |
| Chiron. 1. | 6.94 | 16.15 | $\mathrm{b}=$ mostly terrestrial |
| Other taxa | 26.94a | 26.09b | insects |
| Aufwuchs (index) | 1.80 |  |  |

11(N) 25 July: Salix

|  | Gudgeon | Chub |  |
| :---: | :---: | :---: | :---: |
| No. of fish | 15 | 5 |  |
| No. of prey items | 121 | 43 |  |
| Chydoridae | 52.89 | 34.88 |  |
| Other Cladocera | 4.13 | 18.60 | $\mathrm{a}=$ mostly Ostracoda \& Naidae |
| Copepoda | 15.70 | 20.93 |  |
| Chiron. 1. | 5.79 | 6.98 | $b=$ mostly Naidae |
| Other taxa | 21.49a | 18.60 b |  |



11(P) 19 September: Nuphar

|  | Roach | Perch | Silver Bream |
| :--- | :---: | :---: | :---: |
| No. of fish | 34 | 14 | 15 |
| No. of prey items | 16 | 557 | 35 |


| Chydoridae | 37.50 | 13.64 | 68.57 |
| :---: | :---: | :---: | :---: |
| Sida | 6.25 | 2.69 |  |
| Other Cladocera |  | 0.18 | 8.57 |
| Copepoda | 6.25 | 66.43 |  |
| Chiron. 1. | 6.25 | 5.03 |  |
| Ostracoda | 31.25 | 2.87 | 22.86 |
| Larger crustacea | 6.25 | 2.87 |  |
| Other taxa | 6.25 | 6.28 |  |
| Aufwuchs (index) | 1.47 |  | 1.00 |



### 4.4 Seasonal variations

Information on the seasonal variation in fish diets is presented as the mean number of each prey type per fish. This was found to be a useful measure in similar studies on the River Great Ouse, in which marked differences in prey intake were observed in different years (Mann et al., 1996).

Table 12. Mean numbers of animals per gut; aufwuchs given as the mean index, scale 0-3.

12(A) Roach: Phragmites

|  | $23 / 5$ | $6 / 6$ | $20 / 6$ | $4 / 7$ | $25 / 7$ | $22 / 8$ | $19 / 9$ |
| :--- | :--- | :--- | :--- | :---: | :--- | :--- | :--- |
| No. of fish | 20 | 17 | 20 | 6 | 20 | 21 | 17 |
| Mean SL (mm) | 6.6 | 8.8 | 14.8 | 20.6 | 28.8 | 37.8 | 38.4 |
|  |  |  |  |  |  |  |  |
| Rotifers | 4.95 | 24.41 | 5.75 |  |  |  |  |
| Chydoridae |  | 1.06 | 6.10 | 12.50 | 0.25 | 0.05 | 0.35 |
| Bosmina |  |  | 3.25 |  |  |  |  |
| Polyphemus |  |  | 6.45 | 7.67 |  |  | 0.06 |
| Other Cladocera <br> Chiron. I. |  | 0.05 | 0.67 | 0.40 |  | 0.06 |  |
| Other taxa | 0.06 | 0.30 | 1.00 |  |  |  |  |
|  |  |  |  | 0.50 | 0.19 | 0.35 |  |
| Aufwuchs (index) |  |  |  |  | 2.65 | 2.19 | 1.88 |

12(B) Roach: Mixed reeds

|  | $23 / 5$ | $6 / 6$ | $20 / 6$ | $4 / 7$ | $25 / 7$ | $22 / 8$ | $19 / 9$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 10 | 20 | 20 | 20 | 20 | 20 | 20 |
| No. of fish | 6.7 | 9.8 | 12.9 | 21.5 | 30.6 | 36.6 | 39.9 |
| Mean SL (mm) |  |  |  |  |  |  |  |
|  | 6.65 | 20.10 | 9.20 | 0.40 |  |  |  |
| Rotifers |  | 1.90 | 9.60 | 5.35 | 2.00 | 0.35 |  |
| Chydoridae |  |  | 0.10 | 0.05 | 0.30 |  |  |
| Bosmina |  |  | 0.10 | 2.45 |  |  |  |
| Polyphemus |  | 0.15 | 1.80 | 0.10 |  |  |  |
| Other Cladocera | 0.10 |  | 0.80 | 1.20 |  | 0.15 |  |
| Chiron. l. |  |  |  | 0.40 | 1.20 | 0.45 | 0.20 |
| Other taxa |  |  |  |  | 1.80 | 1.25 | 1.90 |
| Aufwuchs (index) |  |  |  |  |  |  |  |


| 12(C) Roach: Nuphar (combined | seine and electro-fished | samples) |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :--- | :--- | :--- | :--- |
|  | $23 / 5$ | $6 / 6$ | $20 / 6$ | $4 / 7$ | $25 / 7$ | $22 / 8$ | $19 / 9$ |
| No. of fish | 20 | 29 | 40 | 20 | 37 | 27 | 34 |
| Mean SL (mm) | 6.9 | 10.6 | 15.3 | 21.8 | 30.1 | 38.4 | 40.1 |
|  |  |  |  |  |  |  |  |
| Rotifers | 15.75 | 30.28 | 12.50 | 0.60 |  |  |  |
| Chydoridae |  | 1.59 | 4.80 | 6.35 | 0.62 | 0.37 | 0.18 |
| Bosmina |  | 0.31 | 3.10 | 0.25 | 1.84 | 0.04 |  |
| Polyphemus |  |  | 1.20 | 18.50 |  |  |  |
| Other Cladocera |  | 0.07 | 0.03 | 1.70 | 0.27 |  | 0.03 |
| Chiron. 1. | 0.14 | 0.15 | 0.70 | 0.24 | 0.07 | 0.03 |  |
| Other taxa |  | 0.53 | 0.10 | 0.35 | 0.22 | 0.24 |  |
|  |  |  |  |  |  |  |  |
| Aufwuchs (index) |  |  |  |  | 2.68 | 2.78 | 1.47 |

12(D) Gudgeon: All sites combined

|  | $23 / 5$ | $6 / 6$ | $20 / 6$ | $4 / 7$ | $25 / 7$ | $22 / 8$ | $19 / 9$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| No. of fish | 0 | 29 | 39 | 9 | 30 | 0 | 0 |
| Mean SL (mm) |  | 11.4 | 12.5 | 19.9 | 16.8 |  |  |

Rotifers
Chydoridae
Polyphemus
Other Cladocera
Copepoda
Chiron. l.
Ostracoda
Other taxa

|  | 0.23 |  |  |
| ---: | ---: | ---: | ---: |
| 10.21 | 6.90 | 13.33 | 4.20 |
| 0.59 | 0.03 | 1.11 |  |
|  | 0.03 | 0.44 | 0.53 |
| 0.48 | 1.21 | 1.00 | 2.77 |
| 0.03 | 0.33 | 0.67 | 0.70 |
|  | 0.21 | 0.33 | 1.07 |
| 0.03 | 0.05 | 0.23 | 0.97 |

12(E) Chub: All sites combined

|  | $23 / 5$ | $6 / 6$ | $20 / 6$ | $4 / 7$ | $25 / 7$ | $22 / 8$ | $19 / 9$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| No. of fish | 0 | 0 | 44 | 70 | 50 | 15 | 0 |
| Mean SL (mm) |  |  | 10.5 | 11.9 | 15.1 | 12.9 |  |
|  |  |  |  |  |  |  |  |
| Rotifers |  |  | 15.48 | 14.97 | 0.02 | 0.47 |  |
| Chydoridae |  |  | 2.48 | 3.10 | 3.44 | 2.80 |  |
| Polyphemus |  |  | 0.43 | 4.67 |  |  |  |
| Other Cladocera |  |  | 0.61 | 0.26 | 1.84 a | 7.93 b |  |
| Copepoda |  |  | 0.14 | 0.70 | 1.62 | 1.87 |  |
| Chiron. I |  | 0.09 | 0.24 | 0.70 | 0.87 |  |  |
| Other taxa |  |  |  |  |  |  |  |

$\mathrm{a}=$ mostly Bosmina, $\mathrm{b}=$ mostly Scapholeberis, $\mathrm{c}=$ terrestrial insects, $\mathrm{d}=$ terrestrial insects \& Ostracoda.


| 12(G) Perch: All sites combined |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 23/5 | 6/6 | 20/6 | 4/7 | 25/7 | 22/8 | -19/9 |
| No. of fish | 0 | 20 | 4 | 20 | 19 | 15 | 29 |
| Mean SL (mm) |  | 13.4 | 17.3 | 28.5 | 44.2 | 54.9 | 60.2 |
| Chydoridae |  | 4.30 | 0.75 | 2.35 | 1.47 | 4.67 | 3.28 |
| Polyphemus |  | 1.45 |  | 2.15 |  | \% |  |
| Daphnidae |  | 0.15 |  |  | 16.89 | 1.73 | 0.03 |
| Sida- |  |  |  | 12.70 | 6.05 | 4.21 | 0.55 |
| Other Cladocera |  | 0.35 | 2.50 |  | 1.26 | 0.40 |  |
| Copepoda |  | 6.75 | 26.75 | 2.60 | 7.42 | 28.53 | 36.41 |
| Chiron. 1. |  | 0.55 | 0.25 | 0.60 | 0.74 | 7.47 | 4.17 |
| Ostracoda |  | 0.10 |  |  | 0.05 | 1.87 | 0.86 |
| Larger crustacea |  |  |  |  | 0.95 | 2.33 | 1.41 |
| Other taxa |  | 0.15 |  | 0.25 | 1.00 | 1.80 | 1.21 |

APPENDIX 3.1 A full list of the samples held in the macro-invertebrate data-base together with their identifiers and principal environmental descriptors.

| Reach ID | Reach name | $\begin{aligned} & \text { Site } \\ & \text { D } \end{aligned}$ | Site name | Subsidiary ID | Subsidiary code | Season D | Sample <br> ID | Sample date | Easting | Northing | Altitude | Slope | $\begin{aligned} & \text { Disch } \\ & \text { cat } \end{aligned}$ | Dist from source | Water width | Average depth | \%B/c | \%P/g | \%Sa | \%S/e |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH10 | St. Johnns | 38 | A417 Buscot | WA/NRA/E | PUTR.0088 | 3 | 300 | 22/10/80 | 422500 | 198300 | 71 | 0.5 | - 6 | . 35 |  |  |  |  |  |  |
| TH10 | St. Johns | 38 | A417 Buscot | WA/NRA/E | PUTR. 0088 | 2 | 3500 | 10/07/87 | 422500 | 198300 | 71 | 0.5 | 6 | 35 | 20 | 160 | 0 | 70 | 0 | 30 |
| TH10 | St. Johns | 127 | Buscot | LR | BUS86 | 8 | 2 |  |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 127 | Buscot | LR | BUS78 | 8 | 3 |  |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 127 | Buscot | LR | BUS77 | 8 | 1 |  |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 127 | Buscot | LR | 40792 | 2 | 2 | 24/08/92 |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 127 | Buscot | LR | 28492 | 2 | 1 | 24/06/92 |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St Johns | 127 | Buscot | LR | 15192 | 1 | 1 | 08/05/92 |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 143 | Malthouse | FBA | 69110102 | 2 | 7002 | 26/07/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johns | 143 | Malthouse | FBA | 69110203 | 3 | 7006 | 17/10/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johns | 143 | Malthouse | FBA | 69110101 | 1 | 7000 | 09/05/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johins | 143 | Malthouse | FBA | 69110202 | 2 | 7004 | 26/07/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johns | 143. | Malthouse | FBA | 69110201 | 1 | 7001 | 09/05/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johns | 143 | Malthouse | FBA | 69110103 | 3 | 7005 | 17/10/84 | 422500 | 198400 | 71 | 0.5 | 6 | 35 | 27 | 161 | 20 | 52 | 12 | 17 |
| TH10 | St. Johns | 31 | St. Johns | FBA | FBA77 18 | 2 | 1 | 15/08/77 |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 31 | St. Johns | FBA | FBA77 18 | 2 | 18 | 03/08/77 |  |  |  | 0.5 | 6 |  |  |  |  |  |  |  |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PU'TR.0107 | 2 | 7800 | 05/09/90 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRAE | PU'TR. 0107 | 2 | 10400 | 13/07/92 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUIR. 0107 | 1 | 8700 | 08/04/91 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WANRA/E | PU'TR. 0107 | 2 | 9400 | 05/08/91 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR.0107 | 2 | 7200 | 13/07/90 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 3 | 9800 | 14/11/91 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 2 | 12100 | 09/08/94 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 1. | 2100 | 03/02/87 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 10 |  |  |  |  |  |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 1 | 11300 | 29/04/93 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 1 | 13800 | 04/05/95 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WANRA/E | PUTR.0107 | 3 | 16500 | 14/11/95 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR.0107 | 3 | 6400 | 24/04/90 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1. | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WANRA/E | PUTR. 0107 | 1 | 5400 | 01/06/89 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 2 | 4900 | 01/08/88 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 75 | 1 | 54 | 10 | 35 |
| TH10 | St. Johns | 39 | Water Intake Buscot | WA/NRA/E | PUTR. 0107 | 3. | 10800 | 07/10/92 | 422900 | 198100 | 72 | 0.5 | 6 | 36 | 15 | 150 | 0 | 50 | 15 | 35 |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M7 | 2 | 2197 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B7 | 2 | 3197 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B2 | 2 | 3192 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M8 | 2. | 2198 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M6 | 2 | 2196 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |


| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M9 | 2. | 2199 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| THII | Buscot | 37 | Buscot | FBA | FBA77 _19B10 | 2 | 3190 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B1 | 2 | 3191 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B3 | 2. | 3193 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B4 | 2 | 3194 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| THII | Buscot | 37 | Buscot | FBA | FBA77 19B9 | 2 | 3199 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| THII | Buscot | 37 | Buscot | FBA | FBA77 19B6 | 2 | 3196 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B8 | 2 | 3198 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| THII | Buscot | 37 | Buscot | FBA | FBA77 19M5 | 2 | 2195 | 15/08/77 |  |  |  | 0.33 | 6. |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M1 | 2 | 2191 | 15/08/77 |  |  |  | 0.33 . | 6. |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19B5 | 2 | 3195 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V4 | 2 | 1194 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 30 | Buscot | FBA | FBA77 , 17 | 2 | 17 | 03/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V10 | 2 | 1190 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V1 | 2 | 1191 | 15/08/77 |  |  |  | 0.33 | 6. |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77_19M4 | 2 | 2194 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V3 | 2 | 1193 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 30 | Buscot | FBA | FBA77 17 | 2 | 2 | 16/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V5 | 2 | 1195 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V6 | 2 | 1196 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V7 | 2 | 1197 | 15/08/77 |  |  |  | 0.33 | 6. |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V8 | 2 | 1198 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V9 | 2 | 1199 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M10. | 2 | 2190 | 15/08/77 |  |  |  | 0.33 . | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77_19M2 | 2 | 2192 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19M3 | 2 | 2193 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH11 | Buscot | 37 | Buscot | FBA | FBA77 19V2 | 2 | 1192 | 15/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH12 | Orafton | 29 | Gration | FBA | FBA77 16 | 2. | 3 | 16/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH12 | Grafton | 29 | Gratton | FBA | FBA77 16 | 2 | 16 | 03/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH13 | Radcot | 28 | Radcot | FBA | EBA77. 15 | 2 | 4 | 16/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH13 | Radcot | 28 | Radcot | FBA | FBA77 15 | 2 | 15 | 02/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH14 | Rushey | 40 | 0.5 km below Tadpole Bridg | WA/NRA/E | PUTR9993 | 2 | 14900 | 10/08/95 | 434000 | 200300 | 66 | 0.33 | 6 | 52 | 22 | 160 | 0 | 0 | 0 | 100 |
| TH14 | Rushey | 27 | Rushey | FBA | FBA77 14 | 2 | 5 | 16/08/77 |  |  |  | 0.33 | 6. |  |  |  |  | - | , | 1 |
| TH14 | Rushey | 27 | Rushey | FBA | FBA77 14 | 2 | 14 | 02/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH15 | Shifford | 41 | Newbridge | WANRA/E | PUTR. 0099 | 3 | 15400 | 02/10/95 | 440300 | 201400 | 63 | 0.33 | 6 | 60 | 22 | 160 | 0 | 0 | 0 | 100 |
| THIS | Shiftord | 41 | Newbridge | WANRAE | PUTR.0099 | 1 | 14400 | 17/05/95 | 440300 | 201400 | 63 | 0.33 | 6 | 60 | 22 | 160. | 0 | 0 | 0 | 100 |
| TH15 | Shifford | 41 | Newbridge | WA/NRAE | PUTR.0099 | 2 | 3600 | 10/07/87 | 440300 | 201400 | 63. | 0.33 | 6 | 60 | 221 | 160 | 0 | 0 | 0 | 100 |
| TH15 | Shifford | 36 | Shiflord | FBA | FBA77 20B9 | 2 | 3209 | 22/08/77 |  |  |  | 0.33 | 6 |  |  |  |  |  |  |  |
| TH15 | Shifford | 36 | Shifford | FBA | FBA77 20M8 | 2 | 2208 | 22/08/77 |  |  |  | 0.33 | 61 |  |  |  |  |  |  |  |







| TH23 | Sundford | 49 | Abingdon Weir | WANRA/E | PTHR. 0077 | 1 | 2400 | 12/03/87 | 450400 | 197200 | 50 | 0.45 | 8 | 93 | 40 | 300 | 1 | 50 | 34 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH23 | Sundford | '139 | Near Lock Wood | PA | T11 | 2 | 1 | 28/07/92 | 452600 | 197000 | 50 | 0.45 | 8 | 91 | 50 | 150 | 10 | 60 | 24 | $\cdots$ |
| TH23 | Sandiord | 140 | Radley | PA | T12 | 2 | 1. | 20107/92 | 453800 | 199000 | 53 | 0.45 | 8 | 88 | 40 | 150 | 0 | 45 | 45 | 10 |
| TH23 | Sandford | 124 | Radley College Boathouse | RPS | OSP. 16 | 2. | 5006 | 30/06/92 | 453800 | 198800 | 53 | 0.45 | 8 | 88 | 20. | 200 | 30 | 30 | 35 | . |
| TH23 | Sandford | 50 | Sandford | WANRA/E | PTHR 0099 | 1. | 1900 | 12/06/86 | 453000 | 201200 | 54 | 0.45 | 8 | 86. | 40. | 100 |  |  |  |  |
| TH23 | Sandford | 18 | Sandford | FBA | FBA77 5 | 2 | 14 | 01/09/77 |  |  |  | 0.45 | 8 |  |  |  |  |  |  |  |
| TH23 | Sandford | 18 | Sandford | FBA | FBA77 5 | 2 | 5 | 27/07/77 |  |  |  | 0.45 | 8 |  |  |  |  |  |  |  |
| TH23 | Sandford | 141 | Sandford Reach | PA | T13 | 2 | 1 | 28/07/92 | 453300 | 200300 | 54 | 0.45 | 8 | 87 | 40 | 150 | 5 | 63 | 30 | 2 |
| TH24 | Abingdon | 17 | Abingdon | FBA | FBA77. 4 | 2 | 4 | 27/07/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH24 | Abingdon | 17 | Abingdon | FBA | FBA77. 4 | 2 | 15 | 01/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH24 | Abingdon | 1137 | Culham Reach | PA | T7 | 2 | 1. | 20/07/92 | 450000 | 195500 | 50 | 0.26 | 8 | 96 | 55 | 150 | 10 | 60 | 25 | 5 |
| TH24 | Abingdon | 136 | Sutton Pools | PA | T6 | 2 | 1 | 03/08/92 | 450300 | 194500 | 50 | 0.26 | 8 | 99. | 40 | 150 | 0 | 50. | 25 | 25 |
| TH25 | Cutham | 135 | Clifton Reach | PA | T5 | 2 | 1 | 28/07/92 | 452600 | 194200 | 49 | 0.26 | 8 | 102 | 45 | 150 | 5 | 50 | 401 | 5 |
| TH25 | Cutham | 16 | Culham | FBA | FBA77 3 | 2. | 3 | 26/07/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH25 | Culham | 16 | Culham | FBA | FBA77_3 | 2 | 16 | 01/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH25 | Culham | 134 | Long Wittenham | PA | T4 | 2 | 1 | 03/08/92 | 454000 | 193700 | 49 | 0.26 | 8 | 103 | 35 | 150 | 0 | 5 | 75 | 20 |
| TH25 | Culham | 52 | Sutton Bridge, Culham | WANRA/E | PTHR 0105 | 2 | 3200 | 07/07/87 | 4.50900 | 194900 | 49 | 0.26 | 8 | 100. | 5 | 8 | 0. | 0 | 0 | 100 |
| TH25 | Culham | 52 | Sutton Bridge, Culham | WA/NRA/E | PTHR. 0105 | 1 | 13300 | 10/04/95 | 450900 | 194900 | 49 | 0.26 | 8 | 100 | 38 | 190 | 5 | 40 | 40 | 15 |
| TH25 | Culham | 52 | Sutton Bridge, Culham | WA/NRA/E | PTHR. 0105 | 3 | 16400 | 02/11/95 | 450900 | 194900 | 49 | 0.26 | 8 | 100 | 38 | 190 | 2 | 20 | 45 | 33 |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23B4 | 2 | 3234 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifon | FBA | FBA77 23M5 | 2 | 2235 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23M6 | 2 | 2236 | 05/09/77 |  |  |  | 0.26 | 8. |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77. 23M8 | 2 | 2238 | 05/09/77 |  |  |  | 0.26 | 8. |  |  |  |  |  |  |  |
| TH26 | Clitton | 33 | Cliton | FBA | FBA77 23B10 | 2. | 3230 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23B1 | 2. | 3231 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23M4 | 2 | 2234 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77-23B3 | 2 | 3233 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77_23M7 | 2 | 2237 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23B5 | 2 | 3235 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clitton | 33 | Clifton | FBA | FBA77 23B6 | 2 | 3236 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77, 2387 | 2 | 3237 | 05/09777 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clitton | FBA | FBA77, 23B8 | 2 | 3238 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clitton | 33 | Clifton | FBA | FBA77 23B9 | 2 | 3239 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifton | FBA | FBA77 23B2 | 2 | 3232 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 33 | Clifon | FBA | FBA77 23 V 3 | 2 | 1233 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clition | 15 | Clitton | FBA | FBA77. 2 | 2 | 2 | 26/07/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clifton | 15 | Clition | FBA | FBA77 2 | 2 | 17. | 01/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26! | Clitoon | 33 | Clition | FBA | FBA77 23 V 10 | 2 | 1230 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH26 | Clition | 33 | Clitlon | FBA | FBA77 23M9 | 2 | 2239 | 05/09/77 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |



| TH29 | Cleeve | 54 | South Stoke | WANRA/E | PTHR. 0103 | 1 | 13600 | 27/04/95 | 459300 | 183700 | 47 | 0.26 | 8 | 126 | 55 | 300 | 1 | 29 | 50 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH29 | Cleeve | 56 | US Goring Weir | WANRA/E | PTHR.0120 | 1 | 13500 | 27/04/95 | 460000 | 181600 | 46 | 0.26 | 8 | 127 | 52 | 320 | 0 | 40 | 40 | 20 |
| TH29 | Cleeve | 56 | US Goring Weir | WANRA/E | PTHR. 0120 | 2 | 15200 | 13/09/95 | 460000 | 181600 | 46 | 0.26 | 8 | 127 | 52 | 320 | 0 | 40 | 40 | 20 |
| TH30 | Goring | 129 | Whitchurch | LR | 15592 | 1. | - 1 | 08/05/92 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | 13994 | 1 | 2 | 12/05/94 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | 28892 | 2 | 1 | 24/06/92 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | 41192 | 2 | 2 | 24/08/92 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Ooring | 129. | Whitchurch | LR | WH86 | 8 | 3 |  |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | 30594 | 3 | 1 | 12/09/94 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | WH77 | 8 | 1 |  |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Garing | 129 | Whitchurch | LR | WH78 | 8 | 2 |  |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 129 | Whitchurch | LR | 14394 | 2 | 3 | 13/06/94 |  |  |  | 0.26 | 8 |  |  |  |  |  |  |  |
| TH30 | Goring | 55 | Whitchurch Weir | WA/NRA/E | PTHR. 0115 | 3 | 15700 | 16/10/95 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 10 | 50 | 30 | 10 |
| TH30 | Ooring | 55. | Whitchurch Weir | WANRAAE | PTHR. 0115 | 3 | 8300 | 29/11/90 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 5 | 60 | 20 | 15 |
| TH30 | Goring | 55 | Whitchurch Weir | WANRA/E | PTHR. 0115 | 1 | 6600 | 23/05/90 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 5 | 15 | 70 | 10 |
| TH30 | Ooring | 55 | Whitchurch Weir | WA/RRA/E | PTHR. 0115 | 2 | 9000 | 03/07/91 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Goring | 55 | Whitchurch Weir | WANRA/E | PTHR. 0115 | 2 | 3700 | 24/07/87 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 5 | 90 | 5 |
| TH30 | Goring | 55 | Whitchurch Weir | WANRA/E. | PTHR. 0115 | 1 | 13400 | 12/04/95 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Goring | 55 | Whitchurch Weir | WA/NRAE | PTHR.0115 | 2 | 7500 | 23/07/90 | 463300 | 175800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 5 | 90 | 5 |
| TH30 | Goring | 55 | Whitchurch Weir | WANRA/E | PTHR. 0115 | 1 | 8900 | 15/05/91 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Goring | 55 | Whitchurch Weir | WA/NRA/E- | PTHR 0115 | 3 | 9900 | 14/11/91 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50. | 25 | 25 |
| TH30 | Ooring | 55 | Whitchurch Weir | WANRA/E- | PTHR0115 | 1 | 10300 | 28/05/92 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Goring | 55 | Whitchurch Weir | WA/NRA/E. | PTKR.0115 | 2 | 10600 | 11/08/92 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Ooring | 55 | Whitchurch Weir | WA/NRA/E- | PTHR. 0115 | 3 | 11600 | 04/10/93 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH30 | Goring | 55 | Whitchurch Weir | WANRAE | PTHR. 0115 | 2 | 11900 | 28/07/94 | 463300 | 176800 | 42 | 0.26 | 8 | 135 | 50 | 230 | 0 | 50 | 25 | 25 |
| TH32 | Mapledurha | 57 | Caversham Weir | WANRA/E | PTHR. 0080 | 3 | 900 | 31/12/80 | 471700 | 174100 | 41 | 0.26 | 8 | 145 |  |  |  |  |  |  |
| TH32 | Mapledurha | 57 | Caversham Weir | WANRA/E. | PTHR 0080 | 1 | 2300 | 18/02/87 | 471700 | 174100 | 41 | 0.26 | 8 | 145 | 75 | 280 |  |  |  |  |
| TH32 | Mapledurha | 57. | Caversham Weir | WANRRAE | PTHR. 0080 | 2 | 4000 | 12/08/87 | 471700 | 174100 | 41. | 0.26 | 8 | 145 | 75 | 280 | 5 | 60 | 25 | 10 |
| TH32 | Mapledurha | 57. | Caversham Weir | WANRA/E | PTHR 0080 | 1 | 4400 | 09/05/88 | 471700 | 174100 | 41 | 0.26 | 8 | 145 | 75 | 280 | 5. | 60 | 25 | 10 |
| TH32 | Mapledurha | 57. | Caversham Weir | WANRA/E | PTHR 0080 | 2 | 5900 | 19/07/89 | 471700 | 174100 | 41 | 0.26 | 8 | 145 | 75 | 280 | 5 | 60 | 25 | 10 |
| TH32 | Mapledurha |  | Caversham Weir | WANRAE | PTHR. 0080 | 3. | 6000 | 02/10/89 | 471700 | 174100 | 41 | 0.26 | 8 | 145 | 75 | 280 | 5. | 60 | 25 | 10 |

APPENDIX 3.2 A full list of the taxa recorded in one or more samples held in the macro-invertebrate data-base.

