

Article (refereed) - postprint

King-Okumu, Caroline; Jaafar, Hadi; Shehata Aboukheira, Abdrabbo A.A.; Benzaied, Mongi; Obando, Joy; Hannachi, Ahmed. 2019. **Tracing the tradeoffs at the energy-water-environment nexus in drought-prone urbanising regions.**

© Saudi Society for Geosciences 2019

This version available at <u>http://nora.nerc.ac.uk/id/eprint/525905/</u>

Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at https://nora.nerc.ac.uk/policies.html#access

This is a post-peer-review, pre-copyedit version of an article published in *Arabian Journal of Geosciences*, 12 (20), 639. The final authenticated version is available online at: <u>https://doi.org/10.1007/s12517-019-4730-4</u>

There may be differences between this version and the publisher's version. You are advised to consult the publisher's version if you wish to cite from this article.

The definitive version is available at https://link.springer.com/

Contact UKCEH NORA team at <u>noraceh@ceh.ac.uk</u>

The NERC and UKCEH trademarks and logos ('the Trademarks') are registered trademarks of NERC and UKCEH in the UK and other countries, and may not be used without the prior written consent of the Trademark owner.

1	TITLE PAGE
2	
3	Names of the Authors:
4	Caroline King-Okumu ^{1,2,8} , Hadi Jaafar ³ , Abdrabbo A. A. Shehata Aboukheira ⁴ , Mongi Benzaied ⁵ Joy Obando ⁶
5	and Ahmed Hannachi ⁷
6	
7	Affiliations and Addresses of the Authors:
8	1: GeoData Institute, University of Southampton, UK
9	2: The Borders Institute (TBI), Kenya
10	3: American University of Beirut, Lebanon
11	4: National Water Research Center (NWRC), Ministry of Water Resources and Irrigation (MWRI), Republic of
12	Egypt
13	5: Institut des Regions Arides, Tunisia
14	6: Kenyatta University
15	7: Chemical-Process Engineering Department, National Engineering School (ENIG), University of Gabes
16	8: Centre for Ecology and Hydrology
17	
18	Title: 'Tracing the trade-offs at the energy-water-environment nexus in drought-prone urbanizing
19	regions'
20	
21	Email address and telephone number of the corresponding author:
22	<u>caroking@yahoo.com</u> +44 7964548256
23	
24	
25	
26	

27 Abstract

28 Strategies for managing water stress negotiate a complex series of trade-offs and opportunities. Game-changing 29 opportunities for water stressed regions are emerging with new emissions reduction strategies and energy options. 30 These have particular significance for socio-economic development pathways in the marginal drought-prone 31 regions. In this paper, we explore the energy-water-environment nexus in watersheds undergoing acute water 32 stress and energy transitions in the Arab region and the Horn of Africa. A review of published and ongoing 33 scientific activity was used to elaborate four case studies and identify common trade-offs between objectives to 34 reduce water stress, increase productivity and lower energy costs and emissions. The available scientific evidence 35 base for assessment of these trade-offs was then compared via a discursive process amongst review team members.

36

37 Collectively, the case studies present a State of the Art in geoscientific research priorities and methods currently 38 applied in the Arab region to quantify nexus trade-offs for decision-making concerning increasing groundwater 39 use, water harvesting and wastewater reuse across the case studies. The review pursues the wider geographic 40 relevance and scope of this emerging scientific agenda. It identifies global opportunities to boost and progressively 41 enhance the geoscientific information bases for decision-making in the most water stressed regions, as well as 42 direct comparisons to emerging discussion in the Horn of Africa region. Insights for sustainable development 43 decision-makers are highlighted and further scope for the transfer of insights within and beyond the Arab region 44 are discussed.

- 45
- 46
- 47
- 48 Keywords
- 49 Water, energy, water stress, drought, trade-offs
- 50

51

52

54 1. Introduction

55 Strategies for management of water scarcity negotiate a complex system of trade-offs between water, energy, and 56 production of environmental services including food (IPCC 2018). Transitions in the use of energy technologies 57 can boost water supplies in water-stressed environments (IEA 2016). But increasing emissions associated with 58 many such technologies further exacerbate climate change and threaten resource-dependent populations (IPCC 59 2018; WWDR 2014). Analyses of trade-offs at the water-energy-food security nexus have invited decision-60 makers to rethink water resource development strategies that create bloated blue water and emissions footprints 61 in many regions (Endo et al. 2017; Hoff 2011). A range of such analyses have been generated in the Arab Region 62 (Amer et al. 2017; Daher and Mohtar 2015; ESCWA 2016; Hoff et al. 2017), and parts of sub-Saharan Africa 63 (IUCN 2015; Jillo et al. 2016; Wakeford 2017; Yang et al. 2018)¹. This paper emphasizes the implications of 64 energy and emissions trade-offs as a critical entry-point to managing the nexus (i.e. Energy-Water-Environment 65 trade-offs) associated with water management adaptations in the most drought-prone and water stressed regions.

66

67 Increasing surface water shortages and dependence on groundwater to meet domestic, irrigation and other water 68 demands is a growing concern in many of the worlds dry regions (Cherlet et al. 2018; Kaur et al. 2016). Pumping 69 groundwater is about seven times more energy intensive than surface water abstraction (UNWater 2018). On the 70 other hand, in some cases, surface water requires more (energy-intensive) treatment prior to use than groundwater 71 (IEA, 2016a). Many millions of impoverished smallholder farmers across the drylands of sub-Saharan Africa and 72 Asia are reliant on groundwater due to its relatively good quality, ease of access, reliability and flexibility (Shah, 73 2012). But over-extraction of groundwater can have disastrous effects on environmental sustainability (Giordano 74 and Villholth 2007; Velis et al. 2017; Villholth et al. 2017). In many parts of the Arab region, these problems are 75 advanced, and national strategic and scientific capacities are increasingly engaged to address the challenge. This 76 is generating knowledge and insight that is of value to global environmental debates beyond the agricultural sector 77 of the Arabian region.

78

Global interest in the extent and significance of trade-offs and synergies at the land-water-energy nexus in highly
water-stressed contexts is increasing (IPCC 2018; Velis et al. 2017). Geoscientific investigations can shed light
on the stakes involved, and more will be required over the coming years. In this paper, we revisit a previous

¹ see also: https://www.water-energy-food.org/resources/resources-detail/nexus-country-profiles-for-theniger-basin/

analysis of nexus trade-offs in the Arabian region (King and Jaafar 2015) and extend its scope to consider similar
nexus challenges and trade-offs in the drought-prone Horn of Africa. A simplified resource-accounting approach
(after UN and FAO 2014) is used to characterize effects on the production of ecosystem services for human wellbeing associated with different land management technology alternatives and scenarios, based on the information
available in each case. Ecosystem services include food production but can also encompass a range of other
products and services. The simplified resource accounting approach enables analysis of tradeoffs between effects
on water availability and energy consumption required to produce these services.

89

90 The objectives of this paper are to:

- 91 > Identify the critical choices concerning energy technologies, costs and implications for balancing water
 92 stress in the selected water stressed basins;
- 93 > Explore the current state of geoscientific knowledge and tools to support decision-makers' assessments
 94 of trade-offs between available technological options to balance basin-level water stress and objectives
 95 to reduce emissions and energy costs while maximising services to human well-being;
- 96 > Consider the relevance of these insights from geoscientific work in the Arab region for other water
 97 stressed contexts, including those in other drought-prone regions, particularly in Africa.
- 98

99 These objectives are reflected in the organization of the paper. A brief description of the approach and methods 100 used to identify the critical choices concerning energy technologies, costs and implications for balancing water 101 stress in the selected water stressed basins is presented in Section 2, and the four case studies are presented. In 102 Section 3. the trade-offs for sustainable development decision-makers emerging from each of the four case studies 103 are reviewed. Three common trade-offs are identified and discussed. These include increasing demands for energy 104 due to growing dependence on groundwater in the drier regions; the potential of water harvesting and groundwater 105 recharge techniques to slow the increasing demand for energy; and additional trade-offs associated with increased 106 recycling of water. The discussion in Section 4 focuses on knowledge gaps and opportunities for further 107 geoscientific investigation to boost and progressively enhance the information base for decision-making at this 108 critical nexus. The conclusion in Section 5 underlines the broader relevance of geoscientific insights already 109 gained in the Arab region for other water stressed basins in other regions, particularly in Africa.

