



Report

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Severn-Thames Transfer. A Review of Biological Data. Volume 2 - Appendices.

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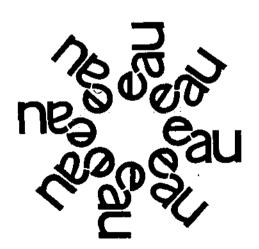
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APPENDIX 2.1 The title page, executive summary and/or conclusions and key tables of references cited in Chapter 2: Fish of the main report.

RIVER THAMES ADULT FISH SURVEY
FINAL REPORT
JULY - OCTOBER 1991



ENVIRONMENTAL ADVISORY UNIT LTD.

The issues address	sed in this report are lere.	dealt with in greate	er detail in later do	cuments and are
not summarised h	iere.			

1992
RIVER THAMES JUVENILE FISH SURVEY
Royal Holloway and Bedford New College
University of London



UNIVERSITY OF LONDON

ROYAL HOLLOWAY AND BEDFORD NEW COLLEGE

1992 RIVER THAMES JUVENILE FISH SURVEY Royal Holloway and Bedford New College University of London

by:

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EXECUTIVE SUMMARY

RIVER THAMES JUVENILE FISH SURVEY 1992

- (1) This reports on a fry survey undertaken in July/August 1992 by Royal Holloway & Bedford New College (RHBNC) in that part of the River Thames between Oxford and Days Weir where the outlet of a new reservoir proposed by Thames Water Utilities is likely to be sited. The location of the thirteen sites sampled were specified and overlapped with a first fry survey conducted in 1991 by the Environmental Advisory Unit Liverpool (EAU). It is noteworthy that 1992 and 1991 represented the fourth and third years of drought and low river flows.
- (2) The RHBNC survey sampled each site with three seine hauls. The three seine hauls per site were used to study the influence of the nature of three kinds of microhabitats on the fry community: (1) shallow sites with water lilies, (2) shallow sites without aquatic plants and (3) deep sites without aquatic plants. The net used was 25 m long made of micromesh with 3 mm mesh size.
- (3) The baseline fry stocks of the River Thames in July/August 1992 compared with July 1991 from the EAU Report were as follows:

	1992	1991	1991
	RHBNC	EAU	EAU
Number of sites	13	13*	36
Number of seinings	38	13	36
Fry density per m'	3.72	4.07	3.07
Catch per unit effort	706	607	490
(= fish per net haul)			

*same sites as in 1992

This shows very similar fry densities and catch per unit effort in the two years.

(4) It is also important to sample different microhabitats because the densities of fry inhabiting them varies greatly:

Nature of microhabitat	Deep	Water-lily	Shallow
Mean 0+ fry density (fish/m²)	0.082	10.07	5.38

This shows that the two very common fry microhabitats in the River Thames (deep and lilies) vary in 0+ fry densities by two orders of magnitude and differ in species diversity and body size. There are probably about seven important microhabitat categories in the River Thames in addition to these three which need to be studied and characterised.

(5) Our recommendations for improving baseline estimates of fish fry stocks are to improve our knowledge of the fish fauna of the River Thames along the lines listed in Recommendations. We offer some comments on impacts of the proposed reservoir and on least damaging designs for reservoir operation.

- 4. The species diversity and composition of fry fish was different in the three microhabitat categories. There were fewer species in the 'deep' category and fry from rheophilous species of fish favoured the 'shallow' category.
- 5. There is some evidence that the sizes of some species of fry differ in the different microhabitats but this needs further analysis.

7. CONCLUSIONS

1. The baseline fry stocks of the River Thames in July/August 1992 compared with those recorded for July 1991 in the EAU Report were as follows:

	1992	1991	1991
Number of sites	13	13*	36**
Number of seinings	38	13	36
Fry CPUE (fish per haul)	706	607	490
Fry density (fish/m2)	3.72	4.07	3.07

^{*} comparable sites to the 1992 survey

This shows very similar levels of catch-per-unit-effort and mean fry densities in the two survey years. The mean netted area was larger in the 1992 survey (190 m²) compared with 149 m² in 1991 which affects the CPUE figures. About 50% of the 36 sampled sites in 1991 were 'shallow with no macrophytes' compared with 33% in 1992 which affects the mean densities for combined microhabitats.

2. It is important to sample different microhabitats because the densities of fry inhabiting varies greatly:

Natu	re of microhabitat	'deep'	water-lily'	'shallow
0+ f	fry density/m ²	0.082	10.07	5.38

^{**} includes the intermediate sites and sites above Oxford.

This shows that the two commonest category of microhabitats in the River Thames ('deep' and 'water-lilies') vary in 0+ fry densities by two orders of magnitude. There is also a difference in species diversity, species composition and body size. There are probably about seven additional fry microhabitats in the River Thames which need a similar study.

- 3. Our main recommendation for improving baseline estimates of fish fry stocks are to improve our knowledge of the fish fauna of the River Thames along the lines listed in section 4.3 Recommendations. A very limited analysis of individual angler catches suggests that angling is a selective mode of sampling the species composition of the river fish fauna.
- 4. We offer some comments on impacts of the proposed reservoir and on least damaging designs for reservoir operation.

Summary of fish densities in different microhabitats of the Sutton Pools compared with the 1991/1992 results for the River Thames

Table 5

Nr.	Site Description Sutton Pools	n/m^2	n/m^3	replicates#	nettec area m^2	netted volume m^3
1	. Scirpus	65.1	76.3	2	92	70.5
11	Scirpus + Typha + little lilies	22.6	33.7	2	89	78.5
III	W. lifies in Sutton Pools	3.6	3.6	1		59.7
IV	Deep site in Sutton Pools	0.13	0.12	;	157	157
٧	Sparse S. emersum	9.5	7.9	!	100	106
VI	Bank with fringing herbs	5.1	4.9	1	160	192
VII	Strong flow, Gravel bottom	0.13		3	69	72.4
VIII	B 5 Gravel spit	_	0.17	1	30	23
IX.	deep base bank	0	0	1	70	28
•	•	0	0	2	96	44
×	Shallow base bank	3.6	6.9	2	58	30.4
	River Thames 1992					
	Deep sites in River	0.94	0.78	13	2260	2719
	Shallow sites in River	4.47	5.08	13	3048	2682
	W. Lily sites in River	8.7	8.57	12	1883	1913
					,,,,,	1010
	River Thames 1991					
	1991 Survey (EAU)	3.07	?	37	• 5515	?

^{*}Our rough estimate of EAU sampled areas

[#]Replicates come from sites belonging to each microhabitat category (Table 3)

Percentage frequency of fish fry species (0+ & 1+) at sites sampled with a fry seine in the River Thames during the 1991 EAU and 1992 RHBNC surveys

Table 6

	1992	1992	1992	1992	1991	1991
	13 sites	11 sites	13 sites	13 sites	36 EAU sites	13 sites*
	Mic	crohabita	ts			
species	Deep	Lily	Shallow	Combined		
pike	0.05	∍ 0.03	0,04	0.04	0.03	0
gudgeon	3.33	8.03	14.86	10.82	22.43	11
silv e r bream	.0.	0	0.01	0.003	0	- ''
bream	1.22	0.29	0.77	0.46	0.63	1.01
bleak	6.39	7.27	17.84	11.69	4.76	6.7
minnow	0	0.76	0.07	0.42	0	0.7
roach	80.18	77.2	53.23	67.13	44.9	72.4
chub	2.58	1.45	7.36	4.03	18,89	3.5
dace	0.33	0.97	3.26	1.9	6.03	3.93
stoneioach	0	0	0.01	0.003	0	0.30
stickleback	0.33	0.14	0.1	0.13	0.34	0.41
perch	5.35	2.27	2.37	2.51	1.9	1,1
ruffe	0.19	0.03	0	0.03	0.012	0.03
bullhead	0	0.01	0.01	0,01	0.094	0.00
Ro/Br, Hyb	0.05	1.56	0.07	0.83	0	0.1
barbel	0	0	0	0	ō	0.013

Percentage frequency was calculated from the sum of fish of all sampled sites.

^{*} This was calculated from only those 1991 sites which were also sampled in 1992

TABLE 7

THE DENSITIES OF FRY BELONGING TO DIFFERENT AGE GROUPS CAUGHT BY SEINING IN DIFFERENT MICROHABITAT CATEGORIES IN SUTTON POOLS

MICRO				METRE SQUARE AGE GROUPS
		0+	1+	>1+
I	Scirpus	63.70	1.10	0.00
11	Scirpus+Typha+lilies	21.50	0.95	0.20
III	Water-lilies only	3.60	0.03	0.01
ıv	Deep bank with Nostoc	0.10	0.00	0.03
v	Sparse Sparganium emersu	m 9.08	0.09	0.31
Vi	Shallow; fringing border	4.70	0.60	0.30
VII	Shallow with strong flow	0.00	0.00	0.13
X	Shallow, sandy + no plant	s 3.30	0.10	0.30

MICROHABITAT CATEGORY		FISH DENSI BELONGING		CUBIC METRE AGE GROUPS
		0+	1+	>1+
i	Scirpus	74.60	1.30	0.00
II	Scirpus+Typha+lilies	32.00	1.40	0.30
ııı	Water-lilies only	3.60	0.03	0.01
17	Deep bank with Nostoc	0.09	0.00	0.03
v	Sparse Sparganium emersu	im 7.57	0.08	0.26
vı	Shallow; fringing border	4.70	0.60	0.30
VII	Shallow with strong flow	0.00	0.00	0.17
X	Shallow, sandy, no plants	6.20	0.10	0.50

No fish were caught in microhabitat categories VIII and IX.

TABLE 9

DENSITIES OF FRY BELONGING TO DIFFERENT AGE GROUPS CAUGHT BY ELECTROFISHING IN DIFFERENT MICROHABITAT CATEGORIES IN SUTTON POOLS

MICROHABITAT CATEGORY	MICROHABITAT ABUNDANCE*			METRE SQUARE AGE GROUPS
	%	0+	1+	>1+
Scirpus	11.1	5.20	0.90	1.60
Water-lilies	11.9	14.60	0.10	0.10
Lilies+S.emersum	12.4	3.10	0.40	0.60
Mixed monocots.	18.5	5.80	0.80	0.60
Overhanging trees	14.4	3.90	0.20	0.00
Stony area	8.2	6.00	0.50	0.30
Weirs	6.6	3.75	1.04	2.90
Shallow, fringing	7.0	24.50	0.00	0.00
Deep, fringing	4.1	0.30	0.00	0.00
Glyceria maxima	3.3	0.00	0.00	1.30
Reed Canary grasse	s 1.7	0.00	0.00	1.70

^{*} Microhabitat abundance was calculated as the number of point samples per microhabitat category as a percentage of the total number of point samples taken. This is possible as the total point samples were spread uniformly throughout the littoral circumference of the Sutton pools at approximately 1 point per 5 m of shoreline.

1992

SUTTON POOLS FISHERY SURVEY
Royal Holloway and Bedford New College
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by:

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EXECUTIVE SUMMARY

- 1. Nine sites in Sutton Pools were sampled quantitatively by a 10x2 m fry seine net and seven other sites by a 25x3 m fry seine net during July/August 1992. In addition, 243 point samples were taken by electrofishing in all four Sutton pools, using a battery-operated portable electrofisher. The total number of fry caught was 11,245.
- 2. The Sutton Pools Report emphasises that it is essential to have a fry microhabitat classification for reliable estimates of the baseline of juvenile fish populations.
- 3. The most important nursery microhabitats were: (a) those with Scirpus beds (65.1 fish/m²); (b) shallow sites with fringing vegetation (24.5 fish/m²) and (c) mixed Monocotyledons Scirpus + Typha with some water lilies (22.6 fish/m²). The roach appears to be ubiquitous as it was always present as over 90 % of the fish catch from the above habitats and between 30-70% of the other categories.
- 4. Other types of microhabitats in Sutton Pools, such as shallow and deep unvegetated sites and sites with water-lilies, support a similar species composition and similar or lower level of fry density to comparable microhabitats in the River Thames.
- 5. The special feature of Sutton Pools is the presence of microhabitats like the three weir pools. These are sites with strong water currents and a much greater relative proportion of shallow bare shores than in the River Thames. None of these microhabitats seems to provide important nursery areas.
- 6. Apart from the weir pools and sites with strong water currents, all other microhabitats are present in both the River Thames and in the Sutton pools. In the River Thames, the areas of each microhabitat category are far more extensive.
- 7. From this brief study during July/August 1992, there is no strong evidence for the Sutton Pools to be considered as an outstanding nursery area for fry living near shore. There is still the possibility that they constitute a good spawning area because of absence of disturbance by navigation which is a major influence upon the ecology of the River Thames and its fish stocks.

portable electrofisher. The total number of fry caught was 11,245.

- 2. Different kinds of fry microhabitats with and without aquatic vegetation were sampled by seining and electrofishing. The highest catches were found in vegetated sites with *Scirpus* beds (65.1 fish/m^2) and in weed beds of mixed aquatic monotylendons with some water-lilies (22.6 fish/m^2) but also in non-vegetated sites with a terrestrial fringing border (24.5 fish/m^2) . The roach was ubiquitous both in the above sites (90% of the fish) and in other sites (30-70%).
- 3. Other microhabitats sampled were shallow and deep non-vegetated sites and sites with water-lilies which were similar to those sampled in the River Thames, both in species composition and density levels.
- 4. The presence of microhabitats such as the weirs and weir pools were a special feature of the Sutton Pools, with strong water currents and hard bottoms. There was also a greater proportion of the littoral with shallow bare shores than in the River Thames.

5. CONCLUSIONS

- 1. From this brief study during July/Augsut 1992, there is no strong evidence for the Sutton Pools to be considered as an outstanding nursery area for fry living near the shore.
- 2. The Sutton Pools study shows that (a) it is essential to develop a fry microhabitat classification for reliable estimates of the baseline stocks of juvenile fish and (b) that this categorization is applied in the River Thames itself.
- 3. There is still the possibility that the Sutton Pools constitute a good spawning area because of the absence of disturbance by navigation which is a major influence upon the

ecology of the River Thames and its fish stocks.

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Royal Holloway and Bedford New College (University of London). 1992. River Thames Juvenile Fish Survey. pp 29 ADULT FISH COMMUNITIES
OF THE
RIVER THAMES
BETWEEN SANDFORD AND BENSON LOCKS
1993



National Rivers Authority
Thames Region

SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY A REPORT FOR NRA THAMES REGION AND THAMES WATER UTILITIES LTD.

ADULT FISH COMMUNITIES OF THE RIVER THAMES BETWEEN SANDFORD AND BENSON LOCKS 1993

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Table 1. Total Values of Density and Biomass by Reach and Phase.

, <u>,</u>	Density	Echo Counting	(l ² ish 100m ⁻³)	Density	Echo Integ.	(Fish 100m ⁻³)	Biomass	(Kg 100m ⁻¹)	
Phase	I	EE .	Mean	1	п	Mean	ı	п	Mean
Reach									
.1	2.198	3.800	2.999	2.052	2.272	2.162	0.159	0.256	0.208
2	5.709	3.961	4.835	6.561	1.746	4.154	0.430	0.388	0.409
3	7.790	6.138	6.964	11.908	3.365	7.637	0.407	0.505	0.456
4	2.371	6.989	4.680	1.232	5.120	3.176	0.206	0.698	0.452
5	2.115	4.550	3.333	1.553	2.653	2.103	0.263	0.286	0.275

1.3 Potential Impacts.

Potential impacts from the reservoir scheme may be divided into two categories:

- a) Direct impacts likely to occur in the short term and involving large changes in water or habitat quality, possibly resulting in significant and rapid changes in fish populations. This type of impact may well be mitigated by reservoir management and operating agreements.
- b) Indirect impacts involving subtle changes to the ecosystem and evinced as a change in selection pressure on fish populations, possibly resulting in changes of fish population abundance and species composition. There is insufficient understanding of the critical pathways involved to model the resulting changes in fish populations.

1.4 Recommendations For Future Work

Continued monitoring of adult and juvenile fish populations comprising similar components as this study, for at least two further consecutive years to begin establishing levels of variation in fish population abundance, including an intensive study of one reach to measure diurnal and seasonal variability in hydroacoustic results.

Identification of critical physical and chemical parameter limits together with food and habitat requirements at different life stages for key fish species identified in this survey.

Further comparisons of hydroacoustic measures of fish population abundance with results from catch depletion methods applied to the same site will add weight to the results of future surveys.

Electric fishing equipment should be modified in time for future surveys to allow effort to be quantified.

Options for gauging the extent and significance of fish migration should be explored.

4.0 RESULTS

4.10 Introduction.

Results from the studies described in this report are presented in this section; the data upon which they are based are appended where appropriate.

4.20 Electric fishing Results.

Electric fishing operations were carried out between 21:00 and 07:30 on 20, 21, 22 and 23 September 1993, from downstream of Sandford Weir to upstream of Benson Lock; a total river length of over 30 km. Twelve principal coarse fish species (Table 4) and 2109 individuals were sampled in the five reaches. One roach/bream hybrid was caught. Table 5 shows the number of fish sampled by electric fishing in each reach.

Table 4. Fish Species Sampled (All Reaches Combined).

Common Name	Generic Name
Barbel	Barbus barbus
Bleak	Alburnus alburnus
Bream (Common)	Abramis brama
Chub	Leuciscus cephalus
Dace	Leuciscus leuciscus
Gudg∞n	Gobio gobio
Perch	Perca Fluvitalis
Pike	Esox lucius
Roach	Rutilus rutilus
Ruffe	Gymnocephalus cernua
Silver Bream	Blicca bjoerkna
Tench	Tinca tinca

the difference in mean FL, which is significant at p = 0.02 (Reach 4) and p = 0.001 (Reach 5).

Roach length frequency for margin zones in Reach 4 (Fig 28) shows a marked bias towards smaller fish (Mean FL = 98.87mm) than for the centre channel zones (Mean FL = 149.51mm). The difference between these means is significant at p = 0.001. This is reflected, although to a lesser extent, by the length frequency for margin zones in Reach 5 (Fig 29), in which mean lengths (centre = 130.13mm, margin = 110.58mm) are significantly different at p = 0.002.

Length frequencies for perch between Reaches 4 & 5 (Figs 30 and 31) appear to be almost identical, with marked modes of 120mm and very similar mean lengths.

Population length frequency also provides a tool for comparison between hydroacoustic and electrofished samples, which is presented elsewhere in this report.

4.22 Species Percentage Abundance in Samples.

Species percentage abundance for all centre channel zone samples in each reach are presented in Figures 32 to 36 and for all margin zone samples in Figures 37 and 38. Table 6 below provides a summary of this information.

Table 6. Species Percent Abundance by Reach and Sample Zone.

Species	Reach 1	Reach 2	Reach 3	Reach	4	Reach	5
Zone	Centre %	Centre %	Centre %	Centre %	Margin %	Centre %	Margin %
Barbel	0	0.66	0	0	0	0	0
Bleak	32.76	51.50	59.58	26.82	13.26	67.97	18.60
Bream (Common)	2.69	3.99	3.75	2.35	0.55	0.78	0.41
Chub	1.96	1.66	10.83	4.47	0	4.30	1.24
Dace	1.22	0.66	1.25	4.71	0.55	0	1.24
Gudgeon	0.73	0.33	0	3.76	2.21	1.17	0.83
Perch	4.89	0.66	1.67	4.00	23.20	3.91	19.42
Pike	3.67	4.65	3.33	1.18	2.21	0.78	1.65
Roach	50.86	35.55	19.58	52.71	56.91	21.09	54.55
Ruffe	0	0	0	0	0.55	0	1.24
Silver Bream	0.98	0.33	0	0	0.55	0	0.41
Tench	0.24	0	0	0	0	0	0.41

Table 35. Total Values of Density and Biomass by Reach and Phase.

	Density	Echo Counting	(Fish 100m ⁻³)	Density	Echo Integ.	(Fish 100m ⁻³)	Biomass	(Kg 100m ⁻³)	
Phase	I	П	Mean	I	П	Mean	1	11	Mean
Reach		<u> </u>							
1	2.198	3.800	2.999	2.052	2.272	2.162	0.159	0.256	0.208
2	5.709	3.961	4.835	6.561	1.746	4.154	0.430	0.388	0.409
3	7.790	6.138	6.964	11.908	3.365	7.637	0.407	0.505	0.456
4	2.371	6.989	4.680	1.232	5.120	3.176	0.206	0.698	0,452
5	2.115	4.550	3.333	1.553	2.653	2.103	0.263	0.286	0.275

These three figures give a good summary of quantitative results collected, and they all show a generally increasing trend in biomass and density towards Reach 3 in phase I and Reach 4 in Phase II, with a decline in values in Reach 5 evident from the results for both phases.

These densities are similar to results for hydroacoustic surveys carried out on the River Thames at Cherstey (3.9 fish 100m³), The River Wey (6.5 fish 100m³), and the River Vltava at Prague (6.6 fish 100m³) (Duncan and Kubecka, 1993).

6.0 CONCLUSIONS

- Populations of similar length distribution were sampled in all five reaches.
- Bleak and roach populations sampled in margin zones from Reaches 4 & 5 have a significantly smaller mean size than populations sampled in centre channel zones for the same reach.
- Reaches 1 & 2 samples contained a greater number of species than Reaches 3 to 5. Margin zone samples contained more species than centre channel zones.
- Populations in Reaches 1, 2 and 4 show a similar relationship between length and weight.
- Pooled results for roach from all reaches shows an age length relationship similar to a nationally derived standard.
- Fish densities show scattered zones of high fish density in all reaches.
- Mean values of fish density and biomass are similarly low in Reaches 1 and 5 and similarly high in reaches 2, 3 and 4, although variation about the mean of results from both phases is relatively high.
- Populations sampled by both methods have similar length frequency distribution.
- Abingdon Marina provides an important habitat for smaller fish.

Figure 2. The Five Study Reaches.

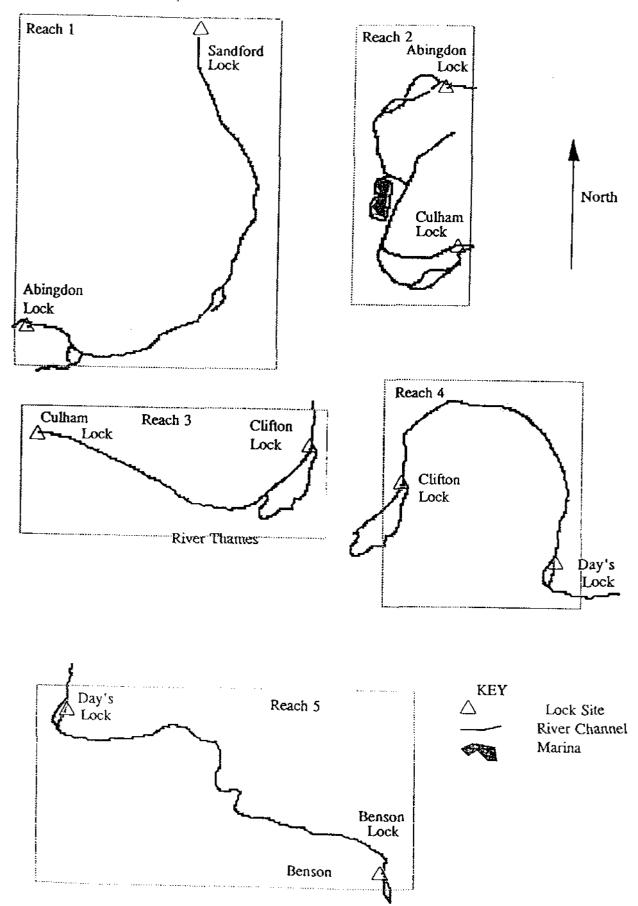


Figure 64. Total Fish Density (Echo Counting) by Reach Phase I, Phase II and Mean of Both.

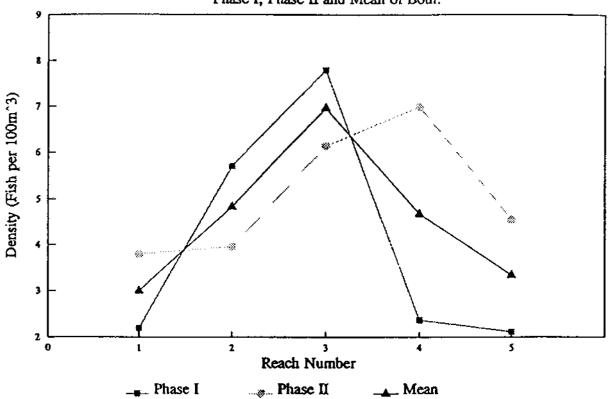


Figure 65. Total Fish Biomass by Reach
Phase I. Phase II and Mean of Both

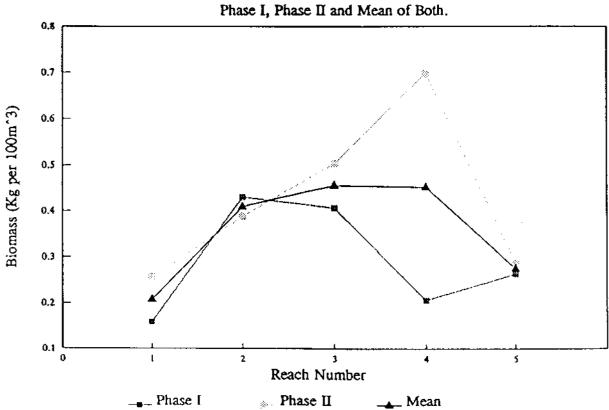
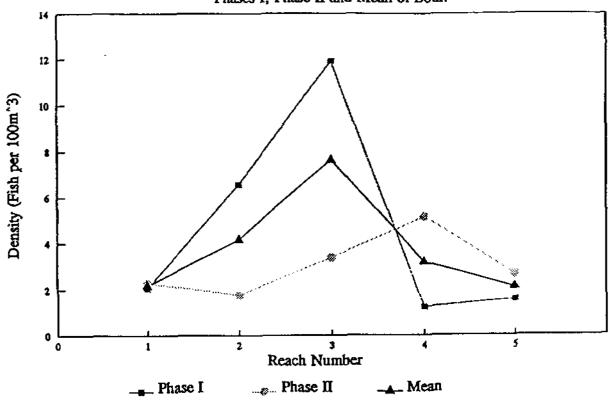


Figure 66. Total EI Fish Density by Reach Phases I, Phase II and Mean of Both.

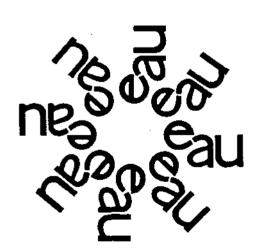


ABINGDON RESERVOIR: FISH STOCK ASSESSMENT

FRY SURVEY

FINAL REPORT

JUNE - SEPTEMBER 1991



ENVIRONMENTAL ADVISORY UNIT LTD. It therefore appears that there may be important differences in the distribution of species which are apparent when 0+ and 1+ age groups of fish are considered.

It was noted during sampling that more fry and older groups seemed to be captured on sunny days and were not present on the shallows in the evening. This possibly reflects a movement on to shallow water habitats to take advantage of the different spatial thermal regime available (O'Hara et al. in prep). Abundance data and field records were of insufficient rigour to test this hypothesis.

5. AGE AND GROWTH

Mean lengths of fish at age 1 were calculated from the June sample on the assumption that their birthdate was 1st June. The analysis of growth in the subsequent samples from July and September concentrated on the size of the O+ fishes since the objective of the study was to investigate the characteristics and vulnerability to capture of fry. Older fishes were generally aged as 1+ and older because the length distributions overlapped. The results are presented in tables 6, 7 and 8. There is perhaps some evidence of variation in the mean length of fish between sites but without detailed statistical analysis and confirmation of ages by scale reading particularly for 1+ and older fish this cannot be confirmed as significant. It should also be noted that variation in measurement by different operatives has not been accounted for, to overcome this potential problem the measuring team should be kept consistent and should be experienced in the identification of juvenile stages.

The mean length of fish achieved at the end of their first year of life was toward the mid-range of results from other rivers in Britain (see Williams 1967, Mann 1976 & 1982, Weatherley 1986). However it should be noted that there is some discussion over the accuracy of some studies because small fishes may not have been representatively sampled or aged accurately. In contrast to other methods of capture, we consider that micromesh seine netting does provide an accurate estimate of fish growth rates for young of the year.

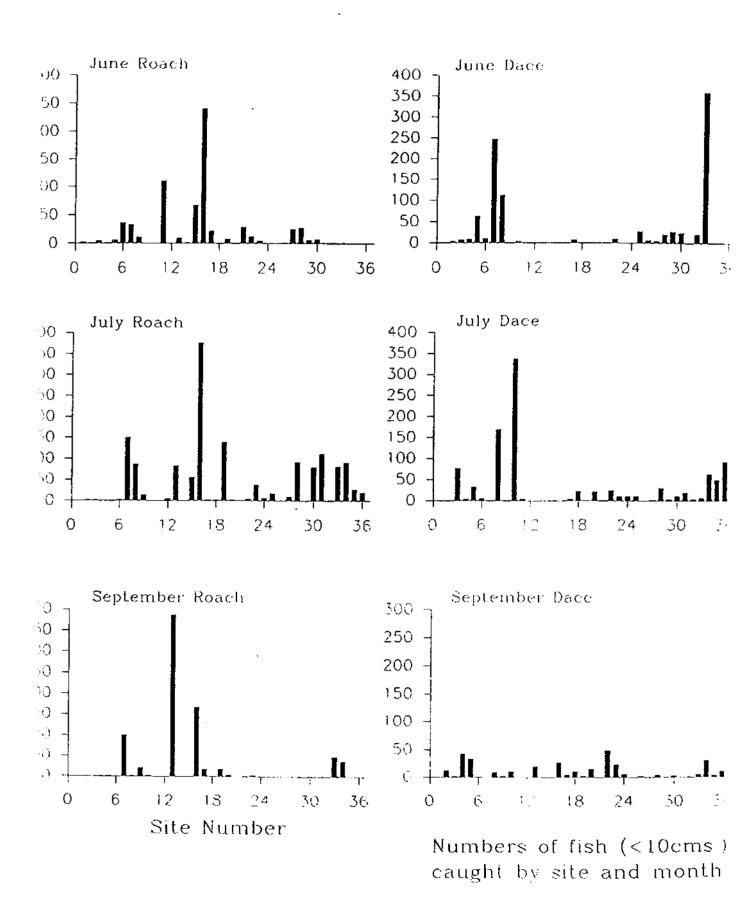
6. CONCLUSIONS AND RECOMMENDATIONS

1. On the basis of this year's work we consider it is possible to capture most of the recreationally important species in their year of hatching by late July. The exception to this appears to be bleak and this may reflect a naturally late spawning time for this species.

- 2. Fish hatched in the previous year remained vulnerable to capture by micromesh seine netting until late July.
- 3. For most species of fish, and assuming normal climatic conditions, it is possible to obtain relative abundance data and thereby information on year class strengths for fish from 2 years by sampling in late July. This would enable not only an estimation of spawning success but also overwinter survival and mortality.
- 4. Ideally a thorough appraisal of the causes of variability in catches between months, between sites and within and between days should be undertaken. However, in the absence of such a rigorous strategy, the use of more samples taken over a restricted single time period, possibly at a limited number of sites, may provide a better database of relative abundance than sampling over three time periods. This would give a more accurate estimate of the relative abundance but would not give such good information on distributions along the length of the river. This may be important, particularly if the observed differing distributions of age classes is a real phenomenon that is repeated in future years.

It was suggested that there is an indication of reduced abundances in the more "channelised" areas of river but this may reflect difficulties in sampling rather than species abundance. Further work would be needed to validate this observation but it may be advantageous to establish a sampling strategy to include areas along the river where the banks shelve gently. It would prove possible with this sampling regime to retain sites above and below the projected location of the reservoir.

- 5. Any sampling programme should be conducted in the most appropriate month to sample and from the present study this appears to be late July possibly extending into August. This period would combine the likely time of warm weather with that when most species are of a sufficient length in their first year to be vulnerable to the capture method.
- 6. Ageing of 1+ fish from scales should be undertaken to validate the age structure of the population. For younger fish, length frequency analysis is a satisfactory method of ageing.
- 7. It is important to maintain a strict quality assurance when measuring and identifying small fish to prevent errors developing. Personnel experienced in the identification of juvenile stages should undertake sample processing.



ANGLER CATCHES IN THE RIVER THAMES BETWEEN SANDFORD LOCK AND DAY'S LOCK

1993 - 1994

S N Hughes.



National Rivers Authority
Thames Region

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SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY A REPORT FOR NRA THAMES REGION AND THAMES WATER UTILITIES LTD

ANGLER CATCHES IN THE RIVER THAMES
BETWEEN SANDFORD LOCK AND DAY'S LOCK

1993 - 1994

S N Hughes.

1.0 EXECUTIVE SUMMARY

The aims of this study were to identify and measure catches of match and non-match anglers during the coarse fishing season within the SWORP study area and to quantify use of the fishery by anglers and thereby gauge the importance of the study reach of the River Thames as an angling resource.

1.1 Methods

The study reach of the River Thames was divided into three subsections, each forming a standard walk of between 3 and 4 km and assigned a colour code:

Blue - Sandford Mill to Radley, Right bank Red - Abingdon Bridge to Culham, Left bank White - Clifton Hampden Bridge to Day's Lock, Right bank

Each reach was walked by a member of NRA Fisheries staff at monthly intervals on Sundays between 20/6/93 and 13/3/94 inclusive. Anglers encountered on the walk were divided into two categories:

Match anglers - Involved in an organised angling match. Non-match anglers - not involved in an angling match.

Non-match anglers were interviewed and the following information recorded:

- a) Distance (m) to the nearest access point.
- b) Distance & time travelled to site.
- c) Length of time spent fishing prior to interview.
- d) Number of fishing rods in use.
- e) Method used (ie float, leger, lure etc).
- f) Bait used.
- g) Species targeted.
- h) Number of days in the week and which days normally fished.
- i) Number of times fished the reach during the current season.
- j) Number of years fishing experience.
- k) Rating of own experience (Novice/Intermediate/Very experienced).

Total weight of any catch was estimated and species composition noted by NRA staff on examination of catches retained in keepnets. A tally of numbers of principal species in 50 mm fork length (FL) size classes were recorded. If an organised match was encountered, anglers were not interviewed but match results were obtained at a later date. Total catch (kg), effort (rod hours) and catch per unit effort (CPUE) in grammes per rod hour were calculated.

Attempts to collect match angler catch information from large matches were unsuccessful due to poor weather and river conditions.

Match CPUE results were classified using the NRA National Fisheries Classification Scheme.

1.2 Results.

Species occurrence in non-match angler catches differ between reaches, bleak are dominant in the Blue reach catch, roach are dominant in catches from the other two reaches.

Maggots are the most common bait, the most common fishing method is the swimfeeder, followed by

float and leger in most reaches.

The majority of non-match anglers do not target specific fish species.

Anglers in the Red and White reaches tend to be within 250m of the nearest access point.

The majority of anglers interviewed had between 15 and 20 years angling experience.

The Red reach is principally used by local anglers, the Blue reach is equally used by local and non local anglers and the White reach is principally used by non local anglers.

Each reach achieves Class A rating (in the upper quartile of national results) for the match fishery.

The greatest total angling effort (Match and Non-match anglers combined) is concentrated on the White reach, followed by the Blue and Red reaches.

1.3 Recommendations for Future Work.

Further studies of this kind are essential to fully describe this important and valuable fishery resource. Future studies should make collection of match angler catch information a priority, and should investigate alternative methods of data collection if necessary.

5.0 DISCUSSION.

5.10 Methods.

5.11 Angler Categories.

The main reasons for the division of anglers into two groups was principally undertaken to reflect the different needs for data collection in the field. There is also some evidence to suggest that match anglers actively target particular species (Steele, pers.comm.) which may influence some of the results.

5.12 Non-match Angler Sampling.

The standard walk creel census is widely used in the USA and is a very successful means of collecting information on both the biological and recreational aspects of the fishery (Bayley, pers.comm.) Building upon the experience of North American exercises, and drawing from work being carried out in the UK for the NRA as part of the research and development programme, a questionnaire was designed to fit the requirements of this investigation. Sunday was selected as the most suitable day on which the maximum number of anglers would be fishing, based upon the experience of the NRA TR Water Bailiffs. As such, all the results of this study relate to data collected on Sundays, and may not accurately reflect use of the resource at other times.

The majority of anglers interviewed were very cooperative and welcomed the opportunity to actively contribute to the study.

Quality assurance checks on fieldwork were carried out by Fisheries Officers, and no problems were identified. A planned exercise to compare estimated to actual catch weights was inadvertently not completed, however this is planned to be carried out during the 1994/5 survey, and the findings retrospectively applied to this study.

5.13 Match Angler Catch Sampling.

A method was developed to collect this information with as little disturbance to anglers as possible. In order for the method to succeed, a large number of Fisheries personnel were required to assemble at a weekend to collect catch data. This inevitably required a long time to organise and resulted in a fairly inflexible sampling strategy. In order to justify this mobilisation of resources, it was planned that only large matches would be targeted in order to collect as large a sample as possible. On each occasion that the numbers of staff were available and sufficiently large matches were planned, the weather conditions deteriorated and very high river flows resulted in extremely low catches.

5.20 Non-match Angler Results.

5.21 Species Composition.

The results presented for the Blue reach are based upon catches from only 10 angler's catches, compared with around 50 in both the other reaches. There is therefore little value in comparing these species occurrence results with those from the other two reaches.

Species composition of catches in any fishery is a function of many different factors, including the method and bait used, species targeted and species composition of the fish population. It is possible to account for some of these factors in this survey.

Results from previous electric fishing surveys (Hughes, 1994) indicate that the species composition of

the fish populations in each reach are broadly similar. Anglers were using similar methods and baits in both reaches, and the majority of anglers were not targeting any particular species.

In general terms, the species occurrence in catches from the Red and White reaches show a similar range of species, with roach as the most abundant and perch and bream contributing a significant proportion. The difference in contribution of other species to the catch make up is difficult to explain and is probably due to a combination of many subtle selection parameters, which may include gear and bait selectivity, time of day fished and river flow, temperature and turbidity conditions.

5.22 Length Distribution.

Length distribution data were collected to provide a guide to the length composition of catches, rather than identify individual year classes, which are likely to be masked by the relatively large class width employed.

The results show fish over a wide range of lengths are caught in both the Red and White reaches. (results for the Blue reach are not presented for the reasons outlined above). The length ranges and clear modes for roach and perch appear to be similar to those described for fish sampled by electric fishing in the same reach (Hughes 1994). This evidence suggests that fish from across the whole population are being caught by anglers.

This is further confirmed by the length selection ogives which demonstrate the similar length (and age) structure of roach perch and bream captures in both reaches. It is apparent that roach and bream do not become vulnerable to angling before reaching an approximate age of 1+.

The apparent difference in length structure of chub catches between the two reaches may well be anomalous due to the small proportion of chub in sampled in the Red reach. If not, it indicates a greater selection for smaller fish in this reach, the reasons for which are unclear.

5.23 Fishing Methods.

Fishing methods were divided into a number of categories to reflect the range of methods used by anglers. Conventional methods suited to the relatively deep, slow flowing River Thames were by far the most popular within each reach, with legers, swimfeeders and float being the three most common. These methods are selective for certain species over others, and are suitable for the most common small fish species found in the river. A similar range and number of each method in each reach reflects the similar nature of the river and fish populations within it, and must be a large factor in determining the catch species content.

Maggots formed the principal bait item, worms and caster were secondary, reflecting the wide availability of these baits and the general acceptance by anglers of their efficacy.

5.24 Reach Use and Angler Experience.

In the Red and White reaches, the number of days fished by the large majority of anglers was two or less and this is probably a reflection of a need to restrict fishing to weekends. The relatively low sample size for the Blue reach may have biassed the results. Anglers who fished most often (5 days a week) were found in this reach.

The distance of anglers from the nearest access point is an indicator of how evenly spread anglers are along a reach, and highlight areas where effort is concentrated. The equipment-conscious modern angler may be less willing to transport a large amount of fishing gear any significant distance along

a river bank, especially where access paths are not good. The red reach results show the greatest amount of clumping around access points, and this may be a reflection of the reach passing through the town of Abingdon and the presence of Abingdon Town Council fishery - a reach of the Thames that is covered by a general licence and free to residents of Abingdon. The White reach shows a similar but less marked concentration of anglers, despite having a excellent bankside path along the whole reach.

The results show that anglers in these reaches tend to remain relatively close to the access points, and similar results have been demonstrated for rivers in the NRA North West Region (Steele, pers. comm.).

In the Blue reach, anglers are much more widely distributed along the whole length. Again this reach has an excellent footpath and access to a large car park at the upstream extent of the reach. There is, however, no evidence to suggest why the pattern of the other two reaches should not be repeated. It is possible that with much more frequent matches on this reach, non-match anglers were forced to travel further before finding a suitable spot.

Information on angler experience is important as it provides an insight into the commitment and ability of anglers utilising a reach. Both of these factors may have some influence on catching ability and therefore on CPUE values. The results show that there is not a great difference in the experience profile (in terms of years involved in the sport) of anglers between reaches, and reflect the vast resource of experience that exists on the river bank. It is interesting to compare this information with a self assessment of experience, which showed a great deal of variation between reaches, and is probably not an objective measure. One angler who had been fishing for over 50 years rated his own experience as "Novice". For the purpose of this type of study, the number of years an angler has been involved in the sport is sufficient, and a self assessment of experience should not be collected in future surveys.

The distance travelled by anglers to a fishery is one means of assessing its financial worth, assuming that there is a relationship between distance travelled and cost to the angler. In other words, a greater distance travelled to the fishery will, in general, reflect a greater cost incurred by the angler to reach that fishery. There are a number of reasons why anglers may wish to travel a greater distance; the lack of suitable venues near to home for example. These results highlight the relative importance of a fishery to local and outside anglers. Information on time spent travelling provides a useful check against these results; the distribution of each should be similar.

Results for the Red reach demonstrate that the fishery is principally used by local anglers, with half of the anglers travelling less than 5 km and for less than 10 minutes to the fishery. This is likely in part to be due to the Abingdon Town fishery described above.

The Blue reach appears to be used principally by anglers living within 15 km of the reach and travelling for less than 30 minutes, and these anglers probably come from Oxford and the surrounding towns. There was, however a fairly large proportion of anglers (about 30%) who had travelled for 80 km or more.

The white reach is a further distance from towns with Oxford, Abingdon and Didcot between 1 - 15 km distant. This is partly reflected in the distance travelled by anglers to the reach, with a greater proportion travelling further than to the other reaches.

5.25 Catch, Effort and CPUE.

On their own, catch and effort results provide an insight into the frequency and scale of angling in a

reach, and therefore a means to gauge its importance as a recreational resource. CPUE gives an indication of anglers' rate of success in a reach, and therefore a means to gauge the quality and value of the biological resource.

Total catch and effort values for each reach place the Red reach highest on terms of effort, meaning that this reach is the most frequently used by non match anglers during the sampling period. The White reach was the second and the Blue reach the least frequently used by this group of anglers.

Peaks in effort were seen in the summer months, between June and September and are probably the result of the generally more comfortable weather conditions at this time of year.

Catch per unit effort results rank the Red reach highest, followed by the Blue reach and the White reach. This does not reflect the different population abundance figures measured in each reach (Hughes 1994), and there is still some question over the validity of angler catches as indicators of fish population abundance in UK rivers (Steele, O'Hara and Aprahamian, 1994).

If these CPUE values were for match anglers, they would place the reaches in class A, B and C of the National Fisheries Classification for match fisheries. This scheme was not developed for non-match angler catches, and there is some evidence to suggest in this study that they are not comparable. In this case, these classifications should therefore not be treated as definitive.

5.30 Match Angler Results.

Match angling represents an important component of the total fishing effort applied and is greater than non-match angling effort in each reach. The distribution of match effort throughout the season varies between reaches, the Blue reach experienced fairly constant and relatively low levels of effort to the White reach which had a few very high levels of effort in the season. Results for the total effort (by both groups of angler combined) applied to each reach place the White reach as the most significant, followed by the Blue and Red reaches.

CPUE results for match anglers consistently place all reaches in the highest category for match fisheries nationally, confirming their excellent quality. CPUE values follow the ranking for total effort above: Red reach with the highest value, followed by Blue and White reaches.

5.31 Historical Match Data.

The CPUE data again confirm the consistently excellent quality of match fisheries in the White and Blue reaches of the Thames since 1989. The lower values noted in 1993 reflect feedback from angling clubs whose members have noted a reduction in match catches over the 1993/4 season, which they largely attributed to the perceived higher flows experienced this year making match angling more challenging.