## Species found in entire database

## Invertebrates only

27-Mar-97

| 02110000 | Spongillidae | 17120000 | Unionidae |
| :---: | :---: | :---: | :---: |
| 03110000 | Hydridae | 17120100 | Uniosp. |
| 03110100 | Hydrasp. | 17120101 | Unio pictorum (L) |
| 05110000 | Planariidae | 17120102 | Unio tumidus Philipsson |
| 05110101 | Planaria torva (Muller) | 17120200 | Anodonta sp. |
| 05110203 | Polycelis tenuis (Ijima) | 17120201 | Anodonta anatina (L) |
| 05110202 | Polycelis nigra group | 17120202 | Anodonta cygnea (L.) |
| 05120103 | Dugesia tigrina (Girard) | 17130000 | Sphaeriidae |
| 05120102 | Dugesia polychroa group | 17130100 | Sphatrium sp. |
| 05130201 | Dendrocoelum lacteum (Muller) | 17130101 | Sphaerium comeam (L.) |
| 05120000 | Planatiidae (incl. Dugesiidac) | 17130102 | Sphatrium lacustre (Muller) |
| 08110100 | Prostoma sp. | 17130103 | Sphaerium rivicola (Lamarck) |
| 10000000 | Nerratoda | 17130105 | Sphacrium transversum (Say) |
| 1400000 | Ectoprocta | 17130200 | Pisidium sp. |
| 16110000 | Neritidac | 17130201 | Pisidium amricum (Muller) |
| 16110101 | Theodoxus fluviatilis (L) | 17130202 | Pisidium casertanum (Poli) |
| 16120101 | Viviparus corriectus (Millet) | 17130204 | Pisidium henslowanum (Sheppard) |
| 16120102 | Viviparus viviparus (L.) | 17130208 | Pisidium moitessierianum Paladilhe |
| 16130000 | Valvatidae | 17130209 | Pisidium nitidum Jenyns |
| 16130100 | Valvata sp. | 17130215 | Pisidium subtruncatum Malm |
| 16130101 | Valvata cristata Muller | 17130216 | Pisidium supinum Schmidt |
| 16130103 | Valvata piscinalis (Muller) | 20000000 | Oligochata |
| 16140000 | Hydrobiidae | 20110000 | Lumbriculidae |
| 16140301 | Potamopyrgus jenkinsi (Smith) | 20110300 | Stylodritus sp. |
| 16160101 | Bithynia leachuii (Sheppard) | 20110302 | Stylodrilus heringianus Claparede |
| 16160102 | Bithynia tentaculata (L) | 20110200 | Lumbriculus group |
| 16120000 | Hydrobiidae (incl. Bithynuidae) | 20310000 | Enchytraeidae |
| 16210000 | Physidae | 20330000 | Naididae |
| 16210200 | Physa sp. | 20330100 | Chaetogaster sp. |
| 16210202 | Physa fontinalis (L) | 20330501 | Uncinais uncinata (Orsted) |
| 16210202 | Physa acuta group | 20330601 | Ophidonais serpentina (Muller) |
| 16220000 | Lymnaeidae | 20330702 | Nais barbata Muller |
| 16220101 | Lymnaea auricularia (L) | 20330703 | Nais bretscheri Michaelsen |
| 16220103 | Lymnaea palustris (Muller) | 20330706 | Nais pardalis Piguet |
| 16220104 | Lymnaea peregra (Muller) | 20330708 | Nais simplex Piguet |
| 16220105 | Lymnaea stagnalis (L.) | $2033070 Y$ | Nais commmis group |
| 16220106 | Lymarea truncatula (Muller) | 20331201 | Stylaria lucustris (L) |
| 16230000 | Planoridae | 20340000 | Tubificidae |
| 16230101 | Planorbis carinalus Multer | 20340102 | Tubifex igrotus (Stolc) |
| 16230102 | Planorbis planorbis (L.) | 20340104 | Tubifex newaensis (Michaelsen) |
| 16230201 | Anisus leucostoma (Millet) | 20340106 | Tubifex tubifex (Muller) |
| 16230202 | Anisus vortex (L.) | 20340201 | Limnodrilus cervix Brinkhurst |
| 16230301 | Bathyomphalus contortus (L.) | 20340202 | Limnodrilus claparedeianus Ratzel |
| 16230401 | Gyraulus acronicus (Ferussac) | 20340203 | Limnodrilus hoffmeisteri Claparede |
| 16230402 | Gyraulus albus (Muller) | 20340204 | Limnodrilus profundicola (Verrill) |
| 16230501 | Arnigar crista (L) | 20340205 | Limnodrilus udekemianus Claparede |
| 16230601 | Hippeutis complanatus (L) | 20340302 | Psammoryctides barbatus (Grube) |
| 16230801 | Planorbarius comeus (L) | 20340402 | Potamothrix hammoniensis (Michaelsen) |
| 16240000 | Ancylidae | 20340404 | Potamothrix moldaviensis (Vejdovsky \& Mrazek) |
| IG240101 | Ancylus fluviatilis Multer | 20340903 | Aulodrilus phuriseta (Piguet) |
| 16250101 | Acroloxus lacustris (L.) | 20341101 | Rhyacodrilus coccineus (Vejdovsky) |
| 162Z0000 | Ancylidae (incl. Acroloxidac) | 20341301 | Branchiura sowerbyi Beddard |
| 16320200 | Succinea sp. | 20420000 | Lumbricidae |

Species found in entire database

## Invertebrates only

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| 22110000 | Piscicolidae | 40120112 | Baetis scambus group |
| :---: | :---: | :---: | :---: |
| 22110101 | Piscicola geometra ( $\mathrm{L}_{-}$) | 40120201 | Centroptilum luteolum (Muller) |
| 22120000 | Glossiphonitdae | 40120300 | Cloeon sp. |
| 22120201 | Theromyzon tessulaturn (Muller) | 40120301 | Cloeon dipterum (L.) |
| 22120301 | Hemiclepsis marginata (Muller) | 40120302 | Cloeen simile Eaton |
| 22120401 | Glossiphonia complanata (L) | 40120401 | Procioeon bifidum Benglsson |
| 22120402 | Glossiphonia heteroclita (L) | 40130000 | Heptagenïdae |
| 22120601 | Boreobdella verrucata (Muller) | 40130201 | Heptagenia fuscogrista (Retzius) |
| 22120701 | Helobdella stagnalis (L) | 40210000 | Leptophlebiidae |
| 22310000 | Erpobdellidae | 40210301 | Habrophlebia fusca (Curtis) |
| 22310100 | Erpobdella sp. | 40320000 | Ephemeridae |
| 22310101 | Efpobdella octoculata (L.) | 40320100 | Ephemera sp. |
| 22310201 | Dina lineata (Muiler) | 40320101 | Ephemera danica Muller |
| 22310302 | Trocheta subviridis Dutrochet | 40320103 | Ephemera vulgata L. |
| 24000000 | Hydracarina | 40410000 | Ephemerellidae |
| 24320107 | Eylais extendens (Mulier) | 40410101 | Ephemerellia ignita (Poda) |
| 24320111 | Eylais infundibulifera Koenike | 40510000 | Cacnidae |
| 24420101 | Hydrodroma despiciens (Muller) | 40510200 | Caenis sp. |
| 24540144 | Lebertia (Pilolebertia) inaequalis (Koch) | 40510201 | Caenis horaria (L.) |
| 24540145 | Lebertia (Pilolebertia) insignis Neuman | 40510203 | Caenis macrura Stephens |
| 24540149 | Lebertia (Pilolebertia) porosa Thor | 40510205 | Caenis rivulorım Eaton |
| 24560111 | Torrenticola (Torrenticola) amplexa (Koenike) | 40510206 | Caenis robusta Eaton |
| 24610113 | Limnesia (Limenesia) koenikei Piersig | 40510202 | Caenis luctuosa group |
| 24610114 | Limnesia (Limmesia) maculata (Muller) | 41110000 | Taeniopterygidae |
| 24610115 | Limnesia (Limnesia) undulata (Muller) | 41120000 | Nemouridae |
| 24620112 | Hygrobates (Hygrobates) fuviatilis (Strom) | 41120400 | Nemotra sp. |
| 24620114 | Hygrobates (Hygrobates) longipalpis (Hermann) | 41130000 | Leuctridae |
| 24630121 | Unionicola (Pentatax) aculeata (Koerike) | 41130102 | Leuctra geniculata (Stephens) |
| 24630211 | Neumania (Neumania) callosa (Koenike) | 41210000 | Perlodidae |
| 24650200 | Piona sp. | 42110000 | Platyerternididae |
| 24650207 | Piona coccinea (Koch) | 42110101 | Platycnmis pennipes (Pallas) |
| 24650208 | Piona conglobata (Koch) | 42120000 | Coenagriidae |
| 24650218 | Piona pusilla (Neuman) | 42120201 | Ischntra elegans (Van der Lixden) |
| 24650913 | Forelia (Forelia) variegator (Koch) | 42120301 | Enallagma cyathigerum (Charpentier) |
| 24730112 | Mideopsis (Mideopsis) orbicularis (Muller) | 42120402 | Coenagrion pueila group |
| 24770100 | Arrenurus sp. | 42120601 | Erythromma najas (Hansemann) |
| 24770113 | Arrenurus (Arrenurus) albator (Muller) | 42140000 | Calopterygidae |
| 30000000 | Ostracoda | 42140100 | Calopteryx sp. |
| 32010100 | Argulus sp. | 42140101 | Calopteryx splendens (Harris) |
| 36110000 | Asellidae | 42140102 | Calopteryx virgo (L) |
| 36110101 | Asellus aquaticus (L.) | 42210000 | Gomphidae |
| 36110104 | Asellus meridianus Racovita | 42210101 | Gorophus vulgatissimus (L) |
| 37110000 | Corophiidae | 42230000 | Aesinidae |
| 37110101 | Corophium curvispinum Sars | 42230200 | Aeshna sp. |
| 37130101 | Crangonyx pseudogracilis Bousfield | 42230202 | Acshna cyanea (Muller) |
| 37140000 | Giammaridae | 42230203 | Aestuna grandis (L.) |
| 37140200 | Gammarus sp. | 42230301 | Anax imperator Leach |
| 37140206 | Gammarus pulex (L.) | 42250000 | Libelatidae |
| 37120000 | Gammaridae (incl. Crangonyctidae \& Niphargidae) | 42250306 | Sympetrum striolatum (Charpentier) |
| 40120000 | Baetidae | 43110000 | Mesovelidae |
| 40120100 | Baetis sp. | 43210000 | Hydrometridac |
| 40120107 | Bactis rhodani (Pictet) | 43210102 | Hydrometra stagnorum (L.) |
| 40120111 | Baetis vernus Curtis | 43220000 | Veliidae |

## Species found in entire database

## Invertebrates only

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| 43230000 | Gerridae | 45150212 | Gyrinus urinator Illiger |
| :---: | :---: | :---: | :---: |
| 43230100 | Gerris sp. | 45150401 | Orectochilus villosus (Multer) |
| 43230114 | Gerris (Gerris) lacustris (L.) | 45120000 | Dytiscidae (incl. Noteridae) |
| 43310000 | Nepidae | 45310000 | Hydrophilidae |
| 43310101 | Nepa cinerea L. | 45310300 | Helophorus sp. |
| 43310201 : | Ranatra linearis (L.) | 45310352 | Helophorus (Atracthelophorus) brevipalpis Bedel |
| 43410000 | Naucoridae | 45311101 | Hydrobrius fuscipes (L.) |
| 43420000 | Aphelocheiridae | 45311301 | Anacaena bipestulata (Marsham) |
| 43420101 | Aphelocheinus aestivalis (Fabricius) | 45311302 | Anacuena globulus (Paykuli) |
| 43510000 | Notonectidae | 45311303 | Anacaena timbata (Fabricius) |
| 43510100 | Notonecta sp. | 45311400 | Laccobius sp. |
| 43510101 | Notonecta glauca L. | 45311412 | Laccobius (Laccobius) mirutus (L.) |
| 43510102 | Notonecta maculata Fabricius | 45311423 | Laccobius (Macrolaccobius) bipunctatus (Fabricius) |
| 43510104 | Notonecta viridis Delcourt | 45311426 | Laccobius (Macrolaccobius) sinuatus Mosschulsky |
| 43610000 | Corixidae | 45311427 | Laccobius (Macrolaccobius) striatulus (Fabricius) |
| 43610100 | Micronecta sp. | 45370000 | Hydrophilidae (incl. Hydraenidae) |
| 43610111 | Micronecta (Dichaetonecta) scholtzi (Scholtz) | 45620200 | Dryops sip. |
| 43610122 | Micronecta (Micronecta) poweri (Douglas \& Scott) | 45630000 | Elmidae |
| 43610900 | Sigara sp. | 45630101 | Elmis aenea (Muller) |
| 43610910 | Sigara (Sigara) sp. | 45630301 | Limnius volckmari (Panzer) |
| 43610921 | Sigara (Subsigara) distincta (Fieber) | 45630600 | Oulimnius sp. |
| 43610922 | Sigara (Subsigara) falleni (Fieber) | 45630601 | Oulitratus major (Rey) |
| 43610924 | Sigara (Subsigara) fossarum (Leach) | 45630604 | Oulimnius tuberculatus (Muller) |
| 45110000 | Haliplidae | 45710200 | Donacia sp. |
| 45110101 | Brychius elevatus (Parzer) | 46110000 | Sialidae |
| 45110300 | Haliplus sp. | 46110101 | Sialis fuliginosa Pictet |
| 45110303 | Haliplus flavicollis Sturm | 46110102 | Sialis lutaria (L) |
| 45110304 | Haliplus fluviatilis Aube | 46110103 | Sialis nigripes Pictet |
| 45110308 | Haliplus immaculatus Gerhardt | 47120100 | Sisyra sp. |
| 45110309 | Haliplus laminatus Schaller | 48130000 | Hydroptilidae |
| 45110311 | Haliplus lineatocollis (Marsham) | 48130101 | Agraylea multipunctata Curtis |
| 45110314 | Hatiplus obliquus (Fabricius) | 48130300 | Hydroptila sp. |
| 45110315 | Haliplus ruficollis (Degeer) | 48130400 | Oxyethira sp. |
| 45130101 | Noterus clavicornis (Degeer) | 48130600 | Ithytrichia sp. |
| 45140000 | Dytiscidae | 48120000 | Rhyacophilidae (incl. Glossosomatidae) |
| 45140100 | Laccophilus sp. | 48220100 | Lypesp. |
| 45140101 | Laccophilus hyalinus (Degeer) | 48220101 | Lype phaeopa (Stephens) |
| 45140301 | Hyphydus ovatus (L) | 48220102 | Lype reducta (Hagen) |
| 45140604 | Hygrotus versicolor (Schaller) | 48220400 | Tinodes sp. |
| 45140800 | Hydroporus sp. | 48220408 | Tinodes waencri (L) |
| 45140812 | Hydroporus incognitus Sharp | 48230101 | Ecnomus tenellus (Rambur) |
| 45140824 | Hydroporus palustris (L) | 48240000 | Polycentropodidae |
| 45141303 | Potamonectes depressus (Fabricius) | 48240101 | Cymus flavidus Mclachlan |
| 45141401 | Stictotarsus duodecimpustulatus (Fabricius) | 48240103 | Cymus trimaculatus (Curtis) |
| 45141901 | Platambus maculatus (L.) | 48240301 | Neureclipsis bimaculata (L.) |
| 45142009 | Agabus didymus (Olivjer) | 48240400 | Plectrocnemia sp. |
| 45142018 | Agabus sturmii (Gyllenhal) | 48240402 | Plectrocnemia conspersa (Curtis) |
| 45142102 | Ilybius ater (Degeer) | 48240403 | Plectrocnemia geniculata Mclachlan |
| 45142104 | Ilybius fuliginosus (Fabricius) | 48240500 | Polycentropus sp. |
| 45142700 | Dytiscus sp. | 48240501 | Polycentropus flavomaculatus (Pitct) |
| 45150000 | Gyrinidae | 48240502 | Polycentropus irroratus (Curlis) |
| 45150200 | Gyrinus sp. | 48250000 | Hydropsychidac |
| 45150204 | Gyrinus distinctus Aube | 48250200 | Hydropsyche sp. |

## Species found in entire database

## Invertebrates only

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| 48250201 | Hydropsyche angustipennis (Curtis) | 50420500 | Procladius sp. |
| :---: | :---: | :---: | :---: |
| 48250207 | Hydropsyche pellucidula (Curtis) | \$0420601 | Psectrotanypus varius (Fabricius) |
| 48220000 | Psychomyïdae (incl. Ecromidae) | 50420800 | Ablabesmyia sp. |
| 48310060 | Phryganeidae | 50420802 | Ablabesmyia monilis ( L ) |
| 48310500 | Phryganea sp. | 50420900 | Arctopelopia sp. |
| 48310501 | Phryganea bipunctata Retzius | 50420Y00 | Thienemannimyia group |
| 48320000 | Brachycentridae | 50421000 | Conchapelopia sp. |
| 48320101 | Brachycentrus subnubilus Curtis | 50421001 | Conchapelopia melanops (Meigen) |
| 48330301 | Lepidostoma hirtum (Fabricius) | \$0421400 | Larsia sp. |
| 48340000 | Limnephilidae | 50421600 | Natarsia sp. |
| 48340501 | Allogamus auricollis (Pictet) | 50421701 | Nilotanypus dubius (Meigen) |
| 48340600 | Halesus sp. | 50421800 | Paramerina sp. |
| 48340602 | Halesus radiatus (Curtis) | 50421900 | Rheopeiopia sp. |
| 48341102 | Potamophylax latipemis (Curtis) | 50422100 | Thienemannimyia sp. |
| 48341401 | Anabolia nervosa (Curtis) | 50422501 | Tanypus punctipennis Meigen |
| 48341700 | Limnephilus sp. | 50440200 | Diamesa sp. |
| 48341719 | Limnephilus lunatus Curtis | 50440300 | Potthastia sp. |
| 48341X00 | Polamophylax group | 5044030Y | Potthastia gaedii group |
| 48350000 | Goeridae | 5044030Z | Potthastia longinana group |
| 48350101 | Goera pilosa (Fabricius) | 50450201 | Odontomesa fulva (Kieffer) |
| 48350200 | Silo sp. | 50450301 | Prodiamesa olivaces (Meigen) |
| 48370000 | Sericostomatidae | 50460000 | Orthocladinae |
| 48390000 | Molanmidae | 50460300 | Brillia sp. |
| 48390101 | Molanma angustata Curtis | 50460301 | Brilia flavifions Johannsen |
| 48410000 | Leptoceridae | 50460402 | Cardiocladius fuscus Kieffer |
| 48410102 | Athripsodes aterrinus (Stephers) | 50460500 | Cricotopas sp. |
| 48410104 | Athripsodes cinereus (Curtis) | 50460510 | Cricotopus (Cricotopus) sp. |
| 48410200 | Ceraclea sp. | 50460513 | Cricotopus (Cricotopus) bicinctus (Meigen) |
| 48410202 | Ceraclea annulicornis (Stephens) | 50460520 | Cricotopus (Cricotopus) sp. |
| 48410203 | Ceraclea dissimilis (Stephens) | 50460540 | Cricotopus (Isocladius) sp. |
| 48410206 | Ceraclea senitis (Burmeister) | 50460549 | Cricotopus (Isocladius) sylvestris (Fabricius) |
| 48410302 | Leptocerus lusitanicus (Mclachlan) | 50460561 | Cricotopus (Nostrocladius) lygropis Edwards |
| 48410400 | Mystacides sp. | 50460800 | Eukiefferiella sp. |
| 48410401 | Mystacides azurea (L) | 50460802 | Eukiefferiella claripennis (Lundbeck) |
| 48410402 | Mystacides longicomis (L) | 50460 Z 00 | Cricotopus group |
| 48410403 | Mystacides nigra (L) | 50461300 | Heterotrissocladius sp. |
| 48410701 | Triaenodes bicolor (Curtis) | 50461800 | Nunocladius sp. |
| 48410900 | Oecetis sp. | 50461801 | Nanocladius balticus Palmen |
| 48410902 | Oecetis lacustris (Pictet) | 50462000 | Orthocladius sp. |
| 48410904 | Oecetis ochracea (Curtis) | 50462010 | Orthocladius (Eudactylocladius) sp. |
| 50000000 | Diptera | 50462021 | Orthocladius (Pogonocladius) consobrinus (Holmgren) |
| 50100000 | Tipulidae | 50462030 | Orthocladius (Euorthocladius) sp. |
| S0t 10300 | Tipula sp. | 50462040 | Orthocladius (Othocladius) sp. |
| 50110412 | Tipula (Yamatotipula) montium group | 50462050 | Orthocladeus (Symposiocladius) sp. |
| 50130900 | Hetius sp. | 30462100 | Paracladius sp. |
| 50.350000 | Ceratopogonidae | 50462101 | Paracladius conversus (Walker) |
| 50360000 | Simuliidae | 50462301 | Paratrichocladius rufiventris (Meigen) |
| 50360361 | Simulium (Boophthora) erythrocephalum (de Geer) | 50462700 | Psectrocladius sp. |
| 50400000 | Chironomidae | 50462721 | Psectrocladius (Allopsectrocladius) obvius (Walker) |
| 50420101 | Clinotanypus nervosus (Meigen) | 50462740 | Psectrocladius (Psectrocladius) sp. |
| 50420201 | Apsectrotanypus trifascipennis (Zetterstedi) | 50462800 | Rheocricotopus sp. |
| 50420400 | Macropelopia sp. | 50462901 | Synorthocladius semivirens (Kiefter) |
| 50420402 | Macropelopia nebulosa (Meigen) | 50463200 | Tvetenia sp. |

## Species found in entire database

## Invertebrates only

27-Mar-97

| 50463512 | Bryophaenocladius subvernalis (Edwards) | 50490900 | Tanytarsus sp. |
| :---: | :---: | :---: | :---: |
| 50463700 | Chactocladius sp. | 50490902 | Tanytarsus brundini Lindeberg |
| 50463800 | Corynoneura sp. | 50490W00 | Micropsectra group |
| 50463901 | Epoicocladius flavens (Malloch) | 50491000 | Virgatanytarsus sp. |
| 50464300 | Limmophyes sp. | 50491101 | Zavrelia pertatoma Kieffer |
| $50464500^{\circ}$ | Metriocnemus sp. | 50630000 | Tabanidae |
| 50464700 | Parakiefferielia sp. | 50630100 | Cirysops sp. |
| 50464701 | Parakiefferiella bathophila (Kieffer) | 50710000 | Empididae |
| 50464900 | Parametriocnemus sp. | 50830000 | Ephydridac |
| 50465201 | Paratrissocladius excerptus (Walker) | 50850000 | Muscidae |
| 50465300 | Pseuderthocladius sp. |  |  |
| 50465400 | Pseudosmittia sp. |  |  |
| 50465900 | Thienemanniella sp. |  |  |
| 50466300 | Paracricotopus sp. |  |  |
| 50470300 | Chiroriomus sp. |  |  |
| 50470400 | Cladopelma sp. |  |  |
| 50470500 | Cryptochironomus sp. |  |  |
| 50470600 | Cryptotendipes sp. |  |  |
| 50470701 | Demeijerea rufipes (L.) |  |  |
| 50470801 | Demicryptochironomus vuineratus (Zetterstedt) |  |  |
| 50470900 | Dicrotendipes sp. |  |  |
| 50470920 | Dicrotendipes (Limnochironomus) sp. |  |  |
| 50470922 | Dicrotendipes (Lirnnochironomus) nervostus (Staeger) |  |  |
| 50470923 | Dicrotendipes (Limnochironomus) notatus (Meigen) |  |  |
| 50471100 | Endochironomus sp. |  |  |
| 50471200 | Glyptotendipes sp. |  |  |
| 50471400 | Hamischia sp. |  |  |
| 50471501 | Kiefferulus tendipediformis (Goetghebuer) |  |  |
| 50471800 | Microchironomus sp. |  |  |
| 50471900 | Microtendipes sp. |  |  |
| 50472300 | Parachironomus sp. |  |  |
| 50472400 | Paracladopelma sp. |  |  |
| 50472401 | Paracladopelma camptolabis (Kieffer) |  |  |
| 5047240 Z | Paracladopelma camptolabis group |  |  |
| 50472500 | Paralauterborniella sp. |  |  |
| 50472501 | Paraluuterborniella nigrohalteralis (Malloch) |  |  |
| 50472600 | Paratendipes sp. |  |  |
| 50472800 | Phaenopsectra sp. |  |  |
| 50472900 | Polypeditum sp. |  |  |
| 50472910 | Polypedilum (Pentapedilum) sp. |  |  |
| 50472920 | Polypedilum (Polypedilum) sp. |  |  |
| 50473100 | Stenochironomus sp. |  |  |
| 50473200 | Stictochironomus sp. |  |  |
| 50473301 | Xenochironomus xenolabis (Kieffer) |  |  |
| 50490100 | Cladotanytarsus sp. |  |  |
| 50490200 | Micropsectra sp. |  |  |
| 50490203 | Micropsectra atrofasciata Kieffer |  |  |
| 50490300 | Neozavrelia sp. |  |  |
| 50490500 | Paratanytarsus sp. |  |  |
| 50490600 | Rheotnnylarsus sp. |  |  |
| 50490701 | Stempellina bausei (Kieffer) |  |  |
| 50490702 | Stempellina almi Brundin |  |  |
| 50490800 | Stempellinella sp. |  |  |

APPENDIX 3.3 A full list of the families recorded in one or more samples held in the macro-invertebrate data-base.