111 2. Approach and methods

112

113 2.1 Case Studies

114 Four basins were selected for discussion as case studies in this article (Table 1). These included a case from the 115 drought-prone Horn of Africa region, as well as three water-stressed basins from the Arab region where 116 geoscientific work has been accelerating over the past few years (as evidenced by an earlier review in King and 117 Jaafar 2015). In each case, transitions are taking place due to increasing water demands and groundwater 118 extraction rates. Based on information available as of early 2018, a rapid review and assessment was made to 119 identify what is presently known or can be inferred about effects on economic production, water stress, energy 120 use and emissions in these four cases, including consideration of available technological options to adapt the 121 current land and water management patterns and scenarios.

122

123 Table 1: Selected case study areas and basins

Study area	Extent (km ²)	Basin / watershed	Countries	Total Extent
				(1000 km ²)
Koutine Watershed	279	Koutine Watershed	Tunisia	0.279
Bekaa Valley	170	Orontes Basin	Lebanon,	37.900
			Syria, Turkey	
Isiolo County	25,336	The Ewaso Ng'iro	Kenya	210.226
		North Catchment Area		
Nile Delta	25,000	Nile Basin	11 countries	3173

124

125 (Figure 1: Location of case study areas and basins)

126

127

129 2.1.1 Case of the Ewaso Ng'iro North Catchment, Kenya

130 The Ewaso Ng'iro North Catchment Area (ENNCA) (surface area: 210,226 km²) (Figure 1) is the largest and 131 most arid catchment in Kenya (WRMA 2014). Beyond Kenya, the catchment continues into the bordering areas 132 of Ethiopia and Somalia, where it forms part of the larger Juba-Shabelle basin²(total basin area: 749 000km²). The 133 Dawa river (58,961km²) forms the border between Kenya and Ethiopia. It joins the Juba River at Dolo, where it 134 then becomes the Somalia-Ethiopia border (Basnyat and Gadain 2009; Woldemariyam and Ayenew 2016a; 135 Woldemariyam and Ayenew 2016b). Although much of the flow from Kenya into Somalia is in the subsurface 136 Merti aquifer system (GIBB 2004; Oord et al. 2014), a seasonal tributary known as the Lagh Dera also joins the 137 Juba river in Somalia (Michalscheck et al. 2016).

138

The upstream flows through ENNCA are reducing because land and water management decisions in the highlands around mount Kenya are resulting in loss of forest cover and increased water extraction for commercial agriculture (Mutiga et al. 2010; WRMA 2013a). This is progressively reducing flows of water to vulnerable populations downstream and also reducing natural flow regulation. This results in increased water stress in the downstream areas during dry seasons, interspersed with disastrous floods that cause economic damage and sometimes loss of human life in urban areas during rainy seasons (GoK 2014; King-Okumu et al. 2017).

145

146 Water managers have observed that the upstream extractions contribute to the economy, and are difficult to 147 prevent - even if they are illegal (Mutiga et al. 2010). Instead of regulating and conserving the flows of surface 148 water, the national water masterplan for the downstream areas focuses on increasing the use of groundwater to 149 supply the downstream populations (WRMA 2013a; 2013b; 2014; 2016a; 2016b). This has implications for 150 energy demand (Harou et al. 2017; King-Okumu et al. 2018). The expansion of water and energy services to 151 improve living standards for the urbanizing downstream populations must rely either on increased use of diesel 152 fuel to pump water, or on renewable energy, such as solar and wind power (Ndirangu 2017). In ENNCA, since 153 the availability of water flow throughout the year is already too low to support hydropower development, this is 154 not an option. However, Kenya is investing heavily in electricity transmission lines to import hydroelectricity 155 from neighbouring Ethiopia (Wakeford 2017).

² See: <u>http://sddr.faoswalim.org/River_atlas/atlas/River_Atlas_Som_Part1_drainage.pdf</u>

156 2.1.2 Case of the Nile Delta, Egypt

157 The Nile Delta extends over around 2,500,000 ha, with maximum annual rainfall of around 200mm at the coast, 158 and less inland. The cultivated area of the delta is supplemented by ongoing land reclamation. The annual volume 159 of surface water inflow at the High Aswan Dam is 55.5 km³/y, and annual outflow to the sea, 12.2 km³/y (MWRI 160 2010). Under climate change and upstream development, the volume of water reaching the Delta from the Nile 161 Basin (Figure 1) will be reduced.

162

163 A recent assessment of the water balance in the Nile Delta (Molle et al. 2016; 2018) found that thanks to the reuse 164 of drainage water, despite increasing volumes of water use, and reduced inflows, the outflows to the sea had 165 remained relatively unchanged since the 1990s. Water demand for drinking in Egypt is 1.8 km³/y, while industry 166 requires 1.4 billion m³ (MWRI 2010). The amount of water used by the agricultural sector in 2010 was about 67 167 km³/y, including leaching requirements and deep percolation to shallow groundwater. To achieve this, and support 168 other sectors, other water sources are used besides surface water. These include deep non-renewable groundwater 169 sources as well as the reuse of drainage water. The major uncertainty concerns the volume of groundwater use 170 across the various sectors, including agriculture and others (including municipal and industrial uses). Water 171 quality challenges including salinity and others are an increasing concern.

172

173 At present, the agricultural sector is estimated to be responsible for 85% of water demand in Egypt, but only 5% 174 of its energy demand (Hoff et al. 2017). These estimates may require further geoscientific investigation. Based on 175 emerging observations of the extent of shallow groundwater reuse in the Delta and the surrounding reclamation 176 areas (El-Agha et al. 2016; King and Salem 2012; 2013; Molle et al. 2018), it appears likely that current estimates 177 of agricultural energy uses may already be underestimated. Irrigation practices can help to increase economic 178 water-use efficiency, but this often creates further increased demand for energy to control the water application 179 (King and Jaafar 2015). Furthermore, the majority of high value horticultural crops that are sufficiently profitable 180 to pay for such systems require irrigation with fresh water, which must still be pumped from deep underground.

181

182 2.1.3 Case of the Upper Orontes, Lebanon

The total area of the Orontes Basin is 37,900 km², including parts of Lebanon, Syria and Turkey. Lebanon's
Beka'aa Valley covers 17,000 ha at the upstream end of the basin and receives average annual rainfall of 150mm

(total volume: 25MCM/y) (Figure 1). The agricultural land cover visible in summertime has expanded by more
than 20% on average since the beginning of the century, mainly in areas not served by the existing open-channel
irrigation scheme, but irrigated using groundwater (Jaafar and Ahmad 2019 in press; Jaafar et al. 2015; SaadéSbeih et al. 2018).