6.0 CONCLUSIONS

- Species occurrence in non-match angler catches differ between reaches.
- Bleak are dominant in the Blue reach catch, roach are dominant in the other two.
- More young chub are caught in the Red reach than the White reach.
- •50% of roach caught in each reach were aged 3+ or less.
- · Maggots are the most common bait.
- The most common fishing method is the swimfeeder, followed by float and leger in most reaches.
- The majority of non-match anglers do not target specific fish species.
- Anglers in the Red and White reaches tend to be within 250m of the nearest access point.
- The majority of anglers in each reach have between 15 and 20 years experience.
- The Red reach is principally used by local anglers.
- The Blue reach is equally used by local and non local anglers.
- The White reach is principally used by non local anglers.
- The order of recreational value for the three reaches to non match anglers is (best first) Red, White, Blue.
- The order of fishery quality for non-match anglers is Red, Blue, White.
- The order of recreational value to match anglers and both groups combined is White, Blue, Red.
- The order of match fishery quality is Red, Blue, White.
- Each reach achieves Class A rating for the match fishery.

ADULT FISH COMMUNITIES OF THE RIVER THAMES BETWEEN SANDFORD AND BENSON LOCKS 1994 VOLUME 1 - EXECUTIVE SUMMARY

Simon Hughes Fisheries Officer



National Rivers Authority
Thames Region

SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY A REPORT FOR NRA THAMES REGION

ADULT FISH COMMUNITIES OF THE RIVER THAMES BETWEEN SANDFORD AND BENSON LOCKS 1994 VOLUME 1 - EXECUTIVE SUMMARY

Simon Hughes Fisheries Officer

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1.0 EXECUTIVE SUMMARY

1.1 Study Area and Methods.

Adult fish populations in five contiguous reaches of the River Thames separated by locks between Sandford and Benson Locks were surveyed as a repeat of the SWORP 1993 Adult Fish Communities investigation. Hydroacoustic methods were used to provide a quantitative assessment of fish populations in two phases of field work during mid July and late November, electric fishing (July only) provided a sample of 3943 fish from all five reaches which provided comparative species specific information. Electric fishing samples were taken separately from the centre channel zone and the margin zone; both electric fishing and hydroacoustic work was carried out at night.

The methods and equipment used were the same as employed in the 1993 study, and catches by each method were validated by comparison of length frequency results; they seem to be sampling statistically similar populations.

Results from Phase II hydroacoustic work were hampered by gear failure and a reduction in the vulnerability of fish to sampling by the equipment through a seasonal reduction in activity observed by other workers in mid autumn.

1.2 Results and Conclusions.

A total of fourteen coarse fish species and roach/bream hybrids were sampled - two more species (carp and eel) were found this year. Bleak and roach were the most abundant species in both centre and margin zone samples. Species composition was similar to samples taken in 1993 with small changes in some reaches, and shows parity with a sample of fish taken from the River Thames at Reading in 1958/59.

Age frequency results provide a comparative index of year class abundance, although for bleak and bream there is evidence to suggest incorrect ageing of some length classes. Results for other species appear to be of good quality and provide key information on their population dynamics.

Recruitment to the adult bleak population appears to be excellent in the past few years, but may be independent of 0+ densities of a given cohort, although possible errors in assigning age to length classes may have masked a relationship. The lack of older year classes may be due to incorrect ageing, but could also be a demonstration of the sensitivity of this population to environmental change, although growth is normal.

The bream population has missing cohorts, which may be a sampling artefact, a problem with the ageing process, or a reflection of recruitment success. Extremely successful cohorts illustrate the potential for change in this population. Results indicate that factors affecting growth do not limit this population's balanced development.

Chub populations are generally stable but appear to be able to exploit favourable circumstances, resulting in some successful cohorts. Growth does not appear to be a limiting factor.

Dace populations have reduced since 1993, but are not widely distributed in the study area in either year.

Pike populations show a sensitivity to change that may be related to a reduced rate of growth,

although food items (ie roach) are plentiful. There is supporting evidence that pike are poorly represented within the study area, which may be a function of the influence of bioaccumulant pollutants upon the balanced development of this population.

The roach population is susceptible to enormous changes in cohort abundance, although in the last four years, recruitment to the adult population has been very good. A depressed growth rate is evident possibly due to intraspecific competition for food, and may be a factor in population change. The variability of growth with age may be a result of changes in preference or availability of diet items. The limited evidence available does not suggest a relationship between 0+ abundance and cohort success.

All species have acceptable levels of parasite infestation that would be expected from a river of this type.

There is evidence to suggest that electric fishing catch per unit effort (CPUE) is closely related to acoustic density for each reach. Even if electric fished samples are not fully quantitative, they do provide an index of relative change between years.

A summary of hydroacoustic density by reach is shown in Table I.

Table I.	Fish	Densities	and	CPUE	by	Reach	and	Year.
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Reach		CPUE (n min ⁻¹)		
	1994 (July)	1993 PI (July)	1993 PII (September)	1994
1	3.87	2.20	3.80	2.77
2	3.38	5.71	3.96	3.30
3	10.36	7.79	6.14	6.78
4	8.27	2.37	6.99	3.80
5	4.83	2.12	4.55	4.81

The results show a degree of consistency between years with a close relationship to electric fishing CPUE. Increases in density over time appear to be related to a reduction in mean FL, suggesting a recruitment of smaller fish to the population. It is difficult to suggest reasons for the difference in fish density between reaches, although it is likely to be a function of available habitat, water and habitat quality.

There is considerable spatial variability in fish density, but little evidence for the key factors influencing fish aggregation about a given point.

The fisheries status of the study area appears to be very good in some reaches, but only moderate to good in others. It is comparable to other reaches of the River Thames, and slightly better than parts of the River Ouse.

Most of the conclusions from this investigation relate to aspects of fish population dynamics and development that are likely to be strongly influenced by interactions within and changes to the trophic links of the river ecosystem, possibly initiated by environmental change. Some parts of the population have been shown to be extremely sensitive and capable of conspicuous change, which could be related to change in growth characteristics. Given the potential for the proposal to impinge upon the ecosystem of the river, it is essential that we understand as much as possible, the mechanisms and links within this system.

1.3 Recommendations.

Further studies of this type are important in allowing us to compare spawning success and subsequent recruitment to adult populations in the study reach; cohorts that were assessed in 1992/3 are only now being fully sampled by the equipment we have available. The comparison of juvenile abundance to cohort success is important in identifying a number of factors, principally the influence of spawning success, environmental factors, juvenile fish growth and survival on cohort success. A description of these relationships may allow prediction of the impact of the proposal on fish populations. The best sampling methods available can only sample a cohort fully about two years after hatching, so this type of survey must continue for a minimum of two years after the final juvenile fish survey.

The results of future studies will allow us to identify any trends in the changes fish populations are undergoing, and will begin to identify the background or natural variability in these populations. A robust statistical investigation of the principal factors affecting observed variance in results will describe the degree of change to the population that we will be able to detect with the methods available.

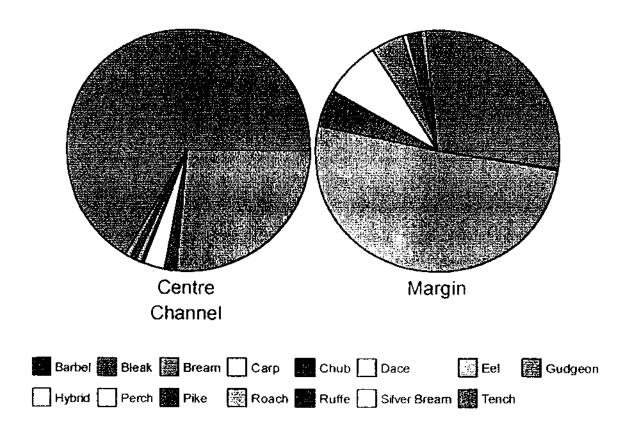
A study of the status of bioaccumulant pollutants in pike (as a top predator) in the study reach should be carried out to assess whether this is a factor affecting the balanced development of this population.

An assessment of the preferred food items of 0+ and 1+ dace (and potential competitor species) should be made to identify whether food availability is a limiting factor to the success of this species in the study reach.

All field work in studies to be compared with this one should be carried out in July to allow more robust comparisons of length and growth data. (A study of the influence of sampling time season on hydroacoustic measures of fish abundance should be made to determine whether July is the optimum sampling period.)

A quality audit of scale age data provided by NRA Anglian Region should be carried out.

Figure II. Species Percent Occurrence.
Reach 1



Combined Reach Total ŝ Reach 5 13 শ នា ᄗ Reach 4 6 17 15 Run No. Reach 3 = Reach 2 = Reach 1

gure VI. Fish Density by Reach.

ADULT FISH COMMUNITIES OF THE RIVER THAMES BETWEEN SANDFORD AND BENSON LOCKS 1994 VOLUME 2 - MAIN REPORT

Simon Hughes Fisheries Officer



National Rivers Authority
Thames Region

SOUTH WEST OXFORDSHIRE RESERVOIR PROPOSAL STUDY A REPORT FOR NRA THAMES REGION

ADULT FISH COMMUNITIES OF THE RIVER THAMES BETWEEN SANDFORD AND BENSON LOCKS 1994 VOLUME 2 - MAIN REPORT

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6.0 CONCLUSIONS

The results presented and discussed in this report lead us to draw a number of conclusions about the status and sensitivity to change of fish populations in the study reach, as well as about how they appear to have changed in the period of one year. We cannot yet identify trends in these changes, but a number of recommended future studies should allow this.

6.1 Methods.

The methods used provide statistically similar samples of the fish population and there is some positive correlation between hydroacoustic densities and electric fishing catch per unit effort.

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Electric fishing is more effective in the shallower margin zones than the deeper centre channel zones, but it is important to sample both areas to minimise sampling bias.

6.2 Species Occurrence, Population Dynamics and Health.

Species diversity provides a basic assessment of fishery quality and the relatively small variability between reaches and years is probably a product of the sampling methods used. The difference in species composition between the margin and centre channel samples is probably due in part to species habitat preference.

Bleak and roach are the most abundant species in all reaches, and perch are the third most abundant species in many margin samples.

Recruitment to the adult bleak population has been good in the past although possible errors in assigning age to length classes may have given misleading results. The lack of older year classes may be a result of this error, but could also be a demonstration of the sensitivity of this population to change.

Missing cohorts in the bream population may be an artefact, however if this is not the case, age frequency results suggest a population that is relatively sensitive to change, but that does not appear to be limited by growth. It will be possible to relate juvenile abundance to year class success in this species from the results of future surveys.

Dace populations have reduced since 1993, but are not widely distributed in the study area in either year.

Chub populations appear generally stable with some variability in year class abundance. Growth does not appear to be a limiting factor, and the relationship between juvenile abundance and cohort success should be possible to describe with results from future surveys.

Pike populations show a sensitivity to change that may be related to growth although their preferred food items (roach) are abundant. There is evidence that pike are poorly represented in the study area compared to other large lowland rivers, which may be a function of the influence of bioaccumulant pollutants upon the balanced development of this population.

The roach population is susceptible to enormous changes in cohort abundance, although in the last several years recruitment to the adult population has been good. A depressed growth rate is evident, possibly due to intra-specific competition for food, and may be a factor in population

change. There is a marked variability in growth rate with age which may be a result of changes in preference for or availability of diet items. The limited evidence available does not suggest a relationship between juvenile abundance and cohort success.

All species have acceptable levels of parasite infestation that would be expected from a river of this type.

6.3 Population Abundance.

There is evidence to suggest that electric fishing catch per unit effort (CPUE) is closely related to acoustic density for each reach. Even if electric fished samples are not fully quantitative, they do provide an index of relative change between years.

The results show a degree of consistency between years with a close relationship to electric fishing CPUE. Increases in density over time appear to be related to a reduction in mean FL, suggesting a recruitment of smaller fish to the population. It is difficult to suggest reasons for the difference in fish density between reaches, although it is likely to be a function of available habitat, water and habitat quality.

There is considerable spatial variability in fish density, but little evidence for the key factors influencing fish aggregation about a given point.

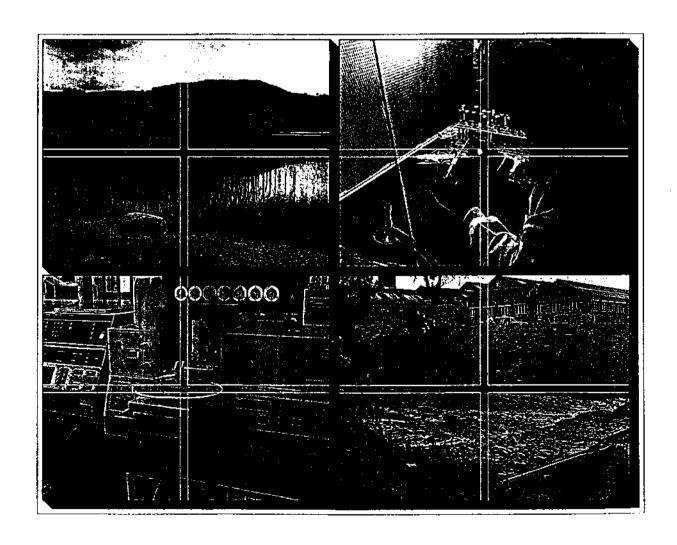
6.4 Denouement.

Most of the conclusions above relate to aspects of fish population dynamics and development that are likely to be strongly influenced by interactions within and changes to the trophic links of the river ecosystem, possibly initiated by environmental change. Some parts of the population have been shown to be extremely sensitive and capable of drastic change, which could be related to change in growth characteristics. Given the potential for the proposal to impinge upon the ecosystem of the river, it is essential that we understand as much as possible, the mechanisms and links within this system.



Strategic Water Resources Assessment: Review of Great Ouse (1988-1993) and River Thames (1958-1973) Fisheries Community Data

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EXECUTIVE SUMMARY

The report reviews studies of coarse fish communities undertaken on the River Thames (Dreadnought Reach) during 1958-73 and on the River Great Ouse since 1988. The results are examined together with a wide range of other relevant studies to identify the main environmental influences on the fish communities of rivers like the Thames, and to comment on the sensitivity of fish to these influences. The review is intended for comparison with current studies on the Thames and to highlight requirements for further, focused studies.

Habitat requirements change as fish grow and develop. Changes are particularly rapid during the first year of life and it is important to maintain habitat diversity in the river. Young fish need cover to provide protection from predation and from high water velocities (over 2 cm s⁻¹). Beds of water lilies and other areas of slack water are important habitats.

Lily beds are also important feeding areas because they contain large populations of small invertebrates during the summer months. Small invertebrates may be a limiting resource for the growth of some species, such as roach, dace and small perch.

Older fish have greater ranges of habitat tolerance but, in all species, the requirements for successful spawning are shown to be more precise than those for feeding and refuge.

Water temperature, current velocity, food availability, refugia and spawning habitats are identified as key factors for the fish community.

It is recommended, in order of priority, that:

- 1) Further studies should be made of the habitat preferences and diel movements of 0 group fish and of the requirements for successful over-wintering.
- 2) Data collected from the Dreadnought Reach of the Thames in 1970-71 should be analyzed more fully to show the effects of a large reduction in the roach population during the 1960s, and to provide information intermediate between 1958-59 and current studies.
- 3) Sonar detection and sonic tags should be used to provide information on the spawning (and other) movements of adult fish, and on the effect of locks and weirs on such movements.
- 4) An assessment should be made of the availability of spawning habitats for selected fish species from the high priority Category A list (barbel, common bream, chub) that appear to have limited recruitment in the Thames.

group fish can feed and find cover from predation and high water current velocities. The young fish need areas with a flow <2 cm s⁻¹. Nuphar beds and other areas of slack water are important habitats, which need to be maintained in the river.

The studies of 0 group fish in the Great Ouse confirm the importance of water temperature in determining growth rates, especially in the first two or three months of life. However, they also showed evidence of food limitation in mid/late summer, when most if not all the fish species had spawned and the eggs had hatched. Changes in river temperatures induced by anthropogenic changes could add to the year-to-year variations that occur naturally though climatic influences.

Much of the food taken by young fish is cladocerans and chironomid larvae known to be in the <u>Nuphar</u> beds. There is evidence of scarcity of invertebrate food for several species of fish. <u>Nuphar</u> beds are the richest areas for invertebrates during the summer months and more food would be available during the main growth season of the fish if there was a greater area of <u>Nuphar</u> in the Thames. Although the data from rivers other than the Thames provide strong indications of the optimum environmental conditions for most species, they also show the need for local knowledge concerning the river and its fish community.

The studies on the Thames, based at the University of Reading, continued from 1958 to 1972. However, the study of the growth, survival and population densities of larger fish were based on samples taken in 1958-59, and those of juvenile fish were based on samples taken in 1967 and 1968. Studies on the energetics and production of the fish and of the ecosystem as a whole are based on these two sets of samples. The study of larger fish was repeated in 1970-71 but the data were never fully analyzed because Dr Berrie left the University to take up another post before this had been achieved. These data are still of considerable interest because of the changes that took place in the structure of the fish community between 1958-59 and 1970-71.

5.2 Summary of key factors

The key environmental factors influencing critical life stages of coarse fish are summarized below. Note that these factors do not operate in isolation; they are often interrelated and may act on fish in a synergistic manner.

5.2.1 Water temperature

Above average temperatures decrease egg incubation periods, increase fish growth rates and improve the swimming ability of 0 group fish. All these elements lead to higher 0 group survival rates and to improved year-class strength. The timing of spawning is partly temperature controlled, and some species require high temperatures (>18°C) before spawning occurs (e.g. tench, carp). Sudden decrease in water temperature during the spawning period can inhibit spawning activity and may cause egg resorption.

5.2.2 Water velocity

Velocities > 2 cm s⁻¹ can lead to the displacement of 0 group fish, especially during their first two or three weeks of life. Hatching of the different fish species may extend from April to July in many rivers, including the Thames. Consequently, newly-hatched fish are present throughout this period and will require suitable refugia.

Most fish species require some flow of water over their spawning substrata to enable the eggs to be well aerated. The flows observed for the target species are indicated in Table II.; they range from near zero in backwaters (tench, pike) to c. 50 cm s⁻¹ on gravel beds (dace, chub, barbel).

5.2.3 Food availability and fish refugia

These two elements are often related, especially with respect to 0 group fish. Marginal vegetation and/or backwater areas provide shelter from high water velocities for 0 group fish, and also constitute feeding areas.

O group fish require small food particles when they start feeding on external food sources (as opposed to feeding on yolk sac supplies). Synchronisation of the phytoplankton and zooplankton cycles with hatching sometimes occurs but is not consistent between species or between years. The phytoplankton cycle is influenced by river discharge rates in the early spring and on the light regime. The timing of this cycle influences that of the zooplankton population (especially the rotifers), which eats the phytoplankton.

There is evidence from the Great Ouse studies that food supplies for 0 group fish can be limited and that this can override the influence of water temperature on growth rates. This may cause some species to switch to other food sources. Thus, roach in the main river channel switch to a detritus diet in July, but continue with a zooplankton diet in marina backwaters. Similarly, roach, dace and young perch in the Thames appeared to be short of small invertebrate food at the time of the studies. This led to a high dependence on organic detritus and to low rates of growth. The effect on survival is not known.

5.2.4 Spawning habitats

Older fish have a wider range of general habitat preferences than 0 group fish, but the spawning requirements for each species are defined more narrowly. Gravel beds washed by fast-flowing water are needed by barbel, dace and chub, whereas most other species spawn on various plant substrata in more slow-flowing areas of the river. These include submerged and emergent aquatic plants, and the submerged tree roots of willow and alder trees. Pike, tench and carp generally prefer vegetated areas with water velocities < 5 cm s⁻¹.

Access to spawning areas is vital, but little is known concerning the extent of coarse fish spawning migrations and how these may be affected by locks, weirs and sluices (see below).

5.3 Recommendations for future research

5.3.1 General

The review highlights the need for more information on the habitat selection and diel movements of different species of 0 group fish, although it is known that habitat preferences vary between species and with the growth and development of individual species. Also, very little is known about the requirements for successful over-wintering, although the Great Ouse studies point to the importance of backwater areas such as marinas. The growth and early survival of 0 group fish are critical aspects in the life stages of coarse fish, and are those most likely to be affected by changes to the river environment and hydrological regime.

Age determinations on the scales and further analyses on the 1970-71 data from the Dreadnought Reach of the River Thames will show whether the large reduction in the roach population during the late 1960s produced other effects on the fish community. The growth rate and food of roach could have changed and there could be consequential effects on other species. This study will also provide growth and age structure information for the fish community that is intermediate in time between the 1958-59 data and any recent studies by the NRA. As the material has already been collected, the cost of the study will be much smaller than if field work was involved.

Further information is needed on the movements of adult fish, particularly for spawning but also at other times of the year. The role of locks and weirs in relation to fish movements is unknown and may be of considerable importance. The use of sonar detection and sonic tags in these respects should be considered, especially for the major fish species.

Information from the 1970-71 studies indicates that, of the nine Category A species, the roach, bleak, dace, gudgeon and perch all spawn successfully in the Thames (Dreadnought Reach). Progeny of barbel, common bream, chub and pike were less numerous, which may point to restrictions in the spawning habitats. Emphasis on the spawning movements of barbel, common bream and chub would be valuable if it is intended to enhance the stocks of these species.

5.3.2 Priorities for River Thames studies

The following recommendations are placed in order of priority, but it should be noted that many aspects are interrelated.

a) Determine the distribution of suitable marginal, vegetated refuge areas for 0 group fish, especially areas where the flow velocities are less than 2 cm s⁻¹ under most summer

Table I. Summary of the spawning habitat requirements of 13 species of freshwater fish. Ind = indifferent.

Species	Depth (cm)	Flow (cm s ⁻¹)	Substratu diam. (cm	m Vegetation)
Category A	A 14-22	35-49	2-5	Absent
Bleak		< 20		Scirpus & filamentous algae.
Bream (common)	15-120	< 20		Glyceria, Sagittaria, Scirpus & filamentous algae, Salix roots, Veronica, Myosotis, Phragmites, Elodea.
Chub	10-30	20-50	> 0.5	Occasional
Dace	25-40	20-50	3-25	Occasional
Gudgeon	1-8	2-80	5-30	<u>Fontinalis</u> , tree roots, <u>Potamogeton pectinatus</u> .
Perch	Variable			Submerged <u>Salix</u> & <u>Alnus</u> roots.
Pike		< 5	Ind.	Myriophyllum, flooded riparian vegetation.
Roach	5-45	> 20	5-15	<u>Fontinalis</u> , <u>Elodea</u> , <u>Scirpus</u> , <u>Salix</u> roots.
Category Carp	<u>B</u> Variable	< 5	Ind.	Submerged riparian vegetation, <u>Carex</u> , <u>Glyceria</u> , <u>Phragmites</u> , <u>Rorippa</u> , <u>Scirpus</u> & filamentous algae.
Ruffe			< 5	Plants/moss.
Bream (silver)	Variable	< 20	Ind.	Variety of aquatic plants.
Tench		< 5	Ind.	Myriophyllum & Lemna

References: Baras 1992, Baras & Philippart 1993, Bastl 1969, Copp & Mann 1993, Diamond 1985, Fedorova & Vetkasov 1971, Grandmottet 1983, Hancock et al., 1976, Holcik & Hruska 1966, Kaufmann et al., 1991, Kennedy 1969, Kennedy & Fitzmaurice 1968, 1972, Kovaleva 1967, Lelek 1987, Mann 1978, 1993, Mills 1981 a,b,c, Sych 1955, Vollestad & L'Abée-Lund 1987.

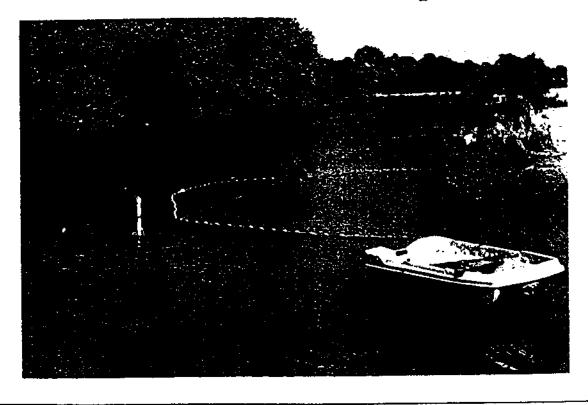
Table II. Published spawning times and water temperatures for 13 species of freshwater fish, and including published (1) and unpublished (2) data from the Thames and unpublished data from the Great Ouse (3).

Species	Months	Temp. °C	References
Category A			· · · · · · · · · · · · · · · · · · ·
Barbel	VI	13.5	Hancock <u>et al</u> . 1976
Bleak (1,3)	V-VII	> 12	R. Stour (Mann, unpublished) Leeming 1963, Mackay & Mann 1969
C. Bream (3)	V-VI	12-18	Hartley 1947, Leeming 1963, Holcik & Hruska 1966, Kennedy & Fitzmaurice 1968
Chub (3)	V-VII	> 12	Cragg-Hine 1963, Leeming 1963, Hellawell 1971, Mann 1976b, Krupka 1988
Dace (2,3)	II-IV	8-12	Cragg-Hine 1963, Kennedy 1969, Mann 1974, Hellawell 1974 Mills 1981 a,b
Gudgeon(1,3)	V-VII	13-17	Hartley 1947, Mathews 1971, Kennedy & Fitzmaurice 1972, Penaz & Prokes 1978, Mann 1980a
Perch (1,3)	IV-V	8-14	Hartley 1947, Williams 1963 Thorpe 1977, Mann 1978, Zeh <u>et al</u> . 1989
Pike (3)	V-III	6-14	Mann 1976a, Raat 1988
Roach (1,3)	V-VI	7-17	Hartley 1947, Mackay & Mann 1969, Hellawell 1972, Mann 1973, Diamond 1985, Vollestad & L'Abée-Lund 1987
<u>Category B</u>			
Carp (3)	VI-VII	> 15	Crivelli 1981
Ruffe (3)	III-VII		Leeming 1963, Bastl 1988
S. Bream (3)	V-VII		Hartley 1947, Leeming 1963
Tench (3)	VI	> 18	Mills 1991, Copp & Mann 1993

South West Oxfordshire Reservoir Proposal (SWORP)

River Thames Juvenile Fish Survey 1993

Volume 1 - Main Report





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Executive Summary

This report presents the results of a juvenile fish survey of the River Thames between Oxford and Days Weir, undertaken in July and August 1993 by King's Environmental Services. It is along this stretch of the Thames that the outlet of the South West Oxfordshire Reservoir proposed by Thames Water Utilities is likely to be sited. The 1993 juvenile fish survey follows on from previous surveys undertaken in 1991 by the Environmental Advisory Unit (EAU), and in 1992 by Royal Holloway and Bedford New College (RHBNC). Fourteen sites were sampled in 1993, thirteen of which corresponded approximately to the 13 sites sampled in 1992.

Each survey site was sampled with three seine net hauls using a $25m \times 3m$ micromesh net with 3mm mesh size. With the exception of the additional site at Abingdon Marina, where three essentially identical net hauls were taken, three contrasting sub-sites were sampled at each site, representing three distinct habitat types, namely:

- i shallow with macrophytes,
- ii shallow without macrophytes,
- iii deep without macrophytes.

A comparison of the baseline juvenile fish stocks in the River Thames in 1993 with those recorded in 1992 and 1991 is as follows:

	1993	1992	1991	1991
	KES	RHBNC	EAU	EAU
Number of sites	14	13	13*	36
Number of seinings	42	38	13	36
Juvenile density(nm ⁻²⁾	3.84	5.81	4.07	3.07
Catch per unit effort	355	706	607	490

^{*} sites corresponding to 1992 & 1993 surveys

Both juvenile density and catch per unit effort were found to be substantially lower in 1993 than for the two previous surveys. This difference in densities is almost entirely due to differences in the numbers of juvenile roach. Although the dominant species in each of the three years, the density of 0+ roach in 1992 was more than twice that of 1993. If roach are subtracted from the overall mean density, the resulting densities for the remaining species are markedly similar: 1.96 nm⁻² in 1993, 1.84 nm⁻² in 1992.

Mean densities of 0+ fry were compared for the three habitat types and found to be as follows:

Habitat type	With macrophytes	Shallow without macrophytes	Deep without macrophytes
		macrophytes	macrophytes
Density 0+ fry 1992(nr	n ⁻²⁾ 8.311	5.359	0.715
Density 0+ fry 1993(nr	n ⁻²⁾ 2.468	3.091	2.317

Thus the very clear trends in overall fry distribution between the habitat types recorded in 1992 were not observed in 1993. In 1992 and 1993, roach densities were found to be highest in macrophyte sites and lowest in deep sites. Relationships between habitat type and density of other species was not the same between survey years.

- 3.4.9 Instantaneous rates of mortality (Z) and survival (S)
- 3.4.9.1 Instantaneous rates of mortality and survival for the 0+ to 1+ year classes in 1993 and 1992 are presented in Table 5.
- 3.4.9.2 The following observations can be made:
 - i) Instantaneous rates of mortality for bleak, chub, perch and roach were all higher in 1992 than in 1993. Only gudgeon had a higher instantaneous rate of mortality in 1993.
 - ii) Survival values for bleak, chub, perch, and roach were all higher in 1993 than in 1992. Only gudgeon had a higher survival rate in 1992.
 - iii) The highest 1993 Z value (and the lowest S value), was calculated for gudgeon, followed by perch, roach, chub, bleak then dace.
 - iv) The highest 1992 Z value (and the lowest S value), was calculated for roach, followed by chub, perch, gudgeon and bleak.
- 3.4.10 Length weight conversions
- 3.4.10.1 Scatter plots (including 95% confidence limits of the regression line), of log length (mm) vs. log preserved weight (g) are presented in Figure 13.
- 3.4.10.2 The equations for length weight conversion are presented below:

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Bleak log w = -4.8654 + 2.9296 (log l)

Chub log w = -4.6662 + 2.8603 (log l)

Dace log w = -4.4374 + 2.7062 (log l)

Gudgeon log w = -4.8433 + 2.9882 (log l)

Perch log w = -4.7558 + 2.9407 (log l)

Roach log w = -4.5888 + 2.8590 (log l)
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- 3.4.11 Biomass (standing crop)
- 3.4.11.1 Overall estimates of biomass (gm⁻²) for the six major species (preserved weight); bleak, chub, dace, gudgeon, perch and roach are presented in decreasing order in the table below (these results are illustrated graphically in Figure 14):

Roach	2.32
Perch	1.12
Gudgeon	0.56
Chub	0.17
Dace	0.14
Bleak	0.07
Total	4.38

- 3.5.9 Instantaneous rates of mortality (Z) and survival (S)
- 3.5.9.1 Values of instantaneous rate of mortality and survival for 1993 and 1992 are presented in Table 5. Observations in both surveys are as follows:
 - i) Instantaneous rate of mortality of chub and perch was greatest in deep sub-sites.
 - ii) Instantaneous rate of mortality of gudgeon was greatest in shallow sub-sites.
 - Instantaneous rate of mortality of perch was lowest in shallow subsites.
 - iv) Instantaneous rate of mortality of gudgeon was lowest in the macrophyte sub-sites.
- 3.5.9.2 Survival values reiterate these results showing a direct inverse relationship between survival and mortality.

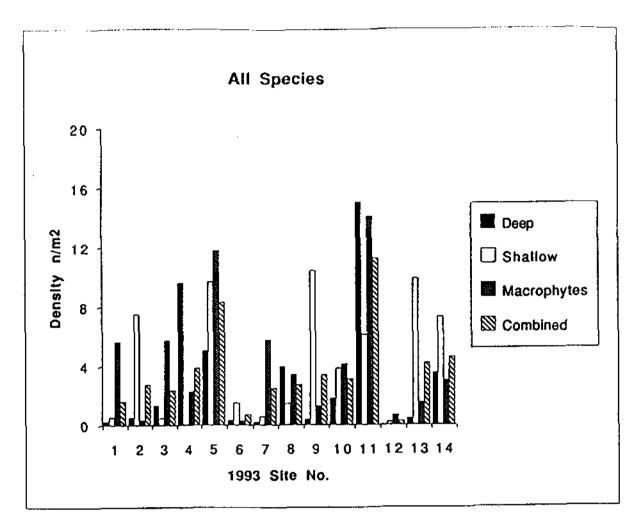
3.5.10 Biomass

3.5.10.1 Values of biomass for the six major fish species in the three different habitat types are ranked in descending order in the following table:

Deep		Shallow Macrophytes		es	
Roach	1.61	Gudgeon	1.03	Roach	3.60
Perch	0.82	Perch	0.52	Perch	1.06
Gudgeon	0.27	Roach	0.50	Gudgeon	0.39
Bleak	0.09	Dace	0.27	Chub	0.19
Chub	0.06	Chub	0.14	Bleak	0.09
Dace	0.04	Bleak	0.03	Dace	0.03
Total	2.89	Total	2.49	Total	5.36

These results are illustrated graphically in Figure 22

- 3.5.10.2 The above table indicates that although total densities were relatively similar in each of the habitat types, biomass in the macrophyte sub-sites appeared considerably higher. The biomass values of roach and perch were particularly high in the macrophytes relative to the other habitat types. Gudgeon biomass was relatively high in the shallow sites, where it ranked first. The biomass of dace was also higher in the shallow sites relative to the other two habitat types.
- 3.5.10.3 Statistical analysis was not carried out on the biomass data, as it was beyond the scope of this investigation. The point raised in 3.5.6.4 should be borne in mind when making inferences from these data.



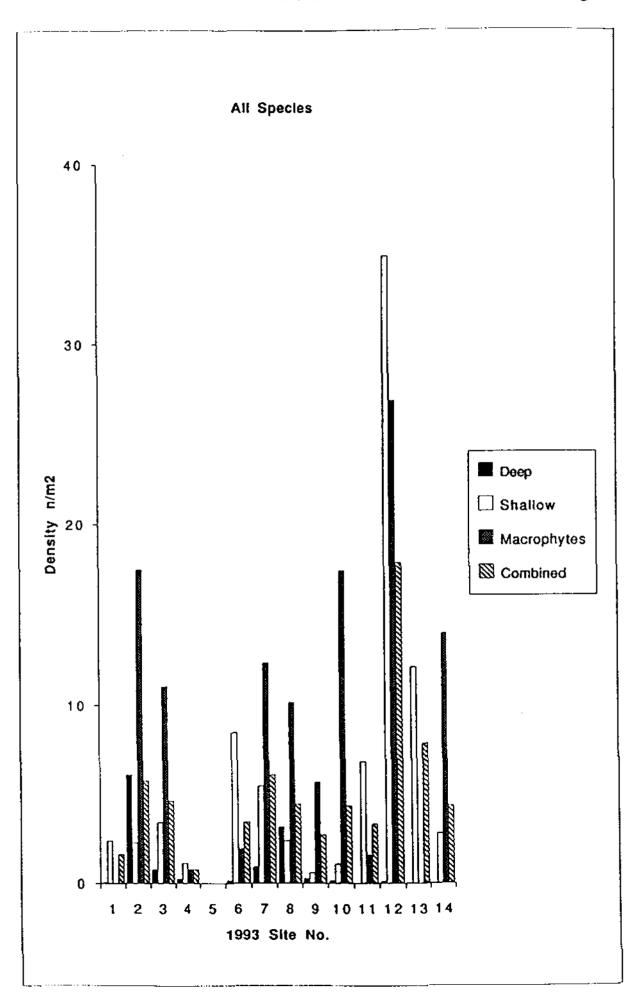
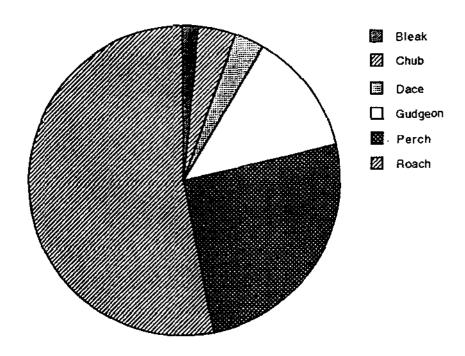


Figure 14

Combined total biomass



Combined total biomass = 4.3844 g/m^2

River Thames Juvenile Fish Survey 1994

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Cover Photograph

Measuring current velocity at Site 13 'shallow without macrophytes' with the Sensa RC2 electromagnetic flow meter. The net has been set and pulled but the net marker buoys and transect posts remain in situ. The buoys and posts define fixed points on the perimeter of the netted arc so that the total netted area can be calculated from the distances and angles of these markers to the centre point on the bank. The posts define the parallel and perpendicular transects along which depth, temperature, and water velocity are measured.

King's Environmental Services November 1994

Executive Summary

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Each survey site was sampled with three seine net hauls using a $25m \times 3m$ micromesh net with 3mm mesh size. With the exception of the Abingdon Marina site, where three essentially identical net hauls were taken, three contrasting subsites were sampled at each site, representing three distinct habitat types:

- i) deep without macrophytes (mean depth 1.37m; 0-3% macrophyte cover),
- ii) shallow without macrophytes (mean depth 0.52m; 0-5% macrophyte cover),
- iii) shallow with macrophytes (mean depth 0.88m; 10-95% macrophyte cover).

A comparison of the baseline juvenile fish stocks in the River Thames in 1994 with those recorded in 1993, 1992 and 1991 is as follows:

	1994	1993	1992	1991	1991
	KES	KES	RHBNC	EAU	EAU
Number of sites	14	14	13	13*	36
Number of seinings	42.	42	38	13	36
Juvenile density(nm ⁻²⁾	13.03	3.84	5.81	4.07	3.07
Catch per unit effort	1095	355	706	607	490

^{*} sites corresponding to 1992 & 1993 surveys

Both juvenile density and catch per unit effort in 1994 were substantially higher than for all previous surveys. This difference in densities is largely due to differences in the numbers of juvenile roach. The mean density of 0⁺ roach was 9.23 nm⁻² in 1994, 1.29 nm⁻² in 1993 and 2.52 nm⁻² in 1992. The densities of other species are less variable between years with combined mean densities for all other species of 3.59 nm⁻² in 1994, 1.96 nm⁻² in 1993 and 1.84 nm⁻² in 1992.

Mean densities of 0+ fry were compared for the three habitat types and found to be as follows:

Habitat type	•	without phytes	Shallow without macrophytes	Shallow with macrophytes
Density 0+ fry 1992	(nm ⁻²⁾	0.72	_	8.31
Density 0+ fry 1993		2.32	3.09	2.47
Density 0+ fry 1994	l(nm-2)	4.17	5.12	32.48

In both 1992 and 1994 there were clear trends in juvenile distribution between habitat types although clear cut trends were fewer in 1993. The preference for macrophyte-rich sub-sites was most marked for roach and was apparent for this species in all three years.

The following relationships between habitat type and juvenile fish species were found in 1994:

- i 0+ dace, chub, gudgeon and perch showed a clear preference for the shallow rather than the deep sites,
- ii 0+ dace and chub showed a preference for the shallow macrophyte-poor sites rather than for the shallow macrophyte-rich sites,
- iii 0+ perch and gudgeon showed a preference for the shallow macrophyte-rich sites rather than the shallow macrophyte-poor sites,
- iv 0+ roach showed a very strong preference for the macrophyte-rich sites but in macrophyte-poor areas appeared to prefer the deep to the shallow sub-sites.

In addition, it was found that:

- i perch, pike and roach density was positively correlated with the percentage macrophyte cover,
- ii chub, dace and gudgeon densities were negatively correlated with depth,
- iii chub and dace densities were positively correlated with the percentage of sand and gravel in the substrate.

There was no apparent correlation between juvenile fish density and water velocity.

This report also discusses the potential impacts to juvenile fish which may arise from construction and operation of the proposed South West Oxfordshire Reservoir and presents proposals for further studies on juvenile cyprinids in the Oxford region of the River Thames.

6 Summary and Conclusions

- The current survey recorded the highest mean density of juvenile fish for the three survey years (1992-1994) in which densities could be calculated. Roach comprised 72.4% of the total 1994 catch of 46,030 fish. Of the remaining species, gudgeon comprised 18.09% of the 1994 catch with chub, bleak, dace and perch together accounting for 8.13%. Roach was also the most abundant species in 1991, 1992 and 1993, although the relative abundance of the other species varied from year to year.
- 6.2 The combined mean density of all species was 5.81 nm⁻² in 1992, 3.84 nm⁻² in 1993 and 13.03 nm⁻² in 1994. The great majority of the density difference between years was due to the varying year class strength of 0+ roach, with densities of 2.52, 1.29, and 9.23 nm⁻² recorded in 1992, 1993 and 1994 respectively. The mean density of 1+ roach showed much smaller variation between years with values of 0.15, 0.26 and 0.19 nm⁻² for 1992-94 respectively.
- 6.3 Z and S values have been calculated for the major species by following cohorts from year to year. Calculated survival rates vary greatly between years and between species and are difficult to interpret because factors other than mortality affect the observed density of an individual species' year class. Calculated S values for 0+/1+ roach, dace, chub and perch are within the range of 0.02 to 0.3 and appear relatively realistic. Calculated first year survival rates for bleak and bream are particularly erratic, ranging over the survey years from 0.05 to >5. The most likely explanation for these results is that only a part of the 0+ cohort is being captured because of the relatively late spawning period of these two species.
- 6.4 Calculated survival values for roach, together with the relative constancy of the 1+ year class strength over the survey years, provides some indication that density-dependent mortality may operate for this species in the first year of life.
- Although total juvenile density varied enormously in the different sites and sub-sites, there was a general overall trend of increasing densities from the top to the bottom of the survey stretch. This general trend was also apparent in 1992 and 1993, although the reasons for the trend are not clear.
- 6.6 Analysis of the measured habitat variables for the 1994 survey confirmed that, despite a small degree of overlap, the three habitat (sub-site) types sampled, namely 'deep without macrophytes', 'shallow without macrophytes' and 'shallow with macrophytes' formed statistically distinct sets with regard to macrophyte cover and depth. Although statistically less distinct, the habitat types also differed with respect to water velocity. Mean velocities were 0.023, 0.31 and 0.34 msec-2 in the 'deep', 'shallow without'

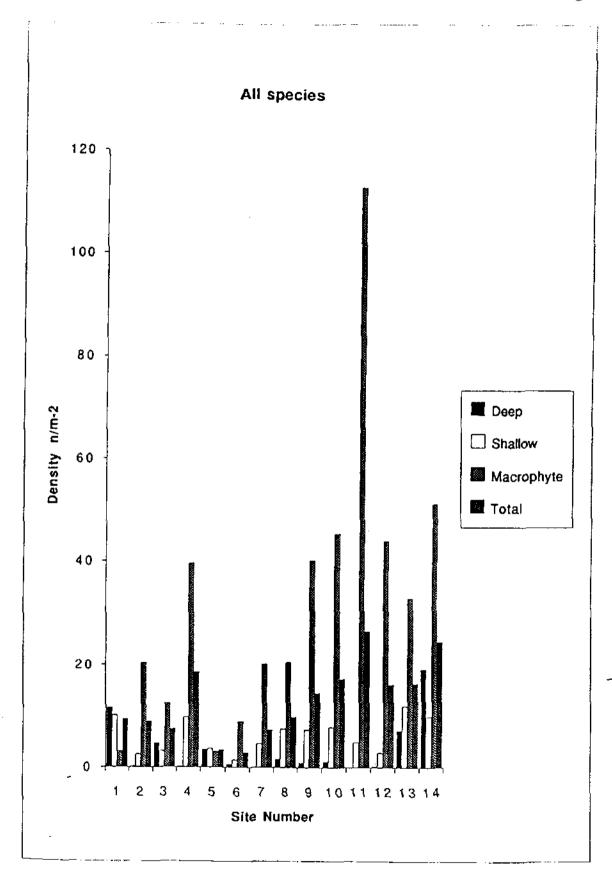
and 'shallow with macrophyte' sub-sites respectively. Although substrate composition varied greatly between sites and between sub-sites, there was no apparent relationship between substrate nature and sub-site type.

- 6.7 The 1994 survey indicated clear relationships between habitat type and fish density and in most cases these relationships were statistically significant. Mean total densities of 0+ fish were 32.48, 5.12 and 4.17 nm⁻² for the 'shallow with macrophyte', 'shallow without macrophyte' and 'deep' subsites respectively. These total densities are heavily influenced by the dominant species, roach.
- 6.8 The 1994 0+ cohorts of perch, chub, dace and gudgeon, together with the 1+ gudgeon cohort, all showed a clear preference for the shallow (with or without macrophytes) rather than the deep sub-sites. Although not statistically significant, bleak appeared to show a similar preference. The 0+ cohorts of gudgeon and perch showed a statistically significant preference for the 'with macrophyte' sub-sites although this trend was not apparent for the 1+ gudgeon. Conversely, the 0+ chub and dace showed a preference for the shallow macrophyte-free sites.
- 6.9 Both 0+ and 1+ roach showed a highly significant association with the macrophyte-rich sites. Unlike the other species however, 0+ roach in the macrophyte free habitats displayed a preference for the deep rather than the shallow sub-sites.
- 6.10 When the habitat variables were considered independently rather than by sub-site groups, a number of significant correlations were obtained:
 - i perch, pike and roach were positively correlated with percentage macrophyte cover,
 - ii chub, dace and gudgeon were negatively correlated with depth,
 - iii chub and dace were positively correlated with the percentage of sand and gravel in the substrate.