|  | $\begin{aligned} & \text { 會 } \\ & \stackrel{y}{n} \\ & \dot{y} \\ & \stackrel{y}{7} \end{aligned}$ |  |  | $\begin{aligned} & \frac{x}{2} \\ & \frac{1}{d} \\ & \frac{1}{2} \\ & \frac{2}{2} \\ & \frac{1}{2} \end{aligned}$ |  |  | TH16－Northmoper |  |  | $\begin{gathered} \infty \\ \dot{\bar{y}} \\ \dot{y} \\ \dot{F} \\ \hline \end{gathered}$ |  | $\begin{aligned} & \text { y } \\ & \frac{1}{\delta} \\ & \frac{1}{9} \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \text { 穿 } \\ & \text { 突 } \\ & \text { E } \end{aligned}$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spongillidate |  |  |  | $+$ | $\pm$ | ＋ | $+$ |  | $+$ |  |  | $+$ | $+$ |  |  |  |  |  |  |  |  |
| Hydridec |  |  |  |  |  |  |  |  |  |  |  |  | $+$ |  |  |  | $+$ | 4 |  |  |  |
| Ftmariidec |  |  |  |  |  |  | 4 | $\pm$ | $\pm$ | 4 |  | $+$ | 4 | $\pm$ |  |  | ＋ |  |  |  |  |
| Duresiidse |  |  |  |  |  | ＋ |  |  | ＋ | 4 | $+$ |  | $+$ | $+$ |  |  | ＋ |  |  | 4 |  |
| Dendroeoclidax |  |  |  |  |  |  |  |  |  |  |  | $\pm$ | ＋ | ＋ |  |  |  |  |  |  |  |
| Planeridxe（incil Dugestidac） |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Teqaticmmatidze |  |  |  |  |  | $\pm$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Nematoda | $\pm$ | ＋ | 4 | $+$ | ＋ | ＋ | $+$ | ＋ | ＋ | ＋ | ＋ | $\pm$ | $+$ | $\pm$ | $+$ | ＋ | $+$ |  |  |  |  |
| Ecoproct |  | ＋ |  |  |  | $\pm$ | ＋ |  | ＋ |  |  | ＋ | ＋ |  |  |  | ＋ |  |  |  |  |
| Neritidse | $\pm$ |  | $+$ |  | $\pm$ | $\pm$ | $\pm$ |  | ＋ | ＋ | $+$ | $+$ | ＋ | $\pm$ | ＋ | $+$ | $+$ | ＋ |  |  | ＋ |
| Vivipandes |  |  |  |  | $\pm$ | $\pm$ | $\pm$ |  | $\pm$ | ＋ | ＋ | ＋ | $+$ | $+$ | $+$ | $+$ | $+$ | $+$ |  | ＋ |  |
| Valvatide | $\pm$ | ＋ |  |  | 4 | $\pm$ | ＋ |  | $\pm$ | $\pm$ |  | $+$ | ＋ | $+$ | $\pm$ | ＋ | ＋ | ＋ | 4 | ＋ | 4 |
| Hydrobiiose | $+$ | $+$ | $+$ | $+$ | $+$ | $\pm$ | ＋ | $\pm$ | $+$ | ＋ | ＋ | $+$ | $+$ | $\pm$ | $+$ | $+$ | ＋ | $\pm$ |  | 4 |  |
| Bithyniidae | $+$ | 4 | 4 | $\pm$ | $\pm$ | $+$ | ＊ | ＋ | ＋ | $\pm$ | $\pm$ | ＋ | ＋ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ |  | $\pm$ |  |
| Hydrobiidae（incl Bidhynitac） | ＋ |  |  |  | $\pm$ | $\pm$ |  |  | ＋ | ＋ |  | $+$ | $+$ | $\pm$ |  | ＋ | ＋ |  | $+$ | $\pm$ | ＋ |
| Physidge | $+$ | $\pm$ |  |  | $\pm$ | $\pm$ | ＋ |  | ＋ | ＋ |  | ＋ | $+$ | $\pm$ | $+$ | ＋ | ＋ | ＋ | $\pm$ | $\pm$ | ＋ |
| Lymnatidae | $+$ | $\pm$ |  |  | $\pm$ | 4 | ＋ |  | $+$ | $+$ | $\pm$ | ＋ | ＋ | $\pm$ | $+$ | $t$ | ＋ | ＋ | $\pm$ | $\pm$ | ＋ |
| Planorbidx | ＋ | ＋ |  |  |  | $\pm$ | $\pm$ | $\pm$ | ＋ | ＋ | $+$ | ＋ | $\pm$ | $+$ | ＋ | $+$ | 4 | 4 | ＋ | $\pm$ | ＋ |
| Ancylidae | $\pm$ | ＋ | 4 |  | 4 | ＋ | $\pm$ |  | ＋ | $\pm$ | ＋ |  | ＋ | ＋ | $+$ | $+$ | $\pm$ | $\pm$ |  |  |  |
| Acroloxidas | 4 |  |  |  |  | ＋ | ＋ |  | $+$ |  |  |  | $\pm$ | $+$ | $+$ | 4 | $\pm$ | ＋ |  |  |  |
| Ancylidae（incl．Acrotoxidae） | 4 |  |  |  | 4 | ＋ |  |  | $+$ | ＋ |  | ＋ | ＋ | $\pm$ |  |  | ＋ |  | 4 | $\pm$ | ＋ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Unionidac | ＋ | $+$ | ＋ | 4 | ＋ | ＋ | $\pm$ | ＋ | $+$ | ＋ | $+$ | $+$ | 4 | $\pm$ | $\pm$ | ＋ | ＋ | ＋ | $\pm$ | $+$ | $\pm$ |
| Sphaeriodme． | $+$ | $+$ | $\pm$ | 4 | $+$ | 4 | $\pm$ | 4 | 4 | ＋ | ＋ | $\pm$ | $+$ | ＋ | $\pm$ | $\rightarrow$ | $+$ | $+$ | 4 | $\pm$ | ＋ |
| Oligochasta． | $+$ |  |  |  | $\pm$ | 4 |  |  | 4 | ＋ |  | $+$ | $\pm$ | $+$ | $+$ | $\pm$ | 4 |  | $+$ | $\pm$ | $+$ |
| Lambriculidae | $\pm$ | $\rightarrow$ | ＋ | 4 | $\pm$ | $\pm$ | 4 | $+$ | ＋ | ＋ | ＋ | ＋ | $+$ | $+$ | $\pm$ | $\pm$ | 4 |  |  | $\pm$ | ＋ |
| Emelay |  |  |  |  |  |  | $\pm$ |  |  |  |  |  |  | 4 |  |  | ＋ | 4 |  |  |  |
| Nasidide： | ＋ | $\pm$ | ＋ |  | ＋ | $+$ | $\pm$ | ＋ |  | ＋ |  |  | $+$ |  |  |  | 4 | ＋ |  |  | ＋ |
| Tubificidae | $+$ | $+$ | $\stackrel{+}{+}$ | 4 | 4 | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $+$ | $\pm$ | ＋ | ＋ | － | 4 | － | 4 |  |  |  |
| Lumbricide |  | $\pm$ |  |  |  |  | $+$ | $+$ | ＋ |  |  | $\pm$ | $+$ | ＋ | $+$ | ＋ |  |  |  |  |  |
| Piscicolidxe | $+$ |  |  |  |  |  | ＋ |  | ＋ | ＋ | ＋ | ＋ | $\pm$ | $+$ | $+$ | ＋ |  | $\pm$ | $\pm$ | $+$ | ＋ |
| Gloasiphoniidae | ＋ | $+$ |  | $\pm$ | ＋ | ＋ | $\pm$ | ＋ | ＋ | ＋ | ＋ | ＋ | $+$ | $+$ | $+$ | ＋ | ＋ | $+$ | ＋ | ＋ | ＋ |
| Erpobdellidx | ＋ | ＋ | $\pm$ |  |  | ＋ | ＋ |  | ＋ | ＋ | ＋ | ＋ | $+$ | ＋ | 4 | $+$ | ＋ | 4 | ＋ | $\pm$ | $\pm$ |
| Hydncrim | $+$ | $+$ |  |  |  | $+$ | 4 |  | ＋ | ＋ |  |  | ＋ |  |  |  | $+$ | $+$ |  |  |  |
| Eylidae |  |  |  |  |  |  |  |  |  |  |  |  | $+$ |  | $+$ |  |  |  |  |  |  |
| Hydrodromidae |  |  |  |  |  | $\pm$ | $\pm$ | $+$ |  | ＋ |  | ＋ |  |  |  |  |  |  |  |  |  |
| Lebertidas | $+$ | ＋ | ＋ |  | 4 | ＋ | $\pm$ | 4 | $\pm$ | $\pm$ | $+$ | ＋ | ＋ |  |  | $\pm$ |  | ＋ |  |  |  |
| Torrenticolidse |  |  |  |  |  | $\pm$ |  |  |  | $\pm$ |  |  | $+$ |  |  |  |  |  |  |  |  |
| Limnesiidre | － |  |  |  | ＋ | ＋ |  | ＋ | ＋ | ＋ |  | ＋ | $\pm$ | $+$ | $\pm$ | $\pm$ | ＋ | $\pm$ |  |  |  |
| Hyprobatidue | ＋ | $\pm$ | $\pm$ | $+$ | $+$ | $+$ | ＋ | 4 | $\pm$ | ＋ | $+$ | ＋ | $+$ |  | $\pm$ | $\pm$ |  | $\pm$ |  |  |  |
| Unionicolidue |  |  |  |  |  |  |  |  | $+$ | ＋ | $+$ | $\pm$ | $+$ | $+$ | ＋ | $\pm$ | ＋ |  |  |  |  |
| Pionidse |  |  |  |  | $+$ | $+$ |  | $+$ | ＋ |  | ＋ | ＋ | $+$ | $\pm$ | $\pm$ |  |  |  |  |  |  |
| Mideoptider | ＋ | ＋ | ＋ | － | ＋ | $\pm$ | $\pm$ | 4 | $\pm$ | ＋ | $+$ | ＋ | ＋ |  | $+$ | 4 | $\pm$ | $+$ |  |  |  |
| Atrenuridic |  |  |  |  |  |  |  |  |  |  |  | ＋ | ＋ |  | $\pm$ |  | ＋ |  |  |  |  |
| Ostracoda | ＋ |  |  |  |  | 4 |  |  | 4 |  |  | $+$ | ＋ | $+$ |  | $\pm$ | ＋ |  |  |  | ＋ |
| Arpulidae | $\pm$ |  |  |  |  | ＋ |  |  |  |  |  | $\pm$ | $+$ | 4 | $\pm$ |  |  |  |  |  |  |
| Axellidx | $\pm$ | $\pm$ | 4 | 4 | 4 | 4 | $\pm$ | $\pm$ | $\pm$ | $+$ | $\pm$ | 4 | $+$ | $+$ | $+$ | 4 | ＋ | ＋ | 4 | 4 | $\pm$ |
| Corophiside | ＋ |  |  |  |  | $+$ | $+$ | $+$ | $\pm$ | $\pm$ | $\pm$ | ＋ | $+$ | $+$ | $+$ | $+$ | 4 | 4 | $+$ | $\pm$ | ＋ |
| Crapgoayctide | ＋ | 4 | ＋ | ＋ | $+$ | $\pm$ | 4 | $\pm$ | $+$ | $+$ | $+$ | $\pm$ | $\pm$ | $+$ | $\pm$ | ＋ | ＋ | ＋ |  | ＋ |  |
| Gemmandre | $\pm$ | 4 | ＋ |  | ＋ | $\pm$ | $\pm$ |  | $+$ | $\pm$ |  | $+$ | $\pm$ | $+$ | $\pm$ | ＋ | ＋ |  |  | ＋ |  |
| Gammaridee（inel．Crangonyetide \＆Niphargidare） | $+$ |  |  |  | ＋ | ＋ |  |  | $\pm$ | $\pm$ |  | $+$ | $\pm$ | $\pm$ |  | ＋ | ＋ |  | $+$ | $\pm$ | $+$ |
| Bacidas | 4 | $+$ | $\stackrel{+}{+}$ | $\pm$ | $\pm$ | $+$ | $+$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $+$ | $\pm$ | $\pm$ | ＋ | ＋ | 4 | $\pm$ | ＋ | $\pm$ | $+$ |
| Heptugeniidae |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Leplophlebiidae | $+$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Ephemeridue | ＋ | ＋ |  |  |  | ＋ | ＋ |  | ＋ | $\pm$ | ＋ | $\pm$ | $\pm$ | ＋ | $+$ | $\pm$ | $\pm$ | $+$ | $+$ | $+$ | $\pm$ |
| Ephemetellidx | $\pm$ | $\pm$ |  |  |  | $+$ | $\pm$ |  | $\pm$ | $+$ | $+$ | $+$ | 4 | $\pm$ | $\pm$ |  | $+$ |  |  | ＋ | $\pm$ |
| Caenidae | $+$ | $\pm$ | $+$ | $\pm$ | $\pm$ | $+$ | $\pm$ | $+$ | $+$ | $+$ | ＋ | ＋ | $+$ | $\pm$ | $+$ | 4 | $+$ | $+$ | 4 | ＋ | $+$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lewtridae | $\pm$ |  |  |  |  |  |  |  | $\pm$ | $+$ |  | ＋ | $+$ |  |  |  |  |  |  | $+$ |  |
| Perlodidae |  |  |  |  |  |  |  |  |  |  |  | $\pm$ |  |  |  |  |  |  |  |  |  |
| Platycnemididae |  |  |  | $+$ |  | ＋ | $+$ |  | $\pm$ | ＋ |  |  | ＋ | $+$ | ＋ | ＋ | $+$ |  |  | $+$ |  |
| Coenagriidae | ＋ |  | $\pm$ |  |  | 4 | $+$ |  | $\pm$ | $\pm$ |  | $\pm$ | ＋ | ＋ | $\pm$ | ＋ | ＋ | $+$ | $+$ | $\pm$ |  |
| Coloplerytide | $+$ |  |  |  |  | ＋ | $+$ | 4 | $\pm$ | $\pm$ |  | $\pm$ | $\pm$ | 4 | 4 | ＋ | ＋ | $\pm$ | 4 | $\pm$ | $\pm$ |
| Gomphiifue |  |  |  |  |  |  | $+$ |  |  |  |  |  | $+$ | ＋ |  |  | $+$ | $+$ |  | $+$ | $+$ |
| Aeshinidse | $+$ |  |  |  |  |  | $\pm$ |  |  | $+$ |  |  | $+$ | $+$ |  | $+$ | $\rightarrow$ |  |  |  |  |



APPENDIX 3.4 The Biological Monitoring Working Party (BMWP) index values of all samples held in the macro-invertebrate data-base.

| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number <br> of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH10 | St. Johns | FBA-1977 survey | 1977 | 31 | St. Johns | 03/08/77 | 2 | 18 | 162 | 5.400 | 30 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 09/05/84 | 1 | 7000 | 158 | 5.852 | 27 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 09/05/84 | 1 | 7001 | 119 | 5.174 | 23 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 26/07/84 | 2 | 7002 | 118 | 4.917 | 24 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 26/07/84 | 2 | 7004 | 128 | 5.565 | 23 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 17/10/84 | 3 | 7005 | 116 | 5.273 | 22 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 143 | Malthouse | 17/10/84 | 3 | 7006 | 126 | 5.727 | 22 |
|  |  | Les Ruse's Chironomid exuviae | 1977 | 127 | Buscot |  | 8 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1978 | 127 | Buscot |  | 8 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1986 | 127 | Buscot. |  | 8 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 127 | Buscot | 08/05/92 | 1 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 127 | Buscot | 24/06/92 | 2 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuriae | 1992 | 127 | Buscot | 24/08/92 | 2 | 2 | 2 | 2.000 | , |
|  |  | Thames WA/NRA/EA routine monitoring | 1980 | 38 | A417 Buscot | 22/10/80 | 3 | 300 | 126 | 5.040 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 38 | A417 Buscot | 10/07/87 | 2 | 3500 | 121 | 4.840 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 39 | Water Intake Buscot | 03/02/87 | 1 | 2100 | 112 | 5.091 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1988 | 39 | Water Intake Buscot | 01/08/88 | 2 | 4900 | 146 | 4.867 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 39 | Water Intake Buscot | 01/06/89 | 1 | 5400 | 103 | 5.150 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 39 | Water Intake Buscot | 24/04/90 | 3 | 6400 | 114 | 4.957 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 39 | Water Intake Buscot | 13/07/90 | 2 | 7200 | 114 | 4.560 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 39 | Water Intake Buscot | 05/09/90 | 2 | 7800 | 153 | 5.100 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 39 | Water Intake Buscot | 08/04/91 | 1 | 8700 | 102 | 5.100 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 39 | Water Intake Buscot | 05/08/91 | 2 | 9400 | 103 | 4.682 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 39 | Water Intake Buscot | 14/11/91 | 3 | 9800. | 98 | 4.667 | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 39 | Water Intake Buscot | 13/07/92 | 2 | 10400 | 112 | 4.667 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 39 | Water Intake Buscot | 07/10/92 | 3 | 10800 | 132 | 5.077 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1993 | 39 | Water Intake Buscot | 29/04/93 | 1 | 11300 | 122 | 4.880 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1994 | 39 | Water Intake Buscot | 09/08/94 | 2 | 12100 | 120 | 5.714 | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 39 | Water Intake Buscot | 04/05/95 | 1 | 13800 | 113 | 5.136 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 39 | Water Intake Buscot | 14/11/95 | 3 | 16500 | 129 | 4.778 | 27 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season <br> D | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH11 | Buscot | FBA-1977 survey | 1977 | 30 | Buscot | 03/08/77 | 2 | 17 | 96 | 4.174 | 23 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1190 | 43 | 4.300 | 10 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1191 | 60 | 4.615 | 13 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1192 | 37 | 3.364 | 11 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1193 | 66 | 5.077 | 13 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2. | 1194 | 34 | 3.400 | 10 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1195 | 40 | 3.636 | 11 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1196 | 17 | 3.400 | 5 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1197. | 58 | 4.833 | 12 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1198 | 28 | 3.500 | 8 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 1199 | 39 | 4.333 | - |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2. | 2190 | 30 | 4.286 | 7 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2191 | 29 | 4.143 | 7 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2192 | 11 | 2.750 | 4 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2193 | 29 | 5.800 | 5 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2194 | 43 | 4.778 | 9 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2195 | 17 | 4.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2196 | 46 | 4.182 | 11 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2197 | 27 | 4.500 | 6 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2198 | 35 | 5.000 | 7 |
|  |  | $\mathrm{FBA}-1977$ survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 2199 | 13 | 3.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2. | 3190 | 18 | 3.000 | 6 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3191 | 45 | 4.500 | 10 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3192 | 56. | 5.091 | 11 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3193 | 16 | 3.200 | 5 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3194 | 21 | 3.000 | 7 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08 77 | 2 | 3195 | 47 | 3.917 | 12 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3196 | 83 | 5.188 | 16 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3197 | 50 | 4.167 | 12 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3198 | 28 | 3.500 | 8 |
|  |  | FBA-1977 survey | 1977 | 37 | Buscot | 15/08/77 | 2 | 3199 | 13 | 4.333 | 3. |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season <br> ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH12 | Grafton | FBA-1977 survey | 1977 | 29 | Grafion | 03/08/77 | 2 | 16 | 123 | 5.125 | 24 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number <br> of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH13 | Radcot | FBA-1977 survey | ! 1977 | 28 | Radcot | 02/08/77 | 2 | 15 | 101 | 5.050 | 20 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH14 | Rushey | FBA-1977 survey | 1977 | 27 | Rushey | 02/08/77 | 2 | 14 | 100 | 4.762 | - 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 40 | 0.5km below Tadpole Brid | 10/08/95 | 2 | 14900 | 109 | 4.739 | 23 |


| Reach <br> ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH15 | Shifford | FBA-1977 survey | 1977 | 26 | Shifford | 02/08/77 | 2 | 13 | 136 | 4.857 | 28 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1200 | 46 | 4.182 | 11 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1201 | 56 | 5.091 | 11 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1202 | 77 | 4.529 | 17 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1203 | 85 | 4.474 | 19 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1204 | 48 | 3.692 | 13 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1205 | 37 | 4.111 | 9 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1206 | 78 | 4.105 | 19 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1207 | 30 | 3.750 | 8 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1208 | 26 | 3.714 | 7 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 1209 | 32 | 3.556 | 9 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2200 | 55 | 5.000 | 11 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2201 | 18 | 3.600 | 5 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2202 | 29 | 4.833 | 6 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2203 | 36 | 4.500 | 8 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2204 | 20 | 4.000 | 5 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2205 | 37 | 4.625 | 8. |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2206 | 88 | 5.867 | 15 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2207. | 29 | 4.833 | 6 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2208 | 9 | 3.000 | 3 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 2209 | 20 | 5,000 | 4 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3200 | 18 | 2.571 | 7 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3201 | 25 | 3.571 | 7 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3202 | 12 | $3.000{ }^{1}$ | 4 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3203 | 93 | 4.650 | 20 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3204 | 20 | 3.333 | 6. |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3205 | 28 | 4.000 | 7 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3206 | 26 | 4.333 | 6 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3207 | 31 | 3.875 | 8 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3208 | 52 | 4.727 | 11 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 3209 | 22 | 3.667 | 6 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 4201 | 12 | $3.000{ }^{\circ}$ | 4 |


| $\begin{array}{\|l} \text { Reach } \\ \text { ID } \\ \hline \end{array}$ | Reach name | Data source name | Year | Site ID | Site name | Sample <br> date | Season <br> ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH15 | Shifford | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 4202 | 29 | 3.625 | 8 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | 2 | 4203 | 29 | 4.833 | 6 |
|  |  | FBA-1977 survey | 1977 | 36 | Shifford | 22/08/77 | - 2 | 4204 | 23 | 3.286 | 7 |
|  |  | Thames WANRA/EA routine monitoring | 1987 | 41 | Newbridge | 10/07/87 | 2 | 3600 | 103 | 5.150 | 20 |
|  |  | Thames WANRA/EA routine monitoring | 1995 | 41 | Newbridge | 17/05/95 |  | 14400 | 91 | 4.789 | 19 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 41 | Newbridge | 02/10/95 |  | 15400 | 122 | 4.8801 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 |  | US Newbridge | 10/08/95 | 2 | 15000 | 122 | 4.880 | 25 |


| Reach <br> ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH16 | Northmoor | FBA-1977 survey | 1977 | 25 | Northmoor | 01/08/77 | - 2 | 12 | 151 | 5.207 | 29 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 09/05/84 | 1 | 7007 | 75 | 4.412 | 17 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 09/05/84 | 1 | 7008 | 114 | 5.429 | 21 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 26/07/84 | 2 | 7009 | 79 | 4.158 | 19 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 26/07/84 | 2 | 7010 | 131 | 5.240 | 25 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 17/10/84 | 3 | 7011 | 121 | 5.261 | 23 |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 144 | Bablock Hythe | 17/10/84 | 3 | 7012 | 136 | 5.440 | 25 |
|  |  | Les Ruse's Chironomid exuviae | 1977 | 131 | Newbridge |  | 8 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1978 | 131 | Newbridge |  | 8 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1986 | 131 | Newbridge |  | 8 | 1 | 2 | 2.000 | , |
|  |  | Les Ruse's Chitonomid exuviae | 1992 | 131 | Newbridge | 08/05/92 | 1 | 1. | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 131 | Newbridge | 24/06/92 | 2 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 131 | Newbridge | 24/08/92 | 2 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 131 | Newbridge | 12/05/94 | 1 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 131 | Newbridge | 13/06/94 | 2 | 31 | 2 | 2.000 | 1 |
|  |  | iles Ruse's Chironomid exuviae | 1994 | 131 | Newbridge | 12/09/94 | 3 | 1 | 2 | 2.000 | 1 |


| Reach <br> ID | Reach name | Data source name | Year | Site 1D | Site name | Sample date | Season <br> ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH17 | Pinkhill | FBA-1977 survey | 1977 | 24 | Pinkhill | 01/08/77 | 2 | 11 | 120 | 5.217 | 23 |
|  |  | Les Ruse's Chironomid exuviae | 1977 | 130 | Swinford |  | 8 | 1 | 2. | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1978 | 130 | Swinford |  | 8. | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1986 | 130 | Swinford |  | 8 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 130 | Swinford | 08/05/92 | 1 | 3 | 2 | 2.000 | $!$ |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 130 | Swinford | 24/06/92 | 2 | 1 | 2 | 2.000 | , |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 130 | Swinford | 24/08/92 | 2 | 2 | 2 | 2.000 | , |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 130 | Swinford | 12/05/94 | 1 | 1. | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 130 | Swinford | 13/06/94 | 2 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 130 | Swinford | 12/09/94 | 3 | 1 | $2 i$ | 2.000 | 1 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP. | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH18 | Eynsham | FBA-1977 survey | 1977 | 23 | Eynsham | 01/08/77 | 2 | 10 | 132 | 5.077 | - 26 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1210 | 31 | 3.875 | 8 |
|  |  | FBA+1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1211 | 55 | 3.667 | 15 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1212 | 39 | 4.333 | 9 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1213 | 43 | 4.300 | 10 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1214 | 45 | 5.000 | 9 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1215 | 28 | 3.500 | 8 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1216 | 41 | 4.100 | 10 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1217 | 42 | 4.200 | 10 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1218 | 75 | 4.412 | 17 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 1219 | 52 | 4.333 | 12 |
|  |  | FBA-1977 Survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2210 | 28 | 4.000 | , |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2211 | 78 | 5.200 | 15 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2212 | 16 | 4.000 | , |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2213 . | 24. | 3.429 | 7 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2214 | 30 | 4.286 | ? |
|  |  | FBA-1977 survey | 1977. | 35 | Eynsham | 24/08/77 | 2 | 2215 | 34 ! | 4.250 | 8 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2216 | 37 | 4.625 | , |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2217 | $26^{\prime}$ | $4.333^{\circ}$ | 6 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2218 | 36 | $4.00{ }^{\circ}$ | 9 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 2219 | 31 | $3.44{ }^{\text {' }}$ | 9 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2. | 3210 | 44 | 4.889 | 9 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 21 | 3211. | 32 | 4.000 | 8 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 3212 . | 24 | 4.000 | 6 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2. | 3213. | 25 | 3.571 | 7 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 3214 | 32 | 4.000 | 8 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2. | 3215 | 6 | 2.000 | 3 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2. | 3216 | 54 | 4.154 | 13 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 3217 | 3 | 1.500 | 2 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season <br> ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH18 | Eynsham | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 3218 | 20. | 3.333 | 6 |
|  |  | FBA-1977 survey | 1977 | 35 | Eynsham | 24/08/77 | 2 | 3219 | 39 | 4.333 | 9 |
|  |  | Oxford Structures Environmental Survey | 1992 | 118 | DS Swinford WTW | 30/06/92 | 2 | 5000 | 132 | 5.077 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1980 | 43 | Water Intake, Swinford | 06/11/80 | 3 | 400 | 121 | 5.261 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 43 | Water Intake, Swinford | 17/02/87 | 1 | 2200 | 115 | 4.600 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 43 | Water Intake, Swinford | 24/07/87 | 2 | 3900 | 124 | 4.593 | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1988 | 43 | Water Intake, Swinford | 15/06/88 | 2 | 4800 | 174 | 5.800 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 43 | Water Intake, Swinford | 10/10/89 | 3 | 6100 | 111 | 4.826 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 43 | Water Intake, Swinford | 10/04/95 | 1 | 13200 | 117 | 4.875 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 43 | Water Intake, Swinford | 09/10/95 | 3 | 15500 | 83 | 4.611 | 18 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH19 | King's | FBA-1977 survey | 1977 | 22 | King's | 25/08/77 | 2 | 9 | 124 | 4.960 | 25 |
|  |  | Oxford Structures Environmental Survey | 1992 | 44 | Trout Inn, Godstow | 30/06/92 | 2 | 5002 | 168 | 5.250 | 32 |
|  |  | Oxford Structures Environmental Survey | 1992. | 119 | West Mead | 30/06/92 | 2 | 5001 | 99 | 4.714 | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1980 | 44 | Trout Inn, Godstow | 01/12/80 | 3 | 800 | 128 | 5.120 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 44 | Trout Inn, Godstow | 30/04/90 | 1 | 6500 | 107 | 5.350 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 44 | Trout Inn, Godstow | 09/08/90 | 2 | 7600 | 181 | 5.171 | 35 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 44 | Trout Inn, Godstow | 08/10/90 | 3 | 8000 | 139 | 5.148 | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 44 | Trout Inn, Godstow | 08/04/91 | 1 | 8600 | 100. | 4.545 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 44 | Trout Inn, Godstow | 03/07/91 | 2 | 9100 | 93 | 4.429 | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 44 | Trout Inn, Godstow | 19/09/91 | 2 | 9500 | 153 | 6.120 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 44 | Trout Inn, Godstow | 26/05/92 | 1 | 10100 | 135 | 5.000 | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 44 | Trout Inn, Godstow | 28/07/92 | 2 | 10500 | 115 | 6.765 | 17 |
|  |  | Thames WA/NRA/EA routine monitoring | 1993 | 44 | Trout Inn, Godstow | 24/02/93 | 1 | 11200 | 136 | 5.667 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1994 | 44 | Trout Inn, Godstow | 28/07/94 | 2 | 12000 | 168 | 6.462 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 44 | Trout Inn, Godstow | 01/05/95 | 1 | 13700 | 148 | 5.481 | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 44 | Trout Inn, Godstow | 30/10/95 | 3 | 16100 | 122. | 5.545 | 22 |