189

190 At present, the energy uses for pumping additional groundwater in the hotspots of water stress are identified as a 191 concern in Lebanon's third national communication to the UNFCCC, which predicts that in the future 192 (MoE/UNDP/GEF 2016 pXII): 'Droughts will occur 15 days to 1 month earlier, and countrywide drought periods 193 will extend 9 days longer by 2040 and 18 days longer by 2090. The already dry regions, such as the Bekaa, 194 Hermel, and the South, will experience the sharpest effects. In addition, cost impacts will be added to irrigation 195 needs, as more pumping hours will be required, therefore consuming more energy.' Where surface water is not 196 available to meet human needs, groundwater is increasingly used (BWE 2015; Jomaa et al. 2015). Water table 197 levels in the Bekaa are falling (MoEW/UNDP 2014). A drop of about 20m since 1970 (average of 60cm/year) has 198 already been observed in the neighbouring Litani area in the Southern Bekaa Neogene-Quaternary Basin. In some 199 parts of the Bekaa, the water table has fallen by more than 15m over the last five years (Jaafar and Ahmad 2019 200 in press).

201

202 Recently, Jaafar et al (2016) examined the water balance in the upper Orontes, and identified the need to consider 203 a more comprehensive range of sectoral water uses and demands affecting water and energy dynamics – not only 204 those of agricultural water uses which have been the focus of previous nexus studies (e.g. King and Jaafar 2015). 205 The analysis responded to popular concerns that population growth since the arrival of refugees from Syria may 206 be creating new water and energy access challenges, and exacerbating climate and drought risks (MoE-EU-UNDP 207 2014). However, investigation of the additional water demands for the refugees (Jaafar 2017) showed that these 208 were still relatively small compared to the agricultural water use volumes. Nonetheless, the spatial and temporal 209 concentration of the demand peaks were creating hotspots of water stress. This was accelerating groundwater 210 extraction and therefore creating additional energy demands due to the increases in pumping depths to 211 groundwater (Jaafar et al. 2016).

212

214 2.1.4 Case of Zeuss Koutine Watershed, Tunisia

215 Several ephemeral rivers feed the Zeuss Koutine Aquifer in the Koutine Watershed, Tunisia (Figure 1). 216 Traditionally, in the upper reaches of the Koutine Watershed, rainfall and water-harvesting structures (known as 217 *jessour*) support agroforestry, intercropping of barley and natural vegetation for grazing by livestock. Further 218 down, in the plain, water-harvesting structures are known as tabia. Harvesting rainwater to support trees and 219 fodder for livestock can simultaneously increase the recharge of groundwater (2009; Ouessar et al. 2004). Where 220 rainwater captured in the soil profile exceeds the water holding capacity, it will percolate downwards and recharge 221 the groundwater reserves flowing through the Zeuss Koutine Aquifer, serving the downstream urban population 222 in the cities of Medenine, Tataouine, Jarzis, Djerba and Benguerdene. The water demands of these cities are 223 growing rapidly and outstripping the available supplies of freshwater.

224

225 There are two main aquifers: the Zeuss Koutine aquifer and the Gres de Trias aquifer both supply domestic water 226 use, tourism, industry and commercial uses. A third aquifer is the saline Jeffara aquifer. The main source of water 227 for the agricultural sector are shallow aquifers supplemented mainly by water from the Zeuss Koutine aquifer and 228 also occasionally to a lesser extent by the Gres de Trias aquifer. Recently a desalination plant was constructed 229 purifying water from the brackish Jeffara aquifer to supply to the urban consumers in Djerba/ Zarzis and for the 230 tourism sector. An additional seawater desalination plant (capacity: 50.000 m³/day) was partially operational in 231 2018 to supply water to the same users. With these developments, pressure on the two aquifers is expected to 232 diminish and plans are underway to develop additional irrigated areas using the conserved water from these 233 aquifers.

234

235 A previous rapid assessment (King and Jaafar 2015) focused on exploring how water harvesting could help to 236 meet urban water demands and reduce the need for energy-intensive and expensive desalination and water 237 transfers in urbanizing catchments such as the Zeuss Koutine watershed. This was based on the view that farmers' 238 use of water harvesting upstream could enhance groundwater recharge and affect the availability of water 239 downstream. This could help to avert the need for increasing use of expensive and energy intensive desalination 240 facilities to supply water to the downstream urban areas of Medenine. However, in light of increasing demand, 241 reliance on desalination has grown over the past years. Water harvesting may substitute for some of the 242 requirements for desalination. But if the population continues to grow and remains concentrated in the coastal 243 cities, demand for desalinated water is likely to continue to increase.

244

245 2.2 Methods for the characterisation of energy-water-environment trade-offs

Trade-offs of interest in this review concern the available technological options to balance basin-level water stress and objectives to reduce emissions and energy costs while maximising services to human well-being. These were identified through a review of published literature and ongoing field studies. This built on and extended a previous review of the three Arabian cases (presented in King and Jaafar 2015). The addition of a case from the Horn of Africa region draws on a body of work by practitioners from that region (Jarso et al. 2017; Jillo et al. 2016; King-Okumu et al. 2018). This sheds light on the wider relevance of the adaptation options under discussion, and of the ongoing geoscientific investigations needed to assess the associated trade-offs.

253

254 A series of common adaptations to water stress were identified amongst the case studies. Their effects on energy 255 use and emissions, and the production of services for human well-being were weighed against the anticipated 256 effects on water stress. This assessment drew on relevant published or ongoing geoscientific work, where available 257 to create a simplified representation of a resource-accounting approach (after UN and FAO 2014). The assessment 258 was presented qualitatively as positive and negative directional indicators. This reflected what is known regarding 259 the characterisation of opening balances, inflows, storage, products, and outflows of water, energy and 260 geochemicals associated with different land management technology scenarios (as explored previously for the 261 three Arabian cases in King and Jaafar 2015). The simplified presentation facilitated discussion and comparison 262 between cases.

263

264 The format for presentation and discussion of the trade-offs in this paper is similar to those used in the future 265 scenarios of the millennium ecosystem assessment (Carpenter et al. 2005), IPCC reports (IPCC 2018) and popular 266 assessments of sustainable land management strategies (WOCAT 2013). This could be intelligible to decisionmakers and non-specialists as well as to researchers. It provides a simple indication of the direction of change as 267 268 either positive or negative. The categories of impacts considered are broad: combing effects on energy use and 269 emissions. Impacts on the production of services for human well-being is also a broad category of impacts that 270 includes consideration of effects on the production of food as well as other services that require water and 271 ecosystem management.

An iterative discursive approach to the review of the available information concerning the nexus trade-offs
encourages reflection by researchers and decision-makers. A preliminary version of the assessment was presented
and discussed during an international conference on Water, Environment, Energy and Societies, held in Zarzis,
Tunisia, in May, 2018. This was supplemented by continuous iterative review and progressive update by members
of the research team. In this way, the review takes into consideration gradual advances in the availability of
geoscientific studies relevant to understanding the nexus trade-offs. It also enables reflection on remaining gaps
and research priorities.
The iterative development and comparison of the four case studies revealed three common trade-offs associated
with adaptation to water stress. These concerned:
- increasing costs and carbon emissions to pump greater volumes of groundwater from depleting
groundwater tables
- slowing the increases in groundwater pumping costs and emissions by recharging water tables (with
different implications for energy requirements)
- increasing water availability by recycling and treating wastewater (with a range of possible implications
for energy use)
Based on this characterisation of the common concerns and challenges amongst the case studies, it was possible
to identify gaps in the available geoscientific knowledge base, and common priorities for further review and
assessment.
2.3 Accounting for water management effects on energy use
In all four of the study areas, increasing reliance on groundwater and falling water tables are reported to have
increased groundwater pumping costs. Previous studies have demonstrated and quantified these effects in
Tunisia (Croitoru et al. 2010), parts of the Nile Delta (King and Salem 2013), and also in ENNCA (King-
Okumu et al. 2018). Quantitative assessments of energy demands and costs for pumping groundwater for the
Bekaa valley case study could not be identified through the review of published literature. However the problem
of the escalating costs is mentioned in Lebanon's third national communication to the UNFCCC, as referred to

301 in the previous section (MoE/UNDP/GEF 2016 pXII). A rapid review of the available strategic documents from

this and the other three cases suggests that the increases in carbon emissions of escalating groundwater pumpingmay not yet have been fully assessed in any of the cases.