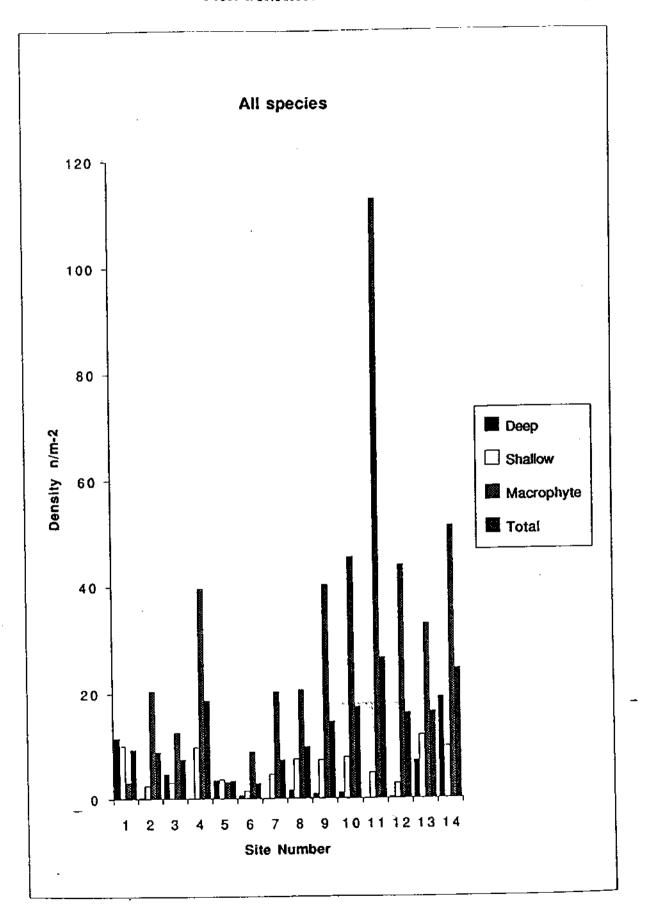
There was no apparent correlation between fish density and water velocity.

- 6.11 Stepwise multiple regression using nine potential habitat predictors suggested that 34.8% of observed variation in total fish density is attributed to % macrophyte cover (22.57%) and water velocity (34.79%). Biomass showed similar trends with 41% of the variation attributed to % cover, water velocity and % sand and gravel. Stepwise multiple regression was not attempted for individual species or year class cohorts.
- 6.12 The associations demonstrated between habitat variables and fish density in the 1994 survey generally confirm and extend observations made in the earlier surveys. The potential value of stepwise multiple regression has been demonstrated. However, the present system of sub-site selection, which seeks uniformity of habitat variables within one of three defined habitat types, may not be optimum for this type of analysis.

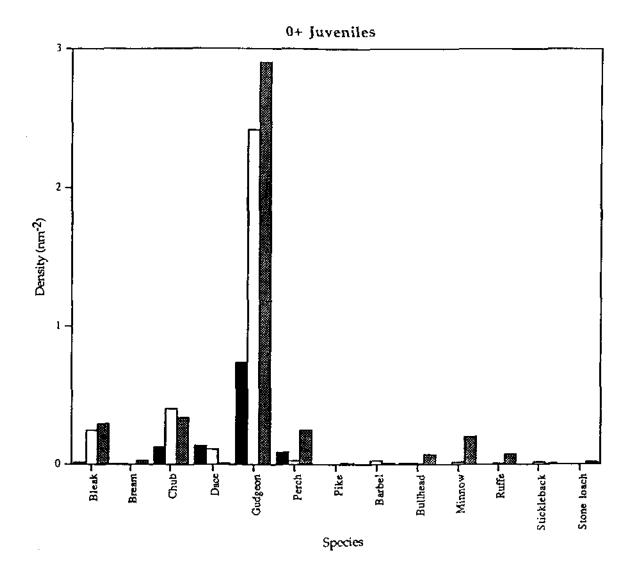
- 6.13 The possible underlying reasons for the variation in 0+ year class strength of the different fish species in the Thames have been discussed in this report. Although direct evidence from the Thames is tenuous, published information for other systems supports the hypothesis that spring and summer water velocities, and spring water temperatures, are likely to be the principal controlling factors. A longer term data set for juvenile fish abundance and more comprehensive seasonal temperature records will be required to adequately test this hypothesis.
- 6.14 Mean lengths and ranges of the various fish species show significant variation over the four survey years, although generally these variations are not great. Growth rates of 0+ to 2+ fish are generally similar to national standards and are thus within the range expected.
- 6.15 Experience gained from the 1991-1994 surveys of the Thames suggests that late July/early August is an appropriate sampling period for juveniles of most species of coarse fish inhabiting the river. However, at least in some years, this date appears to be rather early for the later spawning species such as bleak and bream, leading to under estimation of the 0+ year class strength for these two species
- 6.16 Considering the proposed South West Oxfordshire Reservoir, the most significant impacts to juvenile fish are likely to arise from changes to the flow and temperature regime and from changes in food availability. Changes to flow regime may affect juvenile fish directly, if their velocity tolerance is exceeded, or indirectly, by affecting the distribution and abundance of aquatic macrophyte stands. Water quality impacts would probably be of relatively minor significance if the proposed reservoir were fully mixed but would be potentially more severe if the reservoir were allowed to stratify thermally in summer.
- 6.17 Although of necessity very tentative, the preliminary assessment of potential impacts presented in this report is nevertheless important. It provides a preliminary focus for the development of further studies on both the population dynamics of juvenile cyprinids in the Thames and the vulnerability of these species to the environmental perturbations that may arise from reservoir development.



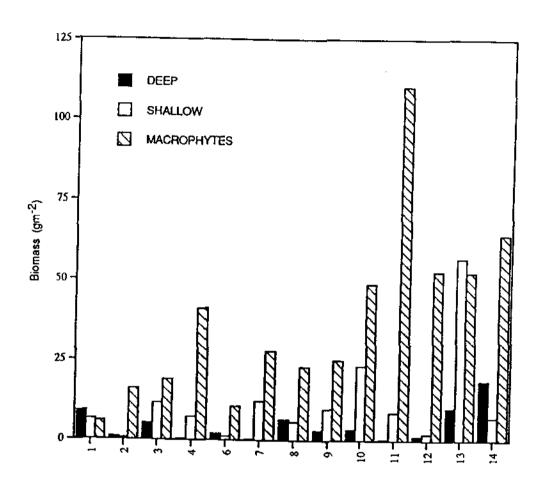
Note: Change of vertical scale



Note: Change of vertical scale



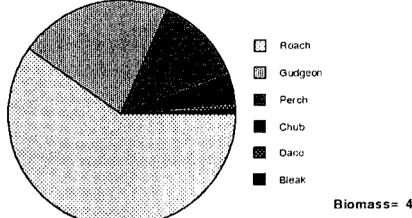
	Deep				Shallow				Macrophy	tes		
	0+	1+	>1+	Total	0+	1+	>1+	Total	0+	1+	>1+	Total
Bleak	0.013	0.002	0.004	0.018	0.244	0.022	0.005	0.270	0.291	0.445	0.024	0.760
Bream	0.000	0.000	0.002	0.002	0.000	0.000	0.000	0.000	0.025	0.009	0.011	0.045
Chub	0.129	0.067	0.001	0.197	0.395	0.031	0.002	0.428	0.338	0.066	0.002	0.405
Dace	0.137	0.004	0.000	0.141	0.401	0.042	0.000	0.443	0.084	0.015	0.000	0.099
Gudgeon	0.732	0.088	0.010	0.830	2.404	0.514	0.341	3.260	2.903	0.382	0.062	3.346
Perch	0.092	0.011	0.008	0.111	0.023	0.005	0.001	0.029	0.247	0.043	0.012	0.302
Pike .	0.002	0.000		0.002	0.000	0.001		0.001	0.006	0.004		0.011
Roach	3,047	0.056	0.016	3.119	1.877	0.045	0.008	1.929	28.364	0.507	0.034	28.905
Barbel	0.001			0.001	0.022			0.022	0.007			0.007
Bullhead	0.011			0.011	0.011			0.011	0.069			0.069
Minnow	0.003			0.003	0.014			0.014	0.199			0.199
Ruffe	0.002			0.002	0.009			0.009	0.071			0,071
Stickleback	0.002			0.002	0.012			0.012	0.011			0.011
Stone loach	0.001			0.001	0.002			0.002	0.014			0.014
Total				4.439			_	6.432	1			34.244



Site number

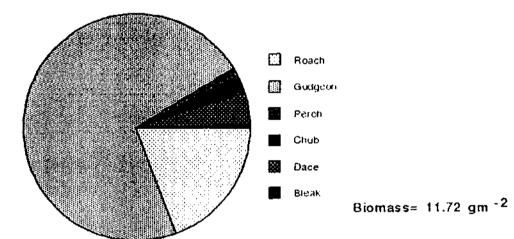
Relative biomass of major fish species in each habitat

Deep without macrophytes

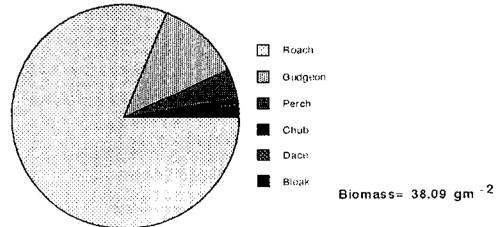


Biomass= 4.68 gm -2

Shallow without macrophytes



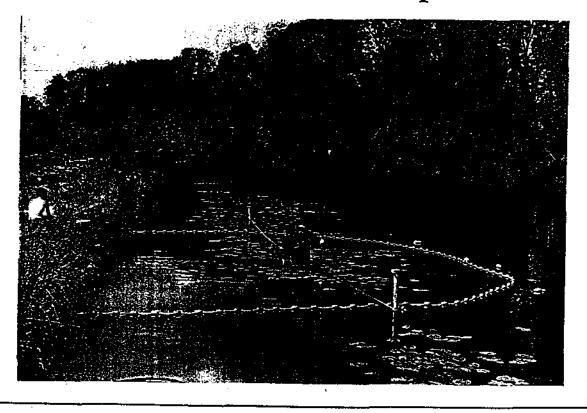
Shallow with macrophytes



Total Biomass all sites= 18.16 gm -2

River Thames Juvenile Fish Survey 1995

Volume 1 - Main Report





November 1995

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Cover Photograph

Measuring current velocity at Site 10 'shallow with macrophytes' with the Sensa RC2 electromagnetic velocity meter. Following velocity measurement, the enclosed macrophytes are cut with a scythe and removed prior to hauling and landing the net. The buoys and posts define fixed points on the perimeter of the netted are so that the total netted area can be calculated from the distances and angles of these markers to the centre point on the bank. Depth, temperature and water velocity are measured along two transects, one parallel to and the other perpendicular to the bank. The two posts define the parallel transect.

King's Environmental Services November 1995

6 Summary and Conclusions

- 6.1 The 1995 survey recorded the highest mean density of juvenile fish for the four survey years (1992-1995) in which densities could be calculated. As in previous years, roach was the dominant species comprising 65.7% of the total 1995 catch of 59,926 fish. Previous total catches were 31,891 in 1992, 14,936 in 1993 and 46,030 in 1994. Of the remaining species, gudgeon comprised 17.19% of the 1995 catch, chub 7.76% and bleak 4.22%, with dace, perch and bream together accounting for 3.58%. Although roach dominated the catch in all survey years, the relative abundance of the other species varied significantly from year to year.
- 6.2 The combined mean density of all species was 4.38 nm⁻² in 1992, 3.52 nm⁻² in 1993, 13.03 nm⁻² in 1994 and 16.05 nm⁻² in 1995. The great majority of the density difference between years was due to the varying year class strength of 0+ roach, with densities of 2.52, 1.29, 9.23 and 10.55 nm⁻² recorded in 1992 to 1995 respectively. The mean density of 1+ roach showed parallel intervear variation with values of 0.15, 0.26 and 0.19 and 1.10 nm⁻² for 1992 to 1995 respectively.
- 6.3 Z and S values were calculated for the major species by following cohorts from year to year over the period 1992-1995. Calculated survival rates vary greatly between years and between species and in many cases are difficult to interpret because factors other than mortality may affect the observed density of an individual species' year class. Calculated S values for 0+/1+ roach, dace, chub and perch are generally within the range of 0.02 to 0.35 and appear relatively realistic. Calculated first year survival rates for bleak and bream, and to a lesser extent for gudgeon, are particularly erratic, ranging over the survey years from 0.05 to > 5. The probable explanation for these results is that only a part of the 0+ cohort was captured. This is due to the relatively late spawning period of bleak and bream and probably because the bottom living habit of gudgeon makes it difficult to net the smallest individuals so that they are under-represented in the catch.
- 6.4 Calculated one year survival values for the 1992, 1993 and 1994 cohorts of roach and chub are markedly similar (mean S roach = 0.12, mean S chub = 0.3), despite substantial variation in initial cohort strength, suggesting that density-dependent mortality does not operate for these two species. In contrast, the calculated S values for perch suggest that density-dependent mortality may be occurring in this species.
- 6.5 Although total juvenile density varied enormously in the different sites and sub-sites in all survey years, a general overall trend of increasing densities from the top to the bottom of the survey stretch was clearly apparent in the 1992-1994 survey years. In contrast, fish densities were much more uniformly distributed along the survey stretch in 1995. The reasons for this trend, and its apparent absence in 1995, remain obscure.

- 6.6 Uniquely amongst the Thames fish species, the mean density of ruffe increased consistently from 1992 to 1995 (0.001, 0.01, 0.05 and 0.17 nm⁻² 1992-1995 respectively) possibly indicating a long term trend.
- 6.7 As in 1993 and 1994, analysis of the measured habitat variables for the 1995 survey confirmed that, despite a small degree of overlap, the three habitat (sub-site) types sampled, namely 'deep without macrophytes', 'shallow without macrophytes' and 'shallow with macrophytes' formed statistically distinct sets with regard to macrophyte cover and depth.
- Mean water velocity in 1995 for the 39 river sub-sites, at 0.012 ms⁻¹, was the lowest recorded for the three years in which velocity was measured (mean velocity 1993 = 0.0623 ms⁻¹; 1994 = 0.030 ms⁻¹). There were no significant differences in water velocity or temperature between the sub-site types in 1995. There was a clear longitudinal trend of increasing water temperature from the top to the bottom of the survey stretch but this was simply a reflection of prevailing weather conditions during the survey period. Although substrate composition varied greatly between sites and between sub-sites, the only clear pattern was the positive relationship between depth and the percentage of bare clay substrate.
- 6.9 As in previous surveys, the 1995 data indicated a clear relationship between habitat type and total fish density. Mean total densities of 0+ fish were 27.49, 10.52 and 3.02 nm⁻² for the 'shallow with macrophyte', 'shallow without macrophyte' and 'deep' sub-sites respectively. These total 0+ densities were heavily influenced by, but not exclusively due to, the dominant species, roach.
- 6.10 Statistical analysis of the 1995 juvenile fish and environmental data established a number of associations and correlations which largely confirmed the findings of the 1993 and 1994 surveys. The more significant findings from the 1995 data set are as follows:
 - i 0+ chub and gudgeon showed a clear preference for the shallow rather than the deep sub-sites;
 - ii 0+ perch and roach showed a very strong preference for the macrophyte-rich sub-sites over either the deep or shallow macrophyte-poor sub-sites;
 - iii 1+ roach showed a preference for the macrophyte-rich sub-sites rather than for the deep macrophyte-poor sub-sites;
 - iv 0+ bleak, perch, roach and ruffe densities were positively correlated with the percentage macrophyte cover;
 - v 0+ bleak, chub and gudgeon densities were negatively correlated with depth and gradient;
 - vi 0+ chub were positively correlated with the percentage of sand and negatively correlated with the percentage of clay in the substrate;
 - vii 1+ bleak, chub and gudgeon and 2+ gudgeon were positively correlated with the percentage of gravel substrate.

- 6.11 When the 1993 to 1995 data sets were pooled to give a larger data set (117 sub-sites), relationships and correlations generally became more clear cut, especially for the 1+ and 2+ age classes. Several additional relationships became apparent that were not evident from the 1995 data set alone, namely:
 - i 0+ chub showed a preference for shallow macrophyte-poor sub-sites over both shallow macrophyte-rich and deep macrophyte-poor sub-sites;
 - ii 0+ dace showed similar preferences to 0+ chub;
 - iii 0+ perch showed a strong preference for macrophyte rich sub-sites
 but in the absence of macrophytes preferred deep to shallow sub-sites;
 - iv both 0+ and 1+ pike showed a preference for macrophyte-rich sites over the other two habitat types.
- 6.12 The associations demonstrated between habitat variables and fish density in the 1992 1994 data sets (RHBNC 1992; KES 1993, 1994) and in the 1995 and combined 1993-95 data sets means that it is now possible to define provisionally the habitat preferences of juveniles of the main fish species occurring in the Oxford region of the Thames. As a general rule, it would appear that for a given species, the habitat preferences of 0+ and 1+ fish are similar, but the usually weaker statistical relationships for 1+ fish suggest that they become more catholic in their habitat requirements with increasing age. Habitat preferences appear to be as follows:
 - i roach show a very strong preference for habitats with macrophyte cover, the preference for cover appearing to eclipse all other habitat requirements;
 - ii for perch, macrophyte cover is the single most important factor governing juvenile distribution. Perch juveniles display a marked aversion to macrophyte free shallows but can occur in significant numbers in deeper cover-free sites;
 - iii the preferred habitat for both chub and dace is gently shelving shallows with a sandy or gravelly bed;
 - iv juvenile pike lurk in weedy shallows, as do the adults;
 - v juvenile bleak appear to prefer shallow water with a gravelly substrate and good macrophyte cover;
 - vi bream show a fairly clear preference for sites containing macrophyte cover.
- 6.13 The possible underlying reasons for the variation in 0+ year class strength of the different fish species in the Thames are discussed in this report. Published information for other river systems supports the hypothesis that spring and summer water velocities and temperatures, are likely to be major controlling factors. Examination of flow records for the 1992-1995 survey years has suggested that there might be a negative relationship between early summer river flows and the 0+ year class strength for roach

and bream. However, a much longer term data set will be required for these or other possible relationships to be confirmed. Evaluation of the potential effects of spring and summer temperature on year class strength has been precluded because of the fragmentary nature of the currently available temperature record.

- 6.14 Mean lengths and ranges of the various fish species showed some variation over the five survey years, although generally these variations were not great. Growth rates of 0+ to 2+ fish were generally similar to national standards and were within the range expected. The 1995 b constants for bleak, chub, dace, perch and roach were higher than those determined in 1994 but the gudgeon b constants were lower.
- 6.15 As shown by the 1992-1995 surveys of the Thames, late July/early August is an appropriate sampling period for juveniles of most species of coarse fish inhabiting the river. However, this date is rather early for the later spawning bleak and bream, leading to under estimation of the 0+ year class strength for these two species. A greater number of 0+ bleak and bream were captured in 1995 than in 1993 or 1994, probably indicating earlier spawning, but it is likely that even in 1995, 0+ individuals of both species were under-represented in the catch.
- 6.16 In the absence of any additional information concerning the design or operation of the proposed South West Oxfordshire Reservoir, no further assessment of potential impacts of the scheme to juvenile fish was undertaken.

Table 4: Comparison of mean densities (nm-2) for all species 1992-1995

1995	Survey	1994 Survey		1993 S	1993 Survey		Зигуөу
Roach	10.55	Roach	9.44	Roach	1.88	Roach	3.97
Gudgeon	2.76	Gudgeon	2.36	Gudgeon	0.99	Bleak	0.68
Chub	1.25	Chub	0.33	Chub	0.27	Gudgeon	0.63
Bleak	0.68	Bleak	0.31	Perch	0.26	Chub	0.23
Perch	0.25	Dace	0.24	Daœ	0.23	Perch	0.13
8ream	0.17	Perch	0.19	Bleak	0.07	Dace	0.08
Dace	0.15	Minnow	0.06	Minnow	0.05	Hybrids	0.03
Ruffe	0.14	Ruffe	0.05	Bullhead	0.02	Bream	0.02
Minnow	0.08	Bullhead	0.03	Bream	0.01	Minnow	0.01
Stickleback	0.013	Bream	0.01	Stickleback	0.01	Ruffe	0.001
Bulihead	0.009	Barbel	0.01	Ruffe	0.001	Pike	0.0009
Stone loach	0.004	Stickleback	0.01	Pike	0.004	Stickleback	0.0009
Barbel	0.0003	Stone loach	0.007	Stone loach	0.001	Bullhead	0.0004
Hybrids	0.0003	Pike	0.004	Tench	0.001	Tench	0.00
Pike	0.00	Tench	0.00	Hybrids	0.00	Stone loach	0.00
Tench	0.00	Hybrids	0.00	Barbel	0.00	Barbel	0.00

Table 5: Comparison of mean densities (nm-2) for all species 1991-1995

Year	n	Mean Density	C.V.
1991*	36	3.07	-
1991*^	13	4.07] -
1992^^	37	4.38	137.18
1993	42	3.52	107.50
1994	42	13.03	148.29
1995	42	16.05	98.59

Table 7: Comparison of mean CPUE 1991-1995

Year	n	Mean CPUE	C.V.
1991*	3 6	470.40	114.40
1991*^	13	607.20	97.50
1992	37	861.92	107.52
1993	42	355.57	104.60
1994	42	1095.95	124.86
1995	42	1426.81	102.21

^{*} Recalculated values (See KES 1993)

^{*} Densities calculated in 1992 survey
^ Sites corresponding to those in 1992 and 1993 surveys

^{^^} Recalculated values

n = number of sites sampled

A Sites corresponding to those in 1992 and 1993 surveys

n = number of sites sampled

%Frequence	· · · · · · · · · · · · · · · · · · ·			<u> </u>					
1995		1994	n=42	1993	n≃42	1992	n=37	1991	n=13°
Roach	65.72	Roach		Roach		Roach		Roach	72.36
Gudgeon	17.19	Gudgeon	18.09	Gudgeon		Bleak		Gudgeon	10.83
Chub	7.76	Chub		Perch	7.75	Gudgeon		Bleak	6.70
Bleak	4.22	Bleak	2.36	Chub		Chub		Dace	3.93
Perch	1.56	Dace	1.82	Dace	6.38	Perch		Chub	3.51
Bream	1.06	Perch	1.43	Bleak	1.94	Dace		Perch	1.08
Dace	0.95	Minnow	0.45	Minnow	1.37	Hybrids	0.82	Bream	1.01
Ruffe	0.85	Rutte	0.38	Bullhead	0.66	Bream	0.19	Stickleback	0.44
Minnow	0.48	Builhead	0.20	Bream	0.54	Minnow	0.16	Bullhead	0.10
Stickleback	0.08	Bream	0.11	Ruffe	0.27	Stickleback		Ruffe	0.03
Bullhead	0.06	Barbel	80.0	Stickleback	0.25	Pike	0.02	Barbel	0.01
Stoneloach	0.03	Stickleback	0.06	Pike	0.12	Ruffe	0.01	Hybrids	0.00
Barbel	0.02	Stone loach	0.04	Tench	0.03	Bullhead		Minnow	0.00
Hybrids	0.02	Pike	0.03	Stone loach	0.03	Stone loach	0.00	Pike	0.00
Tench		Tench	0.00	Barbel	0.00	Barbel	0.00	Stone loach	0.00
Pike	0.00	Hybrids	0.00	Hybrids	0.00	Tench	0.00	Tench	0.00
Relative Imp 1995	ortance V n=42		n=42	1993	n=42	1992	n=37*	1991	n=13*
Roach	165.73		162.90		143.58		157.00		172.36
Gudgeon	4	Gudgeon		Gudgeon	113.02		95.57		101.08
Perch	101.56	_	82.38		100.61		94.47		88.54
Chub	91.10	Chub	76.34	Bleak		Gudgeon		Gudgeon	87.76
Bleak	87.56	Bleak	69.02	Chub	69,30	_	68.93	_	68.24
Bream	51.07	Bullhead	59.73	Dace	46.86	Dace	39.62		65.05
Minnow	50.49	Dace	49.44	Bullhead	45.90	Hybrids	17.04	Stickleback	38.90
Dace	43.81	Minnow	36.17	Mirinow	34.70	Bream	11.00	Bream	24.09
Ruffe	41.32	Ruffe	36.10	Pike	28.69	Pike	10.83	Builhead	23.18
Bullhead	31.01	Pike	21.46	Stickleback	24.08	Minnow	8.27	Ruffe	7.72
Stickleback	19.13	Bream	19.16	Bream	17.21	Stickleback	8.13	Barbel	7.70
Stoneloach	16.69		19.12	Ruffe	9.79	Bullhead	8.12	Hybrids	0.00
Barbel	2.38	Stickleback	16.73	Tench	4.79	Ruffe		Мілпом	0.00
Hybrids	2.38	Stone loach	14.32	Stone loach	4.79	Barbel	0.00	Pike	0.00
Pike	r	Hybrids	0.00	Barbel	0.00	Stone loach		Stone loach	0.00
Tench	0.00	Tench	0.00	Hybrids	0.00	Tench	0.00	Tench	0.00

^{*}Sites corresponding to 1993-1995 surveys n=number of sub-sites sampled

b constants calculated for this and previous studies

Table 9

Species	KES Survey (1995)	KES Survey (1994)	KES Survey (1993)	Williams (1965b)	Mathews (1971)
Bleak	2.99	2.76	2.93	3.60	3.05
Chub	3.11	2.85	2.86	5	9.7
Dace	3.12	2.83	2.71	3.10	90 6
Gudgeon	2.97	3.15	66	<u>;</u>	9 6
Perch	3.07	2.97	294	3.14	0.00
Roach	3.31	3.03	2.86	3.20	3.23

Biomass (gm-2) calculated for all sites in 1995 with overall values for 1993, 1994 and 1995

Table 10

	Sub-sites 1995			Pooled-sites		
Species	Deep	Shallow	Macrophytes	1995	1994	1993
Roach	1.90	8.71	19.11	9.32	11.98	2.32
Gudgeon	8.53	9.05	4.09	7.02	4.68	0.56
Perch	0.40	0.32	1.07	0.55	0.78	1 1 2
Chub	1,12	0.48	0.38	0.72	0.25	. c
Dace	0.11	0.52	0.14	0.25	0.24	; ; ;
Bleak	0.29	1.18	0.38	0.58	0.24	0.07
Bream	0.03	0.12	0.43	0.18		
Total	12.37	20.38	25.61	18.63	18.14	4.38

- data not included in previous reports

Correlations of 1993-5 fish density and environmental variables

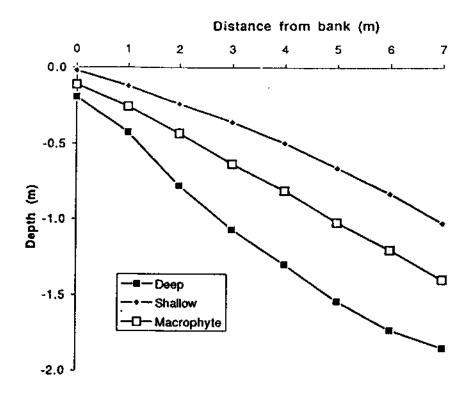
		
Species	r	p
Fish density v macrophyte cover	- · · · ·	
Bream 0+	0.197	< 0.05
Gudgeon 2+	0.197	< 0.05
Perch 0+	0.244	< 0.001
Perch 1+	0.344	< 0.001
Pike 0+	0.373	< 0.001
Pike 1+	0.373	< 0.001
Roach 0+	- ·	< 0.001
Roach 1+	0.449	
1 -	0.298	
Roach 2+		< 0.001
Bullhead	0.457	< 0.001
Ruffe	0.273	< 0.01
Stone loach	0.316	< 0.001
Fish density v % clay		
Chub 0+	-0.246	< 0.01
Perch 1+	-0.203	< 0.05
1 0.001 77	0,200	
Fish density v % silt		
Perch 0+	0.185	< 0.05
Stone loach	0.206	< 0.05
.		
Fish density v % sand		
Chub 0+	0.273	< 0.01
Dace 0+	0.267	< 0.01
8arbel	0.232	< 0.05
Minnow .	0.363	< 0.001
Stickleback	0.184	< 0.05
Fish density v % gravel		
Bleak 1+	0.237	< 0.001
Chub 1+	0.313	< 0.001
Fish density v depth		İ
Bleak 0+	-0.222	< 0.05
Chub 0+	-0.235	< 0.05
Gudgeon 1+	-0.229	< 0.05
Perch 0+	0.252	< 0.01
Perch 2+	0.194	< 0.05
}		
Fish density v gradient		c =
Gudgeon 2+	0.194	< 0.05
Perch 0+	0.249	< 0.01
<u>L</u>		

p>0.05	Not significant
p<0.05	Significant
p<0.01	Highly significant
p<0.001	Very highly significant

Estimated physical habitat preferences of 6 major fish species Table 19 (adapted from Bullock et al. 1991)

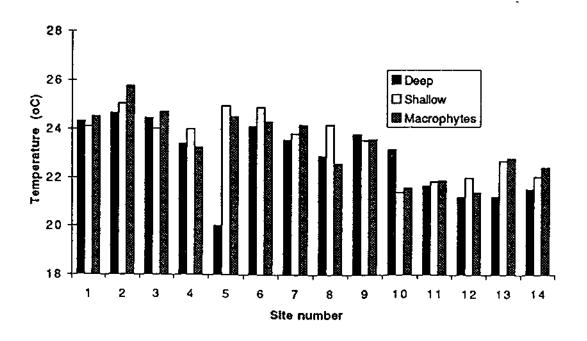
		
Modal value	Velocity (m/s)	Depth (m)
Species		<u> </u>
Bream		
Spawning Fry	0.00-0.10	0.50-1.00
Juveniles	0.00-0.05 0.00-0.10	0.05-0.50 0.50-3.00
Adults	0.00-0.10	1.70-3.00
	0.00 0.10	1770-0.00
Chub		
Spawning	0.25-0.90	0.40-1.70
Fry	0.05-0.30	0.50-0.90
Juveniles Adults	0.30-0.70	0.50-1.60
Adults	0.20-0.60	0.50-1.60
Dace		
Spawning	0.55-1.00	0.20-0.80
Fry	0.05-0.25	0.10-0.30
Juveniles	0.15-0.35	0.30-0.70
Adults	0.20-0.70	0.50-1.00
D1.		
Perch		
Spawning Fry	0.00-0.30 0.00-0.10	0.30-1.50
Juveniles	0.00-0.10	0.10-0.50 0.20-0.80
Adults	0.00-0.40	0.30-2.50
	***************************************	0.50 2.50
Pike		i
Spawning	0.00-0.10	0.20-0.80
Fry	0.00-0.10	0.20-0.90
Juveniles Adults	0.00-0.20	0.10-0.70
Adults	0.00-0.20	0.40-2.90
Roach		1
Spawning	0.40-0.80	0.30-3.00
Fry	0.00-0.20	0.30-3.00
Juveniles	0.00-0.40	1.00-3.00
Adults	0.00-0.40	1.00-3.00

Average depth profiles for shallow and deep sites (1995) Figure 2



Average temperature for sites 1-14 (1995)

Figure 3

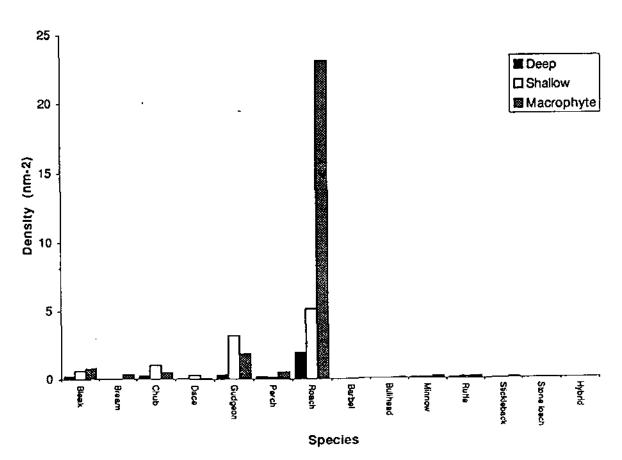


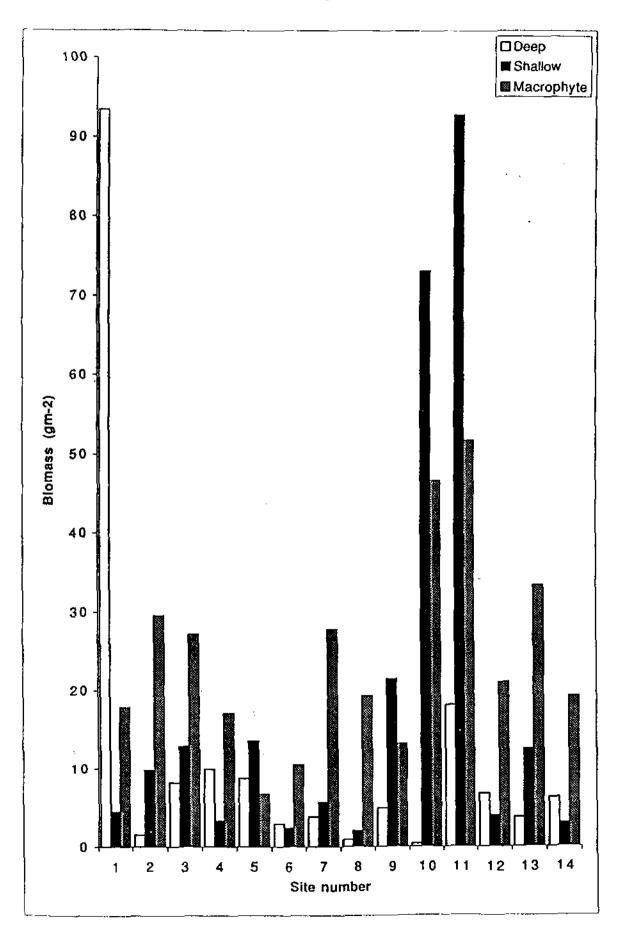
	Deep			Shallow			Macrophyte					
	0+	1+	>1+	Tot.	0+	1+	>1+	Tot.	0+	1+	<u>>1+</u>	Tot.
Bleak	0.220	0.067	0.005	0.292	0,614	0.320	0.072	1.006	0.765	0.136	0.000	0.902
Bream	0.003	0.008	0.000	0.011	0.003	0.027	0.003	0.033	0.341	0.133	0.004	0.477
Chub	0.234	0.204	0.000	0.438	1.027	0.040	0.000	1.067	0.468	0.051	0.000	0.519
Dace	0.046	0.008	0.000	0.054	0.279	0.045	0.000	0.325	0.059	0.014	0.000	0.073
Gudgeon	0.326	1,311	0.007	1.644	3.154	1.039	0.000	4.193	1,853	0.536	0.000	2.389
Perch	0.172	0,007	0.000	0.179	0.102	0.005	0.000	0.107	0.513	-,	0.000	0.525
Roach	1,960	0.221	0.009	2.190	5.125	1.314	0.016	6.456	23.078	1,992	0.001	25.071
Barbel	0.000	0,003		0.000	0.001			0.001	0.000			0.000
Bullhead	0.003			0.003	0.005			0.005	0.022			0.022
Minnow	0.013			0.013	0.052			0.052	0.175			0.175
Ruffe	0.041			0.041	0.141			0.141	0.175			0.175
Stickleback	0.003			0.003	0.003			0.003	0.036			0.036
Stone loach	0.000			0.000	0.009			0.009	0.004			0.004
Hybrid	0.000			0.000	0.000	•		0.000	0.001			0.001
Total	 		_	4.868				13.399				30.368

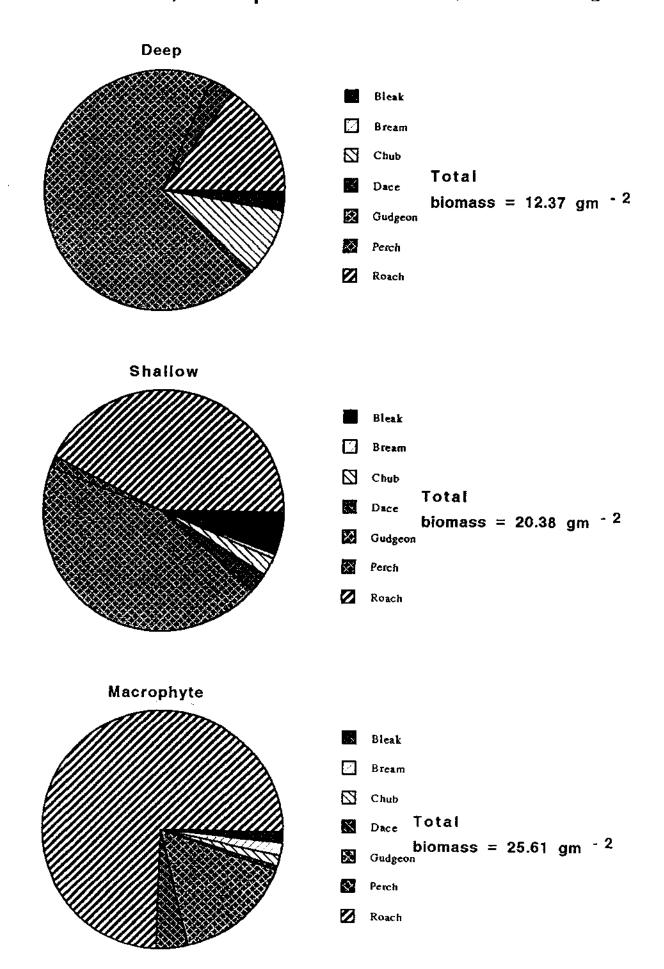
1+ Juveniles

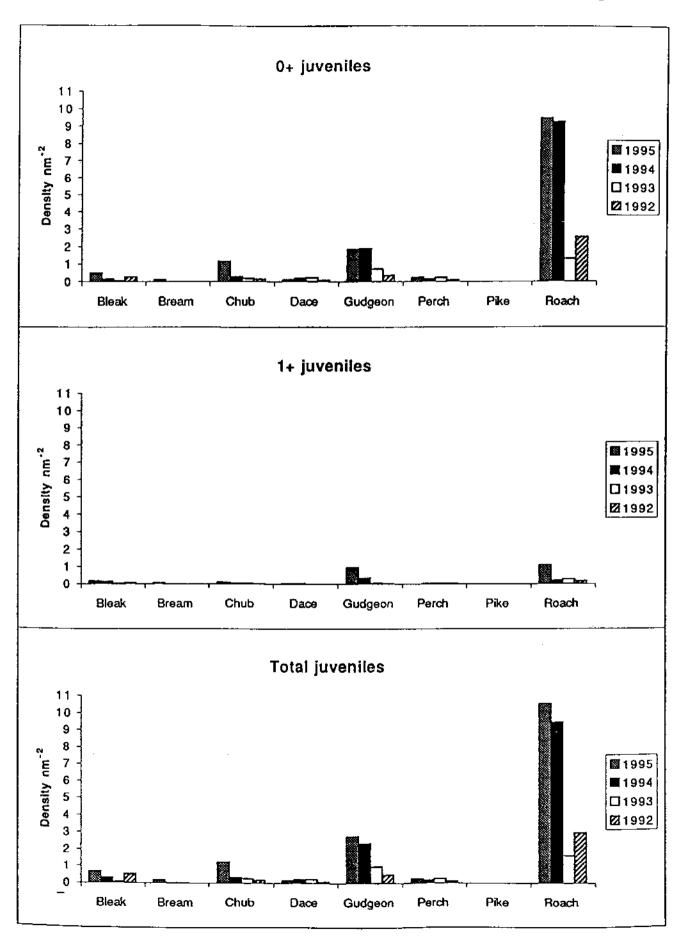


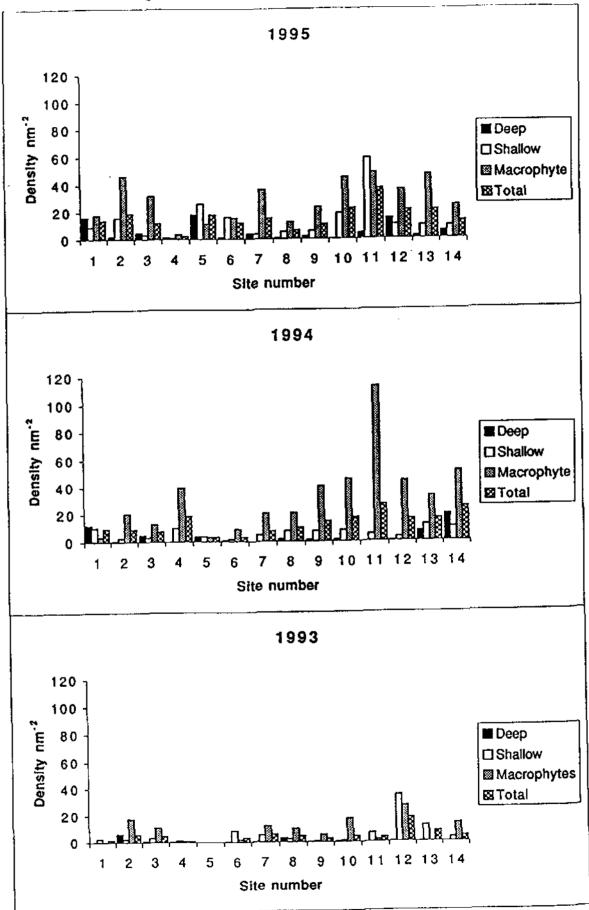
0+ Juveniles













River Thames O Group Fish Gut Contents Study 1995

RHK Mann GD Collett JAB Bass LCV Pinder



RIVER THAMES 0 GROUP FISH GUT CONTENTS STUDY 1995

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EXECUTIVE SUMMARY

This project on the diets of 0 group fish in the River Thames at Abingdon had four main objectives:

- a) Assessment of the between-species, between-habitats and between-season changes in fish diets,
- b) Comparison of the results from the River Thames with those of similar studies by IFE in the River Great Ouse,
- c) Assessment of the likely effects of changes in the flow regime of the River Thames on the food sources of 0 group fish,
- d) Recommendation for future studies that would increase the robustness of the results of the River Thames studies.

The diets were analyzed for five fish species: roach, gudgeon, chub, silver bream, perch. Although differences were observed between the diets of single species caught in different plant habitats, these were small compared with those observed between species. In general, the diet of each species changed from small prey (mostly rotifers) to larger items (mostly microcrustacea and insect larvae) as the fish grew in size. However, larger 0 group perch and gudgeon contained many copepods, whereas the other fish species contained more cladoceran taxa. The most striking difference between species was that, in July, the roach switched from an invertebrate diet to one dominated by detritus (aufwuchs) that accumulates on the underwater surfaces of plants.

The results from the River Thames show a close parallel with those of the same fish species in the River Great Ouse. This similarity adds weight to the reliability of the Thames results, even though they are based on small numbers of fish per sample. However, the Great Ouse data showed marked differences in the numbers and types of prey eaten in different years, which reflected changes in the abundance of these taxa in the river.

The Thames and Great Ouse studies highlight the importance of aquatic plants as feeding and refuge areas for young fish. Any major reduction in their areas, as could occur through large changes in the flow regime, would be detrimental to the growth of the young fish and to the numbers surviving to the adult stage.

Further studies to increase the reliability of the results from the River Thames are:

- a) examination of more of the fish caught in the 1995 study,
- b) collection of additional samples to determine the extent of year-to-year fluctuations.

For the greatest understanding of fish-prey dynamics, such studies should accompanied by assessments of prey availability (planktonic and non-planktonic invertebrates).

6. CONCLUSIONS

- a) As they grew in size the 0 group fish progressed from a diet of small prey items to one that included larger prey taxa. In July, the roach switched from an invertebrate diet to one predominantly of detritus (aufwuchs). All the other species continued to feed on invertebrates.
- b) Between-habitat differences in the diets of the same fish species could be explained by differences in the sizes of fish caught, or by the presence/absence of planktonic Cladocera (Bosmina). However, the results may have been affected by the ability of the larger 0 group fish to move between habitats.
- c) Between-species differences in diets greatly exceeded those observed for the same species in different habitats. These partly reflected where the fish were feeding in the water column, e.g. gudgeon fed on the river bed, whereas roach fed much nearer the water surface. Moreover, as each species spawned and hatched at a different time, their 0 group fish were at different stages of development on any one sampling occasion.
- d) The results closely parallel those obtained from IFE's studies in the Great Ouse. Both show the importance of aquatic plants as feeding and refuge areas for 0 group fish.