| Reach <br> ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH20 | Godstow | FBA-1977 survey | 1977 | 21 | Godstow | 28/07/77 | 2 | 8 | 113 | 4.913 | 23 |
|  |  | Oxford Structures Environmental Survey | 1992 | 121 | Binsey | 30/06/92 | 2 | 5003 | 111 | 4,625 | 24 |




| Reach <br> ID <br> IH22 <br> TH22 | Reach name | Data source name | Year | Site ID | Site name | Sample <br> date | Season ID | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH22 | Iffley | Pond Action SWORDS Survey | 1992 | 48 | Top of Sandford Lock Cut | 09/07/92 | 1-2 | ID | 203 | 5.342 | - 38 |
|  |  | Thames WA/NRA/EA routine monitoring | 1982 | 48 | Top of Sandford Lock Cut | 18/02/82 | 2 | 1000 | 120 | 5.000 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1984 | 48 | Top of Sandford Lock Cut | 09/03/84 | 1 | 1400 | 116 | 5.273 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 48 | Top of Sandford Lock Cut | 22/03/87 | 1 | 2500 | 133 | 5.320 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1988 | 48 | Top of Sandford Lock Cut | 13/06/88 | 1 | 4600 | 156 | 5.200 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 48 | Top of Sandford Lock Cut | 08/05/89 | 1 | 5700 | 165 | 5.690 | 9 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 48 | Top of Sandford Lock Cut | 17/04/90 | 1 | 6300 | 143 | 5.107 | 28 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 48 | Top of Sandford Lock Cut | 20/08/90 | 2 | 7700 | 128. | 4.923 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 48 | Top of Sandford Lock Cut | 08/10/90 | 3 | 7900 | 189 | 5.400 | 35 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991. | 48 | Top of Sandford Lock Cut | 08/04/91 | 1 | 8500 | 107 | 4.652 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 48 | Top of Sandford Lock Cut | 03/07/91 | 2 | 9200 | 115 | 4.792 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 48 | Top of Sandford Lock Cut | 30/10/91 | 3 | 9700 | 119 | 5.409 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 48 | Top of Sandford Lock Cut | 26/05/92 | 1. | 10200 | 146 | $5.407{ }^{1}$ | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 48 | Top of Sandford Lock Cut | 11/08/92 | 2 | 10700 | 136 | 5.231 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1993 | 48 | Top of Sandford Lock Cut | 24/02/93 | 1 | 11100 | 89 | 4.450 | 20. |
|  |  | Thames WA/NRA/EA routine monitoring | 1994 | 48 | Top of Sandford Lock Cut | 08/11/94 | 3 | 12700 | 115 | 5.000 | 23. |


| Reach <br> ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH23 | Sandford | FBA-1977 survey | 1977 | 18 | Sandford | 27/07/77 | 2 | 5 | 93 | 4.429 | 21 |
|  |  | Oxford Structures Environmental Survey | 1992 | 124 | Radley College Boathouse | 30/06/92 | 2 | 5006 | 80 | 5.000 | 16 |
|  |  | Pond Action SWORDS Survey | 1992 | 49 | Abingdon Weir | 28/07/92 | 2 | 1 | 230 | 5.349 | 43 |
|  |  | Pond Action SWORDS Survey | 1992 | 139 | Near Lock Wood | 28/07/92 | 2 | 1 | 171 | 5.344 | 32 |
|  |  | Pond Action SWORDS Survey | 1992 | 140 | Radley | 20/07/92 | 2 | 1 | 193 | 5.361 | 36 |
|  |  | Pond Action SWORDS Survey | 1992 | 141 | Sandford Reach | 28/07/92 | 2 | 1 | 193 | 5.514 | 35 |
|  |  | Thames WA/NRA/EA routine monitoring | 1980 | 49 | Abingdon Weir | 20/11/80 | 3 | 600 | 122 | 4.692 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1982 | 49 | Abingdon Weir | 18/02/82 | 1 | 1100 | 129 | 5.160 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1983 | 49 | Abingdon Weir | 26/01/83 | 1 | 1200 | 118 | 4.720 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1984 | 49 | Abingdon Weir | 09/03/84 | 1 | 1300 | 113 | 4.708 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1986 | 50 | Sandford | 12/06/86 | 1 | 1900 | 50 | 4.545 | 11 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 49 | Abingdon Weir | 12/03/87 | 1 | 2400 | 103 | 4.682 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1988 | 49 | Abingdon Weir | 15/06/88 | 1 | 4700 | 136 | 5.037 | 27 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 49 | Abingdon Weir | 08/05/89 | 1 | 5600 | 143 | 4.767 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 49 | Abingdon Weir | 05/06/90 | 1 | 6700 | 119 | 4.577 | 26 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 49 | Abingdon Weir | 23/11/92 | 3 | 11000 | 89 | 4.944 | 18 |
|  |  | Thames WA/NRA/EA routine monitoring | 1993 | 49 | Abingdon Weir | 24/05/93 | 1 | 11400 | 117 | 5.087 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1994 | 49 | Abingdon Weir | 19/10/94 | 3 | 12400 | 113 | 5.136 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 49 | Abingdon Weir | 09/05/95 | 1 | 14000 | 82 | 5.125 | 16 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 49 | Abingdon Weir | 01/11/95 | 3 | 16200 | 118 | 4.720 | 25 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH24 | Abingdon | FBA-1977 survey | 1977 | 17. | Abingdon | 27/07/77 | 2 | 4 | 132 | 5.077 | 26 |
|  |  | Pond Action SWORDS Survey | 1992 | 136 | Sutton Pools | 03/08/92 | 2 | 1 | 190 | 5.588 | 34 |
|  |  | Pond Action SWORDS Survey | 1992 | 137. | Culham Reach | 20/07/92 | 2 | 1 | 200 | 5.263 | 38 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number <br> of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH25 | Culham | FBA-1977 survey | 1977 | 16 | Culham | 26/07/77 | 2 | 3 | 122 | 5.083 | 24 |
|  |  | Pond Action SWORDS Survey | 1992 | 134 | Long Wittenham | 03/08/92 | 2 | 1 | 187 | 5.054 | 37. |
|  |  | Pond Action SWORDS Survey | 1992 | 135 | Clifton Reach | 28/07/92 | 2 | 1 | 105 | 4.565 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 52 | Sutton Bridge, Culham | 07/07/87 | 2 | 3200 | 13 | 2.600 | 5 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 52 | Sutton Bridge, CuIham | 10/04/95 | 1 | 13300 | 110 | 5.000 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 52 | Sutton Bridge, Culham | 02/11/95 | 3. | 16400 | 91 | 4.550 | 20 |


| Reach <br> ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season <br> ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH26 | Clifton | FBA-1977 survey | 1977 | 15 | Clifton | 26/07/77 | 2 | 2 | 97 | 4.619 | 21 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1230 | 36 | 4.500 | 8 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1231 | 13 | 2.600 | 5 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1232 | 33 | 4.125 | 8 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1233 | 29 | 4.143 | 7 |
|  |  | FBA-1977 survey | 1977 | 33. | Clifton | 05/09/77 | 2 | 1234 | 50 | 4.167 | 12 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1235 | 75 | 4.412 | 17. |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1236 | 50 | 3.846 | 13 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1237 | 14 | 3.500 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1238 | 33 | 4.125 | 8 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 1239 | 27 | 3.857 | 7 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2230 | 1 | 1.000 | , |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2231 | 21 | $3.500:$ | 6 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifion | 05/09/77 | 2 | 2232 | 13 | 3.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifion | 05/09/77 | 2 | 2233 | 25 | 4.167 | 6 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2234 | 3 | $1.500^{\text {- }}$ | 2 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2235 | 10 | 3.333 | 3 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2236 | 10 | 3.333 | 3 |
|  |  | FBA-1977 survey | 1977 | 33. | Clifton | 05/09/77 | 2 | 2237 | 36 | 5.143 | 7 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2238 | 23 | 4.600 | 5 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 2239 | 17 | 4.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifion | 05/09/77 | 2 | 3230 | 14 | 3.500 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3231 | 6 | 2.000 | 3 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2. | 3232 | 1 | 1.000 | 1 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3233 | 23 | 4.600 | 5 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3234 | 28 | 4.667 | 6 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3235 | 13 | 3.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2. | 3236 | 13 | 3.250 | 4 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3237 | 3 | 1.500 | 2 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3238 | 3 | 1.500 | 2 |
|  |  | FBA-1977 survey | 1977 | 33 | Clifton | 05/09/77 | 2 | 3239 | 25 | 4.167 | 6 |
|  |  | Les Ruse's Chironomid exuviae | 1977 | 128 | Clifton Hampden |  | 8. | 2 | 2 | 2.000 | $1)$ |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH26 | Clifton | Les Ruse's Chironomid exuviae | 1978 | 128 | Clition Hampden |  | 8 | 3 | 2 | 2.000 | -1 |
|  |  | Les Ruse's Chironomid exuviae | 1986 | 128 | Clifton Hampden |  | 8 | 1 | 2 | 2.000 |  |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 128 | Clifton Hampden | 08/05/92 | 1 | 1 | 2 | 2.000 |  |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 128 | Clifton Hampden | 24/06/92 | 2 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 128 | Clifton Hampden | 24/08/92 | 2 | 2 | 2 | $2.000^{\circ}$ |  |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 128 | Clifton Hampden | 12/05/94 | 1 | 2 | 2 | 2.0001 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 128 | Clifton Hampden | 13/06/94 | 2 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 128 | Clifton Hampden | 12/09/94 | 3 | 1 | 2 | 2.000 | 1 |
|  |  | Pond Action SWORDS Survey | 1992 | 53 | Day's Lock | 03/08/92 | 2. | 1 | 124 | 5.167 | 24 |
|  |  | Pond Action SWORDS Survey | 1992 | 132 | Days Reach | 28/07/92 | 2 | 1. | 131 | 5.038 | 26 |
|  |  | Pond Action SWORDS Survey | 1992 | 133 | Clifton Bridge | 03/08/92 | 2 | 1 | 147 | 5.250 | 28 |
|  |  | Thames WANRRA/EA routine monitoring | 1980 | 51 | Clifton Hampden Bridge | 06/11/80 | 3 | 500 | 79 | 4.647 | 17 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 51 | Clifton Hampden Bridge | 24/07/87 | 2 | 3800 | 110 | 4.783 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 53 | Day's Lock | 05/05/89 | 1 | 5500 | 179 | 5.594 | 32 |
|  |  | Thames WANRRA/EA routine monitoring | 1990 | 53 | Day's Lock | 14/11/90 | 3 | 8200 | 111 | 4.826 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 53 | Day's Lock | 09/05/95 | 1 | 14100 | 108 | 4.909 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 53 | Day's Lock | 01/11/95 | 3 | 16300 | 81 | 4.500 | 18 |

1984145 Shillingford

| Reach ID | Reach name | Data source name | Year | Site ID | Site name |
| :---: | :---: | :---: | :---: | :---: | :---: |
| TH27 | Day's | FBA-1977 survey | 1977 | 14 | Day's |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |
|  |  | FBA-RIVPACS 1984 survey | 1984 | 145 | Shillingford |


| Reach ID | Reach name | Data source name | Year | Site [D | Site name | Sample date | Season ID | Sample ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH29 | Cleeve | Thames WA/NRA/EA routine monitoring | 1995 | 54 | South Stoke | 27/04/95 | 1 | 13600 | 93 | 4.650 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 54 | South Stoke | 13/09/95 | 2 | 15300 | 100 | 4.762 ! | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 56 | US Goring Weir | 27/04/95 | 1 | 13500 | 105 | 5.250 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 56 | US Goring Weir | 13/09/95 | 2 | 15200 | 112 | 5.091 | 22 |


| Reach ID | Reach name | Data source name | Year | Site ID | Site name | Sanuple date | Season $\mathrm{ID}$ | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH30) | Goring | Les Ruse's Chironomid exuviae | 1977 | 129 | Whitchurch |  | 8 | 1 | 2 | 2.000 | - Ta" |
|  |  | Les Ruse's Chironomid exuviae | 1978 | 129 | Whitchurch |  | 8 | 2 | 2 | 2.000 | , |
|  |  | Les Ruse's Chironomid exuviae | 1986 | 129 | Whitchurch |  | 8 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 129 | Whitchurch | 08/05/92 | 1 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 129 | Whitchurch | 24/06/92 | 2 | 1 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1992 | 129 | Whitchurch | 24/08/92 | 2 | 2 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 129 | Whitchurch | 12/05/94 | 1 | 2 | 2 | $2.000{ }^{\circ}$ | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 129 | Whitchurch | 13/06/94 | 2 | 3 | 2 | 2.000 | 1 |
|  |  | Les Ruse's Chironomid exuviae | 1994 | 129 | Whitchurch | 12/09/94 | 3 | 1 | 2 | 2.000 |  |
|  |  | Thames WA/NRA/EA routine monitoring | 1987 | 55 | Whitchurch Weir | 24/07/87 | 2 | 3700 | 104 | 5.200 | 20 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 55 | Whitchurch Weir | 23/05/90. | 1 | 6600 | 156 | 5.200 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 55 | Whitchurch Weir | 23/07/90 | 2 | 7500 | 83 | 4.882 | 17 |
|  |  | Thames WA/NRA/EA routine monitoring | 1990 | 55 | Whitchurch Weir | 29/11/90 | 3 | 8300 | 97 | 5.105 | 19 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 55 | Whitchurch Weir | 15/05/91 | 1 | 8900 | 123 | 5.348 | 23 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 55 | Whitchurch Weir | 03/07/91 | 2 | 9000 | 154 | 5.310 | 29 |
|  |  | Thames WA/NRA/EA routine monitoring | 1991 | 55 | Whitchurch Weir | 14/11/91 | 3 | 9900 | 92 | 4.600 | 20. |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 55 | Whitchurch Weir | 28/05/92 | 1 | 10300 | 116 | 5.524 | 21 |
|  |  | Thames WA/NRA/EA routine monitoring | 1992 | 55 | Whitchurch Weir | 11/08/92 | 2 | 10600 | 102 | 4.636 | 22 |
|  |  | Thames WA/NRA/EA routine monitoring | 1993 | 55 | Whitchurch Weir | 04/10/93 | 3 | 11600 | 140 | 5.600 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1994 | 55 | Whitchurch Weir | 28/07/94 | 2 | 11900 | 146 | 5.214 | 28 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 55 | Whitchurch Weir | 12/04/95 | 1 | 13400 | 104 | 5.474 | 19 |
|  |  | Thames WA/NRA/EA routine monitoring | 1995 | 55 | Whitchurch Weir | 16/10/95 | 3 | 15700 | 110 | 5.000 | 22 |


| Reach ID | React name | Data source name | Year | Site ID | Site name | Sample <br> date | $\begin{array}{\|l\|} \hline \text { Season } \\ \hline \end{array}$ | Sample <br> ID | BMWP | ASPT | Number of Taxa |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TH32 | Mapledurha | Thames WA/NRA/EA routine monitoring | 1980 | 57 | Caversham Weir | 31/12/80 | 3 | 900 | 119 | 4.760 | 25 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987. | 57 | Caversham Weir | 18/02/87 | 1 | 2300 | 85 | 4.474 | 19 |
|  |  | Thames WA/NRA/EA routine monitoring | 1987. | 57 | Caversham Weir | 12/08/87 | 2 | 4000 | 143 | 5.107 | 28 |
|  |  | Thames WA/NRA/EA routine monitoring | 1988 | 57 | Caversham Weir | 09/05/88 | 1 | 4400 | 159 | 5.300 | 30 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 57 | Cayersham Weir | 19/07/89 | 2 | 5900 | 113 | 4.708 | 24 |
|  |  | Thames WA/NRA/EA routine monitoring | 1989 | 57 | Caversham Weir | 02/10/89 | 3 | 6000 | 116 | 4.8331 | 24 | samples from three distinct zones: margins, mid-channel and vegetation.


| Species name | Margin | Mid-channel | Vegetation |
| :---: | :---: | :---: | :---: |
| Spongilidae | 20\% | 6.0\% | 1.9\% |
| Hydridae |  |  |  |
| Hydra sp. |  |  | 1.9\% |
| Planaria torva (Muller) | 2.0\% |  | 1.9\% |
| Polycelis tenuis (jjima) |  |  | 1.9\% |
| Dugesia tigrina (Girard) | $10.0 \%$ |  | 18.5\% |
| Dugesia polychroa group | 2.0\% |  | 3.7\% |
| Dendrocoelurn lacteum (Muller) | 2.0\% |  |  |
| Prostoma sp. |  |  | 1.9\% |
| Nematoda | 16.0\% | 60.0\% | 3.7\% |
| Ectoprocta | 4.0\% | 10.0\% | 9.3\% |
| Theodoxus fluviatilis (L.) | 6.0\% | 6.0\% | 9.3\% |
| Viviparus viviparus (L.) | $14.0 \%$ | 18.0\% | 9.3\% |
| Valvata sp. | 2.0\% | 2.0\% | 14.8\% |
| Valvata cristata Muiler |  |  |  |
| Valvata piscinalis (Muller) | 4.0\% |  |  |
| Potamopyrgus jenkinsi (Smith) | 68.0\% | 52.0\% | 64.8\% |
| Bithynia leachii (Sheppard) | 6.0\% | 2.0\% | 20.4\% |
| Bithynia tentaculata (L.) | 22.0\% | 6.0\% | 37.0\% |
| Physa sp. |  |  | 3.7\% |
| Physa fontinalis (L) |  |  | 16.7\% |
| Ptysa acuta group |  |  | 3.7\% |
| Lymnaea palustris (Muller) |  |  |  |
| Lymnaea peregra (Muller) | 16.0\% | 2.0\% | 37.0\% |
| Lymnaea stagnatis (L.) |  |  | 3.7\% |
| Planorbidae |  |  | 1.9\% |
| Planorbis carinaus Muller | 2.0\% |  | 3.7\% |
| Planorbis planorbis ( L ) | 2.0\% |  | 1.9\% |
| Anisus vortex (L.) |  |  | 1.9\% |
| Bathyomphalus contortus (L.) | 2.0\% |  |  |
| Gyraulus albus (Moller) | $4.0 \%$ | 4.0\% | 20.4\% |
| Armiger crista (L) |  |  |  |
| Ancylus fluviatilis Muller |  | 6.0\% | 16.7\% |
| Acrotoxus lacustris (L.) |  |  | 18.5\% |
| Succinea sp. |  |  |  |
| Unio sp. |  | 10.0\% |  |
| Unio pictorum (L.) | 18.0\% | 8.0\% |  |
| Unio tumidus Philipsson |  | 14.0\% |  |
| Anodonta sp. |  | 2.0\% |  |
| Anodonta anatina (L.) | 18.0\% | 20.0\% | 1.9\% |
| Anodonta cygnea (L.) |  | 2.0\% |  |
| Sphaerium corneum (L.) | 36.0\% | 16.0\% | 42.6\% |
| Sphaerium lacustre (Muiler) |  | $2.0 \%$ | 3.7\% |
| Sphaerium rivicola (Lamarck) | 18.0\% | 12.0\% | 1.9\% |
| Sphaerium lransversum (Say) |  |  |  |
| Pisidium amnicum (Muller) | 2.0\% | 2.0\% | 1.9\% |
| Pisidium casertanam (Poli) |  | 2.0\% | 5.6\% |
| Pisidium henslowanum (Sheppard) | 14.0\% | 8.0\% | $7.4 \%$ |
| Pisidium moitessierianum Paladilhe |  |  |  |
| Pisidium nitidum Jenyns |  | 2.0\% | 5.6\% |
| Pisidjum subtruncatum Maim |  | 2.0\% |  |
| Pisidium supinum Schmidt | $12.0 \%$ | 10.0\% | $5.6 \%$ |
| Lumbriculidae | 56.0\% | 52.0\% | 9.3\% |
| Stylodrilus sp. |  |  |  |
| Stylodrilus heringianus Claparede | 4.0\% | 14.0\% |  |
| ] umbriculus group |  |  |  |
| Enchytraeidae | 2.0\% |  |  |
| Chaetogaster sp. |  |  |  |
| Uncinais uncinata (Orsted) |  |  |  |
| Ophidonais serpentuna (Muller) |  | - | 5.6\% |
| Nais barbata Muller |  | . |  |


| Species name |  | channel | tation |
| :---: | :---: | :---: | :---: |
| Nais bresscheri Michaelsen |  |  |  |
| Nais pardalis Piguel |  |  |  |
| Nais simplex Piguet |  |  | 5.6\% |
| Nais communis group |  |  |  |
| Stylaria lacustris (L) | 6.0\% |  | 46.3\% |
| Tubificidae | 30.0\% | 52.0\% | 18.5\% |
| Tubifex ignotus (Stolc) | 60\% | 2.0\% |  |
| Tubifex tubifex (Muller) |  | 8.0\% |  |
| Limnodrilus cervix Brinkhursi | 16.0\% | 10.0\% | 5.6\% |
| Limnodrilus claparedeianus Ratzel |  |  |  |
| Limnodrilus hoffmeisteri Claparede | 42.0\% | 34.0\% | 9.3\% |
| Limnodrilus profundicola (Verril!) | $2.0 \%$ |  |  |
| Limmodrilus udekemianus Claparede | 20.0\% | 12.0\% | 3.7\% |
| Psammorycuides barbatus (Grube) | 36.0\% | 44.0\% | 11.1\% |
| Potamothrix hammoniensis (Michaelsen) |  | 6.0\% | 7.4\% |
| Potatnothix moldaviensis (Veidovsky \& Mrazek | 58.0\% | 40.0\% | 20.4\% |
| Aulodrilus pluriseta (Piguet) | 8.0\% |  | 22.2\% |
| Rhyacodrilus coccineus (Vejdovsky) |  | 2.0\% |  |
| Branchiura sowerbyi Beddard | 8.0\% | 2.0\% |  |
| Lumbricidae | 6.0\% |  |  |
| Piscicola geomerra (1.n) | 2.0\% | 4.0\% |  |
| Theromyzon lessulatum (Mulles) | 4.0\% |  | $11.1 \%$ |
| Hemiciepsis marginata (Muller) |  |  | 3.7\% |
| Glossiphonia complanala (LL) | 10.0\% | 14.0\% | 5.6\% |
| Glossiphonia heteroclita (L.) | 2.0\% | 4.0\% | 1.9\% |
| Helobdella stagnalis (L.) | 10.0\% | 16.0\% | 22.2\% |
| Erpobdella sp. - | 4.0\% | 4.0\% | 13.0\% |
| Erpobdella octoculata (2) | 6.0\% | 4.0\% | 7.4\% |
| Dina lineata (Muller) | 4.0\% |  |  |
| Trocheta subviridis Dutrochel | 2.0\% |  |  |
| Hydracatina | $72.0 \%$ | 84.0\% | 64.8\% |
| Argulus sp: |  |  | 7.4\% |
| Asellus aquaticus (L.) | 28.0\% | 26.0\% | 61.1\% |
| Asellus meridjanus Racovitza |  |  | 1.9\% |
| Corophium curvispinum Sars | 12.0\% | 4.0\% | 9.3\% |
| Crangonyx pseudogracilis Bousfield | 24.0\% | 10.0\% | 40.7\% |
| Gammaridae | 2.0\% |  | 1.9\% |
| Gammarus sp. --- |  |  |  |
| Gammarus pulex (L.) | 12.0\% | 2.0\% | 29.6\% |
| Baelidse |  | 2.0\% |  |
|  |  |  | 1.9\% |
| Baelis vermus Curtis | $4.0 \%$ |  | 24.1\% |
| Baetis scambus group | 4.0\% |  | 42.6\% |
| Centroptilum luteolum (Muller) | 36.0\% |  | 81.5\% |
| Cloeon dipterum (L.) | $6.0 \%$ |  | 25.9\% |
| Cloeon simile Eston |  |  | 18.5\% |
| Procloeon bifidum Bengtison | 18.0\% |  | 74.1\% |
| Heptagenia fuscogrisea (Retzius) |  |  |  |
| Habrophlebia fusca (Curtis) |  |  |  |
| Ephemers sp. |  |  |  |
| Ephemera danica Muller |  | 4.0\% |  |
| Ephernera vulgata I.- | 4.0\% | 2.0\% |  |
| Ephemerelta gnita (Poda) |  | 2.0\% | 9.3\% |
| Caenis sp........ | $16.0 \%$ | 24.0\% | 13.0\% |
| Caenis horariz (L.) | 2.0\% |  |  |
| Caenis tuctuosa group | 6.0\% | 24.0\% | 11.1\% |
| Nemouras p . |  | 2.0\% |  |
| Leuctra geniculata (Stephens) |  | 2.0\% |  |
| Platycnemis pennipes (Pallas) |  |  | $1.9 \%$ |
| Coenagriidae |  |  | 9.3\% |
| Ischnura elogans (Van der linden) |  |  |  |