304

In Lebanon, the emissions from combustion of fossil fuels for water pumps for agriculture are assumed to be relatively small compared to those of the other sectors, reaching a total of only 245Gg CO2eq (MoE/UNDP/GEF 2016; MoE/UNDP/GEF 2017). But this is significantly higher than other calculated emissions from the agricultural sector (other than nitrogen emissions from soil)(MoE/UNDP/GEF 2015a). The national reports on emissions calculations for agricultural lands were made using a methodology (based on IPCC 2003; MoE/UNDP/GEF 2015b). At the time of writing, these methods were under review and revision through the IPCC expert process. National and basin-level assessments may soon be able to consider the scope for updates and refinements (see https://www.ince.prgin.ince.or.in/public/index.html.)

- 312 refinements (see: <u>https://www.ipcc-nggip.iges.or.jp/public/index.html</u>).
- 313

For estimating the amounts of energy required to pump the water used in agriculture, basic methods applied previously in Egypt (Attia et al. 2005; ESCWA 2009; Fraenkel 1986), were identified through the review of previous studies. Further methodological contributions were also identified from geoscientific studies published from India (Kaur et al. 2016), due to rising concerns that have emerged in groundwater dependent areas of that country. The following equation can be applied (after Kaur et al. 2016):

Energy E, kWh =
$$\frac{m^*g^*h}{3.6^*10^{6*}\eta_{eff}(\%)^*(1-T\&d losses(\%))}$$
 (equation 1)
22 Where:
23 E is the energy required to lift water in kWh;
24 m is mass of water pumped;
25 g is acceleration due to gravity 9.8 m s_2;
26 h is the total dynamic head, which is a function of initial water level, drawdown, delivery head and
27 losses in pipe) and is calculated as per Equation (2) (Michael et al. 2008):
28
29 h = initial water level (m)
30 + 20% of initial water level (m)

(equation 2)

332

Given the lack of detailed knowledge of the efficiency of individual pumps, Kaur et al (2016) recommended a value of around 30% based on available studies of energy use for irrigation pumping conducted previously in India (Nelson et al. 2009). Overall efficiencies in diesel-powered pumping systems result from the compounded efficiencies of the diesel engine (typically 15–40%), transmission (60–100%), pump (40–80%) and pipes (30– 95%), giving an overall efficiency of 0.5–27% (Fraenkel 1986). Therefore, an approximate midpoint of 15% has been used in previous studies (Shah 2009).

339

Kaur et al (2016) used the all-India average value of 1.4 kg of CO2 kWh⁻¹ at the station (0.406 kg C kWh⁻¹) to estimate the release of C from electric pumps. With 5% transmission losses, an effective C emission rate of 0.427 kg C kWh⁻¹ at the generating facility or 3.87 kg C to lift 1000 m³ to 1 m was used. The emissions from coal-based electricity power generation plants are about 5.82 (3.87/0.665) times higher than the rate of emission with diesel pumps (Nelson et al. 2009). With diesel pumps, the amount of C released to lift 1000 m³ of water to 1 m is 0.665 kg C (0.732*9.08/10.0). The ratio of C emissions to energy content for diesel is 0.073 kg C kWh⁻¹ (Nelson and Robertson 2008).

347

348 The energy and emissions associated with pumping water must be multiplied by the volume of water required, 349 and the pumping depth. Other studies have used a mixture of remote sensing and ground-truthing techniques to 350 identify the proportion of cropped land in the Bekaa that is irrigated, the wells and depths to water and the volumes 351 of water required (Jaafar 2017; Jaafar and Woertz 2016; Jaafar et al. 2015). This information was used to complete 352 the estimations of energy use and emissions in the Bekaa Valley (Table 2), based on a simplified application of 353 Kaur's equations. For this study, the total annual irrigation water demand in the Bekaa for 2016 was estimated at 354 571MCM and the average depth of pumping around 100m (Jaafar 2017). For the other case studies, estimates of 355 pumping costs were available from previous studies. In all cases, further refinements could be made.

356

357

		2016	2017	2018	2019	2020
	total volume of water lifted (mcm)	571	571	571	571	571
	volume of groundwater lifted assuming					
	65% of total is from groundwater (mcm)	371.15	371.15	371.15	371.15	371.15
a	volume of water (1000 m2)	371,150	371,150	371,150	371,150	371,150
	amount of C released to lift 1000 m ³					
c	of water to 1 m (kg C)	0.665	0.665	0.665	0.665	0.665
	a x c (kg C)	246,815	246,815	246,815	246,815	246,815
b	average depth (m)	100	100.6	101.2	101.8	102.4
	Total C emissions to lift groundwater					
	a x b x c (kg C / yr)	24,681,475	24,829,564	24,977,653	25,125,742	25,273,830

359 Table 2: Estimation of total C emissions to lift groundwater each year (2016-2020) in the Bekaa Valley

360

361

362 Alternatives to increasing water supply via additional pumping of groundwater or other means include water 363 conservation and recharge (to slow the demand for water). The extent to which this will slow demand for 364 additional pumping and emissions depends on the volume of additional water that can be recharged. This is 365 calculated based on the volume of rainfall, and context-dependent factors determining the proportion of this water 366 that can infiltrate the soil and percolate down into the aquifer at different locations. The reuse of water presents a 367 second alternative to increasing pumping of groundwater. How much energy is required to treat and reuse different 368 volumes of water is determined by context and interrelated costs. In the case of Zeuss Koutine, rapid quantitative 369 estimates were obtained to explore and illustrate the potential trade-offs that could be associated with these two 370 options.

371

372 3. Energy-water-environment trade-offs under water stress

For review and comparison of the common trade-offs, in this section, the findings from the assessment of theavailable mix of quantitative and qualitative information is presented in a simplified format. This provides a simple

indication of the direction of change as either positive or negative. This presentation device supplements the

376 descriptive text exploring the trade-offs, as observed in each of the various case studies.

377

378 3.1 More water use: more energy for deeper groundwater pumping

In all four of the study areas, increasing reliance on groundwater and falling water tables are reported to have increased groundwater pumping costs. Previous studies have demonstrated this in Tunisia (Croitoru et al. 2010), parts of the Nile Delta (King and Salem 2013), and also in ENNCA (King-Okumu et al. 2018). To our knowledge, the consequences of this trend for increasing carbon emissions have not yet been fully calculated in any of the case study areas. Therefore, these may not yet be fully represented in the national emissions inventories and decision-making scenarios for future planning.

385

Where the water balance is exceeded, and water tables are falling, the depth of pumping will also increase – resulting in further increases in extraction costs and emissions (Table 2). As irrigation requirements and durations increase due to climate change, the total volume of water demand will increase further. Growing domestic, commercial and industrial water demands in the Bekaa will also contribute to the increasing extractions of groundwater. These require additional attention.

391

The methods that have been used to calculate the emissions from pumping groundwater for irrigation in the Bekaa valley could be refined further, as they could in all of the cases. However, it is notable from observations of the nexus in some of the other cases included in this review that there are ways to reduce the energy trade-offs associated with increased groundwater pumping. These require some additional consideration because they would affect the calculation of total energy costs and emission in all four of the cases.