Table 7. 0 group roach: Chi-square analyses showing the statistical significance of differences in the contributions (numbers of animals) by each prey taxon. Ph = Phragmites, N = Nuphar, S = Salix, A = Acorus, Mix = Mixed reeds. Symbols refer to individual chi-square vales: = <3.0, + and o 3.0 - 15.0, ++ and oo >15.0; + values indicate higher than expected contribution, o values indicate lower than expected contribution; e/f = electrofishing.

	Ph	N seine	N e/f	S	A	Mix
6 June (chi-square = 1	18.01, 6 df)					
Rotifers	=	=	=			=
Chydoridae	=	=	==		-	+
Other taxa	O	=	=			=
20 June (chi-square =	555.1, 30 d	lf)				
Rotifers	00	++	-	=	0	=
Chydoridae	=	0	0	=	=	++
Bosmina	=	=	++	=	=	00
Polyphemus	++	O	00	=	+	00
Other Cladocera	=	=	=	=	=	=
Chiron. 1.	=	=	==	+	=	=
Other taxa	=	=	=	o	+	=
4 July (chi-square = 3	30.2, 15 df)					
Rotifers	0	==		+		=
Chydoridae	+-+-	0		00		++
Polyphemus	o	+		++		00
Other Cladocera	o	=		0		++
Chiron, I.	===	=		0		++
Other taxa	=	=		= .		++-
25 July (chi-square =	374.0. 12 di	Ð				
Chydoridae	=	00	=		=	++
Other Cladocera	00	++	=		00	00
Other taxa	+	0	=		=	++
Aufwuchs	++	00	=		++	=
22 August (chi-square	= 14.99, 3	df)				
All taxa	0	=			==	+
Aufwuchs	=	=			=	=
19 September (chi-squ	uare = 17.85	5, 4 df)				
All taxa	=	+	==	=		=
Aufwuchs	=	=	=	=		=

On 25 July comparisons between the same two sites revealed small but statistically significant differences (chi-square \approx 18.98, 6 df), with more chydorids and fewer copepods being found in fish from the *Salix* site, and the reverse at the *Nuphar* site. Nevertheless, the diets were very similar, with considerable overlap.

4.2.4 Chub

Table 9. 0 group chub: Chi-square analyses showing the statistical significance of differences in the contributions (number of animals) by each prey taxon. Key to symbols given in Table 7.

-40100001011111111111111111111111111111				
Habitat	Salix	Mixed	Nuphar	
20 June (chi-square	e = 67.22, 6df) 		
Rotifers	+	0	=	
Chydoridae	00	++	+	
Other Cladocera		=	=	
Other taxa	=	=	=	
	Salix	Mixed	Nuphar	Acorus
4 July (chi-square			•	
Rotifers	+	00	00	++
Chydoridae	===	++	o	00
Polyphemus	00	00	++	00
Other Cladocera	=	+	=	o
Chiron, I.	=	+	=	О
Other taxa	83	+	+	
	Mixed	d/s Acor	us Nuphar	
25 July (chi-square				
Chydoridae	0	=	+	
Other Cladocera	=	+-+-	0	
Copepoda	=	=	<u></u>	
Chiron. 1.	++ `	=	0	
Terrestrial prey	=	=	=	
Other taxa	=	=	=	

In the 4 July sample, the number of *Polyphemus* eaten increased with the length of chub (Figure 2), the data giving the following linear relationship:

Number of *Polyphemus* = 5.113 (SLmm) - 52.57 $r^2 = 0.68$

4.3 Comparisons between fish species

The following tables (11A-11Q)show the percentage number of prey items of various taxa that were eaten by different fish species caught at the same sites on the same sampling occasions.

Table 11. Comparison of prey taken by different fish species caught in the same habitat. Values are the percentage numbers of animals in each prey taxon, the highest contributions (>20%) being indicated in bold type. Aufwuchs values (index scale 0-3) are the means for each sample.

11(A) 6 June: Nuphar (combined seine and electro-fished samples)

	Roach	Gudgeon	Perch	
No. of fish	29	10	20	
No. of prey items	999	132	273	
Rotifers	93.50			= ************************************
Chydoridae	4.90	90.91	31.50	
Polyphemus		6.06	10.62	
Other Cladocera	1.17		3.66	
Copepoda		2.27	49.45	
Chiron. l.	0.43		4.03	
Other taxa		0.76	0.73	

11(B) 20 June: N	uni	har
------------------	-----	-----

LI(D) 20 odner i i	Roach	Gudgeon	Chub	
No. of fish	40	18	13	
No. of prey items	892	214	235	
Rotifers	56.05	3.27	73.62	
Chydoridae	21.52	76.17	17.02	
Polyphemus	5.38	0.47	1.70	
Other Cladocera	14.01	0.47	5.11	
Copepoda	1.23	13.55	0.43	
Chiron. I.	0.67	2.34	1.70	
Other taxa	1.12	3.74	0.43	

11(C) 20 June: Sal	lix		
	Roach	Gudgeon	
No. of fish	20	17	
No. of prey items	382	124	
Rotifers	42.15	:	
Chydoridae	33.51	81.45	
Polyphemus	7.07		·
Other Cladocera	13.61	•	
Copepoda	0.26	12.90	•
Chiron. I.	2.88	4.03	
Other taxa	0.52	1.61	
11(D) 20 June: Mi		CI 1	
	Roach	Chub	
No. of fish	20	12	
No. of prey items	404	121 	~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~
Rotifers	45.54	62.81	
Chydoridae	47.52	31.41	
Polyphemus	0.50		
Other Cladocera	1.24	3.31	
Copepoda	1.24	1.65	
Chiron. I.	1.24		
Other taxa	2.72	0.83	
***************************************		•	
11(E) 20 June: Ac		01 - D	·
	Roach	Silver Bream	
No. of fish	20	19	
No. of prey items	308 	226 ·	
Rotifers	36.69	98.23	
Chydoridae	29.87	1.33	
Polyphemus	12.66		
Other Cladocera	14.94		
Copepoda.	2.60	0.44	
Chiron. l.	0.32		
Other taxa	2.92		

11(F)	4	July:	Nupl	har
-------	---	-------	------	-----

No. of fish	Roach 20	Chub 20	Perch 20	
No. of prey items	564	564	413	
Rotifers	2.13	31.38	B-J	
Chydoridae	22.52	8.16	11.38	
Polyphemus	65.60	50.18	10.41	•
Other Cladocera	6.91	1.77	61.50	
Copepoda	0.18	6.56	12.59	•
Chiron. I.	2.48	1.60	2.91	
Other taxa	0.18	0.35	1.21	

11(G) 4 July: Phragmites

**(~)	0		
	Roach	Silver Bream	
No. of fish	6	20	
No. of prey items	131	361	
Rotifers		57.89	
Chydoridae	57.25	29.09	
Polyphemus	35.11	10.80	
Other Cladocera	3.05	2.21	
Chiron. l.	4.58		

11(H) 4 July: Mixed reeds

•	Roach	Chub	
No. of fish	20	15	·
No. of prey items	235	180	
Rotifers	3.40	13.33	
Chydoridae	45.53	65.00	
Polyphemus	20.85	6.67	
Other Cladocera	15.74a	2.78	a = Sida
Copepoda	0.85	1.11	
Chiron. I.	10.21	3.33	
Other taxa	3.40	7.78	
		_,	

44/100	4	T .	Acorus
	_	1271275	4 CABUS
	4		ALDIU
1-/	-	~ ~~.	

No. of fish No. of prey items	Chub 15 612	Silver Bream 15 107	
Rotifers	96.73	18.69	
Chydoridae	3.10	50.46	
Polyphemus		30.84	· · • •
Chiron. l.	0.16	•	

11(J) 4 July: Salix

Roach	Gudgeon 7	Chub	Silver Bream	
_	•			
	1,3,2	330		
4.68		75.44	82.67	
9.09	78.03	10.36	5.33	
80.99	6.82	9.47	2.00	
4.13	3.79	0.89	10.00	
	3.79	2.89		
0.55	4.55	0.89		
0.55	3.03			
	16 363 4.68 9.09 80.99 4.13 0.55	16 7 363 132 4.68 9.09 78.03 80.99 6.82 4.13 3.79 3.79 0.55 4.55	16 7 20 363 132 338 4.68 75.44 9.09 78.03 10.36 80.99 6.82 9.47 4.13 3.79 0.89 3.79 2.89 0.55 4.55 0.89	16 7 20 11 363 132 338 150 4.68 75.44 82.67 9.09 78.03 10.36 5.33 80.99 6.82 9.47 2.00 4.13 3.79 0.89 10.00 3.79 2.89 0.55 4.55 0.89

11(K) 25 July: Nuphar

TI(IC) 23 Duly. 144	pnui				
	Roach	Gudgeon	Chub	Perch	
No. of fish	37	15	15	19	
No. of prey items	322	186	170	681	
Chydoridae	6.83	33.33	44.12	4.11	
Sida	2.80		1.76	16.89	
Daphnidae		5.91	0.59	47.13	
Other Cladocera	86.96a		9.41	3.52	
Copepoda		34.41	11.18	20.70	
Chiron, I.	1.86	7.52	0.59	2.06	
Large crustacea				2.64	
Other taxa	1.55	18.82b	32.35c	2.94	
Aufwuchs (index)	2.54				

a = Bosmina b = Ostracoda c = Terrestrial insects

11(L) 25 July: Phi	agmites		
	Roach	Silver Br	eam
No. of fish	20	20	
No. of prey items	24	303	·
Rotifers	33.00		
Chydoridae	20.83	18.81	
Other Cladocera	33.33a	24.09a	a = mostly Bosmina
Copepoda	20.83	11.22	
Chiron, 1.	12.50	11.55	
Other taxa	12.50	1.32	
Aufwuchs (index)	2.65	0.30	

11(M) 25 July: Mi	ixed reeds		
	Roach	Chub	
No. of fish	20	15	
No. of prey items	36	161	
Chydoridae	55.56	22.36	
Other Cladocera	11.11	16.15	a = mostly Ostracoda
Copepoda		19.25	
Chiron. 1.	6.94	16.15	b = mostly terrestrial
Other taxa	26.94a	26.09b	insects
Aufwuchs (index)	1.80		

11(N) 25 July: Sal	ix		4
, ,	Gudgeon	Chub	
No. of fish	15	5	
No. of prey items	121	43	
Chydoridae	52.89	34.88	
Other Cladocera	4.13	18.60	a = mostly Ostracoda & Naidae
Copepoda	15.70	20.93	
Chiron. l.	5.79	6.98	b = mostly Naidae
Other taxa	21.49a	18.60b	

11(O) 22 August:	Downstream	of <i>Acorus</i>	site
	Roach	Perch	
No. of fish	20	15	•
No. of prey items	15	791	ii
Chydoridae	13.33	8.85	·
Sida		7.46	
Daphnidae		3.29	
Other Cladocera	6.67	0.76	
Copepoda		54.11	
Chiron. 1.	46.67	14.16	
Ostracoda	26.67	3.54	
Larger crustacea		4.42	
Other taxa	6.67	3.41	
Aufwuchs (index)	2.65		

11(P) 19 Septembe	er: <i>Nuphar</i> Roach	Perch	Silver Bream	
No. of fish	34	14	15	
No. of prey items	16	557	35	
Chydoridae	37.50	13.64	68.57	
Sida	6.25	2.69		
Other Cladocera		0.18	8.57	
Copepoda	6.25	66.43		
Chiron. I.	6.25	5.03		
Ostracoda	31.25	2.87	22.86	
Larger crustacea	6.25	2.87		
Other taxa	6.25	6.28		
Aufwüchs (index)	1.47		1.00	

11(Q) 19	September:	Upstream	of the S	<i>Salix</i> site
	R	oach	Silve	r Bream

Nf.E.	Roach	Silver Bream	•
No. of fish	22	16	
No. of prey items	33	417	
Rotifers	1.44		
Chydoridae	48.48	3.12	
Bosmina		88.73	
Copepoda	3.03	3.36	
Chiron. l.	6.06	0.24	
Ostracoda	36.36	2.88	·
Other taxa	6.06	0.24	
Aufwuchs (index)	1.06	0.19	·

4.4 Seasonal variations

Information on the seasonal variation in fish diets is presented as the mean number of each prey type per fish. This was found to be a useful measure in similar studies on the River Great Ouse, in which marked differences in prey intake were observed in different years (Mann et al., 1996).

Table 12. Mean numbers of animals per gut; aufwuchs given as the mean index, scale 0 - 3.

12(A) Roach: Phragmites							
	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	20	17	20	6	20	21	17
Mean SL (mm)	6.6	8.8	14.8	20.6	28.8	37.8	38.4
Rotifers	4.95	24.41	5.75				
Chydoridae		1.06	6.10	12.50	0.25	0.05	0.35
Bosmina			3.25				
Polyphemus			6.45	7.67			0.06
Other Cladocera			0.05	0.67	0.40		0.06
Chiron. I.			0.30	1.00		·	
Other taxa		0.06	0.40		0.50	0.19	0.35
Aufwuchs (index)					2.65	2.19	1.88

12(B) Roach: Mixed reed	S						
• •	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	10	20	20	20	20	20	20
Mean SL (mm)	6.7	9.8	12.9	21.5	30.6	36.6	39.9
Would DE (min)	···	7.0	12.7				
Rotifers	6.65	20.10	9.20	0.40			
Chydoridae		1.90	9.60	5.35	2.00	0.35	
Bosmina			0.10	0.05	0.30		
Polyphemus			0.10	2.45			
Other Cladocera			0.15	1.80	0.10		
Chiron. I.		0.30	0.25	1.20	-	0.15	
•	0.10	0.50	0.80	0.40	1.20	0.45	0.20
Other taxa	0.10		0.00	0.40	1.20	0.10	•••
Aufwuchs (index)					1.80	1.25	1.90
				.	11		
12(C) Roach: Nuphar (co			d electro	-tished s	amples)	22/0	19/9
	23/5	6/6	20/6	4/7	25/7	22/8	
No. of fish	20	29	40	20	37	27	34
Mean SL (mm)	6.9	10.6	15.3	21.8	30.1	38.4	40.1
D - A f	15.75	30.28	12.50	0.60			
Rotifers	(3.13	1.59	4.80	6.35	0.62	0.37	0.18
Chydoridae		0.31	3.10	0.25	1.84	0.04	•
Bosmina		0.31	1.20	18.50	1.0	0.0.	
Polyphemus		0.07		1.70	0.27		0.03
Other Cladocera		0.07	0.03		0.24	0.07	0.03
Chiron. 1.		0.14	0.15	0.70		0.07	0.24
Other taxa			0.53	0.10	0.35	0.22	0.24
Aufwuchs (index)					2.68	2.78	1.47
12(D) Gudgeon: All sites					15	0010	100
	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	0	29	39	9	30	0	0
Mean SL (mm)		11.4	12.5	19.9	16.8		
Rotifers			0.23				
Chydoridae Chydoridae		10.21	6.90	13.33	4.20		
-		0.59		1.11			
Polyphemus		0.77	0.03	0.44	0.53		
Other Cladocera		0.48		1.00			-
Copepoda				0.67			
Chiron. 1.		0.03		0.37			
Ostracoda		0.03	0.21				
Other taxa		0.03	0.05	0.23	U.71		

12(E) Chub: All sites con	mbined						
	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	0	0	44	70	50	15	0
Mean SL (mm)			10.5	11.9	15.1	12.9	
Rotifers			15.48	14.97	0.02	0.47	
Chydoridae			2.48	3.10	3.44	2.80	
Polyphemus			0.43	4.67			
Other Cladocera			0.61	0.26	1.84a	7.93b	
Copepoda			0.07	0.70	1.62	1.87	
Chiron. 1			0.14	0.27	0.70	0.87	
Other taxa			0.09	0.24	2.00c	1.40d	

a = mostly Bosmina, b = mostly Scapholeberis, c = terrestrial insects, d = terrestrial insects & Ostracoda.

12(F) Silver Bream: All	sites co	mbined					•
	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	0	0	19	46	20	0	31
Mean SL (mm)			7.2	9.3	13.1		29.2
Rotifers			11.68	7.67	5.00		0.19
Chydoridae			0.16	3.63	2.85		1.19
Polyphemus				1.63			
Other Cladocera	4.2			0.50	3.65a		12.03a
Copepoda			0.05		1.70		0.52
Chiron. I.					1.75		0.03
Other taxa					0.45		0.67
Aufwuchs (index)							0.58

a = mostly Bosmina

12(G) Perch: All sites co	mbined		•				
	23/5	6/6	20/6	4/7	25/7	22/8	19/9
No. of fish	0	20	4	20	19	15	29
Mean SL (mm)		13.4	17.3	28.5	44.2	54.9	60.2
Chydoridae		4.30	0.75	2.35	1.47	4.67	3.28
Polyphemus		1.45		2.15		1777	
Daphnidae		0.15			16.89	1.73	0.03
Sida-				12.70	6.05	4.21	0.55
Other Cladocera		0.35	2.50		1.26	0.40	
Copepoda		6.75	26.75	2.60	7.42	28.53	36.41
Chiron. 1.		0.55	0.25	0.60	0.74	7.47	4.17
Ostracoda		0.10			0.05	1.87	0.86
Larger crustacea					0.95	2.33	1.41
Other taxa		0.15		0.25	1.00	1.80	1.21

APPENDIX 3.1 A full list of the samples held in the macro-invertebrate data-base together with their identifiers and principal environmental descriptors.

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144 Danjock Tytile		70106160		2	26/07/84 443500	204700	70	0.33	_	99	33	183	7	200	7.3
144 Bablock Hythe	FBA	69130202	2 7	7010 26	26/07/84 443500	0 204200	L	0.33	1	198	33	183	7	L	L_
131 Newbridge	LR	28592	2	2 24	24/06/92			0.33	7	-	<u> </u> 	_	-		
131 Newbridge	LR	15292	111	1 08	08/05/92			0.33	7				_		_
131 Newbridge	LR	13694	-	2 12	12/05/94			0.33	7	i		-	-	_	ļ
131 Newbridge	LR	40892	2	1 24	24/08/92			0.33	7				-		L-
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131 Newbridge	LR	14094	2	3 13	13/06/94			0.33	7	 			-	_	Ļ.
25 Northmoor	FBA	FBA77_12	7	7 2:	23/08/77		-	0.33	7		 	-	-	<u> </u>	
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24 Piakhill	FBA	FBA77 11	7	8	23/08/77			0.33	7			<u> </u>		<u> </u>	-
24 Pinkhill	FBA	FBA77 11	7	11	01/08/77			0.33	7		_	-		-	_
130 Swinford	I.R	SW77	∞	-				0.33	7	-		-	-		
130 Swinford	I.R.	14194	2	3 13	13/06/94		 	0.33	1		_	_	-	-	-
130 Swinford	LR	SW86	8	71				0.33	7	_				_	
130 Swinford	LR	30394	3	-	12/09/94			0.33	7	<u> </u>	-	<u> </u>	-	-	_
130 Swinford	E.R.	28692	2	1 27	24/06/92		-	0.33	7	 -				 	_
130 Swinford	LR	15392	1	3 08	08/05/92			0.33	7				\vdash	L	ļ
130 Swinford	LR	13794	-	1 12	12/05/94			0.33	7				<u> </u>		_
130 Swinford	LR	40992	2	2 2/	24/08/92			0.33	7	 			 	_	
130 Swinford	LR	SW78	8	3			 	0.33	7				_	_	_
118 DS Swinford WTW	RPS	OSP.01	2 5	5000 30	30/06/92 444900	0 208800	62	0.33	7	71	15	200	0	0	0 100
35 Eynsham	FBA	FBA77_21B5		3215 24	24/08/77			0.33	7	_			 		
35 Eynsham	FBA	FBA77_21M6		2216 2	24/08/77			0.33	7					_	<u> </u>
35 Eynsham	FBA	FBA77 21M7			24/08/77			0.33	7	 				_	
35 Eynsham	FBA	FBA77_21M8	2 2	2218 24	24/08/77			0.33	7	 			_	_	
35 Eynsham	FBA	FBA77_21M9		2219 24	24/08/77			0.33	7				_		
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35 Eynsham	FBA	FBA77_21B1		3211 2	24/08/77			0.33	7			ļ	_		ļ
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35 Eynsham	FBA	FBA77_21B4	2 3	3214 24	24/08/77			0.33	7	 		-			<u> </u>
35 Eynsham	FBA	FBA77_21B6		3216 2	24/08/77			0.33	7			_	_	_	-
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TH19 King's	22 King's	FBA	FBA77_9		10 25/08/77		-		0.45	- 1				$^{+}$	
TH19 King's	44 Trout Inn, Godstow	WA/NRA/E	PTHR.0110	2 10500	28/07/92	448300	209200	59	0.45	-	74 25	220	Ē	5	10
TH19 King's	44 Trout Inn, Godstow	WANNRAJE	PTHR.0110	3 800	01/12/80	_	209200	<u></u>	0.45					1.	
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THIS Kings	44 I rout Inn, Godstow	WA/NRA/E	PTHR.0110				209200	- 28	0.45	7	_	<u>l</u>		↓_	
TUIO VILLE	44 Light Inn, Godstow		PTHR.0110	2 12000			209200		0.45	7	74 25		10		
Salika Aliika	44 HOUR THIS, GOGSEGW	WANNKAVE	PIHK.0110	1 13700	0 01/05/95 448300	- 1	209200	65	0.45	7	74 2		15	40	20 _i 25

44	44 Trout Inn, Godstow	WANTRAJE	PTHR.0110	3 16	16100	30/10/95	448300	209200	99 (0.45	1	74	25	230	15	40 2	20
	Godstow	RPS	OSP.04	2 5	5002	30/06/92		209200	59 (0.45	7	74	40	220	-	L	33
	119 West Mead	RPS	OSP.02	2 5	5001	30/06/92	447100	210200	09	0.45	7	72	25	250	0	0	20
	121 Binsey	RPS	OSP.07	2 5	5003	30/06/92 449400	449400	207800	58 (0.45	7	77	40	250	2	20 4	40
. •	21 Godstow	FBA	FBA77 8	2	œ	28/07/77				0.45	7						
\sim	21 Godstow	- 1	FBA77 8	2	=	25/08/77				0.45	7					 	
<u></u>	46 Donnington Bridge, Oxford WANRA/E		PTHTR,0186	3 16	16800	02/11/95	452400	204500	55	0.45	8	82	28	200	S.	30	35
1	46 Donnington Bridge, Oxford WA/NRA/E		PTHR.0186	1 13	13100	10/04/95	452400	204500	$\overline{}$	0.45	8	82	38	210	5		35
Ī	122 DS Cherwell	RPS	OSP.13	2 5	5004	07/06/92	452100	204800	57	0.45	8	81	30	200	2	18	45
프.	47 Folley Bridge, Oxford	WA/NRA/E	PTHR 0085	1 14	14500	23/05/95	451400	205500	57	0.45	8	80	51,	230	7	8	13
<u> </u>	47 Folley Bridge, Oxford	WA/NRA/E	PTHR.0085	3 16	16000	36/10/08	451400	205500	57	0.45	œ	08	51	230	2	55	15 25
_	20 Osney	FBA	FBA77_7	2	12	30/08/77			_	0.45	∞	-		_	-		ļ —
\sim	20 Osney	FBA	FBA77_7	2	7	28/07/77			_	0.45	00	-	_		-		ŀ
ľ	45 Osney Weir	WANNRAJE	PTHR.0097	2	5800	78/06/89	450600	205600	28	0.45	7	7	22	190	50	04	101
$\boldsymbol{\nu}$	45 Osney Weir	WANRAJE	PTHR.0097	-	2600	22/03/87	450600	205600	58	0.45	7	79	22	190	0	40	09 01
	45 Osney Weir	WANRAE	PTHIR.0097	<u></u>	4500	13/06/88	450600	205600	58	0.45	7	5	00	25	15	80	2
1	34 Iffley	FBA	FBA77 22B5	6	3225	30/08/77			ـــــ	0.45	000	ļ				\vdash	
ات ر	34 Iffley	FBA	FBA77_22M5		2225	30/08/77				0.45	œ	-	_		-	-	-
	34 Iffley	FBA	FBA77_22M6		2226	30/08/77			_	0.45	00				-	-	<u> </u>
	34 Iffley		FBA77_22M7_		2227	30/08/77			_	0.45	8		-		_	L	ļ
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	34 Iffley	FBA	FBA77_22M9	2	2229	30/08/77			-	0.45	8	-	-	_ 	-	 	_
<u> </u>	34 Iffley	FBA	FBA77 22B10		3220	30/08/77]			0.45	90		ļ			-	
ı		FBA	FBA77_22B1		3221	30/08/77				0.45	80		_		-	-	ļ
	34 Iffley	FBA	FBA77_22B2	2	3222	30/08/77				0.45	8		-		-	_	
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ı		FBA	FBA77_22B6		3226	30/08/77				0.45	8					\vdash	<u> </u>
ت ا		FBA	FBA77_22B7	2	3227	30/08/77				0.45	8		_		-	-	
Ľ		FBA	FBA77_22B9	2	3229	30/08/77			_	0.45	∞	-	_		-		-
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	34 Iffley	FBA	FBA77_22B8	2	3228	30/08/77				0.45	∞	<u> </u>	-		\vdash	-	
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1	19 Iffley	FBA	FBA77_6	2	9	28/07/77				0.45	oo.					-	<u> </u>
Ţ	19 Iffley	FBA	FBA77_6	2	13	30/08/77				0.45	8					<u> </u>	¦ —
-1	34 Iffley	FBA	FBA77 22V10		1220	30/08/77				0.45	8		_	_	-	-	-
크		FBA,	FBA77 22V2	~	- 1	30/08/77]	0.45	8					_	
E	24 [fillow	FRA	FBA77 22V3		1223	30/08/77				0.45	∞		_	r	L	<u> </u>	ļ.

TH22 Iffley		34 Iffley	FBA	FBA77 22V4	2	1224 3	30/08/77				0.45	8	-					
TH22 IMey		34 IMey	FBA	FBA77_22M1	2	2221 3	30/08/77	-	-	-	0.45	œ	-					<u>-</u>
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TH22 IMey		48 Top of Sandford Lock Cut	WANNRAJE	PTHR.0109		12700 0	08/11/94 452800	_	202100	22	0.45	∞	85	55 280	2	55	30	10
TH22 IMey	7	48 Top of Sandford Lock Cut	WA/NRA/E	PTHR.0109	-	8500	08/04/91 452800	╙	202100	ـــــ	0.45	90	85	55 280	0	2	20	10
TH22 Iffley	•	48 Top of Sandford Lock Cut	ÞΑ	T14	2	1	09/07/92 4	452800	202100	54	0.45	œ	85	60 150	0	9	20	01
TH22 Iffley	7	48 Top of Sandford Lock Cut	WA/NRA/E	PTHR.0109	_	1100 2	24/02/93 4	452800	202100	22	0.45	00	85	55 280	0	20	20	9
TH22 Iffley	,	48 Top of Sandford Lock Cut	WANTRA/E	PTHR.0109	2 1	10700	11/08/92 4	452800	202100	54	0.45	90	85	55 280	0	5	20	10
TH22 Iffley	,	48 Top of Sandford Lock Cut	WANRAE	PTHR.0109		10200 2	26/02/92 4	L	202100	\$	0.45	90	88	55 280	0	0,	20	10
TH22 (IO)ey	*	48 Top of Sandford Lock Cut	WANRAE	PTHR.0109	3	9700 3	30/10/91 4		202100		0.45	8	85	55 280	0		20	10
THZ2 IMey		48 Top of Sandford Lock Cut	WANNRAKE	PTHR.0109		9200{ 0	03/07/91 4	452800	202100	54	0.45	8	85	55 280	0	70	20	10
TH22 Iffley	1	48 Top of Sandford Lock Cut		OSP.15	2	5005	30/06/92 452800		202100	¥	0.45	8	85	25 150	6		1	O
TH22 Iffley		48 Top of Sandford Lock Cut	WANNRAJE	PTHR.0109		7700 2	20/08/90 452800		202100		0.45	8	88	55 280	0	5	20	01
TH22 IMey	*	48 Top of Sandford Lock Cut		PTHR.0109		6300 1	17/04/90 452800		202100	54	0.45	8	85	55 280	0	20	20	10
TH22 Mey	7	48 Top of Sandford Lock Cut	WANNRAJE	PTHR.0109	1	5700 0	08/05/89 452800		202100		0.45	8	85	55 280	0	70	20	10
TH22 Miley	4	48 Top of Sandford Lock Cut	WANNRAJE	PTHR.0109	1	4600 1	13/06/88 4	452800	202100	¥	0.45	8	85	55 280	5	55	20	2
TH22 IMey	*	48 Top of Sandford Lock Cut	WANRAE	PTHR.0109	-	2500 2		452800	202100		0.45	8	85	55 280	0 (30	10	09
TH22 Iffley	4	48 Top of Sandford Lock Cut	WANRAÆ	PTHR.0109	1	1400 0	09/03/84 4		202100		0.45	8	8.5		_			
TH22 Iffley	7	48 Top of Sandford Lock Cut	WANNRA/E	PTHR.0109	1	1000	18/02/82 4	452800	202100	54	0.45	8	85					
TH22 Iffley		48 Top of Sandford Lock Cut	WANRAE	PTHR.0109		7900 0	08/10/90 452800		202100	54	0.45	80	85	55 280	0	5	20	10
TH23 Sandford		49 Abingdon Weir	WA/NRA/E	WA/NRA/E PTHR.0077	3 1	11000 2	23/11/92 450400		197200	20	0.45	8	93	40 300	1	79	15	S
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TH23 Sandford	-	49 Abingdon Weir	WANNRA/E	PTHR.0077	1	11400 2	24/05/93 450400	_	197200		0.45	0 0	93	40 300	5		15	\$
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TH23 Sandford	-	49 Abingdon Weir	WANRAE		-	4700	15/06/88 450400		197200	20	0.45	8	93	40 300	1 0	50	39	10
TH23 Sandford		49 Abingdon Weir	WANNRATE		-		09/03/84 450400		197200	50	0.45	8	93					
TH23 Sandford	-	49 Abingdon Weir	WANNRA/E	PTHR.0077		1200	26/01/83 450400		197200	20	0.45	8	93	30 80	0			
TH23 Sandford		49 Abingdon Weir	WANRAE	PTHR.0077	-	1100 1	18/02/82 450400		197200	20	0.45	8	93					
TH23 Sandford		49 Abingdon Weir	WA/NRA/E	PTHR.0077	3	600	20/11/80 450400		197200	20	0.45	8	93					
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TH23 Sandford	140 Radiey	PA	T12	2	1 20/07/92	20/07/92 453800	000661	53 (0.45	8	88	40 150	0	45	45	2
TH23 Sandford	124 Radiey College Boathouse	RPS	OSP.16	2 5006	1	30/06/92 453800	198800	53 (0.45	8	88	20 200	30	30	35	\ <u>`</u> \
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TH25 Culham	16 Culham	FBA	FBA77_3		16 01/09/77				0.26	8						<u> </u>
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H H H H	TH30 Goring	129 Whitchurch	LR	15592	1	1 08/05/92			Ĕ	0.26	90					-	
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TH30	TH30 Goring	55 Whitchurch Weir	WANRAE PTHR.0115	PTHR.0115	2 9000	0 03/07/91 463300	L	176800	42 0	0.26	000				: %	× ×	ार
TH30 Goring	Goring	55 Whitchurch Weir	WANNRA/E PTHR.0115	PTHR.0115	2 3700	24/07/87	463300	176800	42 0	0.26		L			1	1 8	1
TH30 Goring	Goring	55 Whitchurch Weir	WANRAE PTHR.0115	PTHR.0115	1 13400	0 12/04/95 463300	L	176800	42 0	0.26					, 8	2 2	75
TH30 Goring	Goring	55 Whitchurch Weir	WA/NRA/E PTHR.0115	PTHTR.0115	2 7500	0 23/07/90 463300	نط	176800	42	0.26	œ				~	8	٠,
TH30 Goring	Goring	55 Whitchurch Weir	WANRAE PTHR.0115	PTHR.0115	1 8900	0 15/05/91 463300		176800	42 0	0.26	∞				8	25	12
TH30 Goring	Goring	55 Whitchurch Weir	WA/NRA/E- PTHR.0115	PTHR.0115	3 9900	0 14/11/91 463300		176800	42 0	0.26	90	135 50		0	ક્ર	22	25
TH30 Goring	Goring	55 Whitchurch Weir	WANRA/E PTHR 0115	PTHR.0115	1 10300	0 28/05/92 463300	<u> </u>	176800	42 0	0.26	80				S.	25	25
TH30 Goring	Goring	55 Whitchurch Weir	WANRAE PTHR.0115	PTHR.0115	2 10600	0 11/08/92 463300		176800	42 0	0.26	∞	135 50		0	S	25	12
TH30 Goring	Goring	55 Whitchurch Weir	WA/NRA/E- PTHR.0115	PTHR.0115	3 11600	0 04/10/93 463300		176800	42 0	0.26	00	135 50		0	8	25	25
TH30 Goring	Goring	55 Whitchurch Weir	WA/NRA/E PTHR.0115	PTHR.0115	2 11900	0 28/07/94 463300		0089/1	42 0	0.26	90	135 50	230	0	8	25	122
1H32	TH32 Mapledurha	57 Caversham Weir	WANRAE PTHR 0080	PTHR.0080	3 900	` -I		174100	410	0.26	00	145				-	T
1H32	TH32 Mapledurha	57 Caversham Weir	WA/NRA/E PTHR 0080	PTHR.0080	1 2300	18/02/87	471700	174100	41 0	0.26	∞	145 75	280		-		
TH32	TH32 Mapledurha	57 Caversham Weir	WA/NRA/E PTHR 0080	PTHR.0080	2 4000	0 12/08/87 471700		174100	41 0	0.26	•	145 75		~	99	25	9
1H32	TH32 Mapledurha	57 Caversham Weir	WA/NRA/E PTHR.0080	PTHR.0080	1 4400	0 09/05/88 471700		174100	41 0	0.26	œ	145 75	_	~	8	25] =
11132	TH32 Mapledurha	57 Caversham Weir	WAARAE PTHR.0080	PTHR.0080			I	174100	41 0	0.26	8	145 75	280	S	8	25	12
1H32 I	IH52 Mapledurha	57 Caversham Weir	WANRA/E PTHR.0080	PTHR.0080	3 6000	0 02/10/89 471700		174100	41 0	0.26	90	145 75	280	2	9	25	2

APPENDIX 3.2 A full list of the taxa recorded in one or more samples held in the macro-invertebrate data-base.

Species found in entire database Invertebrates only

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02110000	Spongillidae	17120000 Unionidae
03110000	Hydridae	17120100 Unio sp.
03110100	Hydra sp.	17120101 Unio pictorum (L.)
05110000	Planariidae	17120102 Unio turnidus Philipsson
05110101	Planaria torva (Muller)	17120200 Anodonta sp.
05110203	Polycelis tenuis (Ijima)	17120201 Anodonta anatina (L.)
0511020Z	· · · · · · · · · · · · · · · · · · ·	17120202 Anodonta cygnea (L.)
05120103		17130000 Sphaeriidae
05120102		17130100 Sphaerium sp.
05130201	Dendrocoelum lacteum (Muller)	17130101 Sphaerium corneum (L.)
051Z0000	Planariidae (incl. Dugesiidae)	17130102 Sphaerium lacustre (Muller)
08110100	Prostoma sp.	17130103 Sphaerium rivicola (Lamarck)
10000000	Nematoda	17130105 Sphaerium transversum (Say)
14000000	Ectoprocta	17130200 Pisidium sp.
16110000	Neritidae	•
16110101		17130201 Pisidium amnicum (Muller)
16120101	Theodoxus fluviatilis (L.)	17130202 Pisidium casertanum (Poli)
	Viviparus contectus (Millet)	17130204 Pisidium henslowanum (Sheppard)
16120102	Viviparus viviparus (L.)	17130208 Pisidium moitessierianum Paladilhe
16130000	Valvatidae	17130209 Pisidium nitidum Jenyns
16130100	Valvata sp.	17130215 Pisidium subtruncatum Malm
16130101	Valvata cristata Muller	17130216 Pisidium supinum Schmidt
16130103	Valvata piscinalis (Muller)	20000000 Oligochaeta
16140000	Hydrobiidae	20110000 Lumbriculidae
16140301	Potamopyrgus jenkinsi (Smith)	20110300 Stylodrilus sp.
16160101	Bithynia leachii (Sheppard)	20110302 Stylodrilus heringianus Claparede
16160102	Bithynia tentaculata (L.)	20110Z00 Lumbriculus group
16120000	Hydrobiidae (incl. Bithyniidae)	20310000 Enchytraeidae
16210000	Physidae	20330000 Naididae
16210200	Physa sp.	20330100 Chaetogaster sp.
16210202	Physa fontinalis (L.)	20330501 Uncinais uncinata (Orsted)
1621020Z	Physa acuta group	20330601 Ophidonais serpentina (Muller)
16220000	Lymnaeidae	20330702 Nais barbata Muller
16220101	Lymnaea auricularia (L)	20330703 Nais bretscheri Michaelsen
16220103	Lymnaea palustris (Muller)	20330706 Nais pardalis Piguet
16220104	Lymnaea peregra (Muller)	20330708 Nais simplex Piguet
16220105	Lymnaea stagnalis (L.)	2033070Y Nais communis group
16220106	Lymnaca truncatula (Muller)	20331201 Stylaria Incustris (L.)
16230000	Planorbidae	20340000 Tubificidae
16230101	Planorbis carinatus Muller	20340102 Tubifex ignotus (Stole)
16230102	Planorbis planorbis (L.)	20340104 Tubifex newaensis (Michaelsen)
16230201	Anisus leucostoma (Millet)	20340106 Tubifex tubifex (Muller)
16230202	Anisus vortex (L.)	20340201 Limnodrilus cervix Brinkhurst
16230301	Bathyomphalus contortus (L.)	20340202 Limnodrilus claparedeianus Ratzel
16230401	Gyraulus acronicus (Ferussac)	20340203 Limnodrilus hoffineisteri Claparede
16230402	Gyraulus albus (Muller)	20340204 Limnodrilus profundicola (Verrill)
16230501	Armiger crista (L.)	20340205 Limnodrilus udekemianus Claparede
16230601	Hippeutis complanatus (L.)	20340302 Psammoryotides barbatus (Grube)
16230801	Planorbarius comeus (L.)	20340402 Potamothrix hammoniensis (Michaelsen)
16240000	Ancylidae	20340404 Potamothrix moldaviensis (Vejdovsky & Mrazel
16240101	Ancylus fluviatilis Muller	20340903 Aulodrilus pluriseta (Piguet)
IOTAGIGI		1
16250101	Acroloxus lacustris (L.)	20341101 Rhyacodrilus coccineus (Veidovsky)
	Acroloxus lacustris (L.) Ancylidae (incl. Acroloxidae)	20341101 Rhyacodrilus coccineus (Vejdovsky) 20341301 Branchiura sowerbyi Beddard

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22110000	Piscicolidae	4012011Z	Baetis scambus group
22110101	Piscicola geometra (L.)	40120201	Centroptilum luteolum (Muller)
22120000	Glossiphoniidae	40120300	Cloeon sp.
22120201	Theromyzon tessulatum (Muller)	40120301	Closon dipterum (L.)
22120301	Hemiclepsis marginata (Muller)	40120302	Clocon simile Eaton
22120401	Glossiphonia complanata (L.)	40120401	Procloeon bifidum Bengtsson
22120402	Glossiphonia heteroclita (L.)	40130000	Heptageniidae
22120601	Boreobdella verrucata (Muller)	40130201	Heptagenia fuscogrisea (Retzius)
22120701	Helobdella stagnalis (L.)	40210000	Leptophlebiidae
22310000	Erpobdellidae	40210301	Habrophlebia fusca (Curtis)
22310100	Erpobdella sp.	40320000	Ephemeridae
22310101	Erpobdella octoculata (L.)	40320100	Ephemera sp.
22310201	Dina lineata (Muller)	40320101	Ephemera danica Muller
22310302	Trocheta subviridis Dutrochet	40320103	Ephemera vulgata L.
24000000	Hydracarina	40410000	Ephemerellidae
24320107	Eylais extendens (Muller)	40410101	Ephemerella ignita (Poda)
24320111	Eylais infundibulifera Koenike	40510000	Cacnidae
24420101	Hydrodroma despiciens (Muller)	40510200	Caenis sp.
24540144	Lebertia (Pilolebertia) inaequalis (Koch)	40510201	Caenis horaria (L.)
24540145	Lebertia (Pilolebertia) insignis Neuman	40510203	Caenis macrura Stephens
24540149	Lebertia (Pilolebertia) porosa Thor	40510205	Caenis rivulorum Eaton
24560111	Torrenticola (Torrenticola) amplexa (Koenike)	40510206	Caenis robusta Eaton
24610113	Limnesia (Limnesia) koenikei Piersig	4051020Z	Caenis luctuosa group
24610114	Limnesia (Limnesia) maculata (Muller)	41110000	Tacniopterygidae
24610115	Limnesia (Limnesia) undulata (Muller)	41120000	Nemouridae
24620112	Hygrobates (Hygrobates) fluviatilis (Strom)	41120400	Nemoura sp.
24620114	Hygrobates (Hygrobates) longipalpis (Hermann)	41130000	Leuctridae
24630121	Unionicola (Pentatax) aculeata (Koenike)	41130102	Leuctra geniculata (Stephens)
24630211	Neumania (Neumania) callosa (Koenike)	41210000	Perlodidae
24650200	Piona sp.	42110000	Platycnemididae
24650207	Piona coccinca (Koch)	42110101	Platycnemis pennipes (Pallas)
24650208	Piona conglobata (Koch)	42120000	Coenagriidae
24650218	Piona pusilla (Neuman)	42120201	Ischnurz elegans (Van der Linden)
24650913	Forelia (Forelia) variegator (Koch)	42120301	Enallagma cyathigerum (Charpentier)
24730112	Mideopsis (Mideopsis) orbicularis (Muller)	4212040Z	Coenagrion puella group
24770100	Arrenurus sp.	42120601	Erythromma najas (Hansemann)
24770113	Arrenurus (Arrenurus) albator (Muller)	42140000	Calopterygidae
30000000	Ostracoda	42140100	Calopteryx sp.
32010100	Argulus sp.	42140101	Calopteryx splendens (Harris)
36110000	Asellidae	42140102	Calopteryx virgo (L.)
36110101	Asellus aquaticus (L.)	42210000	Gomphidae
36110104	Asellus meridianus Racovitza	42210101	Gomphus vulgatissimus (L.)
37110000	Corophiidae	42230000	Aeshnidae
37110101	Corophium curvispinum Sars	42230200	Aeshna sp.
37130101	Crangonyx pseudogracilis Bousfield	42230202	Aeshna cyanea (Muller)
37140000	Gammaridae	42230203	Aeshna grandis (L.)
37140200	Gammarus sp.	42230301	Anax imperator Leach
37140206	Gammarus pulex (L.)	42250000	Libelfulidae
371Z0000	Gammaridae (incl. Crangonyctidae & Niphargidae)	42250306	Sympetrum striolatum (Charpentier)
40120000	Baetidae	43110000	Mesovelidae
40120100	Baetis sp.	43210000	Hydrometridae
40120107	Baetis rhodani (Pictet)	43210102	Hydrometra stagnorum (L.)
40120111	Bactis vernus Curtis	43220000	Veliidae

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43230000		45150212	Gyrinus urinator Illiger
43230100		45150401	Orectochilus villosus (Muller)
43230114	Gerris (Gerris) lacustris (L.)	451Z0000	Dytiscidae (incl. Noteridae)
433100 00	Nepidae	45310000	Hydrophilidae
43310101	•	45310300	Helophorus sp.
33102 01	Ranatra linearis (L.)	45310352	Helophorus (Atracthelophorus) brevipalpis Bede
43410000	Naucoridae	45311101	Hydrobius fuscipes (L)
13420000	Aphelocheiridae	45311301	Anacaena bipustulata (Marsham)
43420101	Aphelocheirus aestivalis (Fabricius)	45311302	Anacsena globulus (Paykuli)
13510000	Notonectidae	45311303	Anacaena limbata (Fabricius)
3510100	Notonecta sp.	45311400	Laccobius sp.
3510101	Notonecta glauca L.	45311412	Laccobius (Laccobius) minutus (L.)
3510102	Notonecta maculata Fabricius	45311423	Laccobius (Macrolaccobius) bipunctatus (Fabric
3510104	Notonecta viridis Delcourt	45311426	Laccobius (Macrolaccobius) sinuatus Motschuls
3610000	Corixidae	45311427	Laccobius (Macrolaccobius) striatulus (Fabricius
3610100	Micronecta sp.	453Z0000	Hydrophilidae (incl. Hydraenidae)
3610111		45620200	Dryops sp.
3610122		45630000	Elmidae
3610900	71	45630101	Elmis aenea (Muller)
3610910		45630301	
3610921	Sigara (Subsigara) distincta (Fieber)		Limnius volckmari (Panzer)
3610922	Sigara (Subsigara) falleni (Fieber)	45630600	Outlimnius sp.
3610924	Sigara (Subsigara) fossarum (Leach)	45630601	Oulimnius major (Rey)
		45630604	Oulimnius tuberculatus (Muller)
5110000	Haliplidae	45710200	Donacia sp.
5110101	Brychius elevatus (Panzer)	46110000	Sialidae
5110300	Haliplus sp.	46110101	Sialis fuliginosa Pictet
5110303	Haliplus flavicollis Sturm	46110102	Sialis lutaria (L.)
5110304	Haliplus fluviatilis Aube	46110103	Sialis nigripes Pictet
5110308	Haliplus immaculatus Gerhardt	47120100	Sisyra sp.
5110309	Haliplus laminatus Schaller	48130000	Hydroptilidae
5110311	Haliplus lineatocollis (Marsham)	48130101	Agraylea multipunctata Curtis
5110314	Haliplus obliquus (Fabricius)	48130300	Hydroptila sp.
5110315	Haliplus ruficollis (Degeer)	48130400	Oxyethira sp.
5130101	Noterus clavicornis (Degeer)	48130600	Ithytrichia sp.
5140000	Dytiscidae	481Z0000	Rhyacophilidae (incl. Glossosomatidae)
5140100	Laccophilus sp.	48220100	Lype sp.
5140101	Laccophilus hyalinus (Degeer)	48220101	Lype phaeopa (Stephens)
5140301	Hyphydrus ovatus (L.)	48220102	Lype reducta (Hagen)
5140604	Hygrotus versicolor (Schaller)	48220400	Tinodes sp.
140800	Hydroporus sp.	48220408	Tinodes waeneri (L.)
5140812	Hydroporus incognitus Sharp	48230101	Ecnomus tenellus (Rambur)
140824	Hydroporus palustris (L.)	48240000	Polycentropodidae
141303	Potamonectes depressus (Fabricius)	48240101	Cymus flavidus Melachlan
141401	Stictotarsus duodecimpustulatus (Fabricius)	48240103	Cymus trimaculatus (Curtis)
141901	Platambus maculatus (L.)	48240301	Neureclipsis bimaculata (L.)
142009	Agabus didymus (Olivier)	48240400	, , ,
142018	Agabus sturmii (Gyllenhal)	48240400	Plectromemia sp.
142102	Ilybius ater (Degeer)		Plectrocnemia conspersa (Curtis)
142104	Ilybius fuliginosus (Fabricius)	48240403	Plectrocnemia geniculata Melachlan
142700	Dytiscus sp.	48240500	Polycentropus sp.
150000	Gyrinidae	48240501	Polycentropus flavomaculatus (Pictet)
150200	· ·	48240502	Polycentropus irroratus (Curtis)
	Gyrimus sp.	48250000	Hydropsychidae
5150204	Gyrinus distinctus Aube	48250200	Hydropsyche sp.