| Species name | Margin | Mid-channe! | Vegetation |
| :---: | :---: | :---: | :---: |
| Enallagma cyathigerum (Charpentier) |  |  |  |
| Coenagrion puella group |  |  |  |
| Calopteryx sp. |  |  | 1.9\% |
| Calopleryx splendens (Harris) |  |  |  |
| Caloptery ${ }^{\text {virgo (L.) }}$ |  |  | 1.9\% |
| Gomphus vuigatissimus ( I. ) |  |  |  |
| Aphelocheirus aestivalis (Fabricius) |  | 2.0\% |  |
| Notonecta sp. |  |  | 1.9\% |
| Notonecta glauca L. |  |  |  |
| Notonecta maculata Fabricius |  |  |  |
| Corixidae | 4.0\% | 2.0\% | 13.0\% |
| Micronecta sp. | 6.0\% | 2.0\% | 1.9\% |
| Micronecta (Microncela) poweri (Douglas \& Scd | $2.0 \%$ |  |  |
| Sigara (Sigara) sp. | 2.0\% |  | 24.1\% |
| Sigara (Subsigara) distincta (Fieber) |  |  | 3.7\% |
| Sigara (Subsigara) falleni (Fieber) |  |  | 14.8\% |
| Sigara (Subsigaral fossarum (Leach) |  |  |  |
| Haliplidae | 26.0\% | 18.0\% | 25.9\% |
| Haliplus sp. |  |  |  |
| Haliplus fluviatilis Aube |  |  |  |
| Haliplus lineatocollis (Marsham) |  |  |  |
| Dytiscidae | 10.0\% |  | 37.0\% |
| Laccophilus sp. |  |  |  |
| Laccophitus hyalinus (Degeer) |  |  |  |
| Hydroporus sp. | 16.0\% | 8.0\% | 5.6\% |
| Potarmonectes depressus (Fabricius) |  |  |  |
| Stictotarsus duodecimpustulatus (Fabricius) |  |  |  |
| Platambus maculatus (L.) |  |  |  |
| Dytuscus sp. |  |  |  |
| Gyrinus distinctus Aube |  |  |  |
| Gyrinus urinator Itliger |  |  |  |
| Orectochilus villosus (Multer) |  |  |  |
| Hydrophilidae | 6.0\% | 4.0\% | 3.7\% |
| Helophorus sp. | $2.0 \%$ |  |  |
| Helophorus (Atractheiophorus) brevipalpis Bedel |  |  |  |
| Dryops sp. |  |  | 1.9\% |
| Elmis aenea (Muller) |  | $2.0 \%$ |  |
| Oulimnius sp. |  | 20\% |  |
| Oulimnius tuberculatus (Muller) | 20.0\% | 24.0\% | 11.1\% |
| Sialis lutaria (L.) | 6.0\% | 12.0\% | 5.6\% |
| Sialis nigripes Pictel |  | 8.0\% |  |
| Sisyra sp. |  |  |  |
| Hydroplilidae |  | 4.0\% | 7.4\% |
| Agraylea multipunctata Curtis | 2.0\% |  |  |
| Hydroptila sp. | 4.0\% | 4.0\% | 13.0\% |
| Oxyethira sp. | 2.0\% |  |  |
| lihytrichia sp. |  |  |  |
| Lype sp. |  |  |  |
| Tinodes waeneri (L.) | 2.0\% |  |  |
| Ecnomus tenellus (Rambur) |  |  |  |
|  |  |  |  |
| Cyrnus flavidus Mclachlan |  |  | 5.6\% |
| Cymus trimaculatus (Curtis) |  | 60.0\% | 13.0\% |
| Neureclipsis bimaculata (L) |  | 2.0\% | 9.3\% |
| Plectrocnemia sp. |  |  | 3.7\% |
| Plectrocnemia geniculata Mclachlan | 2,0\% |  | 1.9\% |
| Polycentropus sp. |  |  | 20.4\% |
| Polycentopus favomaculatus (Pictet) .. ....-- |  |  |  |
| Polycentropus inoratus (Curtis) |  |  |  |
| Hydropsyche pelliucidula (Curtis) |  |  |  |
| Phryganea sp. | 4.0\% | 4.0\% | 11.1\% |


| Species name | Margin | Mid-channel | Vegetation |
| :---: | :---: | :---: | :---: |
| Brachycentrus subnubilus Curtis |  |  |  |
| Lepidostoma hirtum (Fabricius) |  |  |  |
| Allogamus auricollis (Pictet) | 2.0\% |  |  |
| Hatesus sp. |  |  |  |
| Anabolia nervosa (Cuntis) | 4.0\% |  | 3.7\% |
| Limnephilus sp. |  |  |  |
| Limnephitus lunatus Curtis |  |  |  |
| Potamophylax group |  |  |  |
| Goera pilosa (Fabricius) | 2.0\% | 2.0\% |  |
| Silo sp. |  | 2.0\% |  |
| Molannidae |  | 2.0\% |  |
| Molanna angustata Curtis | 12.0\% | 4.0\% | 3.7\% |
| Leptoceridae | 4.0\% | 8.0\% | 3.7\% |
| Athipsodes aterrimus (Stephens) |  |  |  |
| Athripsodes cinereus (Curtis) |  |  |  |
| Ceraclea sp. |  |  | 1.9\% |
| Ceraclea annulicomis (Stephens) |  |  |  |
| Mystacides sp. | 12.0\% | 10.0\% | 9.3\% |
| Mystacides azurea (L) |  |  |  |
| Mystacider fongicornis (L.) |  |  | 1.9\% |
| Mystacides nigra (L) |  |  | 1.9\% |
| Oecetis lacustris (Pieter) |  |  |  |
| Oecetis ochracea (Curtis) |  |  |  |
| Diptera | 16.0\% | 8.0\% | 13.0\% |
| Tippulidae |  |  |  |
| Tipula sp. |  |  |  |
| Tipula (Yamatotipula) montium group |  |  |  |
| Ceratopogonidae |  |  |  |
| Simulium (Boophthora) erythrocephaturn (de Ge | - $2.0 \%$ |  | 19\% |
| Clinotanypus nervosus (Meigen) |  | 4.0\% | 1.9\% |
| Apsectrotarypus rrifascipennis (Zettersted) |  |  |  |
| Macropelopia sp. | 10.0\% | 18.0\% |  |
| Procladius sp. | 26.0\% | $38.0 \%$ | 14.8\% |
| Ablabesmyia sp. | 22.0\% | 24,0\% | 13.0\% |
| Thienermannimyis group | 4.0\% | 6.0\% | 3.7\% |
| Natarsia sp. | 2.0\% |  |  |
| Paramerina sp. | 4.0\% |  |  |
| Potthastia gaedii group |  |  |  |
| Pontbastia longimana group |  | 2.0\% |  |
| Prodiamesa olivacea (Meigen) | 32.0\% | 22.0\% | 11.15 |
|  |  |  |  |
| Brillia sp. |  |  | 3.7\% |
| Cricotopus sp. | 46.0\% | 4.0\% | 68.5\% |
| Cricotopus (Cricotopus) sp. |  |  | 3.7\% |
|  |  |  |  |
|  |  |  |  |
| Nanocladius sp. |  |  | 3.7\% |
| Orhocladius sp. | 16.0\% | 6.0\% | 18.5\% |
| Paracladius sp. |  |  |  |
| Psectrocladius sp. |  |  |  |
| Psectrocladius (Allopsectrocladius) obvius (Walker) |  |  |  |
|  |  |  |  |
| Chaetocladius sp. |  |  |  |
|  |  |  |  |
| Liminophyes sp........ | 6.0\% | 2.0\% | 5.6\% |
| Thienernanniella sp. |  |  | 14.8\% |
| Chionomus sp . | 34.8\% | 70.0\% | 5.6\% |
| Cladopeltna sp. |  |  |  |
| Cryptochironomus sp. | 30.0\% | 34.0\% | 9.3\% |
| Cryptotendipes sp. |  | 14,0\% |  |
| Dicrotendipes sp . |  |  |  |
| Dicrotendipes (Limnochironomus) sp. | 78.0\% | 38.0\% | 35.2\% |


| Species name | Margin | Mid-channel | Vegetation |
| :---: | :---: | :---: | :---: |
| Endochironomus sp. | 8.0\% | - $20 \%$ | $1.9 \%$ |
| Glyptotendipes sp. | 32.0\% | 14.0\% | 1.9\% |
| Itarnischia sp. | 2.0\% |  |  |
| Kiefferulus tendipediformis (Goetghebuer) |  |  |  |
| Microchironomus sp. |  | 6.0\% |  |
| Microtendipes sp. | 16.0\% | 12.0\% | 1.9\% |
| Parachironomus sp. | 2.0\% |  | 13.0\% |
| Paracladopelma sp. | 14.0\% | 20.0\% | 74\% |
| Paralauterborniella sp. |  |  |  |
| Paralauterborniella nigrohalteralis (Malloch) |  |  |  |
| Poratendipes sp. | $18.0 \%$ | 10.0\% | 11.1\% |
| Phaenopsectra sp. |  |  | 7.4\% |
| Polypedilum sp. |  |  |  |
| Polypedilum (Pentapedilum) sp. | 54.0\% | 34.0\% | 24.1\% |
| Polypedilum (Polypedilum) sp. | 8.0\% | 14.0\% |  |
| Stenochironomus sp. |  |  |  |
| Sticlochironomus sp. | 2.0\% | 2.0\% |  |
| Xenochironomus xenolabis (Kieffer) | 4.0\% | 2.0\% | 1.9\% |
| Cladotanytarsus sp. | 34.0\% | 44.0\% | $3.7 \%$ |
| Micropsectra sp. | 2.0\% |  | 1.9\% |
| Paratanytarsus sp. |  |  |  |
| Rheotanytarsus sp. | 12.0\% | 18.0\% | 29.6\% |
| Stempellinelia sp. |  |  |  |
| Tanytarsus sp. | 12.0\% | 10.0\% | 13.0\% |
| Micropsectra group |  |  |  |
| Chrysops sp. |  |  |  |
| Muscidae |  |  | . |
| Total no of samples | 50 | S0 | 54 |

APPENDIX 3.6 The frequency of occurrence of standard macro-invertebrate taxa in samples from eight distinct habitats: clay, silt, gravel, bedrock/concrete, detritus/organic matter, emergent vegetation, submerged vegetation and floating vegetation.

| Species name | HABITAT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\stackrel{\grave{U}}{\mathbf{U}}$ | 彦 | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{y}{5} \end{aligned}$ | 凹 0 0 0 0 0 0 0 0 0 0 | 墨 |  | 으․ 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 |  |
| Spongillidae | 3\％ |  | 4\％ |  | 20\％ |  | 3\％ |  |
| Hydra sp． |  |  |  |  |  |  |  | 6\％ |
| Planaria torva（Muller） |  |  | 2\％ |  |  |  | 3\％ |  |
| Polycelis tenuis（Ijima） |  |  |  |  |  |  | 3\％ |  |
| Dugesia tigrina（Girard） | 12\％ |  | 2\％ |  |  | 50\％ | 19\％ | 12\％ |
| Dugesia polychroa group |  |  | 2\％ |  |  |  | 6\％ |  |
| Dendrocoelum lacteum（Muller） |  |  | 2\％ |  |  |  |  |  |
| Prostoma sp． |  |  |  |  |  |  | 3\％ |  |
| Nematoda | 15\％ | 60\％ | 55\％ |  | 60\％ |  | 6\％ |  |
| Ectoprocta | 6\％ |  | 8\％ |  | 20\％ |  | 10\％ | 12\％ |
| Theodoxus fluviatilis（L．） | 3\％ | 20\％ | 6\％ |  |  |  | 10\％ | 6\％ |
| Viviparus viviparus（L．） | 15\％ | 20\％ | 14\％ |  | 20\％ |  | 10\％ | 12\％ |
| Valvata sp． |  |  | 2\％ |  |  |  | 16\％ | 12\％ |
| Valvata piscinalis（Muller） |  |  | 2\％ |  | 20\％ |  |  |  |
| Potamopyrgus jenkinsi（Smith） | 68\％ | 60\％ | 51\％ | 100\％ | 60\％ | 50\％ | 65\％ | 65\％ |
| Bithynia leachii（Sheppard） | 6\％ |  | 2\％ |  |  |  | 32\％ | 6\％ |
| Bithynia tentaculata（L．） | 18\％ |  | 4\％ | 67\％ | 20\％ |  | 42\％ | $41 \%$ |
| Physa sp． |  |  |  |  |  |  | 3\％ |  |
| Physa fontinalis（L．） |  |  |  |  |  | 50\％ | 10\％ | 24\％ |
| Physa acuta group |  |  |  |  |  |  |  | 12\％ |
| Lymnaea peregra（Muller） | 18\％ |  | 2\％ |  |  | 50\％ | 19\％ | 59\％ |
| Lymnaea stagnalis（L．） |  |  |  |  |  |  | 3\％ | 6\％ |
| Planorbidae |  |  |  |  |  |  | 3\％ |  |
| Planorbis carinatus Muller |  |  | 2\％ |  |  |  | 3\％ | 6\％ |
| Planorbis planorbis（L．） | 3\％ |  |  |  |  |  | 3\％ |  |
| Anisus vortex（L．） |  |  |  |  |  | 25\％ |  |  |
| Bathyomphalus contortus（L．） |  |  | 2\％ |  |  |  |  |  |
| Gyraulus albus（Muller） | $3 \%$ |  | 6\％ |  |  | 50\％ | 10\％ | 24\％ |
| Ancylus fluviatilis Muller |  |  | 6\％ |  |  | 25\％ | 19\％ | 12\％ |
| Acroloxus lacustris（L．） |  |  |  |  |  |  | 29\％ |  |
| Unio sp． |  |  | 10\％ |  |  |  |  |  |
| Unio pictorum（L．） | 9\％ | 40\％ | 12\％ |  | 20\％ |  |  |  |
| Unio tumidus Philipsson |  |  | 14\％ |  |  |  |  |  |
| Anodonta sp． |  |  | 2\％ |  |  |  |  |  |
| Anodonta anatina（L．） | 15\％ | 40\％ | 20\％ |  | 20\％ |  | 3\％ |  |
| Anodonta cygnea（L．） |  |  | 2\％ |  |  |  |  |  |
| Sphaerium corneum（L．） | 32\％ | 40\％ | 18\％ |  | 20\％ |  | 52\％ | 35\％ |
| Sphaerium lacustre（Muller） |  |  | 2\％ |  |  |  | 6\％ |  |
| Sphaerium rivicola（Lamarck） | 12\％ |  | 14\％ |  | 40\％ |  | 3\％ |  |
| Pisidium amnicum（Muller） | 3\％ |  | 2\％ |  |  |  | 3\％ |  |
| Pisidium casertanum（Poli） |  |  | 2\％ |  |  |  | 3\％ | 12\％ |
| Pisidium henslowanum（Sheppard） | 15\％ |  | 10\％ |  | 20\％ |  | 10\％ | 6\％ |
| Pisidium nitidum Jenyns |  | 20\％ |  |  |  |  | 3\％ | 6\％ |
| Pisidium subtruncatum Malm |  |  |  |  | 20\％ |  |  |  |
| Pisidium supinum Schmidt | 6\％ |  | 16\％ |  |  |  | 3\％ | 6\％ |


| Species name | HABITAT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\frac{\pi}{0}$ | $\stackrel{\rightharpoonup}{\theta}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \underset{5}{5} \end{aligned}$ | Bedrock/concrete |  |  |  |  |
| Lumbriculidae | 59\% |  | 57\% | 67\% | 20\% |  | 16\% |  |
| Stylodrilus heringianus Claparede | $3 \%$ | 20\% | 10\% | 33\% | 20\% |  |  |  |
| Enchytraeidae |  |  | $2 \%$ |  |  |  |  |  |
| Ophidonais serpentina (Muller) |  |  |  |  |  | 25\% | 6\% |  |
| Nais simplex Piguet |  |  |  |  |  | 25\% | 3\% | 6\% |
| Stylaria jacustris (L.) | 9\% |  |  |  |  | 50\% | 42\% | 47\% |
| Tubificidae | 29\% | 40\% | 49\% | 67\% | 40\%. | 25\% | 19\% | 18\% |
| Tubifex ignotus (Stolc) | 3\% |  | 4\% |  |  |  |  |  |
| Tubifex tubifex (Muller) |  |  | 6\% |  | 20\% |  |  |  |
| Limnodrilus cervix Brinkhurst | 9\% | 40\% | 12\% |  | 20\% |  | 6\% |  |
| Limnodrilus hoffmeisteri Claparede | 35\% | 60\% | 33\% |  | 60\% |  | 13\% |  |
| Limnodrilus profundicola (Verrill) |  |  | 2\% |  |  |  |  |  |
| Limnodrilus udekemianus Claparede | 9\% | 40\% | 14\% |  | 40\% |  | 3\% |  |
| Psammoryctides barbatus (Grube) | 38\% | 60\% | 39\% | 100\%. | 20\% | 25\% | 16\% |  |
| Potamothrix hammoniensis (Michaelsen) |  |  | 6\% |  |  |  | 13\% |  |
| Potamothrix moldaviensis (Vejdovsky \& Mrazek | 53\% | 40\% | $51 \%$ |  | 60\% |  | 23\% | 18\% |
| Aulodrilus pluriseta (Piguet) | 6\% | 20\% |  |  |  |  | 32\% | 6\% |
| Rhyacodrilus coccineus (Vejdovsky) |  |  | 2\% |  |  |  |  |  |
| Branchiura sowerbyi Beddard | 3\% |  | 6\% |  |  |  |  |  |
| Lumbricidae | 6\% |  | 2\% |  |  |  |  |  |
| Piscicola geometra (L.) | 3\% |  | 4\% |  |  |  |  |  |
| Theromyzon tessulatum (Muller) | 3\% |  | 2\% |  |  | 25\% | 3\% | 24\% |
| Hemiclepsis marginata (Muller) |  |  |  |  |  |  | 3\% | 6\% |
| Glossiphonia complanata (L.) | 12\% | 20\% | 10\% |  | 20\% |  |  | 12\% |
| Glossiphonia heteroclita (L.) |  |  | 2\% |  | 40\% |  | 3\% |  |
| Helobdella stagnalis (L.) | 12\% | 20\% | 14\% |  | 20\% | 25\% | 26\% | 18\% |
| Epobdella sp. |  | 40\% | 4\% |  |  |  | 10\% | 18\% |
| Erpobdella octoculata (L.) | 3\% |  | 6\% |  |  |  | 6\% | 12\% |
| Dina lineata (Muller) | 6\% |  |  |  |  |  |  |  |
| Trocheta subviridis Dutrochet | 3\% |  |  |  |  |  |  |  |
| Hydracarina | $79 \%$ | 80\% | 80\% | 67\% | 40\% | $75 \%$ | 68\% | 65\% |
| Argulus sp. |  |  |  |  |  |  | 3\% | 12\% |
| Asellus aquaticus (L.) | 26\% | 20\% | 24\% | 33\% | 60\% | 50\% | 77\% | 29\% |
| Asellus meridianus Racovitza |  |  |  |  |  |  | 3\% |  |
| Corophium curvispinum Sars | 18\% |  | 4\% |  |  | 25\% | 10\% | 6\% |
| Crangonyx pseudogracilis Bousfield | 26\% |  | 12\% |  | 20\% | 25\% | 58\% | 18\% |
| Gammaridae | 3\% |  |  |  |  |  | 3\% |  |
| Gammarus pulex (L.) | 15\% |  | 2\% |  | 20\% | 25\% | 29\% | 24\% |
| Baetidae |  |  | 2\% |  |  |  |  |  |
| Baetis rhodani (Pictet) |  |  |  |  |  |  | $3 \%$ |  |
| Baetis vernus Curtis | 6\% |  |  |  |  | 25\% | 23\% | 24\% |
| Baetis scambus group | 6\% |  |  |  |  |  | 45\% | 47\% |
| Centroptilum luteolum (Muller) | 47\% |  | 2\% |  |  | 75\% | 90\% | 65\% |
| Cloeon dipterum (L.) | 3\% | 40\% |  |  |  | 25\% | 19\% | 35\% |
| Cloeonsimite Eaton |  |  |  |  |  | 25\% | 6\% | 35\% |


| Species name | HABITAT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { む } \\ \hline \end{gathered}$ | $\overline{\vec{\omega}}$ | $\begin{aligned} & \overline{0} \\ & \stackrel{y}{0} \\ & \hline 0 . \end{aligned}$ | U <br> 0 <br> 0 <br> 0 <br> 0 <br> 0 | 喜 |  |  |  |
| Procloeon bifidum Bengtsson | 24\% |  |  |  |  | 75\% | 81\% | 59\% |
| Ephemera danica Muller |  |  | 4\% |  |  |  |  |  |
| Ephemera vulgata L. | 6\% |  | 2\% |  |  |  |  |  |
| Ephemerella ignita (Poda) |  |  | 2\% |  |  | 25\% | 6\% | 12\% |
| Caenis sp. | 15\% |  | 20\% | 100\% | 20\% | 25\% | 16\% | 6\% |
| Caenis horaria (L.) | 3\% |  |  |  |  |  |  |  |
| Caenis luctuosa group | 9\% | 20\% | 20\% |  |  |  | 10\% | 6\% |
| Nemoura sp. | 3\% |  |  |  |  |  |  |  |
| Leuctra geniculata (Stephens) |  |  | 2\% |  |  |  |  |  |
| Platycnemis pennipes (Pallas) |  |  |  |  |  |  | 3\% |  |
| Coenagriidae |  |  |  |  |  |  | 10\% | 6\% |
| Calopteryx virgo (L.) |  |  |  |  |  |  |  | 6\% |
| Aphelocheirus aestivalis (Fabricius) |  |  | 2\% |  |  |  |  |  |
| Notonecta sp. |  |  |  |  |  | 25\% |  |  |
| Corixidae | 3\% |  | 4\% |  |  |  | 19\% | 6\% |
| Micronecta sp. | 9\% | 20\% |  |  |  |  | 3\% |  |
| Micronecta (Micronecta) poweri (Douglas \& Sco | 3\% |  |  |  |  |  |  |  |
| Sigara (Sigara) sp. | 3\% |  |  |  |  | 25\% | 13\% | 41\% |
| Sigara (Subsigara) distincta (Fieber) |  |  |  |  |  |  |  | 12\% |
| Sigara (Subsigara) falleni (Fieber) |  |  |  |  |  |  | 10\% | 24\% |
| Haliplidae | 21\% | 20\% | 22\% |  | 20\% | 50\% | 23\% | 24\% |
| Dytiscidae | 9\% |  | 2\% |  | 20\% | 25\% | 45\% | 24\% |
| Hydroporus sp. | 18\% |  | 8\% |  | 20\% | 25\% | 3\% | 6\% |
| Hydrophilidae | 6\% |  | 6\% |  |  |  | 3\% | 6\% |
| Helophorus sp. |  | 20\% |  |  |  |  |  |  |
| Dryops sp. |  |  |  |  |  |  | 3\% |  |
| Elmis aenea (Muller) |  |  | 2\% |  |  |  |  |  |
| Oulimnius sp. |  |  | 2\% |  |  |  |  |  |
| Oulimnius tuberculatus (Muller) | 24\% |  | 20\% | 67\% | 20\% |  | 13\% | 12\% |
| Sialis lutaria (L.) | 12\% |  | 8\% |  | 20\% |  | 10\% |  |
| Sialis nigripes Pictet |  |  | 8\%. |  |  |  |  |  |
| Hydroptilidae |  |  | 4\% |  |  |  | 10\% |  |
| Agraylea multipunctata Curtis | 3\% |  |  |  |  |  |  |  |
| Hydroptila sp. | 6\% |  | 4\% |  |  | 25\% | 13\% | 12\% |
| Oxyethira sp. | 3\% |  |  |  |  |  |  |  |
| Polycentropodidae |  |  | 2\% |  |  |  |  |  |
| Cyrnus flavidus Mclachlan |  |  |  |  |  |  |  | 18\% |
| Cyrnus trimaculatus (Curtis) | 59\% | 40\% | 63\% |  |  | 25\% | 6\% | 18\% |
| Neureclipsis bimaculata (L.) |  |  |  | 33\%. |  | 25\% | 6\% | 6\% |
| Plectrocnemia sp. |  |  |  |  |  |  | 3\% | 6\% |
| Plectrocnemia geniculata Mclachlan | 3\% |  |  |  |  |  |  | 6\% |
| Polycentropus sp. | 6\% |  |  |  |  | 25\% | 10\% | 35\% |
| Phryganea sp. |  |  | 4\% |  | 20\% |  | 16\% | 6\% |
| Allogamus auricollis (Pictet) | 3\% |  |  |  |  |  |  |  |
| Anabolia nervosa (Curtis) | 3\% | 20\% |  |  |  |  | 6\% |  |


| Species name | HABITAT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\xrightarrow[0]{3}$ | $\stackrel{\rightharpoonup}{b}$ | 믄 |  |  |  |  |  |
| Goera pilosa (Fabricius) |  |  | 4\% |  |  |  |  |  |
| Silo sp. |  |  | 2\% |  |  |  |  |  |
| Molannidae |  |  | 2\% |  |  |  |  |  |
| Molanna angustata Curtis | 9\% |  | 6\% |  | 20\% |  | 3\% | 6\% |
| Leptoceridae | 6\% |  | 6\% |  | 20\% |  | 6\% |  |
| Ceraclea sp. |  |  |  |  |  |  | 3\% |  |
| Mystacides sp. | 21\% |  | 8\% |  |  |  | 16\% |  |
| Mystacides longicornis (L.) |  |  |  |  |  |  |  | 6\% |
| Mystacides nigra (L.) |  |  |  |  |  |  | 3\% |  |
| Diptera | 21\% | 40\% | 6\% |  |  | 25\% | 10\% | 12\% |
| Simulium (Boophthora) erythrocephalum (de Ged |  |  | 2\% |  |  |  |  | 6\% |
| Clinotanypus nervosus (Meigen) |  |  | 4\% |  |  |  | 3\% |  |
| Apsectrotanypus trifascipennis (Zetterstedt) | 3\% |  |  |  |  |  |  |  |
| Macropelopia sp. | 9\% |  | 16\% |  | 20\% |  |  |  |
| Procladius sp. | 21\% | 60\% | 39\% |  | 40\% | 25\% | 19\% |  |
| Ablabesmyia sp. | 18\% | 20\% | 24\% | 67\% |  | 25\% | 3\% | 24\% |
| Thienemannimyia group |  | 20\% | 6\% | 33\% |  |  |  | 12\% |
| Natarsia sp. | 3\% |  |  |  |  |  |  |  |
| Paramerina sp. | 3\% | 20\% |  |  |  |  |  |  |
| Potthastia longimana group |  |  | 2\% |  |  |  |  |  |
| Prodiamesa olivacea (Meigen) | 24\% |  | 31\% |  | 40\% |  | 13\% | 6\% |
| Brilia sp. |  |  |  |  |  |  | 3\% | 6\% |
| Cricotopus sp. | 50\% |  | 8\% |  | 60\% | 100\% | 68\% | 71\% |
| Cricotopus (Cricotopus) sp. |  |  |  |  |  |  | 3\% | 6\% |
| Eukiefferiella sp. |  |  |  |  |  |  | 3\% |  |
| Nanocladius sp. |  |  |  |  |  |  |  | 12\% |
| Orthocladius sp. | 21\% |  | 6\% | 33\% |  | 50\% | 6\% | 29\% |
| Synorthocladius semivirens (Kieffer) |  |  |  |  |  |  | 3\% |  |
| Corynoneura sp. |  |  |  |  |  |  |  | $12 \%$ |
| Limnophyes sp. | 3\% |  | 6\% |  |  | 25\% | 6\% |  |
| Thienemanniella sp. |  |  |  |  |  |  | 19\% | 12\% |
| Chironomus sp. | 29\% | 60\% | 67\% | 67\% | 60\% |  | 6\% |  |
| Cryptochironomus sp. | 21\% |  | 45\% |  | 20\% |  | 6\% | 12\% |
| Cryptotendipes sp. |  |  | 10\% |  | 40\% |  |  |  |
| Dicrotendipes (Limnochironomus) sp. | 88\% | 60\% | 41\% | 33\% | 40\% | 75\% | 29\% | 35\% |
| Endochironomus sp. |  | 20\% | 6\% |  |  |  | 3\% |  |
| Glyptotendipes sp. | 38\% |  | 16\% |  | 20\% |  | 3\% |  |
| Harnischia sp. | 3\% |  |  |  |  |  |  |  |
| Microchironomus sp. |  | 20\% | 4\% |  |  |  |  |  |
| Microtendipes sp. | 21\% |  | 14\% |  |  |  | 3\% |  |
| Parachironomus sp. | 3\% |  |  |  |  | 25\% | 10\% | 18\% |
| Paracladopelma sp. | 15\% | 20\% | 18\% |  | 40\% |  | 10\% |  |
| Paratendipes sp . | 15\% | 40\% | 12\% |  | 20\% |  | 10\% | $12 \%$ |
| Phaenopsectra sp. |  |  |  |  |  | 25\% | 6\% | 6\% |
| Polypedilum (Pentapedilum) sp. | 53\% | 60\% | 31\% |  | 80\% |  | 19\% | 29\% |


| Species name | HABITAT |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\stackrel{\lambda}{3}$ | 荡 | $\stackrel{\bar{\Delta}}{\stackrel{\rightharpoonup}{0}}$ |  | 总 | E 0 0 0 0 0 0 0 0 0 0 0 0 | 5 0 0 0 0 0 0 0 0 0 0 0 0 0 | 衰 |
| Polypedilum（Polypedilum）sp． | 9\％ |  | 16\％ |  |  |  |  |  |
| Stictochironomus sp． |  |  | 4\％ |  |  |  |  |  |
| Xenochironomus xenolabis（Kieffer） | 3\％ |  |  |  | 20\％ |  |  | 6\％ |
| Cladotanytarsus sp． | 38\％ |  | 47\％ | 33\％ |  |  | 3\％ |  |
| Micropsectra sp． | 3\％ |  |  |  |  |  |  | 6\％ |
| Rheotanytarsus sp． | 15\％ |  | 18\％ |  |  | 25\％ | 19\％ | 47\％ |
| Tanytarsus sp． | 12\％ | 40\％ | 8\％ |  | 20\％ |  | 19\％ |  |
|  |  |  |  |  |  |  |  |  |
| No of samples for habitat | 34 | 5 | 49 | 3 | 5 | 4 | 31 | 17 |