397

According to Egypt's 3rd National Communication to the UNFCCC (MoE 2016), the national water infrastructure system in Egypt is mainly composed of irrigation, drainage, water and wastewater pumping stations, in addition to the water and wastewater treatment facilities. Potential for GHGs emissions mitigation opportunities include solutions for pumping, irrigation control and drainage. For these, considered options include replacement/rehabilitation of inefficient pumps and increased use of renewable energy sources. This is in addition to adaptation options involving water harvesting and reuse of non-conventional water sources. For the time being, scenarios for the potential impact and trade-offs of these adaptation options have not been quantified. However, 405 recently, Egyptian researchers have begun to investigate the energy footprint of agricultural water management406 (El Gafy et al. 2017a; El Gafy et al. 2017b).

407

408 In the case of Egypt, mains electricity has been made available as a more cost-efficient option for pumping 409 groundwater than diesel. The government has made access to this lower-cost energy source conditional upon 410 having a legal permit for water extraction (King and Salem 2012). Furthermore, interestingly, in some areas, solar-411 powered systems are becoming more widely used for groundwater pumping and control of drip irrigation 412 (AbouKheira 1999; IFAD 2016). Similarly, in ENNCA, 70% reductions of operating costs associated with 413 pumping groundwater in the middle section of the catchment have been reported over the period 2013-17 due to 414 conversion of boreholes from diesel to solar power (GoK 2018). These variations in the types of energy that are 415 used for pumping groundwater will affect both the costs of pumping and the emissions that are produced (as 416 explained in the previous section).

417

418 In contrast to mains electricity systems that are used to power deep groundwater extraction in some of Egypt's 419 desert reclamation areas, solar-powered systems are not necessarily controlled or monitored easily by the 420 government, and once installed, they can have lower recurrent costs than diesel-powered systems (Shouman et al. 421 2016). As yet, Egyptian farmers only use these systems to generate power for their own agricultural water 422 management uses and are not able to sell it to local consumers or to the national grid (IFAD 2016). However, in 423 2014, Egypt passed a law promoting the use of renewable energy³. The Law encourages the private sector to 424 produce electricity from renewable energy sources by introducing several development schemes for the private 425 development of renewable energy projects, including competitive bids, feed-in tariff, and independent power 426 production through third party access. Similarly in Kenya, private sector participation in provision of solar-energy 427 powered systems for boreholes and irrigation is sought through public-private partnerships (GoK 2018 p141).

428

The Egyptian government has observed that advances in the capacities of solar-powered technologies to pump increasing volumes of deeper groundwater in the reclaimed desert areas could accelerate extraction and degradation of non-renewable deep groundwater supplies. They have therefore taken control of the solar-powered pumping systems in the desert areas. An alternative tried in India, but still unexplored in any of the case studies, would be to create markets for farmers to sell energy generated through solar technologies back to the national

³ Renewable Energy Law (Decree No 203/2014) see: http://egyptera.org/Downloads/Laws/law2014.pdf

- grid (as described in Shah et al. 2017; Verma et al. 2018). If the sale of energy becomes more profitable than the
 markets for the irrigated crops, farmers with access to technologies to capture solar energy may prefer to do this
 rather than to over-extract and damage their own collective water supplies by increasing water use for irrigation.
- 437
- The various adaptations of systems for pumping groundwater each has different trade-offs for the production of food and other services to human well-being, as well as for water security and emissions reduction (Table 3). In the case from the Bekaa, we can see that increasing irrigated crop production has come at a cost to water security for other uses and has also raised emissions. On the other hand, in Kenya, where boreholes operated by rangeland water resource user associations have been converted from diesel to solar power, this has enabled the WRUAs to provide more reliable services to their members, while also reducing emissions. But this may still come at the cost of an increasing groundwater deficit, as is also the case in Egypt.
- 445
- 446

447 **Table 3:** Trade-offs associated with groundwater pumping and energy options

Adaptation	Service production \$	Water security	Emissions reduction
Increasing groundwater extraction with	1	↓ ↓	↓ ↓
diesel (Bekaa)			
Conversion of existing diesel-powered	1	\downarrow	1
systems to solar power (ENNCA)			
New wells with solar-powered pumps	↑	↓	1
(Western Desert, Egypt)			

448 ↑ positive effect

449 \downarrow negative effect

450

451

452 3.2 Water harvesting and recharge: slow increases in energy needs for pumping

453 In Southern Tunisia, the Bekaa Valley and Northern Kenya, adaptation options involving water harvesting and

454 groundwater recharge have been identified to achieve positive effects on service production, including provision

455 of water supplies to reduce water stress (Table 4). In Egypt, similar options have been explored in the Northern 456 Coastal areas, where there is relatively higher rainfall (Salem 2014). In Southern Tunisia (Ouessar et al. 2009) 457 and in ENNCA (Gies et al. 2014), the effects of water harvesting structures on service production and income 458 generation have been modelled using Soil Water Assessment Tools (SWAT). However, effects on groundwater 459 recharge, labour and energy demands are still poorly understood in both cases. Even less geoscientific work has 460 been published on such questions concerning water harvesting and groundwater recharge practices in the Bekaa 461 and Egypt. On the other hand, in these areas, groundwater recharge associated with irrigation water application 462 has attracted some scientific attention (see following section).

463

464 Table 4: Trade-offs associated with water harvesting and storage options

Adaptation	Service production \$	Water security	Emissions reduction
Micro-catchments (Bekaa)	1	↑	1
Water pans and sand dams (ENNCA)	1	↑	↑
Underground cisterns (Egypt N. Coast)	↑	↑	1
Bunds and terraces (Koutine watershed)	↑	1	1

465 ↑ positive effect

466 \downarrow negative effect

467

468 The human labour-intensity of construction and maintenance required for water-harvesting systems is often seen 469 as a major constraint to their uptake by private landholders (Abdeladhim et al. 2017). In cases where water 470 harvesting will simultaneously boost tree and forage production, this provides some incentives for land-users to 471 implement the water harvesting practices. However, the income generation from trees and forage crops that can 472 be supported through the water harvesting systems is not always considered sufficient to repay the labour 473 investments maintain them. In required to construct and some cases (see e.g. • 474 http://www.vallerani.com/wp/?tag=wocat-2), heavy machinery can reduce the costs of human labour required for 475 the construction of water harvesting systems. As yet, such cases are not common in any of the study areas but 476 investments in promoting such technologies are periodically discussed in international initiatives. The effects in 477 terms of increasing emissions have not yet been assessed.

479 Since increasing groundwater recharge can improve water availability and quality downstream, in theory, 480 downstream water users could be asked to pay upstream farmers for providing water harvesting services to them. 481 At present, the Tunisian government provides support for conservation of water harvesting structures in the upper 482 parts of the Zeuss Koutine water catchment. However, these incentives are still not enough to make water 483 harvesting profitable for the farmers to build and maintain the systems. In all cases, it is difficult to convince 484 downstream communities to provide incentives for improved water harvesting in the upstream areas without clear 485 evidence of the difference that the water harvesting systems will make to their water supplies (see discussion in 486 Swiderska et al. 2018).

487

488 Tunisian researchers have published various studies demonstrating cutting edge scientific tools and methods that 489 are relevant to the quantitative assessment of effects of water harvesting on the water balance (Abdelli et al. 2017; 490 2016a; Adham et al. 2016b). The three aquifers in the Zeuss Koutine watershed are recharged through seepage 491 through the river bed and infiltration into the landscape (Jarray et al. 2017). As there is little information on these 492 processes, the recharge of the aquifers may be estimated based on the known natural recharge rates as a function 493 of the rainfall. The best available estimate for the volume of rainwater recharge to groundwater achieved by the 494 water harvesting structures was reported by Ouessar et al (2009), who observed that 22% of the 12-year average 495 annual rainfall in the Koutine watershed (209mm) recharged groundwater, while 72% was taken up in 496 evapotranspiration of crops, and the outflow was 6%.