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48250201	Hydropsyche angustipennis (Curtis)	50420500	Procladius sp.
18250207	Hydropsyche pellucidula (Curtis)	50420601	Psectrotanypus varius (Fabricius)
182Z0000	Psychomyiidae (incl. Ecnomidae)	50420800	Ablabesmyia sp.
8310000	Phryganeidae	50420802	Ablabesmyia monilis (L.)
8310500	Phryganea sp.	50420900	Arctopelopia sp.
8310501	Phryganea bipunctata Retzius	50420Y00	Thienemannimyia group
8320000	Brachycentridae	50421000	Conchapelopia sp.
832010I	Brachycentrus subnubilus Curtis	50421001	Conchapelopia melanops (Meigen)
8330301	Lepidostoma hirtum (Fabricius)	50421400	Larsia sp.
8340000	Limnephilidae	50421600	Natarsia sp.
8340501	Allogamus auricollis (Pictet)	50421701	Nilotanypus dubius (Meigen)
8340600	Halesus sp.	50421800	Paramerina sp.
8340602	Halesus radiatus (Curtis)	50421900	Rheopelopia sp.
8341102	Potamophylax latipennis (Curtis)	50422100	Thienemannimyia sp.
8341401	Anabolia nervosa (Curtis)	50422501	Tanypus punctipennis Meigen
8341700	Limnephilus sp.	50440200	Diamesa sp.
8341719	Limnephilus lunatus Curtis	50440300	Potthastia sp.
R341X00	•	5044030Y	Potthastia gaedii group
8350000	Goeridae	5044030Z	Potthastia longimana group
8350101	Goera pilosa (Fabricius)	50450201	Odontomesa fulva (Kioffer)
	Silo sp.		Prodiamesa olivacea (Meigen)
8350200	Sericostomatidae	50450301	
8370000	Molannidae	50460000	Orthocladiinae
8390000		50460300	Brillia sp.
R390101	Molanna angustata Curtis	50460301	Brillia flavifrons Johannson
8410000	Leptoceridae	50460402	Cardiocladius fuscus Kieffer
8410102	Athripsodes aterrimus (Stephens)	50460500	Cricotopus sp.
8410104	Athripsodes cinereus (Curtis)	50460510	Cricotopus (Cricotopus) sp.
8410200	Ceraclea sp.	50460513	Cricotopus (Cricotopus) bicinctus (Meigen)
8410202	Ceraclea annulicornis (Stephens)	50460520	Cricotopus (Cricotopus) sp.
R410203	Ceraclea dissimilis (Stephens)	50460540	Cricotopus (Isocladius) sp.
8410206	Ceraclea senilis (Burmeister)	50460549	Cricotopus (Isocladius) sylvestris (Fabricius)
8410302	Leptocerus lusitanicus (Melachlan)	50460561	Cricotopus (Nostrocladius) lygropis Edwards
8410400	Mystacides sp.	50460800	Eukiefferiella sp.
8410401	Mystacides azurea (L.)	50460802	Eukiefferiella claripennis (Lundbeck)
8410402	Mystacides longicomis (L.)	50460Z00	Cricotopus group
8410403	Mystacides nigra (L.)	50461300	Heterotrissociadius sp.
8410701	Triaenodes bicolor (Curtis)	50461800	Nanocladius sp.
8410900	Oecetis sp.	50461801	Nanocladius balticus Palmen
8410902	Oecetis lacustris (Pictet)	50462000	Orthocladius sp.
8410904	Oecetis ochracea (Curtis)	50462010	Orthocladius (Eudactylocladius) sp.
0000000	Diptera	50462021	Orthocladius (Pogonocladius) consobrinus (Holms
0100000	Tipulidae	50462030	Orthocladius (Enorthocladius) sp.
0110300	Tipula sp.	50462040	Orthocladius (Orthocladius) sp.
110412	Tipula (Yamatotipula) montium group	50462050	Orthocladius (Symposiocladius) sp.
130900	Helius sp.	50462100	Paracladius sp.
0350000	Ceratopogonidae	50462101	Paracladius conversus (Walker)
360000	Simuliidae	50462301	Paratrichocladius rufiventris (Meigen)
0360361	Simulium (Boophthora) erythrocephalum (de Geer)	50462700	Psectrocladius sp.
0400000	Chironomidae	50462721	Psectrocladius (Aliopsectrocladius) obvius (Walke
0420101	Clinotanypus nervosus (Meigen)	50462740	Psectrocladius (Psectrocladius) sp.
0420201	Apsectrotanypus trifascipennis (Zetterstedt)	50462800	Rheocricotopus sp.
			•
0420400	Macropelopia sp.	50462901	Synorthocladius semivirens (Kieffer)

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50463512	Bryophaenocladius subvernalis (Edwards)	504
50463700	Chaetocladius sp.	504
50463800	Corynoneura sp.	504
50463901	Epoicocladius flavens (Malloch)	504
50464300	Limnophyes sp.	504
50464500	Metriocnemus sp.	506
50464700	Parakiefferiella sp.	506
50464701	Parakiefferiella bathophila (Kieffer)	507
50464900	Parametriocnemus sp.	508
50465201	Paratrissociadius excerptus (Walker)	508
50465300	Pseudorthocladius sp.	
50465400	Pseudosmittia sp.	
50465900	Thienemanniella sp.	
50466300	Paracricotopus sp.	
50470300	Chironomus sp.	
50470400	Cladopelma sp.	
50470500	Cryptochironomus sp.	
50470600	Cryptotendipes sp.	
50470701	Demeijerea rufipes (L.)	
50470801	Demicryptochironomus vulneratus (Zetterstedt)	
50470900	Dicrotendipes ap.	
50470920	Dicrotendipes (Limnochironomus) sp.	
50470922	Dicrotendipes (Limnochironomus) nervosus (Staeger)	
50470923	Dicrotendipes (Limnochironomus) notatus (Meigen)	
50471100	Endochironomus sp.	
50471200	Glyptotendipes sp.	
50471400	Harnischia sp.	
50471501	Kiefferulus tendipediformis (Goetghebuer)	
50471800	Microchironomus sp.	
50471900	Microtendipes sp.	
50472300	Parachironomus sp.	
50472400	Paracladopelma sp.	
50472401	Paracladopelma camptolabis (Kieffer)	
5047240Z	Paracladopelma camptolabis group	
50472500	Paralauterborniella sp.	
50472501	Paralauterborniella nigrohalteralis (Malloch)	
50472600	Paratendipes sp.	
50472800	Phaenopsectra sp.	
50472900	Polypedilum sp.	
50472910	Polypedilum (Pentapedilum) sp.	
50472920	Polypedilum (Polypedilum) sp.	
50473100	Stenochironomus sp.	
50473200	Stictochironomus sp.	
50473301	Xenochironomus xenolabis (Kieffer)	
50490100	Cladotanytarsus sp.	
50490200	Micropsectra sp.	
50490203	Micropsectra atrofasciata Kieffer	
50490300	Neozavrelia sp.	
50490500	Paratanytarsus sp.	
50490600	Rheotanytarsus sp.	
50490701	Stempellina bausei (Kieffer)	
50490702	Stempellina almi Brundin	
50490800	Stempellinella sp.	

490900 Tanytarsus sp. 490902 Tanytarsus brundini Lindeberg 490W00 Micropsectra group 491000 Virgatanytarsus sp. 491101 Zavrelia pentatoma Kieffer 630000 Tabanidae 630100 Chrysops sp. 710000 Empididae 830000 Ephydridac 850000 Muscidae

APPENDIX 3.3 A full list of the families recorded in one or more samples held in the macro-invertebrate data-base.

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	İ						8		<u>,</u>		ا ۔ ا		İ	_	동						TH32 - Mapleduthum
	THIO - St Johns	8	Ę	ĕ	ķ	E O	TH16 - Northmoos	t)	THI8 · Eynsham	٠	TH20 - Godstow	ৡ	اير	- Sundford	TH24 - Abingdon	tham.	from	ر. ۶	ž,	, <u>12</u>	ş
	S X	THII - Buscot	THI2 - Grafton	TH13 - Radcot	THI4 - Ruchey	TH!5 - Shifford	Z	TH17 - Pinkhill	Eyn	TH19 - King's	S	TH21 - Osney	· 1ffkg	S	¥	TH25 - Cultum	TH26 · Clifton	TH27 - Day's	TH29 - Cheeve	TH30 - Goring	ž
	ė	<u> </u>	2	ė	ż	☆	9.	12	è	<u>à</u>	20.	17	TH22	TH23	ż	B,	92	12	62	용	2
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Spongillidae		<u>_</u>		*	+	+	+		+			+	+			 	-,-	<u> </u>	<u> </u>		
Hydridae		ļ		_	ļ	<u> </u>						+	+	-	\vdash	 	+	+			<u> </u>
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Dugesiidse Dendroeoelidse		┢	\vdash		 	- -			Ť			+	+	+		-	-			Ť	Η
Planariidae (incl. Dugestidae)	 	-			 -		_								+						
Tetrastemmatidae						+				i										ļ	L_
Nematoda	+	+	+	+	+	+	*		+		+	*	*	+	+	•	+				<u> </u>
Ecroprocts		+			<u> </u>	+	*	_	+	1	<u> </u>	*	+	•	+	+	+	+	-	H	<u> </u>
Neritidac	+		+		+	+	*		+	+	+	+	+		+	+	+	+		+	_ *
Vivipandae Valvatidae	+	+	 	-	+	+	+	\vdash	 -	+	١÷	+	+	+	+	+	+	+	+	+	*
Hydrobiidse	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+		+	
Bithyniidae	+	+	+	<u>.</u>	+	+	*	+	+	+	+	+	+	+	+	+	+	+	ļ	+	
Hydrobiidae (incl. Bithyniidae)	+			L	+	+			+	+	<u> </u>	+	+	+	ļ	+	+	<u> </u>	+	+	+
Physidac		*	ļ	ļ	+	+	*	<u> </u>	<u> </u>	+		+	+	+	<u>+</u>	++	+	+	<u> </u>	+	+
Lymnaeidae	+	+	ļ		+	+	+	_	+	+	+	+	+	+	+	+	+	+	+	+	+
Planorbidae	+	+	+		+	+	+	 	+	+	+	Ť	+	+	+	+	+	+	Ť	 	+
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Unionidae	+	+	٠		+	+	Ţ÷.	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Sphaeriidae	+	+	<u> + </u>	<u> </u>	+	+	<u>+</u>	+	+	+	+	+	+	+	+	•	+	+	+	+	+
Oligochaeta	+	-	!	_	+	+	<u> </u>	+	+	+	<u> </u>	+	+	+	÷	*	+		+	+	+
Lumbriculidae	+	-	 +	۰	+	+	+	ļ÷	7	+	+	+	+	+	 	1-	Ť	1	!	Ť	Ť
Enchytraeidae Naididae	+	1	+	- -	-	+	-	+		+			+		⇈	一	+	+	†	1	+
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Lumbrieidae		+					+	+	+			+	+	+	+	+			L		
Piscicolidae	+		i		<u> </u>	L.	+	_	+	+	+	+	*	+	+	+		+	+	<u> +</u>	+
Glossiphoniidae	+	+	ļ	+	 	+	*	+	+	+	+	+	+.	+	+	+	+	+	+	+	+
Erpobdellidae	+	+	+		 	+	+	_	+	+	+	+	+	+	-	┿-	-	+	╁╼	+	. .
Hydracarina Eylaidac	+-	1	} -	⊢	 	 -	1-7-		Ť	+		\vdash	+	H	1	+	 ~~~	 	 	 	
Hydrodromidae	+	 	†	† –	t -	1	+	+	<u> </u>	+	-	+				-				†	
Lebertiidae	+	+	+		+	+	+	+	+	+	+	+	+			+		+			
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Limnesiidac		<u>Ļ</u> .	ļ	ļ	+	+	<u> </u>	+	+	+		+	+	+	+	+	*	+	-	ļ.—.	ļ
Hygrobatidae	<u> </u>	+	+	+	+	+	<u>+</u>	+ .	+	+	+	+	+	+	+	+	+	+	├ -	ļ·	
Unionicolidae Pionidae	+	⊢	 		+	+	├ ─	+	+	†	+	+	+	+	+	1	T	-	┼┈	-	\vdash
Mideoptidae	+	+	+	+	1	+	4	+	+	+	. +	+	+	 	+	╁	+	+	\vdash	 -	†
Acrenuridae	1		1			1-	T.			1	1	+	+		+		+			1	
Ostracoda	+		Ι			+			+			+	+	+		+	+	_	ļ	L	+
Argulidae	+	١		<u> </u>	<u> </u>	+	<u> </u>	<u> </u>	!	<u> </u>	1	+	<u> +</u>	+	\ *	}	}_	}	}	ļ	1
Asellidae	+	+	+	+	+	+	+	+	*	++	+	+	+	+	+	+	<u> </u>	+	+	+	+
Corophidae Craegosyctidae	+	+		+	+	+	+	+	+	+	+	+	+	+	╅	+	1	+	 . * .	+	Ť
Gammaridae	+	1	+	 	+ +	 	7	Ť	·	+	Ť	+	÷	╁	1	1 +	+	Ť	† 	+	\vdash
Gammaridae (incl. Crangonyctidae & Niphargidae)	+	\top	† 		+	+		İ	+	+	ļ	+	+	+		+	+	L	+	+	+
Bactidac	+	٠	+	÷	+	1.±	+	<u>+</u>	+	+	+	+	+	+	+	+	+	+	+	+	+
Heptageniidae		١	<u> </u>	Ĺ	J	ļ	ļ	ļ	ļ	↓	_	_	+	 _		∔	ļ	+	<u> </u>	₩	+
Leptophlebiidae	+	↓ .	 	 	ļ	 	ļ <u>.</u>	┼	┝÷	+	├.	 	 	 	+	+	+	+	+	+	+
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Caenidae		 -	+	+	+	┤╌	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Tecnioptorygidac			<u> </u>	1	İ	<u> </u>	1		Ι.	+	1.			1							
Nemouridae] .]	1	1	+				J		<u> </u>		Ť			ļ <u>.</u>		<u></u>	ļ	Ţ
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Perlodidae			↓	 	 -	╄-	 	 —	+-		↓ _	<u>+</u>	ļ. <u> </u>	↓ -	 	 _	+		-	 	
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Velidae	<u> +</u>	ļ <u>.</u>	ļ	 -	ļ	 -	ļ		<u> </u>	↓	<u> </u>	ļ			 	-	!	<u> </u>	ļ	+	ļ.
Gerridae	<u> </u>	ļ	ļ <u>-</u> -	<u> </u>	ļ <u>.</u> .	<u> +</u> .	Ļ	ļ	+	+	ļ	_	+	+	+	+	+	<u> </u>	ļ	+	ļ.,
Nepidae	+	<u> </u>			l	L.,	L			+			+	+	+	+	+	<u> </u>	1		4-
Naucoridae	<u> </u>	L	<u>L</u>	L		L				<u>L</u> .	L			<u> </u>	<u> </u>	_	<u> </u>	ļ	ļ	1	Ŀ
Aphelocheiridae	i	+	i				+			+		+		ļ	+		ĺ			+	ļ
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Corixidae	+	+	+	+	+	+		+	+	+	Ī	+	+	+	+	+	+	+	+	+	1
Halipfidae		1	•	+	•	+	+	М	+	+	+	+	+	+	+	+	+	+	+	+	٦,
Nateridae	+-	<u> </u>	<u></u> -		<u> </u>	Ė	Ė		Ė	Ė		1	 	+	+	+	Ė	m	Ė	T .	1
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Gyrinidae	•	+	<u> </u>		+-	Ť	+	É	Ť	4	Ť	+	 	+		<u> </u>	+	+	ļ		ļ
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Hydrophilidae	•	<u> </u> •	. ÷.	*		*	<u> </u>	+	+		<u> </u>		_*_	*	+		+	*	·	 	ł
Hydrophilidae (incl. Hydraenidae)	<u> </u>	ļ			 	+	—	<u> </u>	<u> </u>	<u>+</u> .	L		-		ļ	+	⊢	ļ	├—	ļ ·	1
Dryopidae		<u> </u>	<u> </u>	<u> </u>	 	ļ	ļ		+	-	L		L,	- -	ļ	-	┼	-	-	ļ	١.,
Elmidae	+	+	+			+	+		+	+		+	+	+	+	+	<u> +</u>	; +		<u>+</u>	1. 1
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Bialidae	+	+	.+	+	+	+.	+	+	+	+	+	+	+	+	+	+	+	+	+	+	Ŀ
Sisyridae]			+		+	+						Ĺ	Ì			1	L.
Hydroptilidae	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	Γ.
Rhyacophilidae (incl. Glossosomatidae)		ļ			1					1-		+				i		!			[
Psychomyiidae	+			_	 	+	+1	+	+		+		+	+	+	+	+	+			Г
Ecnomidae	 	 			 		 	+		1 -				 		ļ —	+	+	!		Г
Polycentropodidae	+	+	+	+	-		+	+	+	+		1	•	+	+	+	+	+	+	+	t٦
Hydropsychidae	+	 ` -	 - -	Ť	 	+	⊢ -	-	·		-	•	Ť	+	 	+	+	H	Ė		一
		┼─		_	┼	┷.	ļ		-					•	1	<u> </u>	+	 -	+		┢
Psychomyiidae (incl. Ecnomidae)	+		 	├	 	 	 		+	+		+	+		 -		+	 -	, 	+	ŀ
Phryganeidae	+	+	 	*	┧	+	+	+	+	+		+	+	+	+	+	+	+	+	+	ļ
Brachycentridae	-	ļ		_	ļ	_			_	+	<u> </u>	+	_	_	ļ	—	<u> </u>	+		+	<u>ļ </u>
Lepidostomatidae	+	i	ļ	 	ļ	 	ļ	Щ		ļ			ļ	1	L		<u> </u>	⊢	ļ		ļ
Limsephilidae	+	+	L		+	+	+	L	+	+	+	+	+	+	+	+	+	+	+	<u> + </u>	نا
Goeridae	+	+			<u> </u>	+	<u>L.</u>	L	+	+				+	+	<u> </u>	+		<u> </u>	<u> </u>	L
Sericostomatidac							L				L.,	Ĺ					_			+	L
Molannidae	+	+	+		+	+	+		+	+		+	+	+	+	+	+	+	+	+	-
Leptoceridae	+	+	+	+	1	+	+	+	+	+		+	+	+	+	+	+	+	+	+	
Diptera	+	+	+	+	1	+	+	+	+	+ :			+		+	ļ	+				Γ
Tipulidae	+	Ė	<u> </u>		+	1	+	\Box	+	+		\Box	+	+	+	١.	+			+	Γ
Tipelinae	1				t	 -	+	\vdash		1			+			l		+	Ι	ļ	Ĭ
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Simplifidae				 - 	 	+	 			+		_	+	+		+	+	Ť	\vdash	+	
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Chirosomidae	+	-			†	+	 	H	+	+	+	+	+	_	+	+	-		+	+	ļ
Tanypodinac	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+		+	١.
Diamesinae	+	_	4	├ ─	ļ	<u> </u>	+	+	 	<u> </u>	<u> </u>	+	+		<u> </u>	_	+	+	₩	+	-
Prodiamesinae	+	+	+	+	+	+	<u>+</u>	+	+	+		+	+	+	ļ	+	+	<u> </u>	↓ _	+	L
Orthocladiinae	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	ļ	+	L.
Chironomini	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+		+	L
Tunytarsini	+	+	+		1	+	+	+	+	+	+	+	+			+	+	+		+	Ĺ
Tabanidae			T		Ţ <u>.</u>		+		I	Γ.			_		Ī	+]		Γ
Empididae	1		i	1	1				Г				+	+		1	Ţ		1	· · · · ·	ſ
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APPENDIX 3.4 The Biological Monitoring Working Party (BMWP) index values of all samples held in the macro-invertebrate data-base.

Reach			-			Sample	Season	Season Sample			Number
	Reach name	Data source name	Year	Site ID	Site ID Site name	date	Ω	А	BMWP	ASPT	of Taxa
TH10 St	St. Johns	FBA-1977 survey	1977	31	St. Johns	03/08/77	2	+	162	5 400	30
		FBA-RIVPACS 1984 survey	1984	143	143 Malthouse	09/05/84		12	158	\$ 852	27
		FBA-RIVPACS 1984 survey	1984	143	143 Malthouse	09/05/84		7001	119	5.174	23
		FBA-RIVPACS 1984 survey	1984	143	143 Malthouse	26/07/84	2	┖	118	4.917	74
-		FBA-RIVPACS 1984 survey	1984	143	143 Malthouse	26/07/84	2	ļ.,	128	5.565	23
		FBA-RIVPACS 1984 survey	1984	143	143 Malthouse	17/10/84			116	5.273	22
		FBA-RIVPACS 1984 survey	1984	143	143 Maithouse	17/10/84	(n)		126	5.727	22
		Les Ruse's Chironomid exuviae	1977	127	127 Buscot		∞		2	2.000	
+	-	Les Ruse's Chironomid exuviae	1978	127	127 Buscot		000	6	2	2.000	
		Les Ruse's Chironomid exuviae	1986	127	127 Buscot		·	2	7	2.000	1
		Les Ruse's Chironomid exuviae	1992	127	127 Buscot	08/02/92		1	2	2.000	-
+		Les Ruse's Chironomid exuviae	1992	127	127 Buscot	24/06/92	2	1	2	2.000	
		Les Ruse's Chironomid exuviae	1992	127	127 Buscot	24/08/92	2	2	2	2.000	
		Thames WA/NRA/EA routine monitoring	1980	38	38 A417 Buscot	22/10/80	3	300	126	5.040	25
+		Thames WA/NRA/EA routine monitoring	1987	38	38 A417 Buscot	10/07/87	7	3500	121	4.840	25
		I hames WA/NRA/EA routine monitoring	1987	39	39 Water Intake Buscot	03/02/87	-	2100	112	5.091	22
+		Thames WA/NRA/EA routine monitoring	1988	39	39 Water Intake Buscot	01/08/88	7	4900	146	4.867	30
-		Inames WA/NRA/EA routine monitoring	1989	39	39 Water Intake Buscot	01/06/89	1	5400	103	5.150	20
+		Thames WA/NRA/EA routine monitoring	1990	39	39 Water Intake Buscot	24/04/90	3	6400	114	4.957	23
+		Thames WA/NRA/EA routine monitoring	1990	39	39 Water Intake Buscot	13/02//90	2	7200	114	4.560	25
+		Thames WA/NRA/EA routine monitoring	1990	39	39 Water Intake Buscot	06/60/50	2	7800	153	5.100	30
+		I hames WA/NRA/EA routine monitoring	1991	99	39 Water Intake Buscot	08/04/91	1	00/8	102	5,100	20
		I hames WA/NRA/EA routine monitoring	1991	39	39 Water Intake Buscot	05/08/91	2	9400	103	4.682	22
+		Thames WA/NRA/EA routine monitoring	1991	39	39 Water Intake Buscot	14/11/91		0086	86	4.667	21
$\frac{1}{1}$		I hames WA/NRA/EA routine monitoring	1992	39	39 Water Intake Buscot	13/07/92	2	10400	112	4.667	24
-		Thames WA/NRA/EA routine monitoring	1992	39	39 Water Intake Buscot	07/10/92	3	10800	132	5.077	26
1		I names WA/NRA/EA routine monitoring	1993	3	39 Water Intake Buscot	29/04/93	1	11300	122	4.880	25
1		I names WA/NKA/EA routine monitoring	1994	39	39 Water Intake Buscot	09/08/94	2	12100	120	5.714	21
-		I hames WA/NRA/EA routine monitoring	1995	33	39 Water Intake Buscot	04/05/95	1	13800	113	5.136	22
		I hames WA/NRA/EA routine monitoring	1995	8	39 Water Intake Buscot	14/11/95	3	16500	129	4.778	27

Reach					Sample	Season Sample	Sample			Number
£	Reach name	Data source name	Year Site I	Site ID Site name	date	А	, 日	BMWP	ASPT	of Taxa
TH11	Buscot	FBA-1977 survey		30 Buscot	03/08/77	2	17	96	4.174	23
		FBA-1977 survey		37 Buscot	15/08/77	2	1190	43	4.300	10
		FBA-1977 survey		37 Buscot	15/08/77	2	1191	09	4.615	13
		FBA-1977 survey		37 Buscot	15/08/77	2	1192	37	3.364	11
		FBA-1977 survey		37 Buscot	15/08/77	2	1193	99	5.077	13
		FBA-1977 survey		37 Buscot	15/08/77	2	1194	34	3.400	10
		FBA-1977 survey		37 Buscot	15/08/77	2	1195	40	3.636	=
		FBA-1977 survey		37 Buscot	15/08/77	2	1196	17	3.400	2
		FBA-1977 survey		37 Buscot	15/08/77	2	1197	58	4.833	12
		FBA-1977 survey		37 Buscot	15/08/77	2	1198	28	3.500	8
		FBA-1977 survey		37 Buscot	15/08/77	2	1199	39	4,333	6
		FBA-1977 survey		37 Buscot	15/08/77	2	2190	30	4.286	7
		FBA-1977 survey	1977	37 Buscot	15/08/77	2	2191	29	4.143	7
		FBA-1977 survey		37 Buscot	15/08/77	2	2192	11	2.750	4
	-	FBA-1977 survey		37 Buscot	15/08/77	2	2193	29	5.800	3
		FBA-1977 survey		37 Buscot	15/08/77	2	2194	43	4.778	6
		FBA-1977 survey		37 Buscot	15/08/77	2	2195	17	4.250	4
		FBA-1977 survey		37 Buscot	15/08/77	2	2196	46	4.182	11
		FBA-1977 survey		37 Buscot	15/08/77	2	2197	27	4.500	9
		FBA-1977 survey		37 Buscot	15/08/77	2	2198	35	\$.000	-
		FBA-1977 survey		37 Buscot	15/08/77	2	2199	13	3.250	4
		FBA-1977 survey		37 Buscot	15/08/77	2	3190	18	3.000	9
Ī		FBA-1977 survey		37 Buscot	15/08/77	2	3191	45	4.500	10
		FBA-1977 survey		37 Buscot	15/08/77	7	3192	56	5.091	11
		FBA-1977 survey		37 Buscot	15/08/77	2	3193	16	3.200	35
		FBA-1977 survey		37 Buscot	15/08/77	2	3194	21	3.000	7
		FBA-1977 survey		37 Buscot	15/08/77	2	3195	47	3.917	12
[FBA-1977 survey		37 Buscot	15/08/77	2	3196	83	5.188	16
		FBA-1977 survey		37 Buscot	15/08/77	2	3197	20	4.167	12
		FBA-1977 survey		37 Buscot	15/08/77	7	3198	28	3.500	8
		FBA-1977 survey	1977	37 Buscot	15/08/77	2	3199	13	4.333	3

Reach				Sample	Season Sample	ample			Number
<u>e</u>	Reach name	Data source name	Year Site ID Site name	date	<u>П</u>	·Ω	BMWP	ASPT	ASPT of Taxa
TH12	Grafton	FBA-1977 survey	1977 29 Grafton	03/08/77	2	91	123	5.125	24

Reach				Sample Sea	Sample	Season	Sample		-	Number
2	Reach name	Data source name	Year S	Site ID Site name	date	0	<u>e</u>	BMWP	ASPT of	of Taxa
TH13	TH13 Radcot	FBA-1977 survey	1977	7 28 Radcot	02/08/77	2	15	101	5.050	20

Reach					Sample	Season Sa	Sample	•		Number
2	Reach name	Data source name	Year Si	Site ID Site name	date	Ð		BMWP	ASPT	of Taxa
TH14	TH14 Rushey	FBA-1977 survey	1977	27 Rushey	72/08/77	2	14	001	4.762	21
		Thames WA/NRA/EA routine monitoring	1995	40 0.5km below Tadpole Brid	10/08/95	2	14900	109	4.739	23

Reach					Sample	Season	Season Sample	-		Number
<u>a</u>	Reach name	Data source name	Year	Site ID Site name	date	Ω	<u>.</u>	BMWP	ASPT	of Taxa
THIS	Shifford	FBA-1977 survey	1977	26 Shifford	02/08/77	2	13	136	4.857	28
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1200	46	4.182	=
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1071	36	5.091	
:		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1202	77	4.529	17
	i	FBA-1977 survey	1977	36 Shifford	77/08/77	2	1203	85	4.474	19
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1204	48	3.692	13
 	!	FBA-1977 survey	1977	36 Shifford	22/08/77	2	1205	37	4.111	6
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1206	78.	4.105	19
	:	FBA-1977 survey	1977	36 Shifford	22/08/77	2	1207	30	3.750	8
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1208	26	3.714	7
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	1209	32	3.556	6
:		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2200	55	5.000	11
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2201	18	3.600	S
- - - -		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2202	29	4.833	9
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2203	36	4.500	∞
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2204	20	4.000	5
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2205	37	4.625	80
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2206	88	5.867	15
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2207	29	4.833	9
-		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2208	6	3.000	6
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	2209	20	5,000	4
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3200	18	2.571	7
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3201	25	3.571	7
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3202	12	3.000	4
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3203	93	4.650	20
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3204	20	3.333	9
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3205	28	4.000	7
-+-		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3206	26	4,333	9
	 	FBA-1977 survey	1977	36 Shifford	22/08/77	2	3207	31	3.875	8
:		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3208	52	4.727	
1		FBA-1977 survey	1977	36 Shifford	22/08/77	2	3209	22	3.667.	9
		FBA-1977 survey	1977	36 Shifford	22/08/77	2	4201	12	3.000	7

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Keach					Sample	Season	Season Sample			Number
<u>e</u>	Reach name	Reach name Data source name	Year	Site ID Site name		5		China Co	1. Tyti	T. T. C.
				מונס דו מונה	חשונ	3	3	DIVI W	1 100	OI 1 dAd
-H]?	Shifford	FBA-1977 survey	7761	36 Shifford	22/08/77	2	4202	29	3 625	ő
						1		, t	2	,
1		FBA-1977 Survey	1977	36 Shifford	22/08/77	2	4203	29	4 833	9
										3
		FBA-1977 survey	1977	36 Shifford	22/08/77	0	4204	24	3365	-
			П			*	2	3	201.0	_
:		Thames WA/NRA/EA routine monitoring	1987	41 Newbridge	10/07/87	2	3600	103	5 150	20
_		The state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the s	1			í	2		,	1
		I hames w A/NKA/EA routine monitoring	1995	41 Newbridge	17/05/95	-	14400	10	4 780	10
			J.,	İ				. ,		-
		I names WA/NKA/EA routine monitoring	1995	41 Newbridge	02/10/95	~	15400	122	4 880	38
						,				1
		I hames WA/NRA/EA routine monitoring	1995	42 US Newbridge	10/08/95	2	15000	122	4 880	36
				l		ī	,	111	>	1

Reach			-		Comple	Conco	Concor Complet			Minh
<u>e</u>	Reach name	Reach name Data source name	Year	Year Site ID Site name		Scasour T	Sample	D) (1) D	YOU	Double Taxo
TH16	Northmoor	FBA-1977 survey	1977	25 Northmoor	100/17	2 -	3 5	TALIAN I	1 100	01 1 a A a
[,	FRA DIVIDA CC 1004		Commical	01/00/1/	7	77	101	3.407	67
1		LDA-N VFACS 1984 SUIVEY	1,784	144 Bablock Hythe	09/05/84	_	7007	75	4.412	17
!		FBA-RIVPACS 1984 survey	1984	144 Bablock Hythe	09/05/84	П	7008	114	5.429	21
		FBA-RIVPACS 1984 survey	1984	144 Bablock Hythe	26/07/84	2	7009	79	4 158	10
		FBA-RIVPACS 1984 survey	1984	144 Bablock Hythe	26/07/84	,	7010	13.	\$ 240	25
		FBA-RIVPACS 1984 survey	1984	144 Bablock Hythe	17/10/84	1 (*	7011	121	5.261	2 2
		FBA-RIVPACS 1984 survey	1984	144 Bablock Hythe	17/10/84		7012	136	\$ 440	25
		Les Ruse's Chironomid exuviae	1977	131 Newbridge	3	00		ST C	2 000	-
		Les Ruse's Chironomid exuviae	1978	131 Newbridge		000	1 20	1 (2 000	-
:		Les Ruse's Chironomid exuviae	1986	131 Newbridge		00	, -	,	2 000	-
-		Les Ruse's Chironomid exuviae	1992	131 Newbridge	08/05/92	•	: -	2 2	2 000	
; ;		Les Ruse's Chironomid exuviae	1992	131 Newbridge	24/06/92	2	7	2	2.000	7=
	-	Les Ruse's Chironomid exuviae	1992	131 Newbridge	24/08/92	2	-	2	2.000	
	,	Les Ruse's Chironomid exuviae	1994	131 Newbridge	12/05/94	-	7	2	2,000	T-
:		Les Ruse's Chironomid exuviae	1994	131 Newbridge	13/06/94	2	E	2	2,000	
		Les Ruse's Chironomid exuviae	1994	131 Newbridge	12/09/94	3	-	2	2.000	

Reach			- -		Sample	Season	Season Sample			Number
e	Reach name	Reach name Data source name	Year	Site 1D Site name	date	8	£	BMWP	ASPT	of Taxa
TH17	Pinkhill	FBA-1977 survey	1977	24 Pinkhill	01/08/77	2	11	120	5.217	23
		Les Ruse's Chironomid exuviae	1677	130 Swinford		8	1	2	2.000	
		Les Ruse's Chironomid exuviae	1978	130 Swinford		90	3	2	2.000	
<u> </u>		Les Ruse's Chironomid exuviae	1986	130 Swinford		∞	2	2	2.000	=
	 - - - - - - - -	Les Ruse's Chironomid exuviae	1992	130 Swinford	08/05/92	_	3	2	2.000	
 :	i i	Les Ruse's Chironomid exuviae	1992	130 Swinford	24/06/92	2	1	2	2.000	
	!	Les Ruse's Chironomid exuviae	1992	130 Swinford	24/08/92	2	2	2	2.000	-
	 	Les Ruse's Chironomid exuviae	1994	130 Swinford	12/05/94	_	-	2	2.000	
		Les Ruse's Chironomid exuviae	1994	130 Swinford	13/06/94	2	т	2	2.000	_
		Les Ruse's Chironomid exuviae	1994	130 Swinford	12/09/94	3	1	2	2.000	1

Reach					Sample	Scason	Season Sample			Number
<u>e</u>	Reach name	Data source name	Year	Site ID Site name	date	Ω	Ω	BMWP	ASPT	of Taxa
TH18		FBA-1977 survey	1977	23 Eynsham	01/08/77	2	01	132	5.077	26
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1210	31	3.875	∞
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1211	55	3.667	1.5
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1212	39	4.333	6
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1213	43	4.300	10
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1214	45	5.000	<u>م</u>
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1215	28	3.500	30
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1216	41	4.100	2
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1217	42	4.200	0.5
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1218	75	4.412	17
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	1219	52	4.333	12
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2210	28	4.000	7
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2211	78	5.200	15
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2212	91	4.000	त्त
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2213	24	3.429	
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2214	30	4.286	7
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2215	34	4.250	∞
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2216	37	4.625	30
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2217	26	4.333	9
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2218	36	4.000	5
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	2219	31	3.444	6
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3210	44	4.889	9
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3211	32	4.000	∞
		FBA-1977 survey	1977	35 Eynsham	24/08/77	7	3212	24	4.000	9
		FBA-1977 survey	1977	35 Eynsham	24/08/77		3213	25	3.571	7
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3214	32	4.000	
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3215	9	2.000	<u>m</u>
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3216	54	4.154	2
		FBA-1977 survey	1977	35 Eynsham	24/08/77	2	3217	3	1.500	2

Reach						Sample	Season Sample	Sample			Number
	Reach name	Reach name Data source name	Year	Site ID	Site ID Site name	date	В	0	BMWP	ASPT	of Taxa
TH18	Eynsham	FBA-1977 survey	1977	35	35 Eynsham	24/08/77	2	3218	20	3.333	9
		FBA-1977 survey	1977	35	35 Eynsham	24/08/77	2	3219	39	4.333	6
		Oxford Structures Environmental Survey	1992	118	118 DS Swinford WTW	30/06/92	2	2000	132	5.077	26
		Thames WA/NRA/EA routine monitoring	1980	43	43 Water Intake, Swinford	08/11/90	3	400	121	5,261	23
		Thames WA/NRA/EA routine monitoring	1987	43	43 Water Intake, Swinford	17/02/87	1	2200	115	4.600	25
		Thames WA/NRA/EA routine monitoring	1987		43 Water Intake, Swinford	24/07/87	2	3900	124	4.593	27
		Thames WA/NRA/EA routine monitoring	1988		43 Water Intake, Swinford	15/06/88	2	4800	174	5.800	30
		Thames WA/NRA/EA routine monitoring	1989		43 Water Intake, Swinford	10/10/89	3	6100	111	4.826	23
		Thames WA/NRA/EA routine monitoring	1995		43 Water Intake, Swinford	10/04/95	1	13200	117	4.875	24
		Thames WA/NRA/EA routine monitoring	1995	43	43 Water Intake, Swinford	09/10/95	3	15500	83	4.611	18

Reach			_		Sample Sc	Season Sample	umple	<u> </u>		Number
<u>a</u>	Reach name	Data source name	Year Site	Year Site ID Site name	date	D D		BMWP	ASPT	of Taxa
TH19	King's	FBA-1977 survey	1977	22 King's	25/08/77	2	6	124	4.960	25
		Oxford Structures Environmental Survey	1992	44 Trout Inn, Godstow	30/06/92	2	5002	168	5.250	32
		Oxford Structures Environmental Survey	1992 1	119 West Mead	30/06/92	2	5001	66	4.714	21
	_	Thames WA/NRA/EA routine monitoring	1980	44 Trout Inn, Godstow	01/12/80	3	800	128	5.120	25
<u> </u>		Thames WA/NRA/EA routine monitoring	1990	44 Trout Inn, Godstow	30/04/90	1	6500	107	5.350	20
		Thames WA/NRA/EA routine monitoring	1990	44 Trout Inn, Godstow	06/80/60	7	0092	181	5.171	35
}		Thames WA/NRA/EA routine monitoring	1990	44 Trout Inn, Godstow	06/01/80	3	8000	139	5.148	27
		Thames WA/NRA/EA routine monitoring	1661	44 Trout Inn, Godstow	08/04/91	1	8600	100	4.545	22
		Thames WA/NRA/EA routine monitoring	1991	44 Trout Inn, Godstow	03/07/91	2	0016	93	4.429	21
		Thames WA/NRA/EA routine monitoring	1661	44 Trout Inn, Godstow	16/60/61	2	9500	153	6.120	25
		Thames WA/NRA/EA routine monitoring	1992	44 Trout Inn, Godstow	26/02/92	1 1	10100	135	5,000	27
	-	Thames WA/NRA/EA routine monitoring	1992	44 Trout Inn, Godstow	28/07/92	2 1	10500	115	6.765	17
		Thames WA/NRA/EA routine monitoring	1993	44 Trout Inn, Godstow	24/02/93	1 1	11200	136	2.667	24
		Thames WA/NRA/EA routine monitoring	1994	44 Trout Inn, Godstow	28/07/94	2 1	12000	168	6.462	26
		Thames WA/NRA/EA routine monitoring	1995	44 Trout Inn, Godstow	01/05/95	1 1	13700	148	5.481	27
		Thames WA/NRA/EA routine monitoring	1995	44 Trout Inn, Godstow	30/10/95	3 1	3 16100	122	5.545	22

Reach			!		Sample	Season	ason Sample			Number
<u>e</u>	Reach name D	Reach name Data source name	Year	Site ID Site name	date	В		BMWP	ASPT	ASPT of Taxa
TH20	Godstow	FBA-1977 survey	1977	21 Godstow	28/07/77	2	∞	113	4.913	23
		Oxford Structures Environmental Survey	1992	121 Binsey	30/06/92	2	5003	111	4,625	24

Reach					Sample	Season	Season Sample	 		Number
≘	Reach name	Reach name Data source name	Year Si	Year Site ID Site name	date	0	Д	BMWP	ASPT	of Taxa
TH21	Osney	FBA-1977 survey	1977	20 Osney	77/177	2	7	148	4.933	30
:		Oxford Structures Environmental Survey	1992	122 DS Cherwell	07/06/92	2	5004	108	5.143	21
		Thames WA/NRA/EA routine monitoring	1987	45 Osney Weir	22/03/87		2600	87	4.833	18
		Thames WA/NRA/EA routine monitoring	8861	45 Osney Weir	13/06/88		4500	148	5.692	26
: : :	•	Thames WA/NRA/EA routine monitoring	1989	45 Osney Weir	28/06/89	2	5800	152	5.429	28
	(Thames WA/NRA/EA routine monitoring	1995	46 Donnington Bridge, Oxfor	10/04/95	-	13100	138	5.750	24
		Thames WA/NRA/EA routine monitoring	1995	46 Donnington Bridge, Oxfor	02/11/95	3	16800	45	4.091	=
-		Thames WA/NRA/EA routine monitoring	1995	47 Folley Bridge, Oxford	23/05/95	_	14500	122	5.304	23
		Thames WA/NRA/EA routine monitoring	1995	47 Folley Bridge, Oxford	30/10/95	3	16000	89	4.857	14