APPENDIX 3.7 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given emergent macrophyte was present, as the dominant or non-dominant species.

| Taxon | $\frac{\text { 틸 }}{2}$ | $\begin{aligned} & \text { 会 } \\ & 0 \end{aligned}$ |  |  | 苐 | 色 | $\frac{\stackrel{4}{4} 5}{\stackrel{y y y y}{3}}$ | ¢ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spongitlidae |  | 6\％ |  | 14\％ |  |  |  |  |
| Hydra sp： |  |  |  |  |  |  |  |  |
| Planaria torva（Muller） |  |  |  | 14\％ |  |  |  |  |
| Polycetis tenuis（Jima） |  | 6\％ |  | 14\％ |  |  |  |  |
| Dugesia tignina（Girard） | 33\％ | 6\％ | 40\％ | 14\％ | 33\％ | 100\％ |  | 50\％ |
| Dugesia polychroa group |  | 6\％ |  | 29\％ |  |  |  |  |
| Prostoma sp． |  | 6\％ |  | 14\％ |  |  |  |  |
| Nematoda |  | 6\％ |  |  |  |  |  |  |
| Ectoprocta |  | 13\％ |  |  | 33\％ |  |  |  |
| Theodoxus fuviatilis（L．） |  | 13\％ | 20\％ |  | 33\％ | 100\％ |  | 50\％ |
| Viviparus viviparus（L．）． |  | 13\％ |  |  |  |  | 100\％ |  |
| Valvata sp． | 33\％ | 6\％ |  | 14\％ | 67\％ |  | 100\％ |  |
| Potamopyrgus jenkinsi（Smith） | 67\％ | 63\％ | 40\％ | $71 \%$ | 100\％ | 100\％ | 100\％ | 100\％ |
| Bithynia leachii（Sheppard） |  | 38\％ | 20\％ | 29\％ | 67\％ | 100\％ |  | 100\％ |
| Bithynia tentacuiata（L） |  | 25\％ | 20\％ | 29\％ | 100\％ | 100\％ | 100\％ | 50\％ |
| Physa sp． |  | 6\％ |  | 14\％ |  |  |  |  |
| Physa fontinalis（L．） |  | 13\％ |  | 14\％ |  |  | 100\％， |  |
| Physa acuta group |  |  |  |  |  |  |  |  |
| Lymnaea peregra（Muller） |  | 19\％ | 20\％ | 14\％ | 33\％ |  | 100\％ |  |
| Lymnaea stagnalis（L．）． |  |  |  |  |  |  | 100\％ |  |
| Planorbidae |  | 6\％ |  | 14\％ |  |  |  |  |
| Planorbis carinatus Muller |  |  |  |  |  |  | 100\％ |  |
| Planorbis planorbis（L．） |  |  |  |  |  |  | 100\％ |  |
| Anisus vortex（L．） |  |  |  |  |  |  |  |  |
| Gyraulus albus（Muller） | 33\％ | 6\％ |  |  | 33\％ |  | 100\％ |  |
| Ancylus fluviatilis Muller | 67\％ | 25\％ | 20\％ |  | 33\％ |  |  |  |
| Acroloxus lacustris（L．） |  | 25\％ | 20\％ | 14\％ | 67\％ | 100\％ | 100\％ | 100\％ |
| Anodonta anatina（L．） |  |  | 20\％ |  |  |  |  |  |
| Sphaerium corneum（L．） |  | 50\％ | 40\％ | 57\％ | 67\％ | 100\％ |  | 100\％ |
| Sphaerium lacustre（Muller） |  | 6\％ |  | 29\％ |  |  |  |  |
| Sphaerium rivicola（Lamarck） |  |  |  |  |  |  |  |  |
| Pisidium amnicum（Muller） |  | 6\％ |  |  |  |  |  |  |
| Pisidium casertanum（Poli）． |  |  |  |  | 33\％ | 100\％ |  | 50\％ |
| Pisidium henslowanum（Sheppard） |  | 6\％ |  | 29\％ |  |  |  | 50\％ |
| Pisidium nitidum Jenyns |  |  |  |  |  |  |  |  |
| Pisidium supinum Schmidt |  |  | 20\％ |  |  |  |  |  |
| Lumbriculidae |  | 19\％ | 20\％ | 29\％ |  |  |  |  |
| Ophidonais serpentina（Muller） | 33\％ | 6\％ | 20\％ |  | 33\％ |  |  |  |
| Nais simplex Piguet |  | 6\％ |  |  |  |  |  |  |
| Stylaria lacustris（L．） | 67\％ | 50\％ | 60\％ | 14\％ | 67\％ |  |  |  |
| Tubificidae | 33\％ | 19\％ | 20\％ | 14\％ | 33\％ | 100\％ | 100\％ | 50\％ |
| Limnodrilus cervix Brinkhurst |  |  |  |  |  |  |  | 50\％ |
| Limnodrilus hoffmeisteri Claparede |  | 6\％ | 20\％ |  |  |  |  | 50\％ |
| Limnodrilus udekemianus Claparede |  |  |  |  |  |  |  | 50\％ |
| Psammoryctides barbatus（Grube） | 67\％ | 6\％ |  | 14\％ | 33\％ |  | 100\％ | 50\％ |
| Potamothix hammoniensis（Michaeisen） |  | 13\％ |  | 29\％ |  |  |  | 50\％ |
| Potarnothrix moldaviensis（Vejdovsky \＆Mrazek） |  | 13\％ | 20\％ | 29\％ |  |  | 100\％ | 50\％ |
| Autodrilus pluriseta（Piguec） | 33\％ | 19\％ | 20\％ | 43\％ | 33\％ | 100\％ | 100\％ | 100\％ |
| Theromyzon tessulatum（Muller） |  |  |  | 14\％ |  |  |  |  |
| Hemiclepsis marginata（Muller） |  | 6\％ |  | 14\％ |  |  |  |  |
| Glossiphonia complanata（L．） |  |  |  |  |  |  |  |  |
| Glossiphonia heteroclita（L．） |  |  |  |  | 33\％ |  |  |  |
| Helobdella stagnalis（L．） | 33\％ | 19\％ |  | 57\％ | 100\％ | 100\％ |  | 50\％ |
| Erpobdella sp． |  | 6\％ | 20\％ |  |  |  | 100\％ |  |
| Erpobdella octoculata（L．） |  |  |  |  | 67\％ | 100\％ |  | 50\％ |


| Taxon | $\stackrel{y}{2}$ | $\begin{gathered} \text { na } \\ \text { 苞 } \end{gathered}$ |  | E 른 © © | 急 | 鲁 | $\frac{\pi}{4}$ | 氙 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hydracarina | 100\％ | $63 \%$ | 40\％ | 71\％ | 100\％ | 100\％ | 100\％ | 100\％ |
| Argulus sp． |  |  |  |  |  |  | 100\％ |  |
| Asellus aquaticus（L．）． | 100\％ | 63\％ | 80\％ | 71\％ | 100\％ | 100\％ | 100\％ | 100\％ |
| Asellus meridianus Racovitza |  | 6\％ |  | 14\％ |  |  |  |  |
| Corophium curvispinum Sars |  | 6\％ | 20\％ | 14\％ |  |  |  |  |
| Crangonyx pseudogracilis Bousfield | 67\％ | 50\％ | 20\％ | 71\％ | 100\％ | 100\％ | 100\％ | 100\％ |
| Gammaridae |  | 6\％ |  | 14\％ |  |  |  |  |
| Gammarus pulex（L．） |  | 19\％ | 20\％ | 43\％ | 33\％ | 100\％ | 100\％ | 100\％ |
| Baetis rhodani（Pictet） |  | 6\％ |  |  |  |  |  |  |
| Baetis vernus Curtis． | 33\％ | 31\％ | 20\％ | 29\％ |  |  |  |  |
| Baetis scambus group |  | 69\％ | 20\％ | 57\％ |  |  |  |  |
| Centroptilum luteolum（Muller） | 100\％ | 81\％ | 100\％ | 100\％ | 100\％ | 100\％ | 100\％ | 100\％ |
| Cloeon dipterum（L） |  | 19\％ |  | 14\％ | 33\％ | 100\％ | 100\％ | 50\％ |
| Cloeon simile Eaton |  | 6\％ |  |  |  |  | 100\％ |  |
| Procloeon bifidum Bengisson | 100\％ | 75\％ | 80\％ | 71\％ | 100\％ | 100\％ | 100\％ | 50\％ |
| Ephemerella ignita（Poda） |  | 13\％ |  |  |  |  |  |  |
| Caenis sp． | 100\％ | 6\％ |  | 14\％ | 67\％ |  |  |  |
| Caenis luctuosa group |  | 19\％ |  | 14\％ |  |  |  |  |
| Platycnemis pennipes（Pallas） |  | 6\％ |  |  |  |  |  |  |
| Coenagridae |  | 6\％ |  | 14\％ | 67\％ |  |  |  |
| Calopteryx virgo（L） |  |  |  |  |  |  |  |  |
| Notonecta sp． |  |  |  |  |  |  |  |  |
| Corixidae |  | 25\％ | 20\％ |  |  |  |  |  |
| Micronectasp． |  |  |  | 14\％ |  |  |  |  |
| Sigara（Sigara）sp． |  | 6\％ |  | 14\％ | 33\％ | 100\％ | 100\％ | 50\％ |
| Sigara（Subsigara）distincta（Fieber） |  |  |  |  |  |  |  |  |
| Sigara（Subsigara）fatleni（Fieber） |  | 6\％ |  | 29\％ | 33\％ | 100\％ |  | 50\％ |
| Haliplidae |  | 19\％ | 20\％ | 57\％ |  |  | 100\％ |  |
| Dytiscidae |  | 50\％ |  | 71\％ | 67\％ | 100\％ | 100\％ | 50\％ |
| Hydroporus sp． |  | 6\％ |  |  |  |  |  |  |
| Hydrophilidae |  | 6\％ |  |  |  |  |  |  |
| Dryops sp． |  |  |  | 14\％ |  |  |  |  |
| Oulimnius tuberculatus（Muller） |  | $6 \%$ | 20\％ | 29\％ | 33\％ | 100\％ |  | 50\％ |
| Sialis lutaria（L．） |  |  |  | 14\％ |  |  |  | 50\％ |
| Hydroptilidae |  | 6\％ |  | 14\％ | 33\％ |  |  |  |
| Hydroptila sp． |  | 6\％ | 60\％ | 14\％ |  |  |  |  |
| Cyrnus flavidus Mclachlan |  |  |  |  |  |  |  |  |
| Cyrnus trimaculatus（Curtis） | 67\％ | 6\％ |  | 14\％ |  |  |  |  |
| Neureclipsis bimaculata（L．） | 33\％ | 6\％ | 20\％ |  | 33\％ |  |  |  |
| Plectrocnemia sp． |  | 6\％ |  | 14\％ |  |  |  |  |
| Plectrocnemia geniculata Mclachlan |  |  |  |  |  |  |  |  |
| Polycentropus sp． |  | 19\％ |  | 14\％ |  |  |  |  |
| Phryganea sp． |  | 13\％ | 20\％ | 29\％ |  |  | 100\％ |  |
| Anabolia nervosa（Curtis） |  | 6\％ |  | 14\％ | 33\％ |  |  |  |
| Molanna angustata Curtis |  |  |  |  |  |  | 100\％ |  |
| Leproceridae |  | $6 \%$ | 20\％ | 14\％ |  |  |  |  |
| Ceraclea sp． |  |  |  |  | 33\％ | 100\％ |  | 50\％ |
| Mystacides sp． | 33\％ | 6\％ | 20\％ |  |  |  | 100\％ |  |
| Mystacides longicomis（L．） |  |  |  |  |  |  |  |  |
| Mystacides nigra（L．） |  |  |  |  | 33\％ | 100\％ |  | 50\％ |
| Diptera |  | 13\％ |  | 29\％ |  |  |  |  |
| Simulium（Boophthora）erythrocephalum（de Geer） |  |  |  |  |  |  |  |  |
| Clinotanypus nervosus（Meigen） |  |  |  |  |  |  | 100\％ |  |
| Procladius sp． | 33\％ | 13\％ | 20\％ | 14\％ | 33\％ |  | 100\％ |  |
| Ablabesmyia sp． |  |  | 20\％ |  |  |  |  |  |


| Taxon | 哭 | 旁 |  | $\varepsilon$   <br>    <br> W   <br> in   | 翑 | 年 | $\stackrel{\text { U }}{2}$ | 比 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Thienemannimyia＿group |  |  |  |  |  |  |  |  |
| Prodiamesa olivacea（Meigen） |  | 6\％ |  |  |  |  | 100\％ | 50\％ |
| Brillia sp． | 33\％ |  |  |  |  |  |  |  |
| Cricotopus sp． | 100\％ | 75\％ | 40\％ | 57\％ | 100\％ | 100\％ | $100 \%$ | 50\％ |
| Cricotopus（Cricotopus）sp． |  | 6\％ |  | 14\％ |  |  |  |  |
| Eukiefferiella sp． |  | 6\％ |  |  |  |  |  |  |
| Nanocladius sp． |  |  |  |  |  |  |  |  |
| Orthocladius sp． |  | 6\％ |  |  | 33\％ | 100\％ |  | 50\％ |
| Synorthocladius semivirens（Kieffer） |  | 6\％ |  |  |  |  |  |  |
| Corynoneura sp． |  |  |  |  |  |  |  |  |
| Limnophyes sp． | 67\％ | 6\％ |  |  |  |  |  |  |
| Thienemanniella sp． | 33\％ | 31\％ | 20\％ | 29\％ |  |  |  |  |
| Chironomus sp． |  | 6\％ |  | 14\％ |  |  |  | 50\％ |
| Cryptochironomus sp． |  | 6\％ |  | 14\％ |  |  |  | 50\％ |
| Dicrotendipes（Limnochironomus）sp． | 100\％ | 13\％ | 40\％ |  | 33\％ | 100\％ | 100\％ | 50\％ |
| Endochironomus sp． |  |  |  |  |  |  | 100\％ |  |
| Glyptotendipes sp． |  |  |  |  | 33\％ | 100\％ |  | 50\％ |
| Microtendipes sp． |  | 6\％ |  | 14\％ |  |  |  |  |
| Parachironomus sp． |  | 19\％ |  | 14\％ |  |  |  |  |
| Paracladopelma sp． |  | 13\％ |  | 29\％ |  |  |  | 50\％ |
| Paratendipes sp． |  |  | 20\％ |  |  |  |  | 50\％ |
| Phaenopsectra sp． | 33\％ | 6\％ | 20\％ |  | 33\％ |  |  |  |
| Polypedilum（Pentapedilum）sp． |  | 19\％ | 20\％ | 29\％ |  |  |  | 50\％ |
| Xenochironomus xenolabis（Kieffer） |  |  |  |  |  |  |  |  |
| Cladotanytarsus sp． |  |  |  |  |  |  |  |  |
| Micropsectra sp． |  |  |  |  |  |  |  |  |
| Rheotanytarsus sp． |  | 19\％ | 20\％ | 29\％ | 33\％ |  | 100\％ |  |
| Tanytarsus sp． | 33\％ | 13\％ | 20\％ | 14\％ |  |  |  | 50\％ |
|  | 3 | 16 | 5 | 7 | 3 | 1 | 1 | 2 |

APPENDIX 3.8 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given submerged macrophyte was present, as the dominant or non-dominant species.

| Taxa | E E. 曷 ci |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Polycentropus sp. |  |  | 100\% | 100\% |
| Phryganea sp. |  |  |  |  |
| Anabolia nervosa (Curtis) |  |  |  |  |
| Molanna angustata Curtis |  |  |  |  |
| Leptoceridae |  |  |  |  |
| Ceraclea sp. |  |  |  |  |
| Mystacides sp. | 100\% |  |  |  |
| Mystacides Iongicornis (L.) |  |  |  |  |
| Mystacides nigra (L.) |  |  |  |  |
| Diptera |  | 50\% |  | 100\% |
| Simulium (Boophthora) erythrocephalum (de Geer) |  |  |  |  |
| Clinotanypus nervosus (Meigen) |  |  |  |  |
| Procladius sp. | 100\% |  |  |  |
| Ablabesmyia sp. |  |  | 100\% | 100\% |
| Thienemannimyia group |  |  |  | 100\% |
| Prodiamesa olivacea (Meigen) | 100\% |  |  |  |
| Brillia sp. |  |  |  |  |
| Cricotopus sp. |  | 100\% | 100\% | 100\% |
| Cricotopus (Cricotopus) sp. |  |  |  |  |
| Eukiefferiella sp. |  |  |  |  |
| Nanocladius sp. |  |  |  |  |
| Orthociadius sp. |  | 100\% |  | 100\% |
| Synorthocladius semivirens (Kieffer) |  |  |  |  |
| Corynoneura sp. |  |  |  |  |
| Limnophyes sp. |  |  |  |  |
| Thienemanniella sp. |  |  |  |  |
| Chironomus sp. |  |  |  |  |
| Cryptochironomus sp. |  |  |  |  |
| Dicrotendipes (Limnochironomus) sp. | 100\% | 50\% | $100 \%$ | 100\% |
| Endochironomus sp. |  |  |  |  |
| Glyptotendipes sp. |  |  |  |  |
| Microtendipes sp. |  |  |  |  |
| Parachironomus sp. |  |  | 100\% |  |
| Paracladopelma sp. |  |  |  |  |
| Paratendipes sp. | 100\% |  |  |  |
| Phaenopsectra sp. |  |  |  |  |
| Polypedilum (Pentapedihm) sp. | 100\% |  |  | 100\% |
| Xenochironomus xenolabis (Kieffer) |  |  |  |  |
| Cladotanytarsus sp. | 100\% |  |  |  |
| Micropsectra Sp. |  |  |  |  |
| Rheotanytarsus sp. |  |  | 100\% | 100\% |
| Tanytarsus sp. | 100\% |  |  |  |
|  | 1 | 2 | 1 | 1 |


| Taxa |  |  |  | $\begin{aligned} & \text { 펴 } \\ & \text { 응 } \\ & \text { ig } \\ & 0 \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Theromyzon tessulatum (Muller) |  | 50\% |  |  |
| Hemiclepsis marginata (Muller) |  |  |  | 100\% |
| Glossiphonia complanata (L.) |  |  |  |  |
| Glossiphonia heteroclita (L.) |  |  |  |  |
| Helobdella stagnalis (L.) | 100\% |  |  | 100\% |
| Erpobdella sp. |  |  |  | 100\% |
| Erpobdella octoculata (L.) |  |  |  |  |
| Hydracarina | 100\% | 50\% | 100\% | 100\% |
| Argulus sp. |  |  |  |  |
| Asellus aquaticus (L.) |  |  | 100\% | 100\% |
| Asellus meridianus Racovitza |  |  |  |  |
| Corophium curvispinum Sars |  | 50\% |  | 100\% |
| Crangonyx pseudogracilis Bousfield | 100\% |  |  |  |
| Gammaridae |  |  |  |  |
| Gammarus pulex (L.) |  |  | 100\% | 100\% |
| Baetis rhodani (Pictet) |  |  |  |  |
| Baetis vernus Curtis |  |  | 100\% |  |
| Baetis scambus group |  |  |  | 100\% |
| Centroptilum luteolum (Muller) | 100\% | 50\% | 100\% | 100\% |
| Cloeon dipterum (L.) |  |  | 100\% | 100\% |
| Cloeon simile Eaton |  |  | 100\% | 100\% |
| Procioeon bifidum Bengtsson | 100\% | 50\% | 100\% |  |
| Ephemerella ignita (Poda) |  |  | 100\% |  |
| Caenis sp. |  |  |  |  |
| Caenis luctuosa group |  |  |  |  |
| Platycnermis pennipes (Pallas) |  |  |  |  |
| Coenagriidae |  |  |  |  |
| Calopteryx virgo (L.) |  |  |  |  |
| Notonecta sp. |  | 50\% |  |  |
| Corixidae | 100\% |  |  |  |
| Micronecta sp. |  |  |  |  |
| Sigara (Sigara) sp. |  |  | 100\% | 100\% |
| Sigara (Subsigara) distincta (Fieber) |  |  |  |  |
| Sigara (Subsigara) falleni (Fieber) |  |  |  |  |
| Haliplidae |  | 50\% | 100\% | 100\% |
| Dytiscidae |  |  | 100\% | 100\% |
| Hydroporus sp. |  |  | 100\% | 100\% |
| Hydrophilidae |  |  |  |  |
| Dryops sp. |  |  |  |  |
| Oulimnius tuberculatus (Muller) |  |  |  | 100\% |
| Sialis lutaria (L.) | 100\% |  |  |  |
| Hydroptilidae |  |  |  |  |
| Hydroptila sp. |  |  | 100\% | 100\% |
| Cyrnus flavidus Mclachlan |  |  |  |  |
| Cyrnus trimaculatus (Curtis) |  |  |  |  |
| Neureclipsis bimaculata (L.) |  |  |  |  |
| Plectrocnemia sp. |  |  |  |  |
| Plectrocnemia geniculata Mclachlan |  |  |  |  |


| Taxa |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Spongillidae |  |  |  |  |
| Hydra sp. |  |  |  |  |
| Planaria torva (Muller) |  |  |  |  |
| Polycelis tenuis (Ijima) |  |  |  |  |
| Dugesia tigrina (Girard) | 100\% | 50\% |  | 100\% |
| Dugesia polychroa group |  |  |  |  |
| Prostoma sp. |  |  |  |  |
| Nematoda | 100\% |  |  |  |
| Ectoprocta |  |  |  |  |
| Theodoxus fluviatilis (L.) |  |  |  |  |
| Viviparus viviparus (L.) |  |  |  |  |
| Valvata sp. |  |  |  | 100\% |
| Potamopyrgus jenkinsi (Smith) | 100\% | 50\% |  | 100\% |
| Bithynia leachii (Sheppard) | 100\% |  |  |  |
| Bithynia tentaculata (L.) | 100\% |  |  | 100\% |
| Physa sp. |  |  |  |  |
| Physa fontinalis (L.) |  | 50\% | 100\% |  |
| Physa acuta group |  |  |  |  |
| Lymnaea peregra (Muller) |  | 50\% | 100\% | 100\% |
| Lymnaea stagnalis (L.) |  |  |  |  |
| Planorbidae |  |  |  |  |
| Planorbis carinatus Muller |  |  |  |  |
| Planorbis planorbis (L.) |  |  |  |  |
| Anisus vortex (L.) |  | 50\% |  |  |
| Gyraulus albus (Muller) |  | 50\% |  |  |
|  |  |  |  |  |
| Acroloxus lacustris (L.) |  |  |  |  |
| Anodonta anatina (L.) |  |  |  |  |
| Sphaerium corneum (L.) | 100\% |  |  | 100\% |
|  |  |  |  |  |
| Sphaerium rivicola (Lamarck) | 100\% |  |  |  |
|  |  |  |  |  |
| Pisidium casertanum (Poli) |  |  |  |  |
| Pisidium henslowanum (Sheppard) |  |  |  |  |
| Pisidium nitidum Jenyns |  |  |  |  |
| Pisidium supinum Schmidt |  |  |  |  |
| Lumbriculidae |  |  |  |  |
| Ophidonais serpentina (Muller) |  |  |  |  |
| Nais simplex Piguet |  |  | 100\% |  |
| Stylaria lacustris (L.) |  |  | 100\% | 100\% |
| Tubificidae _- |  |  |  |  |
| Limnodrilus cervix Brinkhurst | 100\% |  |  |  |
| Limnodrilus hoffmeisteri Claparede | 100\% |  |  |  |
| Limnodrilus udekemianus Claparede |  |  |  |  |
| Psammoryctides barbatus (Grube) |  |  |  |  |
| Potamothrix hammoniensis (Michaelsen) | 100\% |  |  |  |
| Potamothrix moldaviensis (Vejdovsky \& Mrazek) | 100\% |  |  |  |
| Aulodrilus pluriseta (Piguet) | 100\% |  |  |  |