497

498 A rapid estimation of the volume and value of recharge could therefore be as follows:

499

500 total annual volume of rainfall = $279,000,000 * 0.209 = 58,311,000 \text{ m}^3$

501 total annual volume of recharge = $58,311,000 * 0.22 = 12,828,420 \text{ m}^3$

total annual value of recharge if $1m^3 = 1Euro = \pounds 2,828,420$

The Tunisian government has invested directly in increasing recharge to the aquifer through the construction and maintenance of artificial recharge structures in the upper part of the watershed (Hadded et al. 2013). Two years ago (2016-2017), three more recharge wells were constructed at Wadi Hjar to increase water recharge and two piezometers were installed to monitor ground water levels in the Grés de Trias aquifer with support from the European Commission through its Sustainable Water Integrated Management (SWIM) Programme project on Water harvesting and Agricultural techniques in Dry lands: an Integrated and Sustainable model in MAghrebRegions (WADIS-MAR).

510 Data on water level fluctuations collected over the course of a year (September 2016 - September 2017) by 511 pressure sensors in a data logger registered a 50 cm variation of the water level in the aquifer compared to a 512 reference level, and a negative overall annual balance (Figure 2). The situation was more pronounced at one of 513 the sites (mgarine), where the variation in the water level reached 70 cm and the annual balance was still negative. 514 The total volume of extractions from the aquifer is not known.

515 (Figure 2: Registered water level variation in the Grés de Trias aquifer at Wadi Hjar)

- 516
- 517

518 Similar studies of groundwater recharge effects achievable using land and water management practices have not 519 yet been pursued to the same degree at the other case study sites. There is therefore an opportunity for transfer 520 of experiences both within the Arabian region, and beyond to the Sub-Saharan African case of ENNCA and 521 others.

522

523 3.3 More water recycling: quality threats and energy demands for water treatment

524 In all four of the study areas, wastewater reuse is increasingly recognised as an alternative source of water with 525 potential to boost overall water availability. This includes recycling of irrigation water by farmers who pump 526 water out of the drainage areas in irrigated systems to reapply without treatment – as is commonly practiced in 527 the Nile Delta and parts of the Bekaa (Table 5). It can also include reuse of treated wastewater where treatment 528 plants are able to treat and supply such water to consumers and/or farmers. Reconfiguring the sequences of water 529 uses can enable uses requiring water of higher quality (e.g. for drinking and other domestic needs) to be supplied 530 first, before agricultural water uses that can have lower quality thresholds. However, transferring and treating 531 water entails additional energy costs and emissions.

532

534 Table 5: Trade-offs associated with water treatment and recycling to cope with reduced downstream

535 surface water availability

Adaptation	Service production \$	Water security	Emissions reduction
Desalination (Koutine watershed)	1	Ļ	Ļ
Planning water treatment systems	↑ (1	\downarrow
(ENNCA)			
Reuse for irrigation with- and without	1	\downarrow	\downarrow
treatment (Bekaa)			
Widespread reuse for irrigation often	1	1	↑
without treatment (Nile Delta)			

536 \uparrow positive effect

537 \downarrow negative effect

538

539 In ENNCA, and many other parts of Sub-Saharan Africa, there are relatively few water treatment facilities as yet. 540 However, significant investments are anticipated over coming decades. This provides an opportunity to make 541 well-informed choices and maximise the positive benefits and trade-offs by learning from what has worked 542 elsewhere. Once again, energy choices can make a major difference to costs and emissions from water treatment 543 technologies. They can also shape productive opportunities. Desalination technologies can remove more 544 contaminants than conventional water treatment plants or environmental technologies such as slow-sand filtration 545 units or constructed wetlands. However, they usually cost more and have higher energy requirements. Once again, 546 solar-powered systems are a potential game-changer. But for the time being, these are not widely used in any of 547 the study areas.

548

Lebanon's latest National Communications to the UNFCCC (MoE/UNDP/GEF 2016; MoE/UNDP/GEF 2017) focuses on increases in emissions associated with waste disposal and the treatment of wastewater. Water and air pollution render surface water resources unfit for use and so contribute to increasing demands for groundwater. A medium-term infrastructure investment plan (RoL 2018a) includes proposals for additional water treatment plants. The government's sectoral strategies on water and electricity also aim to increase the share of energy use that is provided from renewable sources of energy (RoL 2018b). In its National Communications to the UNFCCC (EEAA 2016), Egypt has also explored this opportunity and assessed the GHG emissions from wastewater

- handling for domestic wastewater, human sewage, industrial wastewater and the overall total from 1990 to 2005.
- 557 Increasingly, such calculations may be taken into account in water management decision-making.
- 558

559 In Southern Tunisia, a new desalination plant to be constructed at Zarat in the Governorate of Gabes is expected 560 to produce 50,000 m³/d when it begins operating in 2021, increasing to 100,000 m³/d by 2027 (36,525,000 561 m^{3}/yr)(desalination.biz 2018). The cost of the plant, together with three water storage tanks, and a pumping station 562 has required the Tunisian government to commit €14.5 million to the project and to accept a loan from the KfW 563 Development Bank of €32 million. To calculate the cost of production of a cubic metre of water from the 564 desalination plant, in addition to the construction costs, it would be necessary to consider the recurrent costs e.g. 565 for membrane replacement and others. It may also be affected by whether or not brackish groundwater is blended 566 with seawater to increase the volume available for treatment. The source and cost of energy supply will be a major 567 factor determining the cost of clean water production from the plant. If mains electricity is used, and this is mainly 568 generated using natural gas imported from Algeria, then the cost of the water production will depend on this input.

569

570 In other parts of the Arab region, previous studies have identified a unit cost for desalinated water at around 571 US\$3.6/m³ (ESCWA 2009) and a recent published study has compared the capacities and emissions rates of large 572 desalination plants in California (Badiuzzaman et al. 2017). But these costs and emissions are context- and 573 process-specific. In light of this, it is difficult to predict the emissions implications and the future costs of the 574 desalination plant. Overall, the implications of different technologies are likely to change, depending on the 575 volume of research and development attention they receive, as well as in response to commercial factors and 576 economies of scale of production, competition, etc. It is not yet clear what difference could be made by policies 577 intended to encourage the use of renewable technologies, such as solar energy. Little geoscientific work has yet 578 been done to map the scale of the opportunities and priorities for investment in this area across the Arab region or 579 beyond.

580

For smaller scale desalination units, solar power can be used (Salim 2012). In Egypt's desert areas, solar-powered stills have been demonstrated effective for desalination of small amounts of water primarily for domestic uses where other safe supplies are not available (Salem 2014). Where irrigated crops are sufficiently profitable, they can sometimes also justify the use of solar-panels to power reverse osmosis treatment systems that can treat sufficient volumes of water to irrigate high value crops (Ahmed et al. 2019). If they prove viable for pumping and treating shallow groundwater and drainage water, the use of solar-powered technologies that have been observed primarily in the desert reclamation areas could spread inward across the Nile Delta. This could transform individual farmers' income generation strategies, water management options and practices for coping with water scarcity and quality threats – just as the use of diesel-powered shallow wells recycling the drainage waters has been observed to have spread over recent decades (El-Agha et al. 2016; Molle et al. 2018).

591

592 4. Implications for decision-making and geoscientific research needs

593 As global interest in the extent and significance of trade-offs and synergies at the energy-water-environment nexus 594 is increasing (IPCC 2018), geoscientific investigations generated by scientists from across the Arabian region and 595 beyond can shed light on the stakes involved. Much more such scientific work may be required over the coming 596 years. In each of the four cases that we have explored in this paper, geoscientific research is gradually building 597 the robustness of the assessments of the energy and emissions implications of technologies to manage water, as 598 well as the other costs and benefits. Collectively, the cases present a state of the art in available geoscientific 599 assessment methods applied to quantify the stakes at the energy- water-environment nexus in some of the most 600 water stressed and drought-prone basins.