Reach					Sample S.	Season Sample	ple] j	Number
<u>e</u>	Reach name	Data source name	Year Site	Year Site ID Site name	date	<u>Ω</u>	Δ.	BMWP	ASPT	of Taxa
TH22	Iffley	FBA-1977 survey	1977	19 मिल	72/1777	2	9	176	5.176	34
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1220	6	3,000	3
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1221	91	4.789	61
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1222	61	3.167	9
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1223	26	4.333	9
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1224	39	3.900	2
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1225	48	4.000	12
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1226	45	3.750	12
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 12	1227	30	3.750	8
		FBA-1977 survey		34 Iffley	30/08/77	2 12	1228	70	4.375	16
		FBA-1977 survey		34 Iffley	30/08/77	2 12	1229	53	4.077	13
		FBA-1977 survey		34 Ifficy	30/08/77	2 22	2220	35	3.889	6
		FBA-1977 survey	1977	34 [Iffley	30/08/77	2 22	2221	09	5,000	12
		FBA-1977 survey	1977	34 [Iffley	30/08/77	2 22	2222	61	3.800	S
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 22	2223	25	3.571	7
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 22	2224	57	5.182	11
		FBA-1977 survey	1977	34 iffley	30/08/77	2 22	2225	04	3.636	1
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 22	2226	19	3.167	9
		FBA-1977 survey	1977	34 Iffley	30/08/77	L	2227	24	3,429	7
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 22	2228	39	3.900	01
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 22	2229	28	4.462	13
		FBA-1977 survey		34 Iffley	30/08/77	2 32	3220	35	4.375	8
		FBA-1977 survey		34 Iffley	30/08/77	2 32	3221	48	4.364	11
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3222	58	4.833	12
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3223	35	3.500	10
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3224	51	4.636	11
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3225	55	4.583	12
 		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3226	65	4.643	7
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3227	29	4.143	7
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3228	41	4.556	9
		FBA-1977 survey	1977	34 Iffley	30/08/77	2 32	3229	92	4.600	20
		Oxford Structures Environmental Survey	1992	48 Top of Sandford Lock Cut	30/06/92	2 50	5005	108	4.696	23

Reach							-			
Ē	Roach name	Rough name Date courses assess			Sample	Season Sample	nple			Number
TITOS	TO THE HEALT	Data source name	Year	Site ID Site name	date	<u>a</u>	<u>щ</u>	BMWP	ASPT	of Taxa
7781	uriey	Pond Action SWORDS Survey	1992	48 Top of Sandford Lock Cut	09/02/92	2.	-	203	\$ 343	30
		Thames WA/NRA/EA routine monitoring	1982	48 Top of Sandford Lock Cut	18/02/82	\perp	1000	120	2,000	200
•		Thames WA/NRA/EA routine monitoring	1984	48 Top of Sandford Lock Cut	09/03/84	-	1400	114	5 273	4 6
		Thames WA/NRA/EA routine monitoring	1987	48 Top of Sandford Lock Out	22/03/87	, -	2000	123	07.2.2	77
- _		Thames WA/NRA/EA routine monitoring	1988	48 Ton of Sandfard 1 cot Out	10/00/27	1.	36	CCT	3.320	C7
] ! !		Thames WA/NRA/FA fouring monitoring	1000	AO TO SAMMOND LOCK CUI	13/00/88	7	4600	156	5.200	2
		The state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the state of the s	1707	46 10p of Sandrord Lock Cut	08/02/89	1 5	5700	165	5.690	29
		Inames wA/NRA/EA routine monitoring	1990	48 Top of Sandford Lock Cut	17/04/90	1	6300	143	\$ 107	28
		Thames WA/NRA/EA routine monitoring	1990	48 Top of Sandford Lock Cut	20/08/90	, ,	7700	178	4 000	27
		Thames WA/NRA/EA routine monitoring	1990	48 Top of Sandford Lock Cut	00/10/00	⅃₋	2 2	271	7.72	3 1
		Thomas WAATA ATA		יין דיין דיין דיין דיין דיין דיין דיין	06/11/20	5	006/	189	5,400	35
		Thanks WAINKAVEA foutine monitoring	1991	48 Top of Sandford Lock Cut	08/04/91		8500	107	4.652	23
		Thames WA/NRA/EA routine monitoring	1991	48 Top of Sandford Lock Cut	03/07/91	2 9	9200	115	4 702	24
		Thames WA/NRA/EA routine monitoring	1991	48 Top of Sandford Lock Cut	30/10/91	┸	9700	110	\$ 400	٤, د
:		Thames WA/NRA/EA routine monitoring	1992	48 Top of Sandford Lock Cut	26/02/92	-	10200	146	\$ 107	1 6
		Thames WA/NRA/EA routine monitoring	1992	48 Top of Sandford Lock Cut	11/08/92	2 6	10700	135	4 22 3	176
		Thames WA/NRA/EA routine monitoring	1993	48 Top of Sandford Lock Cut	24/02/93		11100	000	4.460	07
		Thames WA/NRA/EA routine monitoring	1994	48 Top of Sandford Lock Cut	08/11/94	1 2	12700	115	0000	07
				110 1100 800	17/17/00		201	611	3,000	73

Reach			•••			Sample So	eason	Season Sample			Number
A	Reach name	Reach name Data source name	Year	Site ID	Year Site ID Site name	date	B	<u>a</u>	BMWP	ASPT	of Taxa
TH23	Sandford	FBA-1977 survey	1977	18	18 Sandford	77/10/17	2	\$	93	4.429	21
		Oxford Structures Environmental Survey	1992	124	124 Radley College Boathouse	30/06/92	2	2006	80	5.000	16
		Pond Action SWORDS Survey	1992	46	49 Abingdon Weir	28/07/92	2	-	230	5.349	43
		Pond Action SWORDS Survey	1992	139	139 Near Lock Wood	28/07/92	7		171	5.344	32
		Pond Action SWORDS Survey	1992	140	140 Radley	20/07/92	2	_	193	5.361	36
		Pond Action SWORDS Survey	1992	141	141 Sandford Reach	28/07/92	2		193	5.514	35
		Thames WA/NRA/EA routine monitoring	1980	49	49 Abingdon Weir	20/11/80	3	009	122	4.692	26
	:	Thames WA/NRA/EA routine monitoring	1982	49	49 Abingdon Weir	18/02/82	1	1100	129	5.160	25
		Thames WA/NRA/EA routine monitoring	1983	49	49 Abingdon Weir	26/01/83	_	1200	118	4.720	25
	:	Thames WA/NRA/EA routine monitoring	1984	49	49 Abingdon Weir	09/03/84		1300	113	4.708	24
		Thames WA/NRA/EA routine monitoring	9861	50	50 Sandford	12/06/86	1	0061	30	4.545	11
		Thames WA/NRA/EA routine monitoring	1987	49	49 Abingdon Weir	12/03/87	1	2400	103	4.682	22
		Thames WA/NRA/EA routine monitoring	1988	49	49 Abingdon Weir	15/06/88	1	4700	136	5.037	27
		Thames WA/NRA/EA routine monitoring	1989	49	49 Abingdon Weir	68/50/80		2600	143	4.767	30
		Thames WA/NRA/EA routine monitoring	1990	49	49 Abingdon Weir	06/90/50	"	0019	119	4.577	97
		Thames WA/NRA/EA routine monitoring	1992	49	49 Abingdon Weir	23/11/92	3	11000	89	4.944	18
		Thames WA/NRA/EA routine monitoring	1993	49	49 Abingdon Weir	24/05/93	Ţ	11400	117	5.087	23
		Thames WA/NRA/EA routine monitoring	1994	49	49 Abingdon Weir	19/10/94		12400	113	5.136	22
		Thames WA/NRA/EA routine monitoring	1995	49	49 Abingdon Weir	56/50/60		14000	82	5.125	16
		Thames WA/NRA/EA routine monitoring	1995	49	49 Abingdon Weir	01/11/95	3	16200	118	4.720	25

Reach				Sample	Season	Season Sample			Number
≘	Reach name	Reach name Data source name	Year Site ID Site name	date	<u>e</u>	A	BMWP	ASPT	of Taxa
TH24	TH24 Abingdon	FBA-1977 survey	1977 17 Abingdon	71/10/172	7 2	4	132	5.077	79
		Pond Action SWORDS Survey	1992 136 Sutton Pools	03/08/92	2	1	130	5.588	34
		Pond Action SWORDS Survey	1992 137 Culham Reach	20/07/92	2	-	200	5.263	38

Reach						Sample	Season	Season Sample		 -	Number
<u>e</u>	Reach name	Reach name Data source name	Year	Site ID	Year Site ID Site name		2	2	BMWP	ASPT	of Taxa
TH25	FH25 Culham	FBA-1977 survey	1977	16	Culham	77170/92	2	3	122	5.083	24
		Pond Action SWORDS Survey	1992		134 Long Wittenham	03/08/92	2	-	187	5.054	37
		Pond Action SWORDS Survey	1992	135	135 Clifton Reach	28/07/92	2	-1	105	4.565	23
		Thames WA/NRA/EA routine monitoring	1987	52	52 Sutton Bridge, Culham	78/10/10	2	3200	13	2.600	5
		Thames WA/NRA/EA routine monitoring			52 Sutton Bridge, Culham	10/04/95	1	13300	110	5.000	22
		Thames WA/NRA/EA routine monitoring			52 Sutton Bridge, Culham	02/11/95	3	16400	91	4.550	20

Reach					Sample	Season	Season Sample			Number
9	Reach name	Data source name	Year	Site ID Site name	date	Ω	Ω	BMWP	ASPT	of Taxa
TH26	Clifton	FBA-1977 survey	1977	15 Clifton	77/777	2	2	76	4.619	21
		FBA-1977 survey	1977	33 Clifton	77/60/50	2	1230	36	4.500	20
		FBA-1977 survey	1977	33 Clifton	12/09/20	2	1231	13	2.600	5
		FBA-1977 survey	1977	33 Clifton	71/60/50	, 2	1232	33	4.125	30
	.	FBA-1977 survey	1977	33 Clifton	71/60/50	2		29	4.143	7
		FBA-1977 survey	1977	33 Clifton	71/60/50	2	1234	20	4.167	12
		FBA-1977 survey	1977	33 Clifton	71/60/50	2	1235	75	4.412	17
		FBA-1977 survey	1977	33 Clifton	77/60/50	2	1236	20	3.846	13
		FBA-1977 survey	1977	33 Clifton	72/60/50	2	1237	14	3.500	7
		FBA-1977 survey	1977	33 Clifton	72/00/50	2	1238	33	4.125	20
		FBA-1977 survey	1977	33 Clifton	05/09/77	2		27	3.857	7
		FBA-1977 survey	1977	33 Clifton	71/60/50	2	2230	I	1.000	
		FBA-1977 survey	1977	33 Clifton	77/60/50	2		21	3.500	9
		FBA-1977 survey	1977	33 Clifton	05/09/77] 2	<u> </u>	13	3.250	च
		FBA-1977 survey	1977	33 Clifton	77/60/50	1 2	2233	25	4.167	9
		FBA-1977 survey	1977	33 Clifton	77/60/50	2	2234	3	1.500	2
		FBA-1977 survey	1977	33 Clifton	77/60/50	, 2	2235	10	3.333	3
		FBA-1977 survey	1977	33 Clifton	05/09/77	2		10	3.333	3
		FBA-1977 survey	1977	33 Clifton	05/09/77	2	2237	36	5.143	7
		FBA-1977 survey	1977	33 Clifton	77/60/50	2	2238	23	4.600	Š
		FBA-1977 survey	1977	33 Clifton	71/60/50	2	2239	11	4.250	→
		FBA-1977 survey	1977	33 Clifton	05/09/77	7		14	3.500	प
		FBA-1977 survey	1977	33 Clifton	05/09/77	7		9	2.000	3
		FBA-1977 survey	1977	33 Clifton	72/09/20	7		-	1.000	1
		FBA-1977 survey	1977	33 Clifton	12/60/50	2	3233	23	4.600	\$
- 1		FBA-1977 survey	1977	33 Clifton	05/09/77	2	_	28	4.667	9
		FBA-1977 survey	1977	33 Clifton	05/09/77	2	3235	13	3.250	-+
		FBA-1977 survey	1977	33 Clifton	05/09/77	2	3236	13	3.250	7
		FBA-1977 survey	1977	33 Clifton	05/09/77	2	3237	5.	1.500	2
		FBA-1977 survey	1977	33 Clifton	05/09/77	, 2		3	1.500	2
	 	FBA-1977 survey	1977	33 Clifton	77/60/50	7	3239	25	4.167	9
		Les Ruse's Chironomid exuviae	1977	128 Clifton Hampden		8	2	2	2.000	1

Reach						Sample	Season	Season Sample			Number
2	Reach name	Reach name Data source name	Year	Site ID	Year Site ID Site name	date	Ω	<u>0</u>	BMWP	ASPT	of Taxa
TH26	Clifton	Les Ruse's Chironomid exuviae	1978	128	128 Clifton Hampden		8	6	2	2.000	
		Les Ruse's Chironomid exuviae	1986	128	128 Clifton Hampden		∞	1	2	2.000	!==
	,	Les Ruse's Chironomid exuviae	1992	128	128 Clifton Hampden	08/02/92	_	_	2	2.000	<u> </u>
	:	Les Ruse's Chironomid exuviae	1992	128	128 Clifton Hampden	24/06/92	2	-	2	2.000	-
		Les Ruse's Chironomid exuviae	1992	128	128 Clifton Hampden	24/08/92	2	2	2	2.000	<u> </u>
		Les Ruse's Chironomid exuviae	1994	128	128 Clifton Hampden	12/05/94	_	2	2	2.000	-
		Les Ruse's Chironomid exuviae	1994	128	128 Clifton Hampden	13/06/94	2	3	2	2.000	-
		Les Ruse's Chironomid exuviae	1994	128	128 Clifton Hampden	12/09/94	50	-	2	2.000	-
		Pond Action SWORDS Survey	1992	53	53 Day's Lock	03/08/92	2	-	124	5.167	24
		Pond Action SWORDS Survey	1992	132	132 Days Reach	28/07/92	2	-	131	5.038	26
		Pond Action SWORDS Survey	1992	133	133 Clifton Bridge	03/08/92	2	-	147	5.250	28
		Thames WA/NRA/EA routine monitoring	1980	51	51 Clifton Hampden Bridge	06/11/80	ω.	200	79	4.647	17
}		Thames WA/NRA/EA routine monitoring	1987	51	51 Clifton Hampden Bridge	24/07/87	2	3800	110	4.783	23
		Thames WA/NRA/EA routine monitoring	1989	53	53 Day's Lock	05/05/89	-	2500	671	5.594	32
		Thames WA/NRA/EA routine monitoring	1990	53	53 Day's Lock	14/11/90	3	8200	=	4.826	23
		Thames WA/NRA/EA routine monitoring	1995	53	53 Day's Lock	56/50/60	1	14100	801	4.909	
	- 4	Thames WA/NRA/EA routine monitoring	1995	53	53 Day's Lock	01/11/95		3 16300	81	4.500	82

2				-	Sample	Season Sample	sample			Number
Keach	1	Date action action	Year Site II	Site ID Site name		<u> </u>	0	BMWP	ASPT	of Taxa
	Keach name	Reach name Uata Source Haine	1977	14 Dav's	72/01/17	2	1	73	4.563	16
/7H [Day s	TDA-1777 Sulvey	1984 145	SiShillingford	09/05/84	-	7013	126	5.250	24
- 1		FDA-KI VEACO 1704 SULVO	4	Shillingford	09/05/84	-	7015	135	5.400	25
		FBA-FLVFACO 1704 Suivey		Chillingford	26/07/84	2	7016	107	4.458	24
		FBA-KIVPACS 1984 survey	1004	-	26/07/84	2	7017	125	5.208	24
		FBA-RIVPACS 1984 survey	- •	-	17/10/84	1 ~	7018	74	4.353	17
<u></u>		FBA-RIVPACS 1984 survey		Smillingiora	10/01/61	7 (0100	130	6 560	3,5
:	·	FBA-RIVPACS 1984 survey	1984 14	145 Shillingtord	1//10/84	7	212	701	2002:1	

Reach	-				Sample Sa	Season Sample	Sample			Number
	Reach name	Reach name Data source name	Year Site	Year Site ID Site name	date		Б	BMWP	ASPT	of Taxa
TH29	TH29 Cleeve	Thames WA/NRA/EA routine monitoring	1995	54 South Stoke	27/04/95	1	1 13600	93	4.650	20
		Thames WA/NRA/EA routine monitoring	1995	54 South Stoke	13/09/95	2	15300	100	4.762	21
		Thames WA/NRA/EA routine monitoring	5661	56 US Goring Weir	27/04/95	-	13500	105	5.250	20
		Thames WA/NRA/EA routine monitoring	1995	56 US Goring Weir	13/09/95	2	2 15200	112	5.091	22

Reach						Sample	Season	Season Sample			Number
9	Reach name	Data source name	Year	Site ID	Site ID Site name	date	Ω		BMWP	ASPT	of Taxa
TH30	Goring	Les Ruse's Chironomid exuviae	1977	129	129 Whitchurch		8	-	2	2.000	
		Les Ruse's Chironomid exuviae	1978	129	129 Whitchurch		8	2	2	2,000	
		Les Ruse's Chironomid exuviae	1986	129	29 Whitchurch		8	3	2	2.000	Г
1		Les Ruse's Chironomid exuviae	1992	129	129 Whitchurch	08/05/92	1	1	2	2,000	_
		Les Ruse's Chironomid exuviae	1992	129	129 Whitchurch	24/06/92	2	-	2	2.000	-
		Les Ruse's Chironomid exuviae	1992	129	29 Whitchurch	24/08/92	2	2	2	2.000	:-
: : i	-	Les Ruse's Chironomid exuviae	1994	129	29 Whitchurch	12/05/94	-	2	2	2.000	
		Les Ruse's Chironomid exuviae	1994	129	29 Whitchurch	13/06/94	2	3	2	2.000	=
:		Les Ruse's Chironomid exuviae	1994	129	29 Whitchurch	12/09/94	'n	-	2	2.000	-
		Thames WA/NRA/EA routine monitoring	1987	55	55 Whitchurch Weir	24/07/87	2	3700	104	5.200	20
1		Thames WA/NRA/EA routine monitoring	1990	55	55 Whitchurch Weir	23/05/90	1	0099	156	5.200	30
		Thames WA/NRA/EA routine monitoring	1990	55	55 Whitchurch Weir	23/07/90	2	7500	83	4.882	17
	_ +	Thames WA/NRA/EA routine monitoring	1990	55	55 Whitchurch Weir	29/11/90	3	8300	66	5.105	61
		Thames WA/NRA/EA routine monitoring	1991	55	55 Whitchurch Weir	18/08/91	1	8900	123	5.348	23
		Thames WA/NRA/EA routine monitoring	1661	55	55 Whitchurch Weir	03/07/91	2	0006	154	5.310	29
		Thames WA/NRA/EA routine monitoring	1991	55	55 Whitchurch Weir	14/11/91	3	0066	92	4.600	2
		Thames WA/NRA/EA routine monitoring	1992	55	55 Whitchurch Weir	28/02/92	1	10300	116	5.524	21
		Thames WA/NRA/EA routine monitoring	1992	55	55 Whitchurch Weir	11/08/92	2	10600	102	1.636	22
		Thames WA/NRA/EA routine monitoring	1993	55	55 Whitchurch Weir	04/10/93	£.	11600	140	5.600	25
:		Thames WA/NRA/EA routine monitoring	1994	55	55 Whitchurch Weir	28/07/94	2	11900	146	5.214	28
		Thames WA/NRA/EA routine monitoring	1995	55	55 Whitchurch Weir	12/04/95	-	13400	104	5.474	61
		Thames WA/NRA/EA routine monitoring	1995	55	55 Whitchurch Weir	16/10/95	3	15700	110	5.000	22

Reach						Sample	Season	Season Sample			Number
е	Reach name	Reach name Data source name	Year	Site ID	Site ID Site name	date	自	а	BMWP	ASPT	of Taxa
TH32	Mapledurha	TH32 Mapledurha Thames WA/NRA/EA routine monitoring	1980	57	57 Caversham Weir	31/12/80	3	006	119	4.760	25
		Thames WA/NRA/EA routine monitoring	1987	57	7 Caversham Weir	18/02/87		2300	85	4.474	[61]
<u> </u>		Thames WA/NRA/EA routine monitoring	1987	57	57 Caversham Weir	12/08/87	2	4000	143	5.107	28
	-	Thames WANRA/EA routine monitoring	1988	57	7 Caversham Weir	09/02/88	-	4400	159	5.300	30
: : ! !		Thames WA/NRA/EA routine monitoring	_	57	7 Caversham Weir	19/0/89	2	\$900	113	4.708	2+
: 		Thames WA/NRA/EA routine monitoring	1989	57	i7 Caversham Weir	02/10/89	3	0009	116	4.833	24

APPENDIX 3.5 The frequency of occurrence of standard macro-invertebrate taxa in samples from three distinct zones: margins, mid-channel and vegetation.

Species name	Margin	Mid-channel	Vegetation
Spongillidae	2.0%	6.0%	1.9%
Hydridae			<u> </u>
Hydra sp.			1.9%
Planaria torva (Muller)	2.0%		1.9%
Polycelis tenuis (ljima)	1		1.9%
Dugesia tigrina (Girard)	10.0%		18.5%
Dugesia polychroa group	2.0%		3.7%
Dendrocoelum lacteum (Muller)	2.0%		
Prostoma sp.		: 	1.9%
Nematoda	16.0%	60.0%	3.7%
Ectoprocta	4.0%	10.0%	9.3%
Theodoxus fluviatilis (L.)	6.0%	6.0%	9.3%
Viviparus viviparus (L.)	14.0%	18.0%	9.3%
Valvata sp.	2.0%	2.0%	14.8%
Valvata cristata Muller		3	
Valvata piscinalis (Muller)	4.0%		ļ <u>.</u>
Potamopyrgus jenkinsi (Smith)	68.0%	52.0%	64.8%
Bithynia leachii (Sheppard)	6.0%	2.0%	20.4%
Bithynia tentaculata (L.)	22.0%	6.0%	37.0%
Physa sp.			3.7%
Physa fontinalis (L.)			16.7%
Physa acuta group			3.7%
Lymnaea palustris (Muller)	i		
Lymnaea peregra (Muller)	16.0%	2.0%	37.0%
Lymnaea stagnalis (L.)			3.7%
Planorbidae		:	1.9%
Planorbis carinatus Muller	2.0%		3.7%
Planorbis planorbis (L.)	2.0%		1.9%
Anisus vortex (L.)			1.9%
Bathyomphalus contortus (L.)	2.0%	:	
Gyraulus albus (Muller)	4.0%	4.0%	20.4%
Armiger crista (L.)			
Ancylus fluviatilis Muller		6.0%	16.7%
Acroloxus lacustris (L.)			18.5%
Succinea sp.		<u>i</u>	T
Unio sp.		10.0%	
Unio pictorum (L.)	18.0%	8.0%	
Unio tumidus Philipsson		14.0%	
Anodonta sp.		2.0%	
Anodonta anatina (L.)	18.0%	20.0%	1.9%
Anodonta cygnea (L.)		2.0%	
Sphaerium corneum (L.)	36.0%	16.0%	42.6%
Sphaerium lacustre (Muller)		2.0%	3.7%
Sphaerium rivicola (Lamarck)	18.0%	12.0%	1.9%
Sphaerium transversum (Say)			
Pisidium amnicum (Muller)	2.0%	2.0%	1.9%
Pisidium casertanum (Poli)	1	2.0%	5.6%
Pisidium henslowanum (Sheppard)	14.0%	8.0%	7.4%
Pisidium moitessierianum Paladilhe		i	Ĭ
Pisidium nitidum Jenyns		2.0%	5.6%
Pisidium subtroncatum Malm		2.0%	
Pisidium supinum Schmidt	12.0%	10.0%	5.6%
Lumbriculidae	56.0%	52.0%	9.3%
Stylodrilus sp.		T	
Stylodrilus heringianus Claparede	4.0%	14.0%	
Lumbriculus group			
Enchytraeidae	2.0%		
Chaetogaster sp.			
Uncinais uncinata (Orsted)			
Ophidonais serpentina (Muller)			5.6%

pecies name	Margin	Mid-channel	Vegetation
Nais bretscheri Michaelsen			
Nais pardalis Piguet	/		
Nais simplex Piguet			5.6%
Nais communis group			
Stylaria lacustris (L.)	6.0%		46.3%
Tubificid a e	30.0%	52.0%	18.5%
Fubifex ignotus (Stole)	6.0%	2.0%	<u> </u>
Tubifex tubifex (Muller)		8.0%	
Limnodrilus cervix Brinkhurst	16.0%	10.0%	5.6%
Limnodrilus claparedeianus Ratzel			
Limoodrilus hoffmeisteri Claparede	42.0%	34.0%	9.3%
Limnodrilus profundicola (Verrill)	2.0%		
Limnodrilus udekemianus Claparede	20.0%	12.0%	3.7%
Psammoryctides barbatus (Grube)	36.0%	44.0%	11.1%
Potamothrix hammoniensis (Michaelsen)		6.0%	7.4%
Potamothrix moldaviensis (Vejdovsky & Mrazek	58.0%	40.0%	20.4%
Aulodrilus pluriseta (Piguet)	8.0%		22.2%
Rhyacodrilus coccineus (Vejdovsky)		2.0%	
Branchiura sowerbyi Beddard	8.0%	2.0%	
Lumbricidae	6.0%		
Piscicola geometra (L.)	2.0%	4.0%	
Theromyzon tessulaturn (Muller)	4.0%		11.1%
Hemiclepsis marginata (Muller)			3.7%
Glossiphonia complanala (L.)	10.0%	14.0%	5.6%
Glossiphonia heteroclita (L.)	2.0%	4.0%	1.9%
Helobdella stagnalis (L.)	10.0%	16.0%	22.2%
Erpobdella sp.	4.0%	4.0%	13.0%
Erpobdella octoculata (L.)	6.0%	4.0%	7.4%
Dina lineata (Muller)	4.0%		
	2.0%		
Trocheta subviridis Dutrochet	72.0%	84.0%	64,8%
Hydracatina	12.070		7,4%
Argulus sp.	28.0%	26.0%	61.1%
Asellus aquaticus (L.)	26.0%		1.9%
Asellus meridianus Racovitza	12.0%	4.0%	9.3%
Corophium curvispinum Sars	24.0%	10.0%	40.7%
Crangonyx pseudogracilis Bousfield	2.0%		1.9%
Gammaridae	1 2070	+	
Gammarus sp.	12.00	2.0%	29.6%
Gammarus pulex (L.)	12.0%	2.0%	27.070
Baetidae		2.0%	1.9%
Baetis rhodani (Pictet)	100		24.1%
Baetis vernus Curtis	4.0%		42.6%
Baetis scambus group	4.0%		81.5%
Centroptilum luteolum (Muller)	36.0%		25.9%
Cloeon dipterum (L.)	6.0%		18.5%
Cloeon simile Eaton	1000		
Procloeon bifidum Bengtsson	18.0%		74.1%
Heptagenia fuscogrisea (Retzius)			
Habrophlebia fusca (Curtis)			
Ephemera sp.		4.00	
Ephemera danica Muller		4.0%	
Ephemera vulgata 1	4.0%	2.0%	0.20
Ephemerella ignita (Poda)		2.0%	9.3%
Caenis sp.	16.0%	24.0%	13.0%
Caenis horaria (L.)	2.0%		
Caenis tuctuosa group	6.0%	24.0%	11.1%
Nemoura sp.		2.0%	
Leuctra geniculata (Stephens)		2.0%	
Platyenemis pennipes (Pallas)			1.9%
	1		9.3%

Species name	Margin	Mid-channel	Vegetation
Enallagma cyathigerum (Charpentier)	• .		:
Coenagrion puella group			
Calopteryx sp.	·		1.9%
Calopteryx splendens (Harris)		:	
Calopteryx virgo (L.)	· · · · · · · · · · · · · · · · · · ·	:	1.9%
Gomphus vulgatissimus (L.)			
Aphelocheirus aestivalis (Fabricius)	· · · · · · · · · · · · · · · · · ·	2.0%	·
Notonecta sp.		:	1.9%
Notonecta glauca L.			
Notonecta maculata Fabricius			1
Corixidae	4.0%	2.0%	13.0%
Micronecta sp.	6.0%	2.0%	1.9%
Micronecta (Micronecta) poweri (Douglas & Sec			† · · · · · · · · · · · · · · · · · · ·
Sìgara (Sigara) sp.	2.0%		24.1%
Sigara (Subsigara) distincta (Fieber)			3.7%
Sigara (Subsigara) falleni (Fieber)			14.8%
Sigara (Subsigara) fossarum (Leach)			1.57
Haliplidae	26.0%	18.0%	25.9%
Haliplus sp.			1 25.570
Haliplus fluviatilis Aube			
Haliplus lineatocollis (Marsham)			
Dytiscidae	10.0%		37.0%
Laccophilus sp.	10.075		31.07
Laccophilus hyalinus (Degeer)			
Hydroporus sp.	16.0%	8.0%	5.6%
Potamonectes depressus (Fabricius)	10.0 %		1
Stictotarsus duodecimpustulatus (Fabricius)			-
Platambus maculatus (L.)			
Dytiscus sp.		······ t·· · ······-···-···············	
Gyrinus distinctus Aube			
Gyrinus urinator Illiger		· · · · · · · · · · · · · · · · · · ·	
Orectochilus villosus (Mulier)			
Hydrophilidae	6.0%	4.0%	3.7%
Helophorus sp.	2.0%	1,070	3,12
Helophorus (Atracthelophorus) brevipalpis Bedei			†
Dryops sp.		<u></u>	1.9%
Elmis aenea (Muller)		2.0%	· · · · · · · · · · · · · · · · · · ·
Oulimnius sp.		2.0%	†
Oulimnius tuberculatus (Muller)	20.0%	24.0%	11.1%
Sialis lutaria (L.)	6.0%	12.0%	5.6%
Sialis nigripes Pictet	0.072	8.0%	3.5%
Sisyra sp.	717		
Hydroptilidae		4.0%	7.4%
Agraylea multipunctata Curtis	2.0%	7,77	1.770
Hydroptila sp.	4.0%	4.0%	13.0%
Oxyethira sp.	2.0%	7.57/0	15.070
Unytrichia sp.	2.070		
Lype sp.		<u>f</u>	
Tinodes waeneri (L.)	2.0%	·····	
Ecnomus tenellus (Rambur)	2.470		1
Polycentropodidae		2.0%	
Cyrnus flavidus Melachlan		2.9.70	5.6%
Cyrnus trimaculatus (Curtis)	50.0%	60.0%	13.0%
Neureclipsis bimaculata (L.)	56.070	2.0%	9.3%
Plectrocnemia sp.		2.0%	3.7%
Plectrocnemia sp.	2.0%		1.9%
Polycentropus sp.	4.0%		20.4%
Polycentropus flavomaculatus (Pictet)	4.070	·····	ZU.476
Polycentropus irroratus (Curtis)		<u> </u>	+
Hydropsyche pellucidula (Curtis)			į ·····

Species name	Margin	Mid-channel	Vegetation
Brachycentrus subnubilus Curtis			
Lepidostoma hirtum (Fabricius)			
Allogamus auricollis (Pictet)	2.0%		
Halesus sp.		- · · · · · · · · · · · · · · · · · · ·	1
Anabolia nervosa (Curtis)	4.0%		3.7%
Limnephilus sp.			<u> </u>
Limnephitus lunatus Curtis			<u>, </u> ,
Potamophylax group			
Goera pilosa (Fabricius)	2.0%	2.0%	
Silo sp.		2.0%	
Molannidae	<u></u>	2.0%	
Molanna angustata Curtis	12.0%	4.0%	3.7%
Leptoceridae	4.0%	8.0%	3.7%
Athripsodes aterrimus (Stephens)	<u></u>		
Athripsodes cinereus (Curtis)	<u>.</u>		1.627
Ceraclea sp.	! !		1.9%
Ceraclea annulicomis (Stephens)			0.27
Mystacides sp.	12.0%	10.0%	9.3%
Mystacides azurea (L.)	ļ		1.9%
Mystacides longicornis (L.)	ļ		
Mystacides nigra (L.)	<u> </u>		1.9%
Oecefis lacustris (Pictet)	·		
Oecetis ochracea (Curtis)	16.00	0.00	13.0%
Diptera	16.0%	8.0%	15.0%
Tipulidae			
Tipula sp.			
Tipula (Yamatotipula) montium group	ļ		-
Ceratopogonidae	2.0%		1.9%
Simulium (Boophthora) erythrocephalum (de Ge	2.070	4.0%	1.9%
Clinotanypus nervosus (Meigen)		2.0%	
Apsectrotanypus trifascipennis (Zetterstedt)	10.0%	18,0%	
Macropelopia sp.	26.0%	38.0%	14.8%
Procladius sp. Ablabesmyia sp.	22.0%	24,0%	13.0%
Thienemannimyia group	4.0%	6.0%	3.7%
Natarsia sp.	2.0%		
Paramerina sp.	4.0%		
Ponthastia gaedii group			
Ponthastia longimana group		2.0%	
Prodíamesa olivacea (Meigen)	32.0%	22.0%	11.1%
Orthocladiinae			
Brillia sp.	1		3.7%
Cricotopus sp.	46.0%	4.0%	68.5%
Cricotopus (Cricotopus) sp.			3.7%
Eukiefferiella sp.			1.9%
Cricotopus group			
Nanociadius sp.			3.7%
Orthocladius sp.	16.0%	6.0%	18.5%
Paracladius sp.			
Psectrocladius sp			
Psectrocladius (Allopsectrocladius) obvius (Wa	ılker)		
Synorthocladius semivirens (Kieffer)	<u></u>		1.9%
Chaetocladius sp.			
Corynoneura sp.			3.7%
Limnophyes sp.	6.0%	2.0%	5.6%
Thienemanniella sp.			14.8%
Chironomus sp.	34.0%	70.0%	5.6%
Ciadopelma sp.		2.00	0.27
Cryptochironomus sp.	30.0%	34.0%	9.3%
Cryptotendipes sp.		14.0%	
Dicrotendipes sp.	78.0%	38.0%	35.2%

Species name	Margin	Mid-channel	Vegetation
Endochironomus sp.	8.0%	2.0%	1.9%
Glyptotendipes sp.	32.0%	14.0%	1.9%
Harnischia sp.	2.0%		
Kiefferulus tendipediformis (Goetghebuer)			:
Microchironomus sp.		6.0%	:
Microtendipes sp.	16.0%	12.0%	1.9%
Parachironomus sp.	2.0%	:	13.0%
Paracladopelma sp.	14.0%	20.0%	7.4%
Paralauterborniella sp.			
Paralauterborniella nigrohalteralis (Malloch)	1		
Paratendipes sp.	18.0%	10.0%	11.1%
Phaenopsectra sp.			7.4%
Polypedilum sp.		[
Polypedilum (Pentapedilum) sp.	54.0%	34.0%	24.1%
Polypedilum (Polypedilum) sp.	8.0%	14.0%	
Stenochironomus sp.			
Stictochironomus sp.	2.0%	2.0%	
Xenochironomus xenolabis (Kieffer)	4.0%	2.0%	1.9%
Cladotanytarsus sp.	34.0%	44.0%	3.7%
Micropsectra sp.	2.0%		1.9%
Paratanytarsus sp.			
Rheotanytarsus sp.	12.0%	18.0%	29.6%
Stempellinelia sp.			:
Tanytarsus sp.	12.0%	10.0%	13.0%
Micropsectra group			
Chrysops sp.		1	
Muscidae		÷	<u> </u> -
Fotal no of samples	50	50	.i

APPENDIX 3.6 The frequency of occurrence of standard macro-invertebrate taxa in samples from eight distinct habitats: clay, silt, gravel, bedrock/concrete, detritus/organic matter, emergent vegetation, submerged vegetation and floating vegetation.

Species name				HAB	ITAT			
				Bedrock/concrete	tus	Emergent vegetation	Submerged vegetation	Floating vegetation
	Clay	Silt	Gravel	Bedr	Detritus	Emei		Float
Spongillidae	3%		4%	<u>.</u>	20%		3%	
Hydra sp.		<u></u>						6%
Planaria torva (Muller)	·		2%				3%	
Polycelis tenuis (Ijima)		L	<u></u>	ļ <u>.</u>			3%	
Dugesia tigrina (Girard)	12%		2%			50%	19%	12%
Dugesia polychroa group			2%				6%	
Dendrocoelum lacteum (Muller)			2%					ĺ
Prostoma sp.							3%	
Nematoda	15%	60%	55%		60%		6%	
Ectoprocta	6%		8%		20%		10%	12%
Theodoxus fluviatilis (L.)	3%	20%	6%				10%	6%
Viviparus viviparus (L.)	15%	20%	14%		20%		10%	12%
Valvata sp.			2%				16%	12%
Valvata piscinalis (Muller)			2%		20%			
Potamopyrgus jenkinsi (Smith)	68%	60%	51%	100%	60%	50%	65%	65%
Bithynia leachii (Sheppard)	6%		2%				32%	6%
Bithynia tentaculata (L.)	18%		4%	67%	20%	· · · · · ·	42%	41%
Physa sp.							3%	
Physa fontinalis (L.)						50%	10%	24%
Physa acuta group					<u> </u>			12%
Lymnaea peregra (Muller)	18%		2%			50%	19%	59%
Lymnaea stagnalis (L.)			- "				3%	6%
Planorbidae		-					3%	
Planorbis carinatus Muller			2%				3%	6%
Planorbis planorbis (L.)	3%						3%	
Anisus vortex (L.)	 					25%		
Bathyomphalus contortus (L.)	 		2%					-
Gyraulus albus (Muller)	3%		6%			50%	10%	24%
Ancylus fluviatilis Muller			6%			25%	19%	12%
Acroloxus lacustris (L.)							29%	
Unio sp.			10%					
Unio pictorum (L.)	9%	40%	12%		20%			
Unio tumidus Philipsson			14%					
Anodonta sp.			2%					
Anodonta anatina (L.)	15%	40%	20%		20%		3%	
Anodonta cygnea (L.)	- 		2%				- 70	
Sphaerium corneum (L.)	32%	40%	18%		20%		52%	35%
Sphaerium lacustre (Muller)	1		2%				6%	
Sphaerium rivicola (Lamarck)	12%		14%		40%		3%	
Pisidium amnicum (Muller)	3%		2%		/-		3%	
Pisidium casertanum (Poli)			2%				3%	12%
Pisidium henslowanum (Sheppard)	15%	·	10%		20%		10%	6%
Pisidium nitidum Jenyns	-	20%	.0,0		2070		3%	6%
Pisidium subtruncatum Malm	++	20 10			20%		570	070
Pisidium supinum Schmidt	6%		16%		20 70		3%	6%

Species name	! 			HABI	TAT			
	Clay]t	Gravel	Bedrock/concrete	Detritus	Emergent vegetation	Submerged vegetation	Floating vegetation
		Silt				_ <u>m</u>		_드
Lumbriculidae	59%		57%	67%	20%		16%	
Stylodrilus heringianus Claparede	3%	20%	10%	33%	20%			
Enchytraeidae			2%			250		-
Ophidonais serpentina (Muller)						25%	6%	
Nais simplex Piguet						25%	3%	6%
Stylaria lacustris (L.)	9%					50%	——	47%
Tubificidae	29%	40%	49%	67%	40%	25%	19%	18%
Tubifex ignotus (Stolc)	3%		4%					
Tubifex tubifex (Muller)			6%		20%			
Limnodrilus cervix Brinkhurst	9%	40%	12%		20%		6%	
Limnodrilus hoffmeisteri Claparede	35%	60%	33%		60%		13%	
Limnodrilus profundicola (Verrill)			2%					
Limnodrilus udekemianus Claparede	9%	40%	14%		40%		3%	
Psammoryctides barbatus (Grube)	38%	60%	39%	100%	20%	25%	16%	
Potamothrix hammoniensis (Michaelsen)			6%				13%	
Potamothrix moldaviensis (Vejdovsky & Mrazek	53%	40%	51%		60%		23%	18%
Aulodrilus pluriseta (Piguet)	6%	20%					32%	6%
Rhyacodrilus coccineus (Vejdovsky)			2%					
Branchiura sowerbyi Beddard	3%		6%					
Lumbricidae	6%		2%					
Piscicola geometra (L.)	3%		4%					
Theromyzon tessulatum (Muller)	3%		2%			25%	3%	24%
Hemiclepsis marginata (Muller)							3%	6%
Glossiphonia complanata (L.)	12%	20%	10%		20%		L	12%
Glossiphonia heteroclita (L.)			2%		40%		3%	<u> </u>
Helobdella stagnalis (L.)	12%	20%	14%		20%	25%	26%	18%
Erpobdella sp.		40%	4%				10%	18%
Erpobdella octoculata (L.)	3%		6%				6%	12%
Dina lineata (Muller)	6%							
Trocheta subviridis Dutrochet	3%							
Hydracarina	79%	80%	80%	67%	40%	75%	68%	65%
Argulus sp.			1				3%	12%
Asellus aquaticus (L.)	26%	20%	24%	33%	60%	50%	77%	29%
Asellus meridianus Racovitza							3%	<u> </u>
Corophium curvispinum Sars	18%		4%			25%	10%	6%
Crangonyx pseudogracilis Bousfield	26%		12%		20%	25%	58%	18%
Gammaridae	3%						3%	
Gammarus pulex (L.)	15%		2%		20%	25%	29%	24%
Baetidae			2%					
Baetis rhodani (Pictet)		Ţ					3%	<u></u>
Baetis vernus Curtis	6%		1			25%	23%	24%
Baetis scambus group	6%			1			45%	47%
Centroptilum luteolum (Muller)	47%		2%	,		75%	90%	65%
Cloeon dipterum (L.)	3%	+	,			25%	19%	35%
Cloeon simile Eaton	1	1			T	25%	6%	35%

Species name				HAB	ITAT	7	_	
	Clay	Silt	Gravel	Bedrock/concrete	Detritus	Emergent vegetation	Submerged vegetation	Floating vegetation
Procloeon bifidum Bengtsson	24%	<u> </u>	-			75%		
Ephemera danica Muller			4%					
Ephemera vulgata L.	6%		2%				†	_
Ephemerella ignita (Poda)			2%			25%	6%	12%
Caenis sp.	15%		20%	100%	20%	25%	-	6%
Caenis horaria (L.)	3%						10.1	
Caenis luctuosa group	9%	20%	20%				10%	6%
Nemoura sp.	3%		-075				1070	
Leuctra geniculata (Stephens)	3 /0		2%					
Platycnemis pennipes (Pallas)			270		<u>_</u>		3%	<u></u>
Coenagriidae					-		10%	6%
Calopteryx virgo (L.)							1070	6%
Aphelocheirus aestivalis (Fabricius)			2%			-		070
Notonecta sp.						25%		
Corixidae	3%		4%			23 70	19%	6%
Micronecta sp.	9%	20%	74 70				3%	070
Micronecta (Micronecta) poweri (Douglas & Sco		40 70					370	
Sigara (Sigara) sp.	3%					25%	13%	41%
Sigara (Subsigara) distincta (Fieber)	270					2370	1376	12%
Sigara (Subsigara) falleni (Fieber)				 -i			10%	24%
Haliplidae	21%	20%	22%		20%	50%	23%	24%
Dytiscidae	9%	2070	2%		20%	25%	45%	24%
Hydroporus sp.	18%		8%		20%	25%	3%	6%
Hydrophilidae	6%	<u> </u>	6%		2076	25%	3%	6%
Helophorus sp.	070	20%	070				370	070
Dryops sp.		2070		· · · · · · · · · ·			3%	
Elmis aenea (Muller)		·	2%				370	_
Oulimnius sp.			2%					·
Oulimnius tuberculatus (Muller)	24%		20%	67%	20%		1207	120
Sialis lutaria (L.)	12%		8%	0/70	20%		13% 10%	12%
Sialis nigripes Pictet	1270		8%		2070		1070	
Hydroptilidae			4%				100%	
Agraylea multipunctata Curtis	3%		470	-			10%	
Hydroptila sp.	6%		4%			2507	1207	1207
Oxyethira sp.	3%		470			2370	13%	12%
Polycentropodidae	370		2%					
Cyrnus flavidus Mclachlan			270			· -		18%
Cyrnus trimaculatus (Curtis)	59%	40%	63%			25%	6%	18%
Neureclipsis bimaculata (L.)	3770	7070	05/10	33%		25%	6%	6%
Plectrocnemia sp.				3370		EJ 70	3%	6%
Plectrocnemia geniculata Mclachlan	3%						نه د	6%
Polycentropus sp.	6%			-		25%	10%	35%
	370		4%	.	20%	23%	16%	35% 6%
FILL VP ALICA SD.					4		112-70	070
Phryganea sp. Allogamus auricollis (Pictet)	3%		770		2070		1070	

Species name	ļ 			HABI	TAT	·		
	Clay	Silt	Gravel	Bedrock/concrete	Detritus	Emergent vegetation	Submerged vegetation	Floating vegetation
Goera pilosa (Fabricius)		<u> </u>	4%					
Silo sp.			2%					
Molannidae			2%					
Molanna angustata Curtis	9%		6%		20%		3%	6%
Leptoceridae	6%		6%		20%		6%	
Ceraclea sp.							3%	
Mystacides sp.	21%		8%				16%	
Mystacides longicornis (L.)								6%
Mystacides nigra (L.)	1	1			Í		3%	
Diptera	21%	40%	6%			25%	10%	12%
Simulium (Boophthora) erythrocephalum (de Ge			2%					6%
Clinotanypus nervosus (Meigen)			4%				3%	
Apsectrotanypus trifascipennis (Zetterstedt)	3%		1					
Macropelopia sp.	9%		16%		20%			
Procladius sp.	21%	60%	39%		40%	25%	19%	
Ablabesmyia sp.	18%	20%	24%	67%		25%	3%	24%
Thienemannimyia group		20%	6%	33%				12%
Natarsia sp.	3%							
Paramerina sp.	3%	20%						
Potthastia longimana group			2%					
Prodiamesa olivacea (Meigen)	24%		31%		40%		13%	6%
Brillia sp.							3%	6%
Cricotopus sp.	50%		8%		60%	100%	68%	71%
Cricotopus (Cricotopus) sp.	1		-				3%	6%
Eukiefferiella sp.	 		-				3%	
Nanocladius sp.	 							12%
Orthocladius sp.	21%		6%	33%		50%	6%	29%
Synorthocladius semivirens (Kieffer)	1		- 0.0				3%	
Corynoneura sp.	 	<u> </u>						12%
Limnophyes sp.	3%		6%			25%	6%	
Thienemanniella sp.	1		0.0	_			19%	12%
Chironomus sp.	29%	60%	67%	67%	60%	<u>-</u> -	6%	
Cryptochironomus sp.	21%	0070	45%		20%		6%	12%
Cryptotendipes sp.	12170		10%		40%			
Dicrotendipes (Limnochironomus) sp.	88%	60%	 	33%	40%		29%	35%
Endochironomus sp.	100,0	20%		20,0	- آ آ آ	· · · · ·	3%	
Glyptotendipes sp.	38%		16%	 -	20%	 -	3%	-
Harnischia sp.	3%		1.070		20,0	 		
Microchironomus sp.	+ 370	20%	4%			 	 	
Microtendipes sp.	21%	20,70	14%				3%	
Parachironomus sp.	3%	 	2170			25%		
Paracladopelma sp.	15%	+	18%		40%		10%	
	15%		+		20%	-	10%	
Paratendipes sp.	1370	4070	12.70		2070	25%		
Phaenopsectra sp.				 _	 		U 70	29%

Species name				HAB	ITAT			
	Clay	Silt	Gravel	Bedrock/concrete	Detritus	Emergent vegetation	Submerged vegetation	Floating vegetation
Polypedilum (Polypedilum) sp.	9%		16%					
Stictochironomus sp.			4%					
Xenochironomus xenolabis (Kieffer)	3%				20%			6%
Cladotanytarsus sp.	38%		47%	33%			3%	
Micropsectra sp.	3%							6%
Rheotanytarsus sp.	15%		18%			25%	19%	47%
Tanytarsus sp.	12%	40%	8%		20%		19%	
No of samples for habitat	34	5	49	3	5	4	31	17

APPENDIX 3.7 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given emergent macrophyte was present, as the dominant or non-dominant species.