APPENDIX 3.9 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given floating macrophyte was present, as the dominant or non-dominant species.

| Taxon | $\begin{aligned} & \text { 䳐 } \\ & \text { 呂 } \\ & \underset{z}{z} \\ & \hline \end{aligned}$ | 0 0 0 0 0 0 0 0 0 0 | 哭 |
| :---: | :---: | :---: | :---: |
| Spongillidae |  |  |  |
| Hydra sp． | 10\％ |  |  |
| Planaria torva（Mulier） |  |  |  |
| Polycelis tenuis（Ijima） |  |  |  |
| Dugesia tigrina（Girard） | 10\％ | 13\％ |  |
| Dugesia polychroa group |  |  |  |
| Prostoma sp． |  |  |  |
| Nematoda |  |  |  |
| Ectoprocta | 10\％ | 13\％ |  |
| Theodoxus fluviatilis（L．） | 10\％ |  |  |
| Viviparus viviparus（L．） | 10\％ | 13\％ |  |
| Valvata sp． |  | 25\％ |  |
| Potamopyrgus jenkinsi（Smith） | 50\％ | 75\％ | 100\％ |
| Bithynia leachii（Sheppard） | 10\％ |  |  |
| Bithynia tentaculata（L．） | 50\％ | 38\％ |  |
| Physa sp． |  |  |  |
| Physa fontinalis（L．） | 10\％ | 38\％ |  |
| Physa acuta group |  | 25\％ | 100\％ |
| Lymnaea peregra（Muller） | 70\％ | 38\％ | 100\％ |
| Lymnaea stagnalis（L．） | 10\％ |  |  |
| Planorbidae |  |  |  |
| Planorbis carinatus Muller |  | 13\％ |  |
| Planorbis planorbis（L．） |  |  |  |
| Anisus vortex（L．） |  |  |  |
| Gyraulus albus（Muller） | 20\％ | 25\％ |  |
| Ancylus fluviatilis Muller | 20\％ |  |  |
| Acroloxus lacustris（L．） |  |  |  |
| Anodonta anatina（L．） |  |  |  |
| Sphaerium coneum（L．） | 10\％ | 75\％ | 100\％ |
| Sphaerium lacustre（Muller） |  |  |  |
| Sphaerium rivicola（Lamarck） |  |  |  |
| Pisidium amnicum（Muller） |  |  |  |
| Pisidium casertanum（Poli） |  | 25\％ |  |
| Pisidium henslowanum（Sheppard） |  | 13\％ |  |
| Pisidium nitidum Jenyns |  | 25\％ |  |
| Pisidium supinum Schmidt |  | 13\％ |  |
| Lumbriculidae |  |  |  |
| Ophidonais serpentina（Muller） |  |  |  |
| Nais simplex Piguet | 10\％ |  |  |
| Stylaria lacustris（L．） | 50\％ | 38\％ |  |
| Tubificidae | 10\％ | 25\％ | 100\％ |
| Limnodrilus cervix Brinkhurst |  |  |  |
| Limnodrilus hoffmeisteri Claparede |  |  |  |
| Limnodrilus udekemianus Claparede |  |  |  |
| Psammoryctides barbatus（Grube） |  |  |  |
| Potamothrix hammoniensis（Michaelsen） |  |  |  |
| Potamothrix moldaviensis（Vejdovsky \＆Mrazek |  | 38\％ | 100\％ |
| Aulodrilus pluriseta（Piguet） |  | 13\％ |  |


| Taxon |  |  | 皆 |
| :---: | :---: | :---: | :---: |
| Theromyzon tessulatum (Muller) | 40\% |  |  |
| Hemiclepsis marginata (Muller) |  | 13\% |  |
| Glossiphonia complanata (L.) |  | 25\% | 100\% |
| Glossiphonia heteroclita (L.) |  |  |  |
| Helobdella stagnalis (L.) | 10\% | 25\% | 100\% |
| Erpobdella sp. | 20\% | 13\% |  |
| Erpobdella octoculata (L.) | 20\% |  |  |
| Hydracarina | 60\% | 63\% | 100\% |
| Argulus sp. |  | 25\% | 100\% |
| Asellus aquaticus (L.) | 10\% | 63\% | 100\% |
| Asellus meridianus Racovitza |  |  |  |
| Corophium curvispinum Sars |  | 13\% |  |
| Crangonyx pseudogracilis Bousfield | 20\% | 13\% |  |
| Gammaridae |  |  |  |
| Gammarus pulex (L.) | 10\% | 50\% |  |
| Baetis rhodani (Pictet) |  |  |  |
| Baetis vernus Curtis | 10\% | 38\% |  |
| Baetis scambus group | 50\% | 50\% |  |
| Centroptilum luteolum (Muller) | 50\% | 88\% | 100\% |
| Cloeon dipterum (L) |  | 88\% |  |
| Ctoeon simile Eaton | 10\% | 63\% |  |
| Procloeon bifidum Bengtsson | 70\% | 50\% | 100\% |
| Ephemerella ignita (Poda) | 10\% | 13\% |  |
| Caenis sp. | 10\% |  |  |
| Caenis luctuosa group |  | 13\% |  |
| Platycnemis pennipes (Pallas) |  |  |  |
| Coenagritdae | 10\% |  |  |
| Calopteryx virgo (L.) | 10\% |  |  |
| Notonecta sp. |  |  |  |
| Corixidae | 10\% |  |  |
| Micronecta sp. |  |  |  |
| Sigara (Sigara) sp. |  | 100\% | 100\% |
| Sigara (Subsigara) distincta (Fieber) |  | 25\% |  |
| Sigara (Subsigara) falleni (Fieber) |  | 50\% | 100\% |
| Haliplidae |  | 50\% |  |
| Dytiscidae |  | 63\% | 100\% |
| Hydroponts sp. |  | 13\% |  |
| Hydrophilidae |  | 13\% |  |
| Dryops sp. |  |  |  |
| Oulimnius tuberculatus (Muller) | 10\% | 13\% |  |
| Sialis lutaria (L.) |  |  |  |
| Hydroptilidae |  | 13\% |  |
| Hydroptila sp. | 10\% | 13\% |  |
| Cyrnus flavidus Mclachlan | 30\% |  |  |
| Cyrnus trimaculatus (Curtis) | 30\% |  |  |
| Neureclipsis bimaculata (L.) | 10\% |  |  |
| Plectrocnemia sp. | 10\% |  |  |
| Plectrocnemia geniculata Mclachlan | 10\% |  |  |


| Taxon | $\begin{aligned} & \frac{\pi}{c} \\ & \stackrel{\rightharpoonup}{2} \\ & \vec{Z} \end{aligned}$ |  | 坒 |
| :---: | :---: | :---: | :---: |
| Polycentropus sp. | 50\% | 13\% |  |
| Phryganea sp. | 10\% |  |  |
| Anabolia nervosa (Curtis) |  |  |  |
| Molanna angustata Curtis |  | 13\% | 100\% |
| Leptoceridae |  |  |  |
| Ceraclea sp. |  |  |  |
| Mystacides sp. |  |  |  |
| Mystacides longicornis (L.) |  | 13\% |  |
| Mystacides nigra (L.) |  |  |  |
| Diptera |  | 38\% |  |
| Simulium (Boophthora) erythrocephalum (de Ge | 10\% |  |  |
| Clinotanypus nervosus (Meigen) |  |  |  |
| Procladius sp. |  |  |  |
| Ablabesmyia sp. | 20\% | 25\% |  |
| Thienemannimyia group | 10\% | 13\% |  |
| Prodiamesa olivacea (Meigen) |  | 13\% | 100\% |
| Brillia sp. | 10\% |  |  |
| Cricotopus sp. | 50\% | 100\% | 100\% |
| Cricotopus (Cricotopus) sp. | 10\% |  |  |
| Eukiefferiella sp. |  |  |  |
| Nanocladius sp. | 20\% |  |  |
| Orthocladius sp. | 10\% | 50\% | 100\% |
| Synorthocladius semivirens (Kieffer) |  |  |  |
| Corynoneura sp. | 20\% |  |  |
| Limnophyes sp. |  |  |  |
| Thienemanniella sp. | 20\% |  |  |
| Chironomus sp. |  |  |  |
| Cryptochironomus sp. | 10\% | 13\% | 100\% |
| Dicrotendipes (Limnochironomus) sp. | 40\% | 38\% |  |
| Endochironomus sp. |  |  |  |
| Glyptotendipes sp. |  |  |  |
| Microtendipes sp. |  |  |  |
| Parachironomus sp. | 30\% |  |  |
| Paracladopelma sp. |  |  |  |
| Paratendipes sp. |  | 25\% |  |
| Phaenopsectra sp. | 10\% |  |  |
| Polypedilum (Pentapedilum) sp. | 10\% | 50\% | 100\% |
| Xenochironornus xenolabis (Kieffer) | 10\% |  |  |
| Cladotanytarsus sp. |  |  |  |
| Micropsectra sp. |  | 13\% |  |
| Rheotanytarsus sp. | 40\% | 50\% |  |
| Tanytarsus sp. |  |  |  |
|  | 10 | 8 | 1 |

APPENDIX 3.10 The occurrence of each taxon with national conservation status in the macro-invertebrate data-base.

Notable taxa recorded within database
03-1pr-97

Species Gyraulus acronicus (Ferussac) has notability code(s) RDB 2 and is found in the following samples:

| Reach ID: TH23 | Reach name: Sandford |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Site ID | Site name | Easting | Northing | Sample date | Scason ID | Sample ID |
| 49 | Abingdon Weir | 450400 | 197200 | $28 / 07 / 92$ | 2 | 1 |
| 140 | Radley | 453800 | 199000 | $20 / 07 / 92$ | 2 | 1 |

Species Pisidium moitessierianum Paladilhe has notability code(s) NB and is found in the following samples:

| Reach ID: THl6 | Reach name: Northmoor |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| I44 | Bablock Hythe | 443500 | 204200 | $09 / 05 / 84$ | 1 | 7007 |
| 144 | Bablock Hythe | 443500 | 204200 | $09 / 05 / 84$ | 1 | 7008 |
| 144 | Bablock Hythe | 443500 | 204200 | $26 / 07 / 84$ | 2 | 7010 |
| 144 | Bablock Hythe | 443500 | 204200 | $17 / 10 / 84$ | 3 | 7012 |
| Reach ID: TH27 | Reach name: Day's |  |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| I45 | Shillingford | 459000 | 193200 | $09 / 05 / 84$ | 1 | 7013 |
| I45 | Shillingford | 459000 | 193200 | $17 / 10 / 84$ | 3 | 7019 |

## Notable taxa recorded within database

03-Apr-97

Species Pisidium supinum Schmidt has notability code(s) NB and is found in the following samples:

| Reach ID: TH10 |  |
| :--- | :--- |
| Site ID | Site name |
| 143 | Matthouse |
| 143 | Malthouse |
| 143 | Malthouse |
| 143 | Malthouse |
| Reach ID: THl 1 |  |
| Site ID | Site name |
| 37 | Buscot |
| 37 | Buscot |
| 37 | Buscot |
| Reach ID: | TH14 |
| Site ID | Site name |
| 27 | Rushey |
| Reach ID: | TH15 |
| Site ID | Site name |
| 26 | Shifford |
| 36 | Shifford |
| Reach ID: | TH16 |

Site ID Site name

144 Bablock Hythe
144 Bablock Hythe
144 Bablock Hythe
Reach name: St. Johns

| Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- |
| 422500 | 198400 | $09 / 05 / 84$ | 1 | 7001 |
| 422500 | 198400 | $26 / 07 / 84$ | 2 | 7004 |
| 422500 | 198400 | $17 / 10 / 84$ | 3 | 7005 |
| 422500 | 198400 | $17 / 10 / 84$ | 3 | 7006 |

Reach name: Buscot
Easting Northing Sample date Season ID Sample ID

| $15 / 08 / 77$ | 2 | 2199 |
| :--- | :--- | :--- |
| $15 / 08 / 77$ | 2 | 2196 |

Reach name: Rushey
Easting Northing Sample date Season ID Sample ID 02/08/77 214
Reach name: Shifford
Easting Northing Sample date Season ID Sample ID
02/08/77 $2 \quad 13$
22/08/77 21204

Reach name: Northmoor

| Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- |
| 443500 | 204200 | $09 / 05 / 84$ | 1 | $70(18$ |
| 443500 | 204200 | $26 / 07 / 84$ | 2 | 7010 |
| 443500 | 204200 | $17 / 10 / 84$ | 3 | 7011 |

Reach ID: TH18

| Site ID | Site name |
| :--- | :--- |
| 35 | Eynsham |
| 35 | Eynsham |
| 35 | Eynsham |

Reach ID: TH19
Site ID Site name
22 King's
Reach ID: TH20
Site ID Site name
21 Godstow
Reach ID: TH22
Site ID Site name

| 34 | lffley |
| :--- | :--- |
| 34 | Iffley |
| 34 | Iffley |
| 34 | Iffley |
| 34 | Iffley |
| 34 | Iffley |
| 34 | [ffley |

Reach name: Eynsham

| Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- |
|  | $24 / 08 / 77$ | 2 | 1213 |  |
|  | $24 / 08 / 77$ | 2 | 3214 |  |
|  | $24 / 08 / 77$ | 2 | 2219 |  |

Reach name: King's
Easting Northing Sample date Season ID Sample ID 25/08/77 29
Reach name: Godstow
Easting Northing Sample date Season ID Sample 1D $\begin{array}{lll}28 / 07 / 77 & 2 & 8\end{array}$
Reach name: [fflcy

| Easting | Northing | Sample date | Season ID |
| :---: | :--- | :--- | :--- |
|  | Sample ID |  |  |
|  | $30 / 08 / 77$ | 2 | 3220 |
| $30 / 08 / 77$ | 2 | 1229 |  |
| $30 / 08 / 77$ | 2 | 3222 |  |
|  | $30 / 08 / 77$ | 2 | 3224 |
| $30 / 08 / 77$ | 2 | 3229 |  |
| $30 / 08 / 77$ | 2 | 2225 |  |
|  | $30 / 08 / 77$ | 2 | 3228 |

Notable taxa recorded within database
$03-4 p r .97$

| Reach ID: | TH23 | Reach name: Sandford |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 18 | Sandford |  |  | 27/07/77 | 2 | 5 |
| Reach [D: | TH27 | Reach name: Day's |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 145 | Shillingford | 459000 | 193200 | 09/05/84 | 1 | 7013 |
| 145 | Shillingford | 459000 | 193200 | 09/05/84 | 1 | 7015 |
| 145 | Shillingford | 459000 | 193200 | 26/07/84 | 2 | 7016 |
| 145 | Shillingford | 459000 | 193200 | 17/10/84 | 3 | 7018 |

Species Heptagenia fuscogrisea (Retzius) has notability code(s) N and is found in the following samples:
Reach ID: TH27 Reach name: Day's

| Site ID | Site name |
| :--- | :--- |
| 145 | Shillingford |


| Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- |
| 459000 | 193200 | $09 / 05 / 84$ | 1 | 7013 |

Species Gomphus vulgatissimus (L.) has notability code(s) N and is found in the following samples:

| Reach ID: | :TH16 R | Reach name: Northmoor |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 144 | Bablock Hythe | 443500 | 204200 | 26/07/84 | 2 | 7010 |
| 144 | Bablock Hythe | 443500 | 204200 | 17/[0/84 | 3 | 7012 |
| Reach ID; | ; TH26 R | Reach name: Clifton |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 133 | Clifton Bridge | 454700 | 195400 | 03/08/92 | 2 | 1 |
| Reach ID: | :TH27 R | Reach name: Day's |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample 1D |
| 145 | Shillingford | 459000 | 193200 | 26/07/84 | 2 | 7017 |
| 145 | Shitlingford | 459000 | [93200 | [7/10/84 | 3 | 7019 |
| Reach ID: | TH30 R | Reach name: Goring |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Scason ID | Sample ID |
| 55 | Whitchurch Weir | ir 463300 | 176800 | 23/05/90 | 1 | 6600 |



Species Gyrinus distinctus Aube has notability code(s) RDB 3 and is found in the following samples:
Reach ID: TH27 Reach name: Day's

| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 145 | Shillingford | 459000 | 193200 | $09 / 05 / 84$ | 1 | 7013 |

## Notable taxa recorded within database

03-Apr-97

| Reach ID: | : TH10 | Reach name: St. Johns |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site ID | Site name |  | Easting | Northing | Sample date | Season ID | Sample ID |
| 143 | Malthouse |  | 422500 | 198400 | 17/10/84 | 3 | 7005 |
| Reach ID: | : TH22 | Reach name: Iffley |  |  |  |  |  |
| Site ID | Site name |  | Easting | Northing | Sample date | Season ID | Sample ID |
| 48 | Top of San | rd Lock Cut | 452800 | 202100 | 08/10/90 | 3 | 7900 |

Species Anacaena bipustulata (Marsham) has notability code(s) NB and is found in the following samples:

| Reach ID: TH 22 | Reach name: Ifflcy |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Site ID | Site name |  | Easting | Northing | Sample date | Season ID | Sample ID |
| 48 | Top of Sandford Lock Cut | 452800 | 202100 | $09 / 07 / 92$ | 2 | I |  |



Species Oulimnius major (Rey) has notability code(s) NA and is found in the following samples:

| Reach ID: TH19 | Reach name: King's |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 44 | Trout Inn, Godstow | 448300 | 209200 | $09 / 08 / 90$ | 2 | 7600 |

## Notable taxa recorded within database

03-Apr-97

Species Sialis nigripes Pictet has notability code(s) NB and is found in the following samples:

| Reach ID: TH10 |  | Reach name: St. Johns |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site 1D | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 143 | Malthouse | 422500 | 198400 | 09/05/84 | 1 | 7000 |
| 143 | Malthouse | 422500 | 198400 | 09/05/84 | 1 | 7001 |
| 143 | Malthouse | 422500 | 198400 | 26/07/84 | 2 | 7004 |
| 143 | Malthouse | 422500 | 198400 | 17/10/84 | 3 | 7006 |
| Reach 10: TH11 |  | Reach name: Buscot |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season 1D | Sample ID |
| 37 | Buscot |  |  | 15/08/77 | 2 | 2191 |
| 37 | Buscot |  |  | 15/08/77 | 2 | 2190 |
| Reach ID: TH16 |  | Reach name: Northnoor |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 144 | Bablock Hythe | 443500 | 204200 | 26/07/84 | 2 | 7010 |
| 144 | Bablock Hythe | 443500 | 204200 | 17/10/84 | 3 | 7012 |
| Reach ID: | : TH18 | Reach name: Eynsham |  |  |  |  |
| Site ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |
| 35 | Eynsham |  |  | 24/08/77 | 2 | 2219 |
| 35 | Eynsham |  |  | 24/08/77 | 2 | 2218 |

Species Ceraclea senilis (Burmeister) has notability code(s) $\mathbf{N}$ and is found in the following samples:

| Reach ID: | TH23 | Reach name: Sandford |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Site ID | Site name |  | Easting | Northing | Sample date | Season ID | Sample ID |
| 124 | Radley College | Boathouse | 453800 | 198800 | $30 / 06 / 92$ | 2 | 5006 |
| Reach ID: | TH30 | Reach name: | Goring |  |  |  |  |
| Ste ID | Site name | Easting | Northing | Sample date | Season ID | Sample ID |  |
| 55 | Whitchurch Weir | 463300 | 176800 | $23 / 05 / 90$ | 1 | 6600 |  |

Species Leptocerus lusitanicus (Mclachlan) has notability code(s) RDB 2 and is found in the following samples:
Reach ID: TH30 Reach name: Goring

| Site ID | Site name |
| :--- | :--- |
| 55 | Whitchurch Weir |


| Easting | Northing | Sample date | Season ID | Sample ID |
| :--- | :--- | :--- | :--- | :--- |
| 46,3300 | 176800 | $23 / 05 / 90$ | 1 | 66010 |

APPENDIX 4.1 The executive summary and key supporting tables and figures from references cited in Chapter 4: Macrophytes of the main report.

## BUNDED RESERVOIR STORAGE STUDIES 970VB5

## RIVER CORRIDOR, PHASE ! AND PHASE 2 SURVEYS: RIVER THAMES

## EXECUTIVE SUMMARY

Ecosurveys Ltd has been contracted by Thames Water Utilities Ltd to undertake a River Corridor survey, and Phase 1 and Phase 2 surveys on the Thames Floodplain between Abingdon Lock and Benson Lock, Oxfordshire. This study is part of the SWORDS Project and is intended to provide a baseline understanding of the nature conservation interest of this reach of the River Thames and its floodplain, such that it can help;

- identify any impacts on the catchment which could potentially arise from the reservoir operation and the construction of abstraction and discharge structures:
- the design of a least damaging development proposal and operating regime; and
- bring forward recomrandations for enhancement of the river corridors, particularly around and adjacent to any proposed abstraction and discharge structures.

The study has three integrated components:

- a River Corridor Survey on approximately 30 km of statutory main river;
- a Phase 1 Ecological Land-use Survey of the designated Thames floodplain; and
- detailed Phase 2 botanical surveys of water level dependant habitats of conservation interest in the designated Thames floodplain.

The field survey for the various componeats was undertaken between May and August 1992, according to standard NRA and English Nature methodologies.

The River Corridor Survey sbows the Thames to be a broad deep navigable river, with its water level maintained by locks and weirs. The large amount of boat traffic precludes extensive growth of channel vegetation, which is therefore virtually limited to strands of tall marginal species. The river banks are of varied profile and are frequently tree and shrub lined, with tall herbs and coarse grasses underneath and where open.

The River Corridor sections were assessed according to the London Ecology Unit River Reach Evaluation Criteria. The majority of the 61 sections were graded good, with four graded poor, eight graded impertant and one graded critical on account of the adjacent Culham Brake, a wetland site of Special Scientific Interest.

The Phase I Survey reveals the designated floodplain of the Thames between Abingdon Lock and Benson Lock to be an intensively managed agricultural area, largely arable land with some semi-improved and improved grasslands, managed as pasture and hayfieids. There
consequently is a gencrally low level of nature conservation interest; that which exists does so as rather isolated habitats (eg including marshy pastures, open standing waters, swamps, small woodlands supporting the nationally rare summer snowflake, but otherwise unremarkable) scattered at intervals along the floodplain, separated by large tracks of arable land.

16 sites for detailed Phase 2 botanical survey were identified by the Phase 1 survey. These include flushes, marshy pastures, swamps, open water sites, wet woodlands and tall herb marshes. Plant communities identified by the Phase 2 surveys were 2 open water types, 3 woodland types, 10 swamp types and 4 mesotrophic grassland types. With the exception of the flood meadow habitat, MG 4, on site 9 (Long Wittenham Nature Reserve), all are common and widespread in lowland Britain; the flood meadow community is of local occurrence in Britain but this site holds a slightly modified example, especially when compared to the large areas further upstream (out of this Study Area) around Oxford. No nationally or regionally rare plant species were recorded by the Phase 2 surveys, which took place too late in the year for confirmation of the summer snowflake records for sites 15 and 16. The Phase 2 survey sites' general wetland species interest was identified to be vulnerable to agricultural improvements and lack of management, in addition to permanent and marked reductions in the mean water level of the Thames, but as long as the river periodically floods, the seed dispersal requirements of the summer soowflake will be met.

## SUMMARY

## RIVER THAMES: ABINGDON LOCK TO BENSON LOCK

## INIRODUCLION

Ecosurveys Ltd has been contracted by Thames Water Utilities Lid to undertake, as part of the SWORDS Project in Oxfordshire:

- a River Corridor Survey on 30 km of statutory main river of the River Thames between Abingdon Lock and Benson Lock;
a Phase 1 Ecological Habitat Survey of the designated Thames Floodplain between Abingdon Lock and Benson Lock; and
detailed Phase 2 botanical surveys of water level dependent habitats of conservation interest in the designated floodplain.

The field survey was undertaken in June, July and August 1992 using standard NRA and English Nature methodologies. The results of the survey are summarized below and are detailed in the report, which comprises this volume (river corridor survey data), a volume of Phase 1 and Phase 2 survey data and a volume of additional photographs.

## RIVER CORRIDQR HABITATS

The River Thames between Abingdon Lock and Benson Lock is a broad, deep, navigable river, gently meandering and with its water level maintained by locks and weirs. The river carries a large amount of boat traffic, especially in summer. The flood plain is largely arable land with semi-improved and improved pastures, incorporating the occasional woodland, marsby grassland and built-up area.

The river banks, composed of clay, sand and gravel (and locally artificially reinforced), are generally well vegetated and of varying profile, from vertical cliffs to gently graded sections. Trees and shrubs often line the river, crack willows (as standards, old coppice or poliards) are especially frequent, with hawthorn, alder (often as old coppice), ash, elm, field maple, oak and sycamore. Various ornamental trees occur through built-up areas and other less common native shrubs present are spindle, guelder-rose, dogwood and buckthora, in addition to shrub willows: grey, goat and osier. Tall herbs and coarse grasses dominate the herbaceous vegetation under the trees, and where open, and it is generally unremarkable and comprises ubiquitous species but with the addition of wild onion, escaped Brassica and introduced Aster species. Grazed sections of bank are grass dominated. Crevices in artificial banks support small populations of pellitory-of-the-wall and stonecrops.

Marginal vegetation is generally tall and dominated by great willowherb, comfrey, meadowsweet, bard rush, bittersweet and purple loosestrife. Also present are soft rush, water chickweed, hemp agrimony, marsh yellow-cress, water figwort and false fox sedge. Gently graded banks support gipsywort, water mint, brooklime, marsh woundwor, water forget-meoot, watercress and fool's watercress. Other less common marginal species include yellow loosestrife, skullcap, ragged robin and Himalayan Balsam, the latter especially near built-up areas. The marginal vegetation frequently grades into the emergent aquatics, except where precluded by sbade, water depth and boat moorings. Common species in the channel are branched bur-reed, common club rush, sweet flag, reed canary grass, greater pond sedge, reedmace and common reed, forming species-pure or mixed stands, often showing zonation from shallower to deeper water. Also present are flowering rush, great water dock, common water plantain, narrow-leaved water plantain, yellow water lily and unbranched bur-reed (the latter two suffering from damage caused by boat traffic).