601

602 The analyses and comparison between cases also highlight knowledge gaps and opportunities to boost and 603 progressively enhance the information base for decision-making at this critical nexus. Countries are proceeding 604 at different rates to develop emissions accounts and some other environmental accounting systems (e.g. energy 605 accounting in Kenya (KNBS 2018) and other countries⁴. In this paper, a simplified illustrative approach was used 606 to characterize the effects on economic productivity and demands for water and energy associated with each 607 adaptation option that decision-makers could select. This can assist decision-makers to consider the relevant trade-608 offs and available information needed to evaluate and prioritize them. Where they may wish to obtain more 609 detailed assessments, they can then commission the geoscientific research community to provide information to 610 decision-makers, as needed.

⁴ See: https://www.iea.org/Sankey/#?c=World&s=Balance

Some of the trade-offs and synergies including costs and implications for carbon emissions as well as water conservation are better understood than others. In the water-stressed regions of the world, many knowledge gaps remain and will not be filled until decision-makers commission the necessary studies to be done. Researchers can highlight the availability of relevant assessment methods to work toward filling these gaps in sustainability decision-making. Based on the review of the case study experiences, the remainder of this section discusses three common themes and concerns requiring geoscientific investigation at this critical nexus in the drought-prone regions.

619

620 4.1 Modelling approaches to quantify groundwater recharge as a low-emission nexus

621 solution

The growing concern in many of the world's dry regions regarding increasing dependence on groundwater to meet domestic, irrigation and other water demands emerges strongly from all four case studies. On the other hand, careful management and use of groundwater can maximise the benefits from natural filtration, storage and distribution of water supplies in the subsurface. In regions that are prone to high temperatures, these are sustainable options. Conservation and use of groundwater can also help to avoid the construction and operating costs and emissions associated with man-made infrastructure for water treatment and transfers that require energy to pump water supplies from one area to another.

629

630 The cases that are explored in this paper provide insights demonstrating that geoscientific methods can be applied 631 to assess the volumes of groundwater recharge and to test strategies to enhance recharge processes, as in the 632 Tunisian case. This offers a useful example of an approach that could also be applied in other parts of the Arabian 633 region, such as the Lebanese case, and also in Sub-Saharan Africa, as in the case from ENNCA. Mapping and 634 characterising the processes affecting groundwater recharge and transport is also important in order to understand 635 the vulnerability of water supplies that are stored in the ground. This can include vulnerability to contamination 636 and increasing levels of salinity, which are critical concerns in areas where populations must depend on these 637 sources for their drinking water supplies.

639 4.2 Integrating assessment of agricultural water and energy demands with other

640 sectors

Across Africa, the extent of irrigated areas is expected to increase over the coming decades. The IEA (2017) predicts that by 2030 the combined additional electricity demand for water pumping, full mechanisation and postprocessing will amount to 12 terawatt-hours (TWh), which is equal to a further 6% on top of the additional energy demands foreseen for increases in household access to energy. This review has identified examples of the application of geoscientific investigation to assess the emissions generated from different agricultural water management strategies. It has also explored available methods to compare these to previous studies of the energy implications of increasing groundwater use for irrigation in India.

648

The case study from Lebanon underlined the importance of balancing agricultural water uses with other demands across sectors at different times of the year. Insights emerging from this case study that could contribute to the nexus debate in other regions include the relevance of geoscientific techniques such as mapping of hotspots for observation of the uneven spatial and temporal distribution of peaks in energy demand. These tend to be concentrated during the drier seasons, and to exacerbate droughts. These inter-annual spatial and temporal variations should be captured in national energy and emissions accounts. They could be used to inform strategies for water and energy management at different times of the year during drought and non-drought periods.

656

4.3 Catchment scale water stress assessments to include demands of transitory

658 populations

In the case studies from water stressed catchments that are explored in this paper, some relatively sophisticated modelling work has been done to assess water allocation planning for the settled populations. This includes studies carried out in the sub-Saharan African case (GoK 2012), as well as the cases from the Arabian region (as in Ouessar et al. 2009). In Southern Tunisia and also in Sub-Saharan Africa, assessments of water availability and demand are already providing a justification for large investments in desalination plants. The review of the future economic significance of water allocation decisions and their associated energy demands in the case from southern Tunisia suggests that some of these assessments and anticipated trade-offs may require further analysis.

667 The assessment of water needs for transitory as well as permanent populations that we observed in the Orontes 668 case is very relevant elsewhere, including in the contexts of the Horn of Africa. In the ENNCA case study, the 669 presence of a large refugee camp at Dadaab has caused periodic concerns about the additional environmental 670 stress that can be created by addition demands for water supply and waste disposal. Water demands for temporary 671 and seasonal residents and refugee communities are also relevant in Egypt. Seasonal fluctuations of population 672 numbers and their implications for resource use can be high in many water scarce catchments – especially during 673 drought periods. This has considerable implications for water and energy resource planning. It is an ongoing 674 challenge also in various parts of Sub-Saharan Africa. Although it has not yet been possible to take the needs of 675 the transitory populations into account, recent experiences from the Arabian region in coping with these water 676 demands could offer instructive examples of how the geoscientific challenges involved in assessing water 677 demands could be approached elsewhere. 678

- 679
- 680
- 681

682 5. Conclusions

683 Emerging options to make more use of low energy or renewable energy-powered technologies for water 684 management offer opportunities for water managers across Africa and the Arabian region to reduce water stress 685 and also maximise other positive trade-offs and synergies. These have implications both for short-term economic 686 productivity and also for the longer-term effects of environmentally damaging emissions. Critical choices 687 concerning investments in research and development of solar-powered technologies have particular implications 688 for development in the driest and most water-stressed regions. In the four cases explored in this paper, geo-689 scientific investigations are contributing to the quantification of benefits and trade-offs for sustainable 690 development. Common themes of interest for further geoscientific investigation identified concern recharge 691 modelling, cross-sectoral energy and emissions accounting; and water and energy demand assessment for 692 integrated catchment management planning.

693

694 The insights gained from geoscientific investigations and experiences presented in this paper from the selected 695 case studies from the Arab region are of broader relevance also for decision-makers in other water stressed basins 696 and regions. This is demonstrated in the case study from the Horn of Africa, and the wider discussion. There is an 697 opportunity for decision-makers in the most water-stressed and drought-prone regions to improve their analyses 698 of critical trade-offs by supporting further geoscientific investigation and exchanges on their priority questions. 699 The implication for scientists in the Arabian region and beyond concerns the need to demonstrate the range of 700 suitable geoscientific methods that are available and their relevance to addressing the challenges of drought and 701 water stress. Decision-makers could then support their application to shed light on the trade-offs at the energy-702 water-environment nexus across the drought-prone regions.

703

704 Acknowledgements

705

The authors are grateful to Mohamed Ouessar and the organizers of the international Conference on
Water, Environment, Energy and Society for encouraging the preparation of this article. We also wish to
thank Ahmed Abdelkhalek and several anonymous reviewers, who provided helpful and constructive
reviews. Research on the three Arabian cases by the National Agricultural and Extension Systems was
partially supported by USAID through ICARDA's Water and Livelihoods Initiative 2012-2014. The Sub-

- 711 Saharan African case study drew on studies that were partially funded by DfID. Backgrounds of figures in
- 712 the online supplementary material were derived from a base map available from Esri
- 713 http://goto.arcgisonline.com/maps/World_Topo_Map).