Tavon	Typha	Scirpus	Phragmites	S.erectum	Acorus	Iris	Olyceria	Carex
Taxon		6%	<i></i>	14%		:		
Spongillidae				·			[
Hydra sp.		-		14%				
Planaria torva (Muller)		6%	-	14%			:	
Polycelis tenuis (Ijima)	33%	6%	40%	14%	33%	100%		50%
Dugesia tigrina (Girard)	1 300	6%	- 	29%		.3 7=		• • • • • • • • • • • • • • • • • • • •
Dugesia polychroa group		6%		14%	\- 	-		
Prostoma sp.		6%		1770			†	
Nematoda					33%	 -		
Ectoprocta		13%	200		33%	100%		50%
Theodoxus fluviatilis (L.)		13%	20%		2370	100 70	100%	
Viviparus viviparus (L.)		13%		1.40	(70			
Valvata sp.	33%	6%		14%	67%	1000	100%	1000
Potamopyrgus jenkinsi (Smith)	67%	63%	40%	71%	 	100%	100%	100%
Bithynia leachii (Sheppard)	<u> </u>	38%	20%	29%	67%	100%	100-	100%
Bithynia tentaculata (L.)	_	25%	20%	29%	100%	100%	100%	50%
Physa sp.		6%	<u></u>	14%	ļ <u></u>			
Physa fontinalis (L.)		13%		14%			100%	
Physa acuta group				L	<u>.</u>			
Lymnaea peregra (Muller)		19%	20%	14%	33%		100%	
Lymnaea stagnalis (L.)							100%	
Planorbidae		6%	1	14%				
Planorbis carinatus Muller							100%	
Planorbis planorbis (L.)							100%	
Anisus vortex (L.)			<u> </u>					
Gyraulus albus (Muller)	33%	6%		i	33%		100%	
Ancylus fluviatilis Muller	67%	25%	20%	· · · · ·	33%	-		
Acroloxus lacustris (L.)		25%	20%	14%	67%	100%	100%	100%
Anodonta anatina (L.)		23.0	20%					
Sphaerium corneum (L.)		50%	40%	57%	67%	100%		100%
Sphaerium lacustre (Muller)		6%	10.0	29%	1			
Sphaerium rivicola (Lamarck)		UN		2,50	<u></u>			
Pisidium amnicum (Muller)		6%			 			<u></u>
		0.70			33%	100%		50%
Pisidium casertanum (Poli)		COL		29%	3370	100 %		50%
Pisidium henslowanum (Sheppard)		6%		2970	 			3070
Pisidium nitidum Jenyns		 	200	 	 	<u> </u>	ļ	
Pisidium supinum Schmidt			20%	200	 			ļ
Lumbriculidae		19%	20%	29%				
Ophidonais serpentina (Muller)	33%	6%	20%	<u> </u>	33%			
Nais simplex Piguet		6%						·
Stylaria lacustris (L.)	67%	50%	60%	14%	67%			
Tubificidae	33%	19%	20%	14%	33%	100%	100%	50%
Limnodrilus cervix Brinkhurst		}	}	.	}		ļ	50%
Limnodrilus hoffmeisteri Claparede		6%	20%		ļ]	50%
Limnodrilus udekemianus Claparede			ļ	ļ				50%
Psammoryctides barbatus (Grube)	67%	6%	<u> </u>	14%	33%	<u></u>	100%	50%
Potamothrix hammoniensis (Michaelsen)		13%	<u> </u>	29%	 	<u> </u>	_	50%
Potamothrix moldaviensis (Vejdovsky & Mrazek)		13%	20%	29%		<u> </u>	100%	50%
Autodrilus pluriseta (Piguet)	33%	19%	20%	43%	33%	100%	100%	100%
Theromyzon tessulatum (Muller)				14%		ļ		. <u> </u>
Hemiclepsis marginata (Muller)		6%		14%		<u></u>	<u> </u>	L
Glossiphonia complanata (L.)	Ţ	1	1				ļ	[
Glossiphonia heteroclita (L.)		1		[33%		1	
Helobdella stagnalis (L.)	33%	19%		57%	100%	100%		50%
Erpobdella sp.		6%	20%	I	T		100%	
Erpobdella octoculata (L.)	1	†- <u></u>	1	1	67%	100%		50%

			nite.	£				
_	Typha	Scirpus	Phragmites	S.erectum	Acorus	Ti.	Glyceria	Carex
Taxon			,	↑	}	ŧ 		
Hydracarina	100%	63%	40%	71%	100%	100%	100%	100%
Argulus sp.		722			1000	1000	100%	
Asellus aquaticus (L.)	100%	_	80%	71%	100%	100%	100%	100%
Asellus meridianus Racovitza		6%	200	14%		į	<u> </u>	
Corophium curvispinum Sars		6%	20%	14%	1000	1000	1000	1000
Crangonyx pseudogracilis Bousfield	67%	50%	20%	71%	100%	100%	100%	100%
Gammaridae		6%	200	14%	220	1000	1000	1000
Gammarus pulex (L.)		19%	20%	43%	33%	100%	100%	100%
Baetis rhodani (Pictet)		6%	000	200	ļ			
Baetis vernus Curtis	33%	31%	20%	29%	ļ			
Baetis scambus group	+	69%	20%	57%	1000	1000	1000	100~
Centroptilum luteolum (Muller)	100%	81%	100%	100%	_	-	100%	100%
Closon dipterum (L.)	_	19%		14%	33%	100%	100%	50%
Cloeon simile Eaton		6%					100%	
Procloeon bifidum Bengtsson	100%	75%	80%	71%	100%	3001	100%	50%
Ephemerella ignita (Poda)	<u> </u>	13%	L					
Caenis sp.	100%	6%		14%	67%			
Caenis luctuosa group		19%	,	14%				
Platycnemis pennipes (Pallas)		6%						
Coenagriidae		6%		14%	67%			
Calopteryx virgo (L.)								
Notonecta sp.								
Corixidae		25%	20%					
Micronecta sp.				14%				
Sigara (Sigara) sp.		6%		14%	33%	100%	100%	50%
Sigara (Subsigara) distincta (Fieber)								
Sigara (Subsigara) falleni (Fieber)		6%		29%	33%	100%		50%
Haliplidae		19%	20%	57%			100%	
Dytiscidae		50%		71%	67%	100%	100%	50%
Hydroporus sp.		6%						
Hydrophilidae		6%		···				
Dryops sp.	<u> </u>			14%				
Oulimnius tuberculatus (Muller)		6%	20%	29%	33%	100%		50%
Sialis lutaria (L.)				14%				50%
Hydroptilidae	-	6%		14%	33%			
Hydroptila sp.		6%	60%	14%				
Cyrnus flavidus Mclachlan								
Cyrnus trimaculatus (Curtis)	67%	6%		14%				
Neureclipsis bimaculata (L.)	33%	6%	20%		33%			
Plectrocnemia sp.	1 20 10	6%		14%	3373			
Plectroenemia geniculata Melachlan				1110				
Polycentropus sp.		19%		14%	1			
Phryganea sp.		13%	20%	29%			100%	
Anabolia nervosa (Curtis)		6%	2070	14%	33%		10070	
Molanna angustata Curtis	-	0 70		1470	3370		100%	
Leproceridae		6%	20%	14%			10070	
Ceraclea sp.		070	20 %	1770	33%	100%		50%
Mystacides sp.	33%	6%	20%		עודעני	100770	100%	30 10
Mystacides longicornis (L.)	1070	0.70	4070	 			100 10	
Mystacides nigra (L.)					320	1000		50%
Diptera		100		200	33%	100%		3070
·		13%		29%				
Simulium (Boophthora) erythrocephalum (de Geer)				 	 -		100~	· · · ·
Clinotanypus nervosus (Meigen)	 		200	1.40	220		100%	
Procladius sp.	33%	13%	20%	14%	33%		100%	
Ablabesmyia sp.			20%	<u> </u>	•	ļ	<u> </u>	

Тахоп	Typha	Scirpus	Phragmites	S.erectum	Acorus	Iris	Glyceria	Carex
Thienemannimyia group			ļ		:	.	† . 	 .
Prodiamesa olivacea (Meigen)		6%		<u> </u>	9 · · ·	† · · · · ·	100%	50%
Brillia sp.	33%		T	1		<u> </u>	*	
Cricotopus sp.	100%	75%	40%	57%	100%	100%	100%	50%
Cricotopus (Cricotopus) sp.		6%	1	14%			•	<u></u>
Eukiefferiella sp.	· ·-	6%		!	İ	<u> </u>	j	··••
Nanocladius sp.			1	<u></u>	1	f 		
Orthocladius sp.		6%		<u> </u>	33%	100%		50%
Synorthocladius semivirens (Kieffer)		6%				1		
Corynoneura sp.		1				<u> </u>		
Limnophyes sp.	67%	6%]			
Thienemanniella sp.	33%	31%	20%	29%				
Chironomus sp.		6%	1	14%				50%
Cryptochironomus sp.		6%		14%	1	<u> </u>		50%
Dicrotendipes (Limnochironomus) sp.	100%	13%	40%		33%	100%	100%	50%
Endochironomus sp.							100%	
Glyptotendipes sp.					33%	100%		50%
Microtendipes sp.		6%		14%				
Parachironomus sp.	······	19%		14%	<u> </u>	 		
Paracladopelma sp.		13%		29%				50%
Paratendipes sp.			20%					50%
Phaenopsectra sp.	33%	6%	20%		33%			
Polypedilum (Pentapedilum) sp.		19%	20%	29%				50%
Xenochironomus xenolabis (Kieffer)		,					<u> </u>	
Cladotanytarsus sp.							;	
Micropsectra sp.								
Rheotanytarsus sp.		19%	20%	29%	33%		100%	• • • • • • • • • • • • • • • • • • • •
Tanytarsus sp.	33%	13%	20%	14%			ĺ	50%
	3	16	5	7	3	1	1	2

APPENDIX 3.8 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given submerged macrophyte was present, as the dominant or non-dominant species.

Taxa	S.erectum	S.emersum	Myriophyllum	Ciadophora
Polycentropus sp.			100%	100%
Phryganea sp.	1		100.0	
Anabolia nervosa (Curtis)	- 			
Molanna angustata Curtis	+			
Leptoceridae				
Ceraclea sp.				
Mystacides sp.	100%			
Mystacides longicornis (L.)	10070			
Mystacides nigra (L.)				
Diptera Diptera		50%		100%
Simulium (Boophthora) erythrocephalum (de Geer)	+	3070		10070
Clinotanypus nervosus (Meigen)	-			
Procladius sp.	100%			
Ablabesmyia sp.	10070		100%	100%
Thienemannimyia group			10070	100%
Prodiamesa olivacea (Meigen)	100%			10070
Brillia sp.	100%			
Cricotopus sp.		100%	100%	100%
Cricotopus (Cricotopus) sp.	-	100%	10070	10070
Eukiefferiella sp.	╂			
Nanocladius sp.	- -			
Orthociadius sp.		100%		100%
Synorthocladius semivirens (Kieffer)		10070		10070
Corynoneura sp.				
Limnophyes sp.	-			
Thienemanniella sp.	-			
Chironomus sp.			 	
Cryptochironomus sp.				
Dicrotendipes (Limnochironomus) sp.	100%	50%	100%	100%
Endochironomus sp.	100%	5070	10075	10070
Glyptotendipes sp.	-			
Microtendipes sp.				
Parachironomus sp.	+		100%	- ·
Paraciadopelma sp.			10070	
Paratendipes sp.	100%			
Phaenopsectra sp.	- 3076			
Polypedilum (Pentapedilum) sp.	100%			100%
Xenochironomus xenolabis (Kieffer)	10070			10070
Cladotanytarsus sp.	100%	. 		<u> </u>
Micropsectra sp.	10070			
Rheotanytarsus sp.			100%	100%
Tanytarsus sp.	100%		10070	10070
ronimana ap.	100%	2	1	1

	, -	!	F	
		Ę	Myriophyllum	ē
	S.erectum	S.emersum	phy	Cladophora
	5	meı	rio	ę g
Taxa	S.el	S.el	My	Cla
Theromyzon tessulatum (Muller)	1	50%		
Hemiclepsis marginata (Muller)	1			100%
Glossiphonia complanata (L.)				
Glossiphonia heteroclita (L.)				
Helobdella stagnalis (L.)	100%			100%
Erpobdella sp.	1	-		100%
Erpobdella octoculata (L.)				
Hydracarina	100%	50%	100%	100%
Argulus sp.	1			
Asellus aquaticus (L.)			100%	100%
Asellus meridianus Racovitza				
Corophium curvispinum Sars		50%		100%
Crangonyx pseudogracilis Bousfield	100%		-	
Gammaridae	1.55,0			
Gammarus pulex (L.)			100%	100%
Baetis rhodani (Pictet)			100.0	100.0
Baetis vernus Curtis			100%	
Baetis scambus group	1		10070	100%
Centroptilum luteolum (Muller)	100%	50%	100%	
Cloeon dipterum (L.)	10070	2070	-	100%
Cloeon simile Eaton				100%
Procloeon bifidum Bengtsson	100%	50%	100%	10070
Ephemerella ignita (Poda)	100%	30 70	100%	
	 		100%	
Caenis sp. Caenis luctuosa group				
Platycnemis pennipes (Pallas)	+	-		
Colombia (I.)	- 			
Calopteryx virgo (L.)		50%	<u></u>	-
Notonecta sp.	100%	30%		
Corixidae	100%			
Micronecta sp.			1000	1000
Sigara (Sigara) sp.			100%	100%
Sigara (Subsigara) distincta (Fieber)	-		ļ	
Sigara (Subsigara) falleni (Fieber)	-	500	1000	1000
Haliplidae	 	50%	_	100%
Dytiscidae	-		100%	
Hydroporus sp.	-		100%	100%
Hydrophilidae			}	}
Dryops sp.				1007
Oulimnius tuberculatus (Muller)	1000		 	100%
Sialis lutaria (L.)	100%	-		
Hydroptilidae	-		1000	1000
Hydroptila sp.			100%	100%
Cyrnus flavidus McIachlan	1			<u> </u>
Cyrnus trimaculatus (Curtis)				
Neureclipsis bimaculata (L.)	 			<u> </u>
Plectrocnemia sp.	-	 	1	<u> </u>
Plectrocnemia geniculata Mclachlan	<u> </u>	1	<u>i</u>	<u> </u>

, <u>, , , , , , , , , , , , , , , , , , </u>				
Taxa	S.erectum	S.emersum	Myriophyllum	Cladophora
Spongillidae	- S	N N	 ~	1 -
Hydra sp.		<u> </u>	 	
Planaria torva (Muller)		<u> </u>	 	├
			 	-
Polycelis tenuis (Ijima)			 	
Dugesia tigrina (Girard)	100%	50%	-	100%
Dugesia polychroa group			ļ	_
Prostoma sp.			ļ	ļ
Nematoda	100%	_	ļ	ļ. <u></u> .
Ectoprocta			ļ	
Theodoxus fluviatilis (L.)				
Viviparus viviparus (L.)			<u> </u>	<u> </u>
Valvata sp.		!		100%
Potamopyrgus jenkinsi (Smith)	100%	50%		100%
Bithynia leachii (Sheppard)	100%			
Bithynia tentaculata (L.)	100%			100%
Physa sp.				
Physa fontinalis (L.)		50%	100%	
Physa acuta group	_			
Lymnaea peregra (Muller)		50%	100%	100%
Lymnaea stagnalis (L.)				100.0
Planorbidae	`			-
Planorbis carinatus Muller	- -			
Planorbis planorbis (L.)				
Anisus vortex (L.)		50%	-	
Gyraulus albus (Muller)	-+	50%		
Ancylus fluviatilis Muller		5070		
Acroloxus lacustris (L.)				
Anodonta anatina (L.)				
Sphaerium corneum (L.)	100%			100%
Sphaerium lacustre (Muller)	100%			100%
Sphaerium rivicola (Lamarck)	1000			
Pisidium amnicum (Muller)	100%	i		
Pisidium casertanum (Poli)				
Pisidium henslowanum (Sheppard)				
Pisidium nitidum Jenyns				
Pisidium supinum Schmidt				
Lumbriculidae				
Ophidonais serpentina (Muller)				
Nais simplex Piguet			100%	
Stylaria lacustris (L.)			100%	100%
Tubificidae				
Limnodrilus cervix Brinkhurst	100%]
Limnodrilus hoffmeisteri Claparede	100%			
Limnodrilus udekemianus Claparede				
Psammoryctides barbatus (Grube)				
Potamothrix hammoniensis (Michaelsen)	100%			
Potamothrix moldaviensis (Vejdovsky & Mrazek)	100%			
Aulodrilus pluriseta (Piguet)	100%			

APPENDIX 3.9 The frequency of occurrence of standard with macro-invertebrate taxa in samples where a given floating macrophyte was present, as the dominant <u>or</u> non-dominant species.

	Nuphar	Pot.pectinatus	Lemna
Taxon	Ž	<u>\$</u>	=
Spongillidae		1	
Hydra sp.	10%	<u> </u>	
Planaria torva (Muller)	1		
Polycelis tenuis (Ijima)	 -		
Dugesia tigrina (Girard)	10%	13%	
Dugesia polychroa group	†		
Prostoma sp.			
Nematoda			
Ectoprocta	10%	13%	
Theodoxus fluviatilis (L.)	10%	_	1
Viviparus viviparus (L.)	10%		
Valvata sp.	1070	25%	
Potamopyrgus jenkinsi (Smith)	50%	75%	100%
Bithynia leachii (Sheppard)	10%	1370	100%
Bithynia tentaculata (L.)	50%	38%	
Physa sp.	30%	3070	
Physa fontinalis (L.)	10%	38%	
	10%	25%	1000
Physa acuta group	700	_	100%
Lymnaea peregra (Muller)	70%	38%	100%
Lymnaea stagnalis (L.) Planorbidae	10%	1	
Planorbis carinatus Muller			
		13%	
Planorbis planorbis (L.) Anisus vortex (L.)			
	20~	0.504	
Gyraulus albus (Muller)	20%	25%	
Ancylus fluviatilis Muller	20%		
Acroloxus lacustris (L.)			
Anodonta anatina (L.)	400		
Sphaerium corneum (L.)	10%	75%	100%
Sphaerium lacustre (Muller)			
Sphaerium rivicola (Lamarck)			
Pisidium amnicum (Muller)			
Pisidium casertanum (Poli)		25%	
Pisidium henslowanum (Sheppard)		13%	
Pisidium nitidum Jenyns		25%	
Pisidium supinum Schmidt		13%	
Lumbriculidae			
Ophidonais serpentina (Muller)			
Nais simplex Piguet	10%		
Stylaria lacustris (L.)	50%	38%	i
Tubificidae	10%	25%	100%
Limnodrilus cervix Brinkhurst			
Limnodrilus hoffmeisteri Claparede			
Limnodrilus udekemianus Claparede]
Psammoryctides barbatus (Grube)	<u> </u>]
Potamothrix hammoniensis (Michaelsen)]
Potamothrix moldaviensis (Vejdovsky & Mrazek		38%	100%
Aulodrilus pluriseta (Piguet)		13%	

	<u> </u>		
		Pot.pectinatus	
		ina	
	ᇣ	သွ	52
_	Nuphar	ç	emna
Taxon	<u> Ž</u>	<u> </u>	<u> </u>
Theromyzon tessulatum (Muller)	40%		
Hemiclepsis marginata (Muller)		13%	
Glossiphonia complanata (L.)		25%	100%
Glossiphonia heteroclita (L.)			
Helobdella stagnalis (L.)	10%	25%	100%
Erpobdella sp.	20%	13%	
Erpobdella octoculata (L.)	20%		
Hydracarina	60%	63%	100%
Argulus sp.		25%	100%
Asellus aquaticus (L.)	10%	63%	100%
Asellus meridianus Racovitza			
Corophium curvispinum Sars		13%	
Crangonyx pseudogracilis Bousfield	20%	13%	
Gammaridae	1		
Gammarus pulex (L.)	10%	50%	
Baetis rhodani (Pictet)	1.072		
Baetis vernus Curtis	10%	38%	
Baetis scambus group	50%	50%	
Centroptilum luteolum (Muller)	50%	88%	100%
	30%	88%	100%
Cloeon dipterum (L.) Cloeon simile Eaton	10%	63%	
			1000
Procloeon bifidum Bengtsson	70%	50%	100%
Ephemerella ignita (Poda)	10%	13%	
Caenis sp.	10%	500	
Caenis luctuosa group		13%	
Platycnemis pennipes (Pallas)	100		
Coenagriidae	10%		
Calopteryx virgo (L.)	10%		
Notonecta sp.			
Corixidae	10%		
Micronecta sp.	_}-		
Sigara (Sigara) sp.	<u> </u>	100%	100%
Sigara (Subsigara) distincta (Fieber)		25%	
Sigara (Subsigara) falleni (Fieber)		50%	100%
Haliplidae		50%	
Dytiscidae		63%	100%
Hydroporus sp.		13%	
Hydrophilidae		13%	
Dryops sp.			
Oulimnius tuberculatus (Muller)	10%	13%	
Sialis lutaria (L.)			
Hydroptilidae	1	13%	
Hydroptila sp.	10%	13%	
Cyrnus flavidus Mclachlan	30%		
Cyrnus trimaculatus (Curtis)	30%		
Neureclipsis bimaculata (L.)	10%	-	
Plectrocnemia sp.	10%		
Plectrocnemia geniculata Mclachlan	10%		\
B	1.0,0		<u> </u>

Taxon	Nuphar	Pot. pectinatus	emna
Taxon	50%	13%	<u> </u>
Polycentropus sp.		13%	
Phryganea sp.	10%		
Anabolia nervosa (Curtis)			
Molanna angustata Curtis		13%	100%
Leptoceridae			<u> </u>
Ceraclea sp.			
Mystacides sp.		 	ļ .
Mystacides longicornis (L.)		13%	
Mystacides nigra (L.)			
Diptera		38%	<u> </u>
Simulium (Boophthora) erythrocephalum (de Ge	10%		
Clinotanypus nervosus (Meigen)			[
Proctadius sp.			
Ablabesmyia sp.	20%	25%	
Thienemannimyia group	10%	13%	
Prodiamesa olivacea (Meigen)		13%	100%
Brillia sp.	10%		
Cricotopus sp.	50%	100%	100%
Cricotopus (Cricotopus) sp.	10%		
Eukiefferiella sp.		4	
Nanocladius sp.	20%		
Orthocladius sp.	10%	50%	100%
Synorthocladius semivirens (Kieffer)		30,70	10070
Corynoneura sp.	20%		
Limnophyes sp.			
Thienemanniella sp.	20%		
Chironomus sp.	2070		
Cryptochironomus sp.	10%	13%	100%
Dicrotendipes (Limnochironomus) sp.	40%	38%	10070
Endochironomus sp.	4070	30 %	
Glyptotendipes sp.			
Microtendipes sp.			
Parachironomus sp.	30%		
Paracladopelma sp.	3070		
Paratendipes sp.		25/4	
Phaenopsectra sp.	100	25%	
	10%	507	1000
Polypedilum (Pentapedilum) sp.	10%	50%	100%
Xenochironomus xenolabis (Kieffer)	10%		
Cladotanytarsus sp.			
Micropsectra sp.	10	13%	
Rheotanytarsus sp.	40%	50%	
Tanytarsus sp.	4.0		
	10	8	1

APPENDIX 3.10 The occurrence of each taxon with national conservation status in the macro-invertebrate data-base.

03-Apr-97

Species Gyraulus acronicus (Ferussac) has notability code(s) RDB 2 and is found in the following samples:

Reach ID	: TH23	Reach name: Sa	ındford				
Site (D	Site name		Easting	Northing	Sample date	Season ID	Sample ID
49	Abingdon Wei	r	450400	197200	28/07/92	2	1
140	Radley		453800	199000	20/07/92	2	t

Species Pisidium moitessierianum Paladilhe has notability code(s) NB and is found in the following samples:

Reach ID:	:TH16	Reach name: North	nmoor				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
144	Bablock Hythe		443500	204200	09/05/84	i	7007
144	Bablock Hythe		443500	204200	09/05/84	1	7008
144	Bablock Hythe		443500	204200	26/07/84	2	7010
144	Bablock Hythe		443500	204200	17/10/84	3	7012
Reach ID	: TH27	Reach name: Day's	5				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
145	Shillingford		459000	193200	09/05/84	1	7013
145	Shillingford		459000	193200	17/10/84	3	7019

03-Apr-97

Species Pisidium supinum Schmidt has notability code(s) NB and is found in the following samples:

Reach ID	: TH10	Reach name: St	lohns				
Site ID	Site name	teach mane, or,	Easting	Northing	Sample date	Season ID	Sample ID
143	Malthouse		422500	198400	09/05/84	1	7001
143	Malthouse		422500	198400	26/07/84	2	7004
143	Malthouse		422500	198400	17/10/84	3	7004
143	Malthouse		422500	198400	17/10/84	3	7006
Reach ID	Sec. 2	Reach name: Bus			1	-	7000
Site (D	Site name		Easting	Northing	Sample date	Season ID	Sample ID
37	Buscot		•		15/08/77	2	2199
37	Buscot				15/08/77	2	2196
37	Buscot				15/08/77	2	2194
Reach ID	: TH14	Reach name: Rusi	hey				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample (D
27	Rushey				02/08/77	2	14
Reach ID	:TH15	Reach name: Shif	ford				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
26	Shifford			-	02/08/77	2	13
36	Shifford				22/08/77	2	1204
Reach ID	:TH16	Reach name: Nort	hmoor				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
144	Bablock Hythe		443500	204200	09/05/84	1	7008
144	Bablock Hythe		443500	204200	26/07/84	2	7010
144	Bablock Hythe		443500	204200	17/10/84	3	7011
Reach ID:	:TH18	Reach name: Eyns	sham				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
35	Eynsham				24/08/77	2	1213
35	Eynsham				24/08/77	2	3214
35	Eynsham				24/08/77	2	2219
Reach ID:		Reach name: King	•				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
22	King's				25/08/77	2	9
Reach ID:		Reach name: Gods					
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample 1D
21	Godstow				28/07/77	2	8
Reach ID:		Reach name: Iffley					
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
34	lffley				30/08/77	2	3220
34	Iffley				30/08/77	2	1229
34	Iffley				30/08/77	2	3222
34	Iffley				30/08/77	2	3224
34	lffley				30/08/77	2	3229
34 34	Iffley				30/08/77	2	2225
24	Ifficy				30/08/77	2	3228

Notable taxa recorded within database	Notable	taxa	recorded	within	database
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$03-A_{i}$	pr.S	27
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0.5 mg 2	·	- · - · · · · · · · · · · · · · · · · ·					
Reach [I	D: TH23	Reach name: Sa	ndford				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
18	Sandford			_	27/07/77	2	5
Reach [[): TH27	Reach name: Da	ıy's				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample (D
145	Shillingford		459000	193200	09/05/84	1	7013
145	Shillingford		459000	193200	09/05/84	1	7015
145	Shillingford		459000	193200	26/07/84	2	7016
145	Shillingford		459000	193200	17/10/84	3	7018
		fuscogrisea (F	Retzius) ha	ıs notabili	ty code(s) N	and is fou	nd
	ollowing san	•	_ •				
Reach ID Site ID	Site name	Reach name: Da	-	N- 41.			
			Easting	_	Sample date		-
145	Shillingford		459000	193200	09/05/84	1	7013
		ulgatissimus (l	L.) has no	tability co	de(s) N and	is found ir	1
Reach (D	owing sampl			,		•	
	Site name	Reach name: No				C 70	
			Easting	Northing	•		-
144 [44	Bablock Hythe Bablock Hythe		443500	204200	26/07/84	2	7010
Reach ID	•	e Reach name: Cli	443500	204200	17/10/84	3	7012
Site ID	Site name	Reach name: Ch		Nouthing	Cample data	Caran ID	Cample ID
133			Easting	Northing	-		Sample ID
Reach ID	Clifton Bridge	Reach name: Da	454700	195400	03/08/92	2	I
Site ID	Site name	Reacti traine: Da	ys Easting	Northina	Cample data	Canana ID	Camala ID
145			•	Northing	Sample date		Sample 1D
145	Shillingford Shillingford		459000	193200	26/07/84	2	7017
Reach ID	=	Reach name: Go:	459000	193200	17/10/84	3	7019
	Site name	Reacti traine. 00.	Easting	Northing	Sample date	Sangan III	Samula ID
55	Whitchurch W	/eir	463300	176800	23/05/90	i	6600
S:						_	
the follo	nanpius ian wing sampl	ninatus Schalle es:	r has nota	ability cod	e(s) NB and	is found i	n
Reach ID	•	Reach name: Cul	lham				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample 1D
	I 37774 1						•
134	Long Wittenha	am	454000	193700	03/08/92	2	Ī

Species Gyrinus distinctus Aube has notability code(s) RDB 3 and is found in the following samples:

Reach ID	: TH27	Reach name: Day's					
Site ID	Site name	E	Casting	Northing	Sample date	Season ID	Sample ID
145	Shillingford	4	59000	193200	09/05/84	1	7013

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Species Gyrinus urinator Illiger has notability code(s) NB and is found in the following samples:

Reach ID: TH10

Reach name: St. Johns

Site ID Site name Easting Northing Sample date Season ID Sample ID

143 Malthouse 422500 198400 17/10/84 3 7005

Reach ID: TH22 Reach name: Iffley

Site ID Site name Easting Northing Sample date Season ID Sample ID
48 Top of Sandford Lock Cut 452800 202100 08/10/90 3 7900

Species Anacaena bipustulata (Marsham) has notability code(s) NB and is found in the following samples:

Reach ID: TH22

Reach name: Iffley

Site ID Site name Easting Northing Sample date Season ID Sample ID
48 Top of Sandford Lock Cut 452800 202100 09/07/92 2 1

Species Laccobius (Macrolaccobius) sinuatus Motschulsky has notability code(s) NB and is found in the following samples:

Reach ID: TH22 Reach name: Iffley

Site ID Site name Easting Northing Sample date Season ID Sample ID

48 Top of Sandford Lock Cut 452800 202100 09/07/92 2 1

Reach ID: TH26 Reach name: Clifton

Site ID Site name Easting Northing Sample date Season ID Sample ID

132 Days Reach 456600 195500 28/07/92 2 1

Species Oulimnius major (Rey) has notability code(s) NA and is found in the following samples:

Reach ID: TH19 Reach name: King's

Site ID Site name Easting Northing Sample date Season ID Sample ID
44 Trout Inn, Godstow 448300 209200 09/08/90 2 7600

03-Apr-97

Species Sialis nigripes Pictet has notability code(s) NB and is found in the following samples:

Reach ID	: TH10	Reach name: St. Jo	ohns				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
143	Malthouse		422500	198400	09/05/84	1	7000
143	Malthouse		422500	198400	09/05/84	1	7001
143	Malthouse		422500	198400	26/07/84	2	7004
143	Malthouse		422500	198400	17/10/84	3	7006
Reach ID	:TH11	Reach name: Busc	:01				
Site ID	Site name		Easting	Northing	Sample date	Season 1D	Sample ID
37	Buscot				15/08/77	2	2191
37	Buscot				15/08/77	2	2190
Reach ID	:TH16	Reach name: Norti	hmoor				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample 1D
144	Bablock Hythe		443500	204200	26/07/84	2	7010
144	Bablock Hythe		443500	204200	17/10/84	3	7012
Reach ID	:TH18	Reach name: Eyns	ham				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
35	Eynsham				24/08/77	2	2219
35	Eynsham				24/08/77	2	2218

Species Ceraclea senilis (Burmeister) has notability code(s) N and is found in the following samples:

Reach ID	: TH23	Reach name: Sand	dford				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
124	Radley College	Boathouse	453800	198800	30/06/92	2	5006
Reach ID	:TH30	Reach name: Gori	ing				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
55	Whitchurch W	eir	463300	176800	23/05/90	1	6600

Species Leptocerus lusitanicus (Mclachlan) has notability code(s) RDB 2 and is found in the following samples:

Reach ID	:TH30	Reach name: Gori	ng				
Site ID	Site name		Easting	Northing	Sample date	Season ID	Sample ID
55	Whitchurch W	Veir	463300	176800	23/05/90	1	6600

APPENDIX 4.1 The executive summary and key supporting tables and figures from references cited in Chapter 4: Macrophytes of the main report.

BUNDED RESERVOIR STORAGE STUDIES 9701/B5

RIVER CORRIDOR, PHASE I AND PHASE 2 SURVEYS: RIVER THAMES

EXECUTIVE SUMMARY

Ecosurveys Ltd has been contracted by Thames Water Utilities Ltd to undertake a River Corridor survey, and Phase 1 and Phase 2 surveys on the Thames Floodplain between Abingdon Lock and Benson Lock, Oxfordshire. This study is part of the SWORDS Project and is intended to provide a baseline understanding of the nature conservation interest of this reach of the River Thames and its floodplain, such that it can help;

- identify any impacts on the catchment which could potentially arise from the reservoir operation and the construction of abstraction and discharge structures:
- the design of a least damaging development proposal and operating regime;
 and
- bring forward recommendations for enhancement of the river corridors, particularly around and adjacent to any proposed abstraction and discharge structures.

The study has three integrated components:

- a River Corridor Survey on approximately 30km of statutory main river;
- a Phase 1 Ecological Land-use Survey of the designated Thames floodplain;
 and
- detailed Phase 2 botanical surveys of water level dependant habitats of conservation interest in the designated Thames floodplain.

The field survey for the various components was undertaken between May and August 1992, according to standard NRA and English Nature methodologies.

The River Corridor Survey shows the Thames to be a broad deep navigable river, with its water level maintained by locks and weirs. The large amount of boat traffic precludes extensive growth of channel vegetation, which is therefore virtually limited to strands of tall marginal species. The river banks are of varied profile and are frequently tree and shrub lined, with tall herbs and coarse grasses underneath and where open.

The River Corridor sections were assessed according to the London Ecology Unit River Reach Evaluation Criteria. The majority of the 61 sections were graded good, with four graded poor, eight graded important and one graded critical on account of the adjacent Culham Brake, a wetland site of Special Scientific Interest.

The Phase I Survey reveals the designated floodplain of the Thames between Abingdon Lock and Benson Lock to be an intensively managed agricultural area, largely arable land with some semi-improved and improved grasslands, managed as pasture and hayfields. There

consequently is a generally low level of nature conservation interest; that which exists does so as rather isolated habitats (eg including marshy pastures, open standing waters, swamps, small woodlands supporting the nationally rare summer snowflake, but otherwise unremarkable) scattered at intervals along the floodplain, separated by large tracks of arable land.

16 sites for detailed Phase 2 botanical survey were identified by the Phase 1 survey. These include flushes, marshy pastures, swamps, open water sites, wet woodlands and tall herb marshes. Plant communities identified by the Phase 2 surveys were 2 open water types, 3 woodland types, 10 swamp types and 4 mesotrophic grassland types. With the exception of the flood meadow habitat, MG 4, on site 9 (Long Wittenham Nature Reserve), all are common and widespread in lowland Britain; the flood meadow community is of local occurrence in Britain but this site holds a slightly modified example, especially when compared to the large areas further upstream (out of this Study Area) around Oxford. No nationally or regionally rare plant species were recorded by the Phase 2 surveys, which took place too late in the year for confirmation of the summer snowflake records for sites 15 and 16. The Phase 2 survey sites' general wetland species interest was identified to be vulnerable to agricultural improvements and lack of management, in addition to permanent and marked reductions in the mean water level of the Thames, but as long as the river periodically floods, the seed dispersal requirements of the summer snowflake will be met.

SUMMARY

RIVER THAMES: ABINGDON LOCK TO BENSON LOCK

INTRODUCTION

Ecosurveys Ltd has been contracted by Thames Water Utilities Ltd to undertake, as part of the SWORDS Project in Oxfordshire:

- a River Corridor Survey on 30km of statutory main river of the River Thames between Abingdon Lock and Benson Lock;
- a Phase 1 Ecological Habitat Survey of the designated Thames Floodplain between Abingdon Lock and Benson Lock; and
- detailed Phase 2 botanical surveys of water level dependent habitats of conservation interest in the designated floodplain.

The field survey was undertaken in June, July and August 1992 using standard NRA and English Nature methodologies. The results of the survey are summarized below and are detailed in the report, which comprises this volume (river corridor survey data), a volume of Phase 1 and Phase 2 survey data and a volume of additional photographs.

RIVER CORRIDOR HABITATS

The River Thames between Abingdon Lock and Benson Lock is a broad, deep, navigable river, gently meandering and with its water level maintained by locks and weirs. The river carries a large amount of boat traffic, especially in summer. The flood plain is largely arable land with semi-improved and improved pastures, incorporating the occasional woodland, marsby grassland and built-up area.

The river banks, composed of clay, sand and gravel (and locally artificially reinforced), are generally well vegetated and of varying profile, from vertical cliffs to gently graded sections. Trees and shrubs often line the river, crack willows (as standards, old coppice or pollards) are especially frequent, with hawthorn, alder (often as old coppice), ash, elm, field maple, oak and sycamore. Various ornamental trees occur through built-up areas and other less common native shrubs present are spindle, guelder-rose, dogwood and buckthorn, in addition to shrub willows: grey, goat and osier. Tall herbs and coarse grasses dominate the herbaceous vegetation under the trees, and where open, and it is generally unremarkable and comprises ubiquitous species but with the addition of wild onion, escaped Brassica and introduced Aster species. Grazed sections of bank are grass dominated. Crevices in artificial banks support small populations of pellitory-of-the-wall and stonecrops.

Marginal vegetation is generally tall and dominated by great willowherb, comfrey, meadow-sweet, bard rush, bittersweet and purple loosestrife. Also present are soft rush, water chickweed, hemp agrimony, marsh yellow-cress, water figwort and false fox sedge. Gently graded banks support gipsywort, water mint, brooklime, marsh woundwort, water forget-menot, watercress and fool's watercress. Other less common marginal species include yellow loosestrife, skullcap, ragged robin and Himalayan Balsam, the latter especially near built-up areas. The marginal vegetation frequently grades into the emergent aquatics, except where precluded by shade, water depth and boat moorings. Common species in the channel are branched bur-reed, common club rush, sweet flag, reed canary grass, greater pond sedge, reedmace and common reed, forming species-pure or mixed stands, often showing zonation from shallower to deeper water. Also present are flowering rush, great water dock, common water plantain, narrow-leaved water plantain, yellow water lily and unbranched bur-reed (the latter two suffering from damage caused by boat traffic).

TARGET SPECIES

The River Corridor Survey of the River Thames included an element of searching for, and acurately recording the location of, a number of Target Species, which are those in any of the following three categories (from Palmer & Newbold, 1983, Wetland & Riparian Plants in Great Britain, NCC Focus on Nature Conservation No 1):

- Aquatic plants recorded from 100 or fewer 10x10km squares in Great Britain and which need special-protection in the NRA Thames Region;
- Riparian and non-aquatic wetland plants recorded from 100 or fewer 10x10km squares
 in Great Britain and which need special protection in the NRA Thames Region; and
- Aquatic plants recorded from more than 100 10x10km squares in Great Britain but which need special protection in the NRA Thames Region.

No Target Species were recorded from the River Thames River Corridors.

CONSERVATION MANAGEMENT

The management of the riparian tree cover is the principal target of management of the existing wildlife habitat resource. Single, and lines or groups of, old pollarded and coppiced trees are important in wildlife and landscape terms and management recommendations are made for repollarding and recoppicing. A certain number of trees should be worked each year out of the total resource, in order to provide variety of structure and age of regrowth. This will also act as a precautionary measure, in case the trees do not respond to management, so avoiding the situation of having a river bank lined with dead stumps and pollards.

As a spin-off from pollarding and coppicing, the increased light let onto the river bank should promote the growth and spread of channel vegetation, so protecting the bank from erosion by boat wash. Fallen trees are to be removed unless it can be demonstrated that they function as Kingfisher perches and do not significantly impede flow. Occasional standing dead trees, unless a danger, are best left in situ for woodpeckers and invertebrates.

A second target of conservation management is tall wetland vegetation and relatively speciesrich drier neutral grassland in the floodplain adjacent to the river. These areas are typically unmanaged and with time, this leads to loss of species diversity and succession to different habitats of lower conservation interest. To prevent this and hence to maintain speciesrichness, the initiation of suitable mowing regimes is required, combined with the removal of cut material and eradication of invading scrub.

A third and equally simple conservation management operation involves thinning out of trees and shrubs along the south sides of channels, to let in more light to the water's edge to promote the growth of channel vegetation.

HABITAT ENHANCEMENT

The principal opportunities for habitat enhancement centre on the areas of tall wetland vegetation, habitats which are important in their own right and are of value for birds and invertebrates. The recommendations are to increase the extent of these areas by excavating the surrounding ground down to, or near, water level, enabling the wetland species to spread.

Areas of ruderal vegetation, unmanaged arable or pasture represent opportunities for the planting of native tree and shrub species, especially where adjacent to existing areas of woodland.

PHASE 1 SURVEY

The designated floodplain of the River Thames between Abingdon Lock and Benson Lock extends for 20km, covers approximately 15km² and varies in width from 400m to over 1000m.

The survey reveals that the floodplain is predominantly intensively agricultural - dominated by arable land with occasional improved and semi-improved pastures, built-up areas, woodlands (ornamental, broadleaved, mixed, coniferous), amenity grasslands, tall herbs, caravans and mineral workings.