## TARGET SPECIES

The River Corridor Survey of the River Thames included an element of searching for, and acurately recording the location of, a number of Target Species, which are those in any of the following three categories (from Palmer \& Newbold, 1983, Wetland \& Riparian Plants in Great Britain, NCC Focus on Nature Conservation No 1):

- Aquatic plants recorded from 100 or fewer $10 \times 10 \mathrm{~km}$ squares in Great Britain and which need special-protection in the NRA Thames Region;
- Riparian and non-aquatic wetland plants recorded from 100 or fewer $10 \times 10 \mathrm{~km}$ squares in Great Britain and which need special protection in the NRA Thames Region; and
- Aquatic plants recorded from more than $10010 \times 10 \mathrm{~km}$ squares in Great Britain but which need special protection in the NRA Thames Region.

No Target Species were recorded from the River Thames River Corridors.

## CONSERYATION MANAGEMENT

The management of the riparian tree cover is the principal target of management of the existing wildlife habitat resource. Single, and lines or groups of, old pollarded and coppiced trees are important in wildlife and landscape terms and management recommendations are made for repollarding and recoppicing. A certain number of trees should be worked each year out of the total resource, in order to provide variety of structure and age of regrowth. This will also act as a precautionary measure, in case the trees do not respond to management, so avoiding the situation of baving a river bank lined with dead stumps and pollards.

As a spin-off from pollarding and coppicing, the increased light let onto the river bank should promote the growth and spread of channel vegetation, so protecting the bank from erosion by boat wash. Fallen trees are to be removed unless it can be demonstrated that they function as Kingfisher perches and do not significantly impede flow. Occasional standing dead trees, unless a danger, are best left in situ for woodpeckers and invertebrates.

A second target of conservation management is tall wetland vegetation and relatively speciesrich drier neutral grassland in the floodplain adjacent to the river. These areas are typically unmanaged and with time, this leads to loss of species diversity and succession to different habitats of lower conservation interest. To prevent this and hence to maintain speciesrichness, the initiation of suitable mowing regimes is required, combined with the removal of cut material and eradication of invading scrub.

A third and equally simple conservation management operation involves thinning out of trees and shrubs along the south sides of channeis, to let in more light to the water's edge to promote the growth of channel vegetation.

## HABITAT ENHANCEMENT

The principal opportunities for habitat eabancement centre on the areas of tall wetland vegetation, habitats which are important in their own right and are of value for birds and invertebrates. The recommendations are to increase the extent of these areas by excavating the surrounding ground down to, or near, water level, enabling the wetland species to spread.

Areas of ruderal vegetation, unmanaged arable or pasture represent opportunities for the planting of native tree and shrub species, especially where adjacent to existing areas of woodland.

## PHASE 1 SURVEY

The designated floodplain of the River Thames between Abingdon Lock and Beason Lock extends for 20 km , covers approximately $15 \mathrm{~km}^{2}$ and varies in width from 400 m to over 1000 m .

The survey reveals that the floodplain is predominantly intensively agricultural - dominated by arable land with occasional improved and semi-improved pastures, built-up areas, woodlands (omamental, broadleaved, mixed, coniferous), amenity grasslands, tall berbs, caravans and mineral workings.

## PHASE 2 SURVEY

Fourteen sites were identified by the Phase 1 survey for detailed Phase 2 surveys, along with one site notified by a landowner which was deemed to meet the criteria for Phase 2 survey, and a second additional site notified by Thames Region NRA, giving a total of 16 sites. Eleven sites were surveyed in late June/early July with three of these revisited in August, and five were surveyed only in August. The habitats surveyed included willow cars and other woodiands, tall herb swamps, old gravel pits and other open water bodies, inundation grasslands and marshy pastures.

The majority of sites supported only one or two water level dependent plant community types, with one diverse tall herb swamp holding four and one large mosaic of habitats holding six. With the exception of Site 9 (a relatively herb rich flood meadow), none of the plant communities ideatified were those other than common types known to be widespread in lowland Britain and typical of wetlands in general. No rare species were recorded although two of the sites ( 15 and 16 ) are reliably reported to be localities for a nationally rare plant, probably not seen in the present survey because of the time of year.

RTYER THAMES SECIION EVALUATION

| SECTION | CRITICAL | IMPORTANT | GOOD | POOR |
| :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & 001 \\ & 002 \\ & 003 \\ & 004 \\ & 005 \end{aligned}$ |  |  | $\begin{aligned} & \checkmark \\ & \checkmark \\ & 8 \end{aligned}$ | $\checkmark$ |
| $\begin{aligned} & 006 \\ & 007 \\ & 008 \\ & 009 \\ & 010 \\ & \hline \end{aligned}$ |  |  | $\begin{aligned} & 8 \\ & 6 \\ & 6 \\ & 6 \\ & 6 \end{aligned}$ |  |
| $\begin{aligned} & 011 \\ & 012 \\ & 013 \\ & 014 \\ & 015 \end{aligned}$ |  | $\begin{aligned} & \checkmark 1 \\ & \checkmark 2 \end{aligned}$ | $\begin{aligned} & \checkmark \\ & \checkmark \\ & \\ & \hline \end{aligned}$ |  |
| $\begin{aligned} & 016 \\ & 017 \\ & 018 \\ & 019 \\ & 020 \end{aligned}$ |  |  | $\begin{aligned} & 6 \\ & 6 \\ & 6 \\ & 6 \\ & 6 \end{aligned}$ |  |
| $\begin{aligned} & 021 \\ & 022 \\ & 023 \\ & 024 \\ & 025 \\ & \hline \end{aligned}$ |  | $\begin{array}{r} \checkmark 3 \\ \checkmark 4 \end{array}$ | $\checkmark$ <br> $\checkmark$ | $\checkmark$ |
| $\begin{aligned} & 026 \\ & 027 \\ & 028 \\ & 029 \\ & 030 \end{aligned}$ | - | $\checkmark 5$ | $\begin{aligned} & \checkmark \\ & \checkmark \\ & 6 \\ & 6 \end{aligned}$ |  |
| $\begin{aligned} & 031 \\ & 032 \\ & 033 \\ & 034 \\ & 035 \\ & \hline \end{aligned}$ |  | $\checkmark 6$ | $\begin{aligned} & 6 \\ & \checkmark \\ & 6 \\ & 6 \end{aligned}$ |  |
| $\begin{aligned} & 036 \\ & 037 \\ & 038 \\ & 039 \\ & 040 \end{aligned}$ |  |  | $\begin{aligned} & \checkmark \\ & \checkmark \\ & \checkmark \\ & \checkmark \\ & \checkmark \end{aligned}$ |  |


| SECTION | CRITICAL | IMPORTANT | GOOD | POOR |
| :---: | :---: | :---: | :---: | :---: |
| 041 |  |  | $\checkmark$ |  |
| 042 |  |  | $\checkmark$ |  |
| 043 |  |  | $\checkmark$ |  |
| 044 |  |  | $\checkmark$ |  |
| 045 |  |  | $\checkmark$ |  |
| 046 |  |  | $\checkmark$ |  |
| 047 |  |  | $\checkmark$ |  |
| 048 |  | $\checkmark 7$ |  |  |
| 049 |  | $\checkmark 8$ |  |  |
| 050 |  |  | $\checkmark$ |  |
| 051 |  |  | $\checkmark$ |  |
| 052 |  |  |  | $\checkmark$ |
| 053 |  |  | $\checkmark$ |  |
| 054 |  |  | $\checkmark$ |  |
| 055 |  |  | $\checkmark$ |  |
| 056 |  |  | $\checkmark$ |  |
| 057 |  |  | $\checkmark$ |  |
| 058 |  |  | $\checkmark$ |  |
| 059 |  | . | $\checkmark$ |  |
| 060 | $\checkmark 9$ |  |  |  |
| 061 |  |  | $\checkmark$ |  |







## OXEQRDFLOOQPLAINENYIRONMENTAL SURYEY 1292

## EXECUTIVE SUMMARY

Ecosurveys Lid has been contracted by the National Rivers Authority Thames Region to undertake a study of the nature conservation interest of a part of the floodplain of the River Thames between Eynsham and Sandford-on-Thames, Oxfordshire. This study is part of a roodelling process providing information on floodplain envelopes and flow routes. It is to supply baseline information to the NRA to hetp them assess the effects of any proposed modification in the operation of flood control structures and to optimise their operation.

The study has several integrated componens:

- a Desk Study to collate and review information on areas of known conservation importance in the Study Area;
- a River Corridor Habitat Survey of a designated 74 km of the River Thames, its tributaries, secoodary channels and drains in the Study Area;
- a Phase 1 Ecological Habitat Survey of about $25 \mathrm{kro}^{2}$ of the designated Oxford Floodplain Environmental Survey Study Area; and
- detailed Phase 2 botanical surveys of sites of water-level dependence within the study Area for which adequate information does not already exist.

The tield survey for the various components was undectaken berween May and August 1992 using standard NRA and English Nature methodologies.

The data collection and review, on the whole, yielded information of limited use, due mainly to its brevity. The River Corridor Habitat Survey shows that, in general, the river corridor habitats are evaluated as critical or good, according to the London Ecology Unit Evaluation with the exception of urban Oxford where they are evaluated as poor on the whole. The large number of river corridor sections evaluated as critical is in large part due to their proximity to the extensive SSSI's in the floodplain.

The Phase 1 Survey reveals that the designated floodplain is largely agricultural with the exception of the of the urban area of Oxford and is repeatedly criss crossed by large tree and scrub lined watercourses.

From the Desk Study, River Corridor and Phase I Surveys, 34 sites as requiring Phase 2 survey were identified and each received detailed botanical investigations with one $2 \times 2 \mathrm{~m}$ quadrat being recorded in each water level dependent plant community and with full species lists recorded for those sites.


## Desk Sudy - Review of Data un Sites of Nature Cunservation Sipnificance in the Scudy_Ared

The collection and review of existing information on sites of nature conservation importance was principally carried out at the commencement of the study though continued throughout as further sources, sites or information came to light. Ecosurveys Lid believe that all sites of nature conservation significance in the Study Area have been identified and have been studied hy one, some or all of the datid review, River Corridor, Phase 1 and Phase 2 surveys. Whilst we believe that we have collated the vast majority of the available data, some sources were unwilling to release data or advised us that the data we had from elsewhere exceeded their data in terms of quality, and was therefore either not released or collated.

One purpose of the data collection was to produce a document which held in one place the available information, from previous work, on sites of nature conservation significance in the Study Area. Part 5 of this report constitutes that document.

A second purpose was to use the available data to identify known sites and to determine which water-level dependent sites in the Study Area required additional fieldwork due to the lack, antiquity or unreliability of the existing information. The data from the surveys carried out this year was then to be compared with previous information on past site management and nature conservation value.

### 2.3.2 River Curtidor

The designated River Corridor was divided into 148 sections each of approximately 500 m length and surveyed between May and August using the standard NRA River Corridor Habitat survey Methodology. The information for each section thus comprises photograph(s), a short written descriptive text and a habitat map. Aquatic vegetation was sampled using a grapnel from the banks of namower channels and from a boat on the River Thames. Management recommendations to maintain and/or enhance the nature conservation value of the sections are also supplied.

The River Corridor habitat survey results are compiled in Pant 2 of this report with the locations of the River Corridor sections shown on the supporting maps and tabulated with channe! names and start and end point grid references

The River Corridor sections have been assessed for ther nature conservation imporance according to the London Ecology Unil River Reach Evaluation Criteria, see Appendix Two, and this information is discussed and analysed in this document (Part I) atong with the general management recommendations required to maintain and enhance the nature conservation interest of the River Corridor sections.

### 2.3.3 Phase 1

The purpose of the Phase I survey was to identify habitats within the Study Area, but outside designated SSSI's, which are of value to wildlife. The Phase I survey was undertaken in June 1992 using a standard methodology (NCC 1990) in order to supply a readily accessible visual representation of the entire floodplain in the Study Area. For this reason, the colour coding system rather than alphanumeric was adopted.

The results of this survey (Par 3 of this document) comprise a series of A4 and A3 maps which cover the entire survey area. The maps are supported by target notes which indicate;

- the location of sites deemed to be of nature conservation interest which received a Phase 2 detailed botanical survey.
- features of interest or importance to wildife which were identified in the review of previous available data, or
- features observed in the course of the fieldwork components, and
- site specific information for researched data on notable hird species or populations breeding within or regularly using the floodplain.


## $2.34 \quad$ Phase 2

Sites identified from the earlicr elements of the study which were described as water-level dependent, at least in part, for which there was only old, or unreliable, or no avaitable information and which were apparently of nature conservation interest, received a detailed botanical survey equivalent to NCC (now Engtish Nature) Phase 2 survey level, (Smith et al, 1985).

In the event, the data collection yielded information of rather variable quality, which for mose stes was rather brief, and the data gathered by Ecosurveys Lid fieldwork greatly exceeding the previous data in terms of the time spent on the site and the detailed botanical information gathered. In consequence a formalised review of the previous data was not carried out as no sensible comparison could be made for the majority of the sites. However, the information from the data collection has been analysed and integrated into the report as a whole in the form of Target Notes on the Phase 1 maps and as part of the Conservation Assessment of the Phase 2 site reports. Thus, all of the available previous information on a site may be found in Part 5 of the report, as an abridged version in a Target Note on the Phase 1 map and list, and possibly analysed within a Phase 2 site report. with all information being referenced eg to it corresponding Target Note or previous survey data, for easy accessibility.

## 4. RIVER CORRIDOR SURYEY RESULTS AND DISCUSSION

## 4.I INTRODUCTION

The results of the River Corridor Habitat Survey are compiled in Part 2 of this report, in three volumes. The results for each section consist of a brief descripuive sext of the adjacent land, bank and channel vegetation, photographs, cross sections and management recommendations.

The 148 River Cortidor Habitat Survey sections studied comprise a considerable resource for wildlife at least in a local context. The watercourses vary from large slow flowing deep rivers such as the River Thames, through fast flowing streams to narrow drains providing habitat for a wide variety of wildlife. Generally the various watercourses are largely tree and scrub lined and the adjacent habitats include a complex mosaic of urban, arable and pasture land, tall herb marshes and swamps and woodland. Despite their proximity to Oxford, many of the watercourses outside the urban area are relatively undisturbed, a factor which enhances their value to wildife. A notable exception, however, is the River Thames which, in particular with regard to marginals and aquatics, appears to be rather poorer in species than would otherwise be expected principally due to the large amount of boat traffic especially in summer and its associated problems with wash, and sediment disturbance etc. It is, however, still of high local importance as a wildlife restource.

The watercourses of the floodplain which were studied can be conveniently but arificially grouped as large rivers, smaller rivers and streams and drains. The table iverleaf provides information on the evaluation of the ecological interest of the River Corridor sections according to the London Ecology Unit Ecological Evaluation Criteria for River Reaches which are attached in Appendix Two.

The catcerorics are as follows:
Crtheally Imporiant lor Widelife - The mose ecologically trabide sections of river or adjacen habitats which depend on the mamenance of the preacon hydrological conditions. This category includes most examples of regionally rare habitats and sections with significant communities of rare flora and fauna.

Important for Wildife - Sections of high wildife value but not necessarily closely associated with the river, however potentially vulnerable, and typically composed of either a mixture of well structured habitat types but with lew species of particular note or a good example of a single habitat type.

Good for Wildife - Sections which provide particutar local interest and typically comprising habitats of a more robust nature less volnerable (1), and possibly restorable after, engineering works.

Poor for Wildlife - Sections of river and adjacent land with little semi-natural vegetation and of low intrinsic wildlife interest. Enhancement opportunities may present themselves.

Must "typical" River Corridor sections would thus fall into the category of good and in the following table of evaluated sections justifications are only noted for critical, important and poor evaluations with supplementary notes.

The evaluations of the River Corridor Sections were carried out by two members of Ecosurveys Lid technical staff who had been involved in the project and were familiar with the Study Area (Tim Harvey and Tim Smith). The evaluations are a subjective judgement arising from a assessment of each sections conservation value represented by its habitats and physical features, The columns are colour coded to match the following series of maps with;

| Critical | $=$ Red |
| :--- | :--- |
| Important | $=$ Green |
| Good | $=$ Blue |
| Poor | $=$ Brown |

TABLE 1
RIVER CORRIDOR NATURE CONSERVATION EVAIUATION BASED ON
L.E.U EVALUATION CATEGORIES

| SECTION |  | $\rho$ |  |  | NOTES |
| :---: | :---: | :---: | :---: | :---: | :---: |
| No | Name |  |  |  |  |
| 1 | River Thames |  | $\checkmark$ |  | Create berm on the left in midsection. |
| 2 | Sandford <br> Ditch | $\checkmark$ |  |  | Hottonia palustris locality. Repollarding of shading trees. |
| 3 | River Thames |  | $\checkmark$ |  |  |
| 4 | River Thames |  | $\checkmark$ |  |  |
| 5 | River Thames | $\checkmark$ |  |  | Fiddler's Elbow Island adjacent. (Phase 2 Site 2). |
| 6 | River Thames | $\checkmark$ |  |  | Adjacent to Rose Isile. (Phase 2 Site 4). |
| 7 | River Thames | $\checkmark$ |  |  | Unimproved wet readow adjacent to Heyford Hill Lave Pasture. (Pbase 2 Site 5). |
| 8 | River Thames |  | $\checkmark$ |  |  |
| 9 | River Thames | 1 |  |  | Adjacent to Fiddier's Elbow Island. (Phase 2 Site 2). |
| 10 | River Thames | $\checkmark$ |  |  | Adjacent to Wetland and Woodland south of Iffley Meadows SSSI. (Pasture 2 Site 10). |
| 11 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI, but channel poor. |
| 12 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI, but channel poor., |
| 13 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI, but chaninel poor. |
| 14 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 15 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 16 | River Thames |  |  | $\checkmark$ | Artificial. Boat traffic. Plant trees upstream on right. Limited scope for enhancement. |
| 17 | River Thames |  |  | $\checkmark$ | Artificial. Mostly urban. Plant trees and shrubs. Limited scope for onhancement. |
| 18 | River Thames |  |  | $\checkmark$ | Artificial. Urban. Plant trees and strubs. Little scope for eahancement. |
| 19 | Castle Mill Stream |  | $\checkmark$ |  | Urban. Create riffe and pool. |
| 20 | Castle Mil <br> Stream |  |  | $\checkmark$ | Artificial: Urban. Lacrease tree and shrub cover. Limited scope for enhancement. |
| 21 | Castle Mill <br> Stream |  | $\checkmark$ |  | Poor channel, but wooded. Limited scope for eabancement. |


| SECTION |  |  | $G$ | NOTES |
| :---: | :---: | :---: | :---: | :---: |
| No | Name |  |  |  |
| 22 | Castle Mill <br> Stream |  | $\checkmark$ | Poor channel, but wooded. Limited scope for enhancement. |
| 23 | Railway Drain |  | $\checkmark$ |  |
| 24 | Railway Drain |  | $\checkmark$ |  |
| 25 | Railway Drain |  | $\checkmark$ |  |
| 26 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 27 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 28 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 29 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 30 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 31 | Port Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 32 | Por Meadow Drain | $\checkmark$ |  | Adjacent to SSSI. |
| 33 | Hinksey <br> Stream | $\checkmark$ |  | Adjacent to SSSI. |
| 34 | Weirs Mill <br> Stream | $\checkmark$ |  | Adjacent to SSSI. |
| 35 | Weirs Mill <br> Stream | $\checkmark$ |  | Adjacent to SSSI. |
| 36 | River Tharmes | $\checkmark$ |  | Adjacent to SSSI. |
| 37 | River Thames |  | $\checkmark$ |  |
| 38 | New Hinksey <br> Drain |  | $\checkmark$ | Poor as wetland, but wooded. |
| 39 | New Hinksoy <br> Drain |  | $\checkmark$ | Poor as wetland, but wooded. |
| 40 | Shire lako <br> Ditsh |  | $\checkmark$ | Channel poor, good wooded corridor. |


| SECTION | N | Name |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |



| SECTION |  | $C$ |  |  | NOTES |
| :---: | :---: | :---: | :---: | :---: | :---: |
| No | Name |  |  |  |  |
| 79 | Bulstake Stream |  | $\checkmark$ |  |  |
| 80 | Bulstake Stream |  | $\checkmark$ |  |  |
| 81 | Bulstake Stream |  |  |  |  |
| 82 | Bulstake Stream |  |  |  | Create pools, deepe |
| 83 | River Thames |  | $\checkmark$ |  | Boats. Wooded. |
| 84 | Fiddler's Island Stream |  |  |  | Wooded. |
| 85 | Fiddler's <br> Island Stream | $\checkmark$ |  |  | Adjacent to SSSI. |
| 86 | River Thames |  |  |  |  |
| 87 | Fiddler's <br> Island Stream | $\checkmark$ |  |  | Adjaceat to SSSSI. |
| 88 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 89 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 90 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 91 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 92 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 93 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 94 | River Thames |  |  |  |  |
| 95 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 96 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 97 | Duke's Cut |  |  |  |  |
| 98 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 99 | River Thames |  |  |  |  |
| 100 | River Thames | $\checkmark$ |  |  | Adajcent to SSSI. |
| 101 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 102 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 103 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |
| 104 | River Thames | $\checkmark$ |  |  | Adjacent to SSSI. |




In summary, for the River Corridor mabiat survey of 148 sections (c. 74 km ),

| S1 | sections were evaluated as | critical |
| :--- | :--- | :--- |
| 1 | section was evaluated as | important |
| 77 | were evaluated as | geod |
| 19 | were evaluated as | poos |

The following series of 3 coded maps provide an easiiy accessible visual representation of the River Corridor Section Evaluations and their relationships.

Outlined below are brief descriptions of the watercuurses and information on their evaluations, management recommendations and enhancement opportunities. Whilst some of the management recommendations may apply only to specific sections or parts, there are some general recommendations that are applicable to the majority of the sections.

The existing management of channels bordering or passing through SSSI's should continue through close liaison and consultation with English Nature.

The management of the riparian tree cover is one goal of management of the existing wildlife habitat resource. Single, lines or groups of old pollarded and coppiced trees are important in wildlife and landscape terms and recommendations are made for repollarding and recoppicing where appropriate. A certain number of trees should be worked each year out of the total resource in order to provide variety of structure and age of regrowth. This will also act as a precautionary measure in case the trees do not respond to management, so avoiding the situation of having a river bank lined with dead stumps and pollards.

As a spin-off from pollarding and coppicing, the increased light let on to the river bank should promote the growth and spread of channel vegetation, so protecting, where appropriate, the bank from erosion by boat wash and, in any case, potentially increasing botanical diversity over time. Fallen trees are to be removed from channels unless it can be demonstrated that they function as Kingfisher perches and do not significantly impede now or navigation. Occasional standing dead trees, unless a danger to the public, are best left in situ for woodpeckers, invertebrates and bats. Wooded areas could have bird and bat boxes sited.




## The Riven Thames (Sections 001.003-018.036-037.071-076.083.086.088 096.098-110, and 148)

The River Thames is mosely gently meandering, wide and deep with islands, locks and weirs and slow flowing through agricultural land, on either side of urban Oxford. Its banks are principally clay and sand of varying profile with large parts being vertical and artificial through the buith-up area. The channel carties a large amount of boat traffic especially in summer.

Of the 49 sections of the River Thames 29 are evaluated as critical by virtue of the presence of SSSI or valuable water-level dependent sites such as Fiddler's Elbow Istand, Rose Isle and Heyford Hill Lane Pasture adjacent to one or both sides of the river for at least a part of the section. Sections 011 . 013 are evaluated as critical by virtue of their flowing alongside lffley Meadows SSSI but the nature conservation value of the river is relatively low due to it urban, artificial nature, Iffley Lock and environs, and use by man. Unfortunately there is also limited scope for enhancement of the river in these sections. It is recommended that the best option for these 29 critical sections be a continuation of the present management in order to maintain their nature conservation value.

13 sections of the River Thames are evaluated as good with the recommended conservation management being, in the main, to retain the section as at present coupled with sympathetic future management of, for example, stands of emergent vegetation. Section 007 would benefit from enhancement measures, for example the creation of a wide berm to permit the growth of large stands of fringing vegetation. Section 083 and 086 would be enhanced by practical measures to improve the extent and diversity of emergent and aquatic vegetation given the heavy use of these sections by boat traffic.

7 sections are evaluated as poor, all of these being urban or lock sections with mostly arificial banks and heavily used by man. All of the existing seminatural habitat in these sections should be retained. In most of these sections there is little scope for enhancement beyond cosmetic tree and shrub planting with the exception of 075 in which it is recommended that the channel profile be changed to create marginal berms and riffles and pools.

## CONDITIONS

Surveyed left bank. Dry and sunny. Flow - slow. One photograph from left side.

## PHYSICAL FEATURES OF CHANNEL

i Broad nature
ii Dimensions
iii Substrate
iv Bank type

LAND USE

BANK
VEGETATION

## CHANNEL VEGETATION

FEATURES OF INTEREST

POTENTIAL THREATS
MANAGEMENT
RECOMMENDATIONS

A gently meandering section entering the outskirts of Oxford. The section is crossed by the A420 road bridge in midsection.

Channel $20-25 \mathrm{~m}$ wide, $>2 \mathrm{~m}$ deep.
Clay and mud.
Artificial on the left; vegetated entirely, except for downstream, on the right. Banks $0.5-0.75 \mathrm{~m}$ high, $45-90^{\circ}$ slopes.

LS Houses, gardens, hotel and factory.
RS Allotment gardens, houses and gardens.
LB Urrica dioica, Eupatorium cannabinum, Scrophularia aquatica, Epilobium hirsuṭm, Lolium perenne and Dacrylis glomerata occur upstream with Sambucus nigra, Acer pseudoplatanus and Salix cinerea saplings. Near the road bridge, a few standards of Salix alba, Acer pseudoplatanus, Comus sanguine and Laburnum anagyroides overhang the bank: Below the bridge, Salix fragilis and Aesculus hippocastanum overhang the bank. Downstream the bank is artificial.
RB A mixture of Lolium perenne grassland with patches of Utica dioica, Eupatorium cannabinum and Epilobium hirsurum occur upstream and in midsection with recently pollarded Salix fragilis stumps. Above the road bridge a pollarded Populus sp overhangs the channel. Below the bridge the artificial bank has Lolium perenne and Bellis perennis.

Channel vegetation is sparse. Nuphar lute and Sparganium emersum occur scattered along both sides upstream and in midsection. Polygonum amphibium occurs on the right in midsection with Iris pseudacorus, a few patches of Carex riparia and Schoenoplectus lacustris.

A number of Crack Willow on the right bank have recently been repollarded. Mute Swan.

Continued speeding by boats.
Retain the scrub and trees. Plant trees and shrubs on the right bank upstream. Retain existing marginal vegetation.

$L$
$R$
Nar 503061



[^0]:    * Recalculated values (See KES 1993)
    ^ Sites corresponding to those in 1992 and 1993 surveys
    $n=$ number of sites sampled

[^1]:    *Sites corresponding to 1993-1995 surveys