PHASE 2 SURVEY

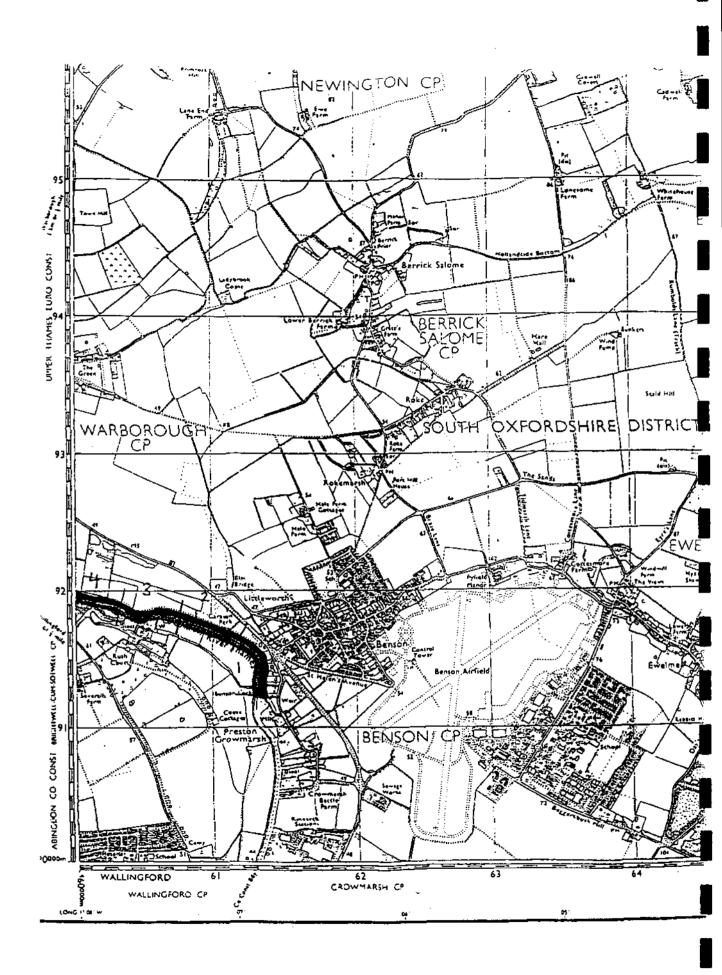
Fourteen sites were identified by the Phase 1 survey for detailed Phase 2 surveys, along with one site notified by a landowner which was deemed to meet the criteria for Phase 2 survey, and a second additional site notified by Thames Region NRA, giving a total of 16 sites. Eleven sites were surveyed in late June/early July with three of these revisited in August, and five were surveyed only in August. The habitats surveyed included willow cars and other woodlands, tall herb swamps, old gravel pits and other open water bodies, inundation grasslands and marshy pastures.

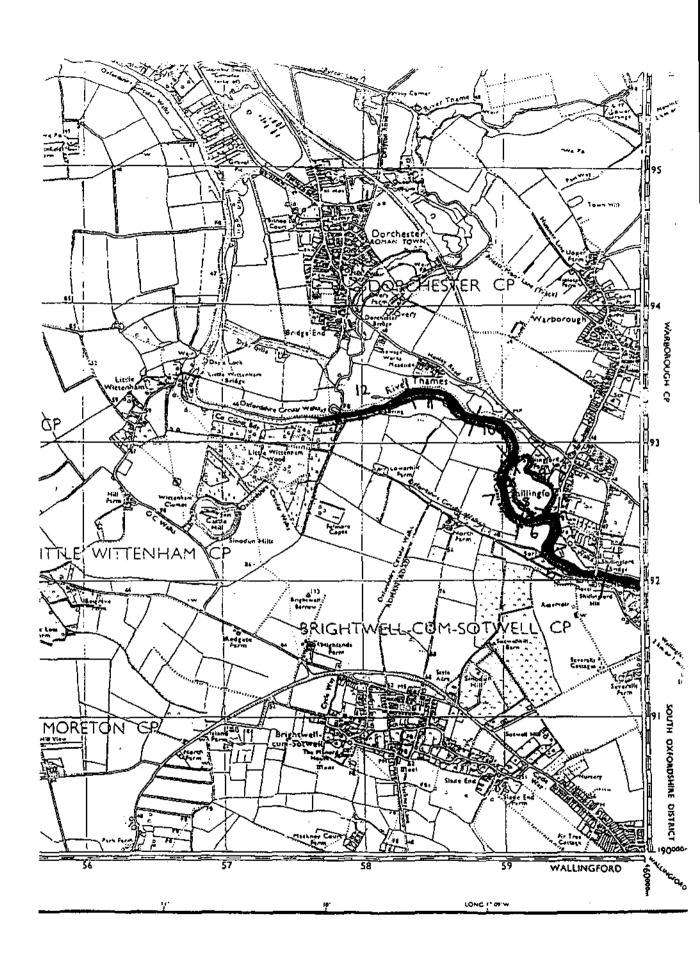
The majority of sites supported only one or two water level dependent plant community types, with one diverse tall herb swamp holding four and one large mosaic of habitats holding six. With the exception of Site 9 (a relatively herb rich flood meadow), none of the plant communities identified were those other than common types known to be widespread in lowland Britain and typical of wetlands in general. No rare species were recorded although two of the sites (15 and 16) are reliably reported to be localities for a nationally rare plant, probably not seen in the present survey because of the time of year.

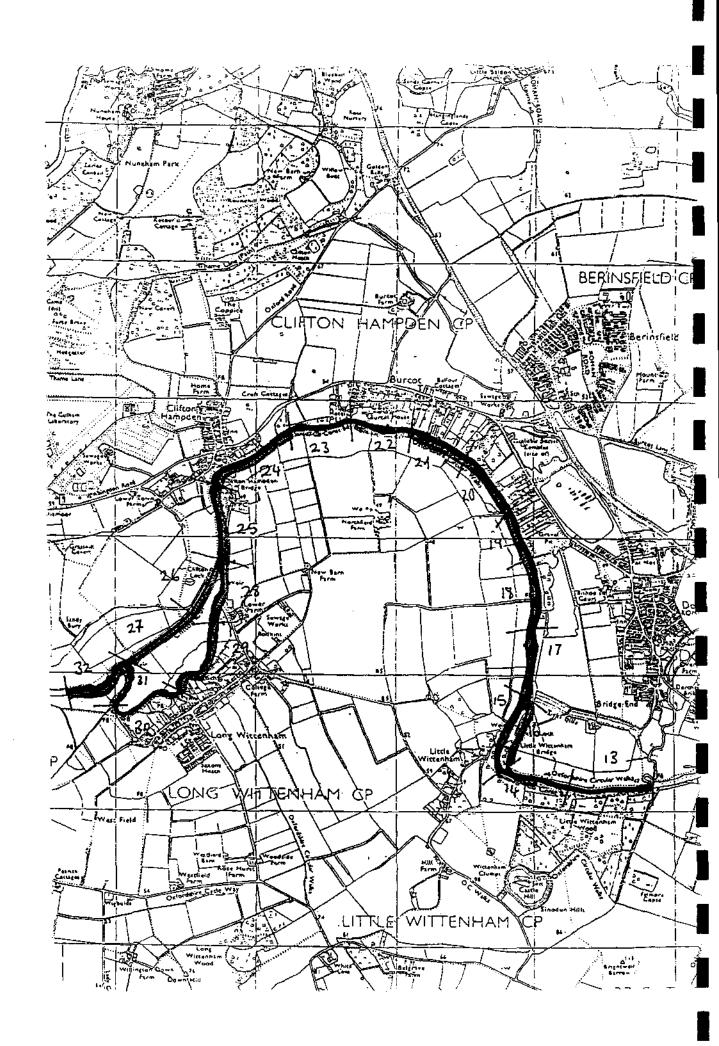
RIVER THAMES SECTION EVALUATION

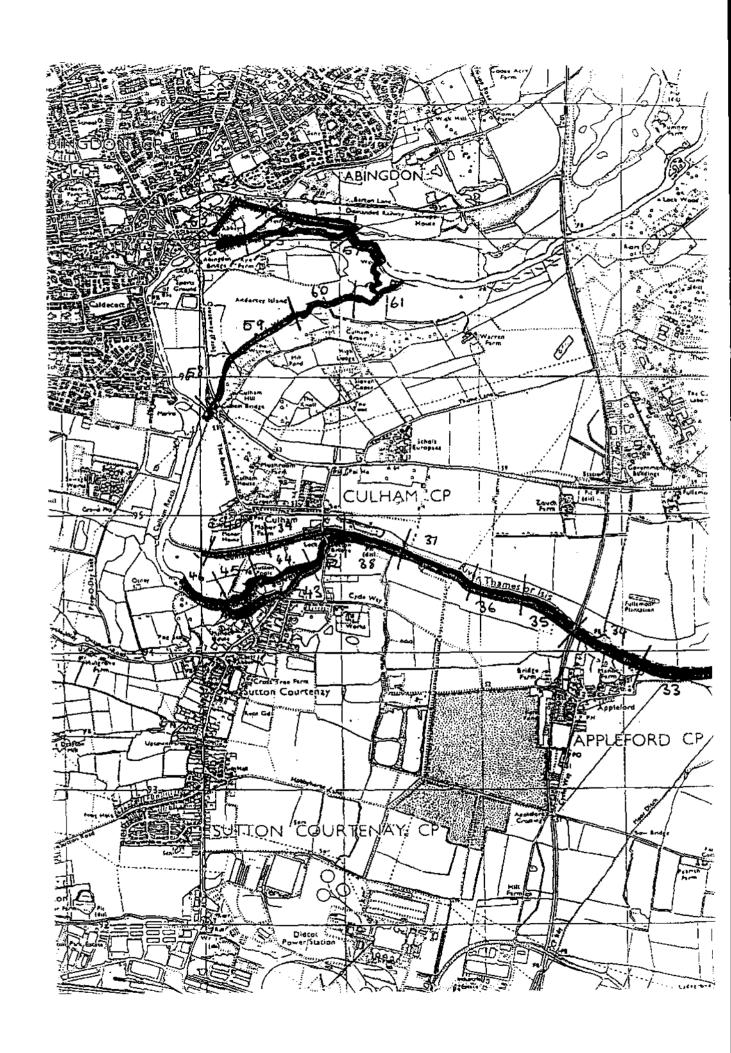
SECTION	CRITICAL	IMPORTANT	GOOD	POOR
001 002 003 004 005			4 4	1
006 007 008 009 010			/ / / / / /	
011 012 013 014 015		✓ 1 ✓ 2	1	
016 017 018 019 020			<i>y y y y</i>	
021 022 023 024 025		√3 √4	1	/
026 027 028 029 030	•	✓ 5	<i>y y y y</i>	
031 032 033 034 035		√ 6	<i>y y y y</i>	
036 037 038 039 040			1 1 1	

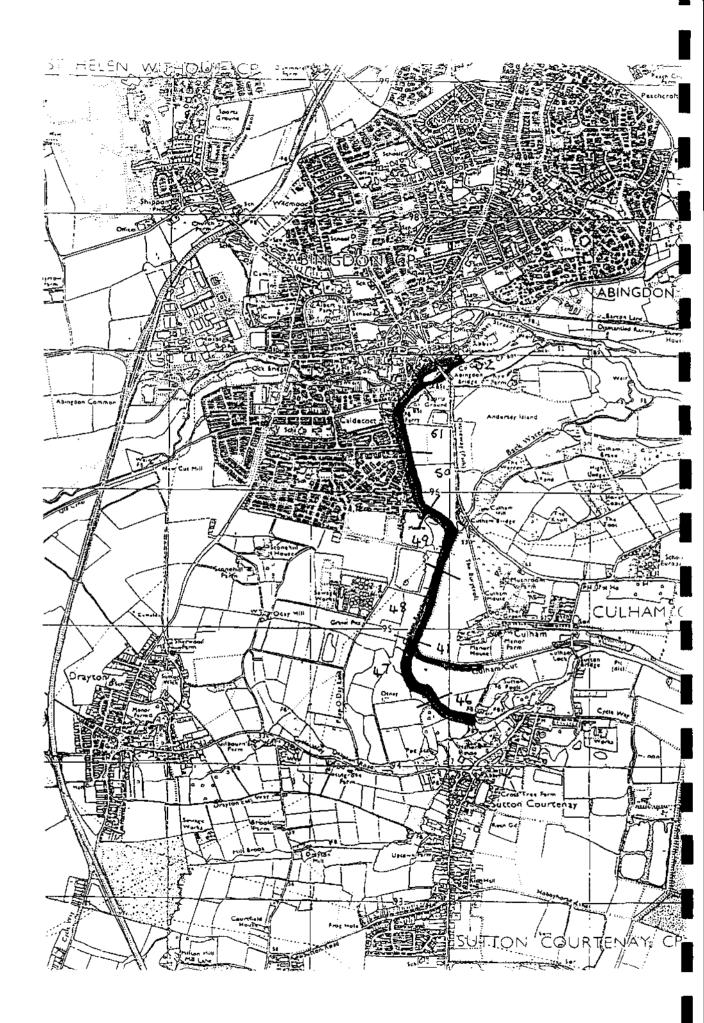
SECTION	CRITICAL	IMPORTANT	GOOD	POOR
041			/	
042	}		1	
043			1	
044			1	
045			/	
046			1	
047			1	
048 049		17	}	ļ
050		√8]
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051			1	}
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060	1 9			
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OXFORD FLOODPLAIN ENVIRONMENTAL SURVEY 1992

EXECUTIVE SUMMARY

Ecosurveys Ltd has been contracted by the National Rivers Authority Thames Region to undertake a study of the nature conservation interest of a part of the floodplain of the River Thames between Eynsham and Sandford-on-Thames, Oxfordshire. This study is part of a modelling process providing information on floodplain envelopes and flow routes. It is to supply baseline information to the NRA to help them assess the effects of any proposed modification in the operation of flood control structures and to optimise their operation.

The study has several integrated components:

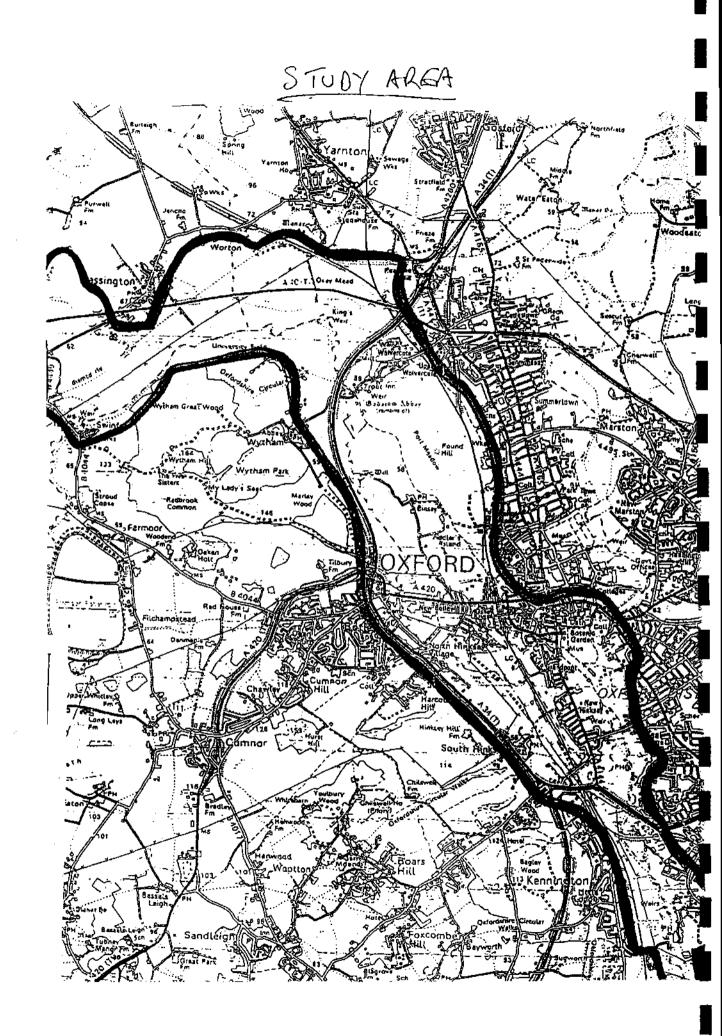
- a Desk Study to collate and review information on areas of known conservation importance in the Study Area;
- a River Corridor Habitat Survey of a designated 74km of the River Thames, its tributaries, secondary channels and drains in the Study Area;
- a Phase I Ecological Habitat Survey of about 25 km² of the designated Oxford Floodplain Environmental Survey Study Area; and
- detailed Phase 2 botanical surveys of sites of water-level dependence within the study
 Area for which adequate information does not already exist.

The field survey for the various components was undertaken between May and August 1992 using standard NRA and English Nature methodologies.

The data collection and review, on the whole, yielded information of limited use, due mainly to its brevity. The River Corridor Habitat Survey shows that, in general, the river corridor habitats are evaluated as critical or good, according to the London Ecology Unit Evaluation with the exception of urban Oxford where they are evaluated as poor on the whole. The large number of river corridor sections evaluated as critical is in large part due to their proximity to the extensive SSSI's in the floodplain.

The Phase 1 Survey reveals that the designated floodplain is largely agricultural with the exception of the of the urban area of Oxford and is repeatedly criss crossed by large tree and scrub lined watercourses.

From the Desk Study, River Corridor and Phase I Surveys, 34 sites as requiring Phase 2 survey were identified and each received detailed botanical investigations with one 2 x 2m quadrat being recorded in each water level dependent plant community and with full species lists recorded for those sites.



2.3 APPROACH TO THE STUDY COMPONENTS

2.3 Desk Study - Review of Data on Sites of Nature Conservation Significance in the Study Area

The collection and review of existing information on sites of nature conservation importance was principally carried out at the commencement of the study though continued throughout as further sources, sites or information came to light. Ecosurveys Ltd believe that all sites of nature conservation significance in the Study Area have been identified and have been studied by one, some or all of the data review, River Corridor, Phase 1 and Phase 2 surveys. Whilst we believe that we have collated the vast majority of the available data, some sources were unwilling to release data or advised us that the data we had from elsewhere exceeded their data in terms of quality, and was therefore either not released or collated.

One purpose of the data collection was to produce a document which held in one place the available information, from previous work, on sites of nature conservation significance in the Study Area. Part 5 of this report constitutes that document.

A second purpose was to use the available data to identify known sites and to determine which water-level dependent sites in the Study Area required additional fieldwork due to the lack, antiquity or unreliability of the existing information. The data from the surveys carried out this year was then to be compared with previous information on past site management and nature conservation value.

2.3.2 River Corridor

The designated River Corridor was divided into 148 sections each of approximately 500m length and surveyed between May and August using the standard NRA River Corridor Habitat survey Methodology. The information for each section thus comprises photograph(s), a short written descriptive text and a habitat map. Aquatic vegetation was sampled using a grapnel from the banks of narrower channels and from a boat on the River Thames. Management recommendations to maintain and/or enhance the nature conservation value of the sections are also supplied.

The River Corridor habitat survey results are compiled in Part 2 of this report with the locations of the River Corridor sections shown on the supporting maps and tabulated with channel names and start and end point grid references.

The River Corridor sections have been assessed for their nature conservation importance according to the London Ecology Unit River Reach Evaluation Criteria, see Appendix Two, and this information is discussed and analysed in this document (Part 1) along with the general management recommendations required to maintain and enhance the nature conservation interest of the River Corridor sections.

2.3.3 Phase_L

The purpose of the Phase I survey was to identify habitats within the Study Area, but outside designated SSSI's, which are of value to wildlife. The Phase I survey was undertaken in June 1992 using a standard methodology (NCC 1990) in order to supply a readily accessible visual representation of the entire floodplain in the Study Area. For this reason, the colour coding system rather than alphanumeric was adopted.

The results of this survey (Part 3 of this document) comprise a series of A4 and A3 maps which cover the entire survey area. The maps are supported by target notes which indicate;

- the location of sites deemed to be of nature conservation interest which received a Phase 2 detailed botanical survey,
- features of interest or importance to wildlife which were identified in the review of previous available data, or
- features observed in the course of the fieldwork components, and
- site specific information for researched data on notable hird species or populations breeding within or regularly using the floodplain.

2.3.4 Phase 2

Sites identified from the earlier elements of the study which were described as water-level dependent, at least in part, for which there was only old, or unreliable, or no available information and which were apparently of nature conservation interest, received a detailed botanical survey equivalent to NCC (now English Nature) Phase 2 survey level, (Smith et al., 1985).

RESULTS OF REVIEW OF EXISTING INFORMATION

In the event, the data collection yielded information of rather variable quality, which for most sites was rather brief, and the data gathered by Ecosurveys Ltd fieldwork greatly exceeding the previous data in terms of the time spent on the site and the detailed botanical information gathered. In consequence a formalised review of the previous data was not carried out as no sensible comparison could be made for the majority of the sites. However, the information from the data collection has been analysed and integrated into the report as a whole in the form of Target Notes on the Phase 1 maps and as part of the Conservation Assessment of the Phase 2 site reports. Thus, all of the available previous information on a site may be found in Part 5 of the report, as an abridged version in a Target Note on the Phase 1 map and list, and possibly analysed within a Phase 2 site report, with all information being referenced eg to its corresponding Target Note or previous survey data, for easy accessibility.

4. RIVER CORRIDOR SURVEY RESULTS AND DISCUSSION

4.1 INTRODUCTION

The results of the River Corridor Habitat Survey are compiled in Part 2 of this report, in three volumes. The results for each section consist of a brief descriptive text of the adjacent land, bank and channel vegetation, photographs, cross sections and management recommendations.

The 148 River Corridor Habitat Survey sections studied comprise a considerable resource for wildlife at least in a local context. The watercourses vary from large slow flowing deep rivers such as the River Thames, through fast flowing streams to narrow drains providing habitat for a wide variety of wildlife. Generally the various watercourses are largely tree and scrub lined and the adjacent habitats include a complex mosaic of urban, arable and pasture land, tall herb marshes and swamps and woodland. Despite their proximity to Oxford, many of the watercourses outside the urban area are relatively undisturbed, a factor which enhances their value to wildlife. A notable exception, however, is the River Thames which, in particular with regard to marginals and aquatics, appears to be rather poorer in species than would otherwise be expected principally due to the large amount of boat traffic especially in summer and its associated problems with wash, and sediment disturbance etc. It is, however, still of high local importance as a wildlife resource.

The watercourses of the floodplain which were studied can be conveniently but artificially grouped as large rivers, smaller rivers and streams and drains. The table overleaf provides information on the evaluation of the ecological interest of the River Corridor sections according to the London Ecology Unit Ecological Evaluation Criteria for River Reaches which are attached in Appendix Two.

The categories are as follows:

<u>Critically Important for Wildlife</u> - The most ecologically fragile sections of river or adjacent habitats which depend on the maintenance of the present hydrological conditions. This category includes most examples of regionally rare habitats and sections with significant communities of rare flora and fauna.

Important for Wildlife - Sections of high wildlife value but not necessarily closely associated with the river, however potentially vulnerable, and typically composed of either a mixture of well structured habitat types but with few species of particular note or a good example of a single habitat type.

Good for Wildlife - Sections which provide particular local interest and typically comprising habitats of a more robust nature less vulnerable to, and possibly restorable after, engineering works.

Poor for Wildlife - Sections of river and adjacent land with little semi-natural vegetation and of low intrinsic wildlife interest. Enhancement opportunities may present themselves.

Most "typical" River Corridor sections would thus fall into the category of good and in the following table of evaluated sections justifications are only noted for critical, important and poor evaluations with supplementary notes.

The evaluations of the River Corridor Sections were carried out by two members of Ecosurveys Ltd technical staff who had been involved in the project and were familiar with the Study Area (Tim Harvey and Tim Smith). The evaluations are a subjective judgement arising from a assessment of each sections conservation value represented by its habitats and physical features. The columns are colour coded to match the following series of maps with:

Critical = Red
Important = Green
Good = Blue
Poor = Brown

TABLE 1

RIVER CORRIDOR NATURE CONSERVATION EVALUATION BASED ON

L.E.U EVALUATION CATEGORIES

SECTION		C	I.	G	P	NOTES
No	Name		i-,			
1	River Thames			1		Create berm on the left in midsection.
2	Sandford Ditch	\				Hottonia palustris locality. Repollarding of shading trees.
3	River Thames			1		
4	River Thames			1		
5	River Thames	1				Fiddler's Elbow Island adjacent. (Phase 2 Site 2).
6	River Thames	1				Adjacent to Rose Isle. (Phase 2 Site 4).
7	River Thames	s				Unimproved wet meadow adjacent to Heyford Hill Lane Pasture. (Phase 2 Site 5).
8	River Thames		-	1		
9	River Thames	1				Adjacent to Fiddler's Elbow Island. (Phase 2 Site 2).
10	River Thames	1				Adjacent to Wetland and Woodland south of Iffley Meadows SSSI. (Pasture 2 Size 10).
11	River Thames	1				Adjacent to SSSI, but channel poor.
12	River Thames	1				Adjacent to SSSI, but channel poor.
13	River Thames	1				Adjacent to SSSI, but channel poor.
14	River Thames	1				Adjacent to SSSI.
15	River Thames	1				Adjacent to SSSI.
16	River Thames				1	Artificial. Boat traffic. Plant trees upstream on right. Limited scope for enhancement.
17	River Thames				1	Artificial. Mostly urban. Plant trees and shrubs. Limited scope for enhancement.
18	River Thames				1	Artificial. Urban. Plant trees and shrubs. Little scope for enhancement.
19	Castle Mill Stream			1		Urban. Create riffle and pool.
20	Castle Mill Stream				1	Artificial: Urban. Increase tree and shrub cover. Limited scope for enhancement.
21	Castle Mill Stream			1		Poor channel, but wooded. Limited scope for enhancement.

SECT	ION	c	ĩ	G-	٩	NOTES
No	Name					
2 2	Castle Mill Stream			1		Poor channel, but wooded. Limited scope for enhancement.
23	Railway Drain			1		
24	Railway Drain			1		
25	Railway Drain			1		
26	Port Meadow Drain	1				Adjacent to SSSI
27	Port Meadow Drain	>				Adjacent to SSSI.
28	Port Meadow Drain	>				Adjacent to SSSI.
29	Port Meadow Drain	1				Adjacent to SSSI
30	Port Meadow Drain	\$				Adjacent to SSSI.
31	Port Meadow Drain	V				Adjacent to SSSI.
32	Port Meadow Drain	1				Adjacent to SSSI.
33	Hinksey Stream	1				Adjacent to SSSI
34	Weirs Mill Stream	1				Adjacent to SSSI.
35	Weirs Mili Stream	1				Adjacent to SSSI.
36	River Thames	1				Adjacent to SSSI.
37	River Thames			1		
38	New Hinksey Drain			1		Poor as wetland, but wooded.
39	New Hinksey Drain			1		Poor as wetland, but wooded.
40	Shire Lake Ditch			1		Channel poor, good wooded corridor.

SEC	TION	С	I	G-	Ρ	NOTES	
Νo	Name						
41	Shire Lake Ditch	1				Channel poor, good corridor. Ajacent wetland habitat. Jackdaw Lane Marsh (Phase 2 Site 16).	
42	River Cherwell			1		Artificial banks. Heavy use. Adjacent to wet meadow, Christchurch Pasture (Phase 2 Site 15).	
43	Christchurch Drain	1				Adjacent to wet meadow, Christchurch Pasture (Phase 2 Site 15).	
44	Trill Mill Stream				1	Urban - artificial concrete channel. Limited scope for enhancement.	
45	River Cherwell			✓		Wooded. Create berms and profile channel. Adjacent to Christchurch Pasture (Phase 2 Site 15).	
46	Hinksey Stream	1				Adjacent to SSSI.	
47	Hinksey Stream			1			
48	Coldharbour Drain			1		Urban sections - much rubbish in channel.	
49	Coldharbour Drain			1		Urban sections - much rubbish in channel.	
50	Railway Drain		~			Part large pond. (Railway Pond). Contiguous with Sections 051-053.	
51	Railway Drain and Pond		į	1		All one large pond with emergents and aquatics; an undisturbed backwater.	
52	Railway Pond	1				All one large pond with emergents and aquatics; an undisturbed backwater.	
53	Railway Pond	1				All one large pond with emergents and aquatics; an undisturbed backwater.	
54	South Hinksey Drain			\ .		Scallop berm midsection.	
55	South Hinksey Drain			\			
56	South Hinksey Drain				1	Poor as wetland, channel drying out.	
57	Hinksey Stream			1			

SEC	SECTION CIGP				P	NOTES
No	Name				1 1 42 1	
58	Hinksey Stream			1		
59	Hinksey Stream			1		
60	Hinksey Stream			1		
61	Hogacre Ditch			1		
62	Hogacre Ditch				1	Woodland, but poor as wetland.
63	Eastwyke Drain				1	Urban Woodland, but poor as wetland.
64	Eastwyke Drain			1		
65	Eastwyke Drain			V		
66	Hogacre Ditch			1		
67	Hogacre Ditch			1		
68	Hogacre Ditch				1	Urban Poor as wetland, but wooded.
69	Bulstake Stream			•		
70	Bulstake Stream			1		
71	River Thames			1		
72	Mill Stream				1	Heavy use - boats moorings. Some artificial bank. Plant trees.
73	River Thames				1	Urban. Artificial. Includes Lock. Some artificial bank. Plant trees and shrubs.
74	River Thames				1	Urban. Artificial. Plant trees and shrubs.
75	River Thames				1	Urban. Artificial. Create riffles, pools, berms on margins.
76	River Thames				1	Urban. Artificial. Plant trees.
77	Rewley Abbey Stream				1	Urban. Artificial. Abused. Create berm for marginals.
78	Osney Drain			/		Channel poor, but wooded. Rubbish in channel.

SECT	SECTION CI		G	ρ	NOTES	
No	Name	,				
79	Bulstake Stream			1		
80	Bulstake Stream			\		
81	Bulstake Stream			1		
82	Bulstake Stream				V	Create pools, deepen channel, plant trees and shrubs.
83	River Thames			1		Boats. Wooded.
84	Fiddler's Island Stream			1		Wooded.
85	Fiddler's Island Stream	V				Adjacent to SSSI.
86	River Thames			1		
87	Fiddler's Island Stream	∢				Adjacent to SSSI.
88	River Thames	1				Adjacent to SSSI.
89	River Thames	1				Adjacent to SSSI.
90	River Thames	1				Adjacent to SSSI.
91	River Thames	1				Adjacent to SSSI.
92	River Thames	1				Adjacent to SSSI.
93	River Thames	1				Adjacent to SSSI.
94	River Thames			1		
95	River Thames	1				Adjacent to SSSI.
96	River Thames	1				Adjacent to SSSI
97	Duke's Cut			1		
98	River Thames	1			<u> </u>	Adjacent to SSSI.
99	River Thames			1		
100	River Thames	1				Adajcent to SSSI.
101	River Thames	1				Adjacent to SSSI.
102	River Thames	1			<u> </u>	Adjacent to SSSI.
103	River Thames	1				Adjacent to SSSI.
104	River Thames	1				Adjacent to SSSI.

SECT	TON	С	ユ	G	ρ	NOTES
No	Name					
105	River Thames	1				Adjacent to proposed extension to SSSI.
106	River Thames		<u> </u>			
107	River Thames			1		
108	River Thames	1				Adjacent to SSSI.
109	River Thames	1				Adjacent to SSSI.
110	River Thames			1		
111	River Evenlode			1		Plant trees and shrubs on north bank.
112	River Evenlode			✓		
113	Old Canal			1		Potential Otter holt site.
114	Wharf Stream			\		
115	Wharf Stream			1		
116	Weir Stream			✓		
117	Seacourt Stream			1		
118	Seacourt Stream			1		
119	Seacourt Stream			1		
120	Seacourt Stream			✓		Plant trees and shrubs on parts of right bank.
121	Seacourt Stream			1		Create riffle and pool system.
122	Seacourt Stream			*		
123	Seacourt Stream			1		
124	Seacourt Stream			V		
125	Seacourt Stream			V		
126	Seacourt Stream			1		
127	Seacourt Stream			1		

SEC	TION	C	ユ	G	P	NOTES
No	Name					
128	Seacourt Stream			1		Create riffle and pool system.
129	Seacourt Stream	1				Adjacent to SSSI.
130	Botley Ditch				1	Poor as wetland/watercourse.
131	Botley Ditch			1		
132	Botley Ditch			1		
133	Binsey Drain			1		
134	Binsey Drain			1		
135	Binsey Drain			1		Good but poor as wetland. Limited scope for enhancing wetland features.
136	Binsey Drain			1		Good but poor as wetland. Limited scope for enhancing wetland features.
137	Binsey Drain			1		
138	Binsey Drain			1		
139	Binsey Drain			1		Long-term thinning?
140	Biosey Drain			1		Long-term thinning?
141	Binsey Drain			1		Poor for wetland. Birthwort site.
142	Binsey Drain		1	1		
143	Binsey Drain		'	1		
144	Binsey Drain			1		
145	Godstow Holt Drain				1	Poor as wetland. Part destroyed.
146	Binsey Drain			1		Not good wetland, but wooded. Long-term thinning?
147	Duke's Lock Drain			•	1	
148	River Thames	1				Adjacent to SSSI.

In summary, for the River Corridor habitat survey of 148 sections (c.74km),

51	sections were evaluated as	critical
1	section was evaluated as	important
7 7	were evaluated as	good
19	were evaluated as	1000

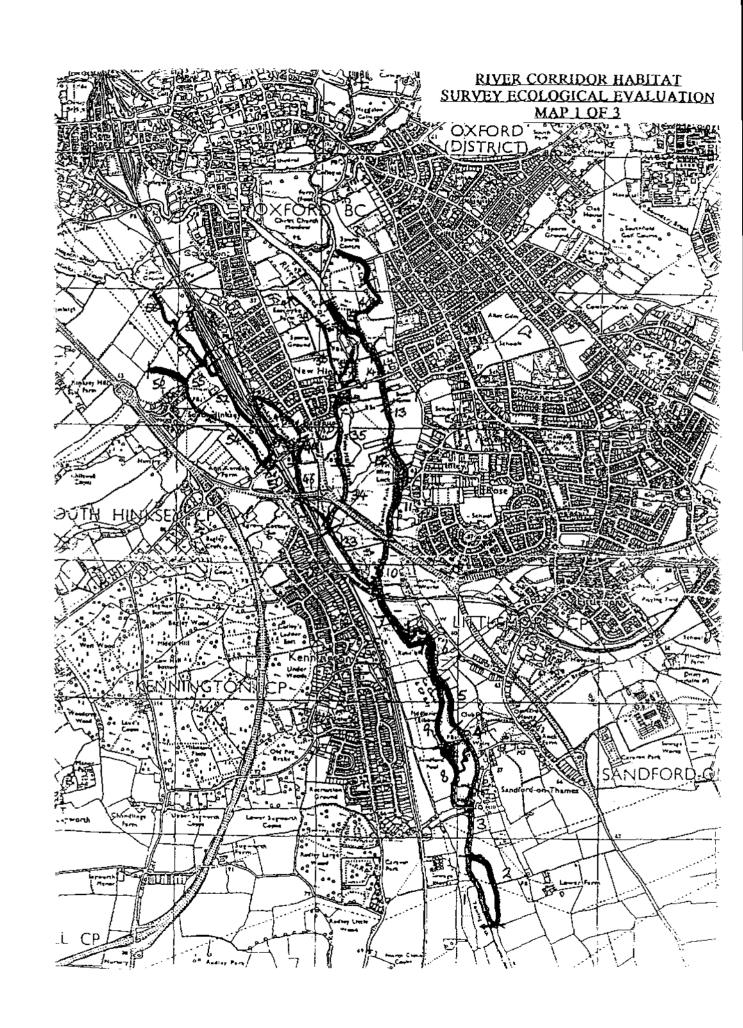
The following series of 3 coded maps provide an easily accessible visual representation of the River Corridor Section Evaluations and their relationships.

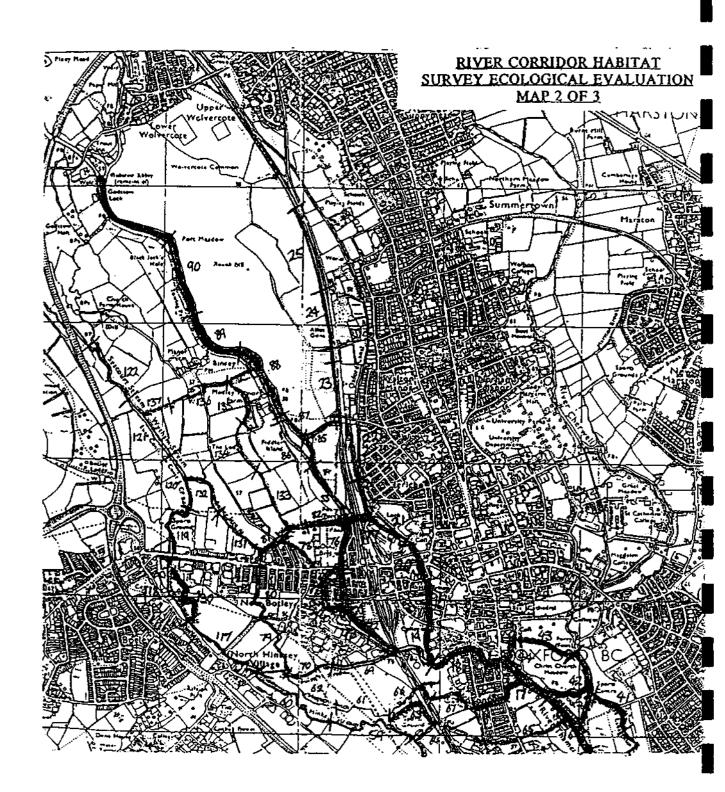
Outlined below are brief descriptions of the watercourses and information on their evaluations, management recommendations and enhancement opportunities. Whilst some of the management recommendations may apply only to specific sections or parts, there are some general recommendations that are applicable to the majority of the sections.

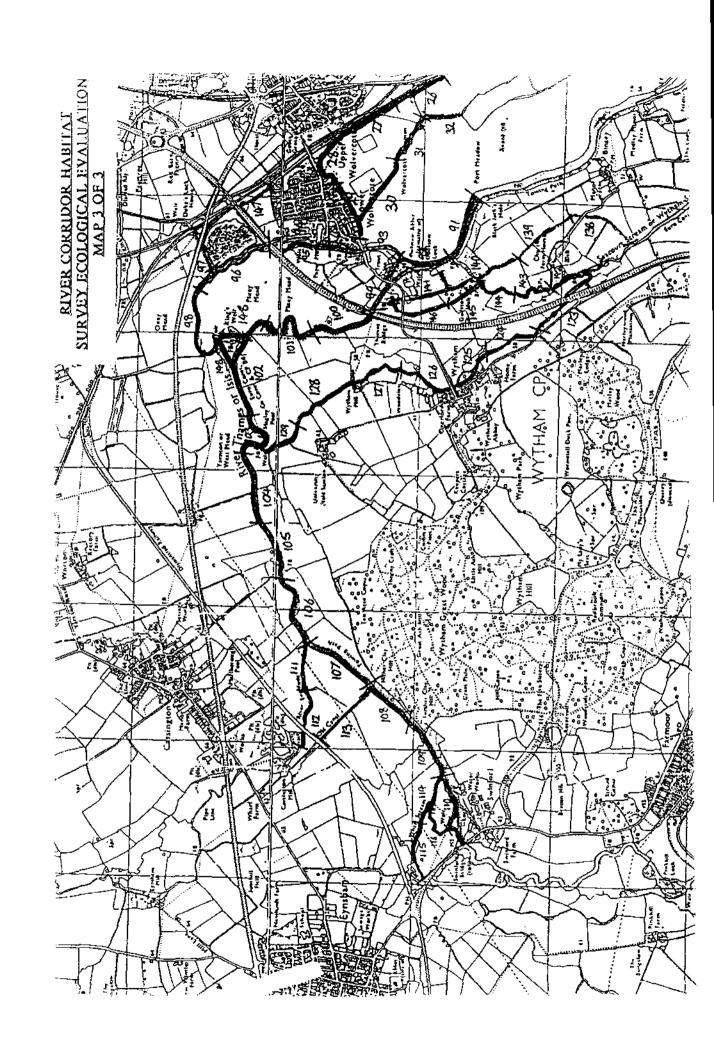
The existing management of channels bordering or passing through SSSI's should continue through close liaison and consultation with English Nature.

The management of the riparian tree cover is one goal of management of the existing wildlife habitat resource. Single, lines or groups of old pollarded and coppiced trees are important in wildlife and landscape terms and recommendations are made for repollarding and recoppicing where appropriate. A certain number of trees should be worked each year out of the total resource in order to provide variety of structure and age of regrowth. This will also act as a precautionary measure in case the trees do not respond to management, so avoiding the situation of having a river bank lined with dead stumps and pollards.

As a spin-off from pollarding and coppicing, the increased light let on to the river bank should promote the growth and spread of channel vegetation, so protecting, where appropriate, the bank from erosion by boat wash and, in any case, potentially increasing botanical diversity over time. Fallen trees are to be removed from channels unless it can be demonstrated that they function as Kingfisher perches and do not significantly impede flow or navigation. Occasional standing dead trees, unless a danger to the public, are best left in situ for woodpeckers, invertebrates and bats. Wooded areas could have bird and bat boxes sited.







4.2 LARGE RIVERS

The River Thames (Sections 001 003-018 036-037 071-076 083 086 088-096 098-110, and 148)

The River Thames is mostly gently meandering, wide and deep with islands, locks and weirs and slow flowing through agricultural land, on either side of urban Oxford. Its banks are principally clay and sand of varying profile with large parts being vertical and artificial through the built-up area. The channel carries a large amount of boat traffic especially in summer.

Of the 49 sections of the River Thames 29 are evaluated as <u>critical</u> by virtue of the presence of SSSI or valuable water-level dependent sites such as Fiddler's Elbow Island, Rose Isle and Heyford Hill Lane Pasture adjacent to one or both sides of the river for at least a part of the section. Sections 011-013 are evaluated as <u>critical</u> by virtue of their flowing alongside Iffley Meadows SSSI but the nature conservation value of the river is relatively low due to its urban, artificial nature, Iffley Lock and environs, and use by man. Unfortunately there is also limited scope for enhancement of the river in these sections. It is recommended that the best option for these 29 <u>critical</u> sections be a continuation of the present management in order to maintain their nature conservation value.

13 sections of the River Thames are evaluated as good with the recommended conservation management being, in the main, to retain the section as at present coupled with sympathetic future management of, for example, stands of emergent vegetation. Section 007 would benefit from enhancement measures, for example the creation of a wide berm to permit the growth of large stands of fringing vegetation. Section 083 and 086 would be enhanced by practical measures to improve the extent and diversity of emergent and aquatic vegetation given the heavy use of these sections by boat traffic.

7 sections are evaluated as <u>poor</u>, all of these being urban or lock sections with mostly artificial banks and heavily used by man. All of the existing seminatural habitat in these sections should be retained. In most of these sections there is little scope for enhancement beyond cosmetic tree and shrub planting with the exception of 075 in which it is recommended that the channel profile be changed to create marginal berms and riffles and pools.



EXAMPLE

i

DATE: 10 July 1992

CODE: OFP 076

River Thames

CONDITIONS

Surveyed left bank. Dry and sunny. Flow - slow. One

photograph from left side.

PHYSICAL FEATURES OF CHANNEL

Broad nature

A gently meandering section entering the outskirts of Oxford. The section is crossed by the A420 road bridge in midsection.

ii Dimensions

Channel 20-25m wide, >2m deep.

iii Substrate

Clay and mud.

iv Bank type

Artificial on the left; vegetated entirely, except for downstream, on the right. Banks 0.5-0.75m high, 45-90° slopes.

LAND USE

LS Houses, gardens, hotel and factory.
RS Allotment gardens, houses and gardens.

BANK VEGETATION LB Urtica dioica, Eupatorium cannabinum, Scrophularia aquatica, Epilobium hirsurum, Lolium perenne and Dactylis glomerata occur upstream with Sambucus nigra, Acer pseudoplatanus and Salix cinerea saplings. Near the road bridge, a few standards of Salix alba, Acer pseudoplatanus, Cornus sanguinea and Laburnum anagyroides overhang the bank. Below the bridge, Salix fragilis and Aesculus hippocastanum overhang the bank. Downstream the bank is artificial.

RB A mixture of Lolium perenne grassland with patches of Urtica dioica, Eupatorium cannabinum and Epilobium hirsutum occur upstream and in midsection with recently pollarded Salix fragilis stumps. Above the road bridge a pollarded Populus sp overhangs the channel. Below the bridge the artificial bank has Lolium perenne and Bellis perennis.

CHANNEL VEGETATION

Channel vegetation is sparse. Nuphar lutea and Sparganium emersum occur scattered along both sides upstream and in midsection. Polygonum amphibium occurs on the right in midsection with Iris pseudacorus, a few patches of Carex riparia and Schoenoplectus lacustris.

FEATURES OF INTEREST

A number of Crack Willow on the right bank have recently been repollarded. Mute Swan.

POTENTIAL THREATS

Continued speeding by boats.

MANAGEMENT RECOMMENDATIONS Retain the scrub and trees. Plant trees and shrubs on the right bank upstream. Retain existing marginal vegetation.