714 References

715 Abdeladhim MA, Sghaier M, Fleskens L, Ouessar M (2017) An Integrated Cost-Benefit and Livelihood Approach for 716 Assessing the Impact of Water Harvesting Techniques (WHTs) on Livelihoods: A Case Study in the Oum Zessar 717 Watershed, South-East Tunisia. In: Ouessar M, Gabriels D, Tsunekawa A, Evett S, Editors (eds) Water and Land 718 Security in Drylands - Response to Climate Change. Springer, Cham, Switzerland, pp 303-316 719 Abdelli F, Ouessar M, M'Hemdi S, Guied M, Khatteli H (2017) Monitoring Soil Moisture Content of Jessour in the Watershed 720 of Wadi Jir (Matmata, Southeast Tunisia). In: Ouessar M, Gabriels D, Tsunekawa A, Evett S, Editors (eds) Water 721 and Land Security in Drylands - Response to Climate Change. Springer, Cham, Switzerland, pp 97-110 722 AbouKheira AAE-AAS (1999) The possibility of using solar energy in driving irrigation pumps. Minufiya University, Shebin El-723 Kom, Egypt 724 Adham A, Riksen M, Ouessar M, Ritsema C (2016a) Identification of suitable sites for rainwater harvesting structures in arid 725 and semi-arid regions: A review International Soil and Water Conservation Research 4:108-120 726 Adham A, Wesseling JG, Riksen M, Ouessar M, Ritsema CJ (2016b) A water harvesting model for optimizing rainwater 727 harvesting in the wadi Oum Zessar watershed, Tunisia Agricultural Water Management 176:191-202 728 Ahmed FE, Hashaikeh R, Hilal N (2019) Solar powered desalination – Technology, energy and future outlook 729 Desalination:54-76 doi:10.1016/j.desal.2018.12.002 730 Amer K, Adeel Z, Böer B, Saleh W (eds) (2017) The Water, Energy, and Food Security Nexus in the Arab Region. Springer, 731 Attia F, Fahmi H, Gambarelli J, Hoevenaars R, Slootweg R, AbdelDayem S (2005) WDWCIARP drainframe analysis, main 732 report. Unpublished. 733 Badiuzzaman P, McLaughlin E, McCauley D (2017) Substituting freshwater: Can ocean desalination and water recycling 734 capacities substitute for groundwater depletion in California? Journal of Environmental Management 203:123-735 135 doi:10.1016/j.jenvman.2017.06.051 736 Basnyat DB, Gadain HM (2009) Hydraulic behaviour of the Juba and Shabelle Rivers: Basic analysis for irrigation and flood 737 management purposes. FAOSWALIM (GCP/SOM/047/EC), Nairobi, Kenya: 738 BWE (2015) Water Supply and Wastewater Systems Master Plan for the Bekaa Water Establishment. 739 Carpenter SR, Pingali PL, Bennett EM, Zurek MB (2005) Ecosystems and human well-being: Scenarios vol 2. Island Press, 740 New York, NY 741 Cherlet M, Hutchinson C, Reynolds J, Hill J, Sommer S, von Maltitz G (eds) (2018) World Atlas of Desertification. Publication 742 Office of the European Union, JRC, Luxembourg 743 Croitoru L, Sarraf M, Ghariani F, Matoussi MS, Daly-Hassen H (2010) Water Degradation: The Case of Tunisia. In: Croitoru L, 744 Sarraf M (eds) The cost of environmental degradation : case studies from the Middle East and North Africa 745 (English). vol Directions in development ; environment. . World Bank, Washington, DC: , pp 11-36 746 Daher BT, Mohtar RH (2015) Water-energy-food (WEF) Nexus Tool 2.0: guiding integrative resource planning and decision-747 making Water International 40:748-771 doi:10.1080/02508060.2015.1074148 748 desalination.biz (2018) KfW finances €82 million Tunisian desalination project. 749 EEAA (2016) EGYPT THIRD NATIONAL COMMUNICATION Under the United Nations Framework Convention on Climate 750 Change. Egyptian Environmental Affairs Agency (EEAA), Cairo, Egypt 751 El-Agha DE, Closas A, Molle F (2016) Below the radar: the boom of groundwater use in the central part of the Nile Delta in 752 Egypt Hydrogeology doi:10.1007/s10040-017-1570-8 753 El Gafy I, Grigg N, Reagan W (2017a) Dynamic Behaviour of the Water-Food-Energy Nexus: Focus on Crop Production and 754 Consumption Irrigation and Drainage 66:19-33 755 El Gafy I, Grigg N, Reagan W (2017b) Water-food-energy nexus index to maximize the economic water and energy 756 productivity in an optimal cropping pattern Water International 42:495-503 757 Endo A, Tsurita I, Burnett K, Orenciod PM (2017) A review of the current state of research on the water, energy, and food 758 nexus Journal of Hydrology: Regional Studies 11:20-30 759 ESCWA (2009) ESCWA water development report 3: role of desalination in addressing water scarcity. United Nations. 760 http://bit.ly/2FuCq97, 761 ESCWA (2016) Developing the Capacity of ESCWA Member Countries to Address the Water and Energy Nexus for Achieving 762 Sustainable Development Goals. ESCWA, Beirut 763 Fraenkel PL (1986) Water lifting. Rome. United Nations Food and Agriculture Organization, Rome, Italy 764 GIBB (2004) The Merti Aquifer Study. GIBB Africa Ltd. UNICEF Kenya Country Office, 765 Gies L, Agusdinata DB, Merwade V (2014) Drought adaptation policy development and assessment in East Africa using 766 hydrologic and system dynamics modeling Natural Hazards 74:789-813 767 Giordano M, Villholth K (eds) (2007) The agricultural groundwater revolution: opportunities and threats to development. 768 Comprehensive Assessment of Water Management in Agriculture Series 3. CABI, Wallingford, UK 769 GoK (2012) Master Plan for Conservation and Sustainable Management of Water Catchment Areas in Kenya. Ministry of 770 Environment and Natural Resources, Republic of Kenya, 771 GoK (2014) Republic of Kenya Project on Capacity Development for Effective Flood Management in Flood Prone Area 772 ISIOLO RIVER BASIN INTEGRATED FLOOD MANAGEMENT PLAN. Japan International Cooperation Agency, NEWJEC 773 Inc. - Ministry of Environment, Water and Natural Resources - Water Resources Management Authority, 774 GoK (2018) ISIOLO COUNTY INTEGRATED DEVELOPMENT PLAN, CIDP 2018-2022. Isiolo County Government, Nairobi, Kenya

775 Hadded R et al. (2013) A Decision Support System to Manage the Groundwater of the Zeuss Koutine Aquifer Using the 776 WEAP-MODFLOW Framework Water Resources Management 27:1981-2000 777 Harou J, King-Okumu C, Orindi V (2017) Mitigating and adapting to 'arid' climate extremes: Locally prioritized ecosystem-778 based adaptations at the water-energy nexus (PPT). Paper presented at the Symposium Droughts Resilience and 779 Climate Change in Africa, Nairobi, Kenya, 7th-9th November 2017, Nairobi, 780 Hoff H (2011) Understanding the Nexus. Background Paper for the Bonn 2011 Conference: The Water, Energy and Food 781 Security Nexus. . Stockholm Environment Institute, Stockholm 782 Hoff H, Al-Zubari W, Mansour L, Abaza H, Biad M, Al Ouran N, Ulrich A (2017) Mainstreaming the Water-Energy-Food 783 Security Nexus into Policies and Institutions in the MENA Region: Nexus Evidence Base. GIZ, GFA, 784 IEA (2016) Water Energy Nexus. In: The World Energy Outlook 2016. International Energy Agency (IEA) Paris: OECD/IEA, p 785 63 786 IFAD (2016) SAIL Supervision Report 2016 Supervision report. IFAD, Rome 787 IPCC (2003) Good Practice Guidance for Land Use, Land-Use Change and Forestry. Institute for Global Environmental 788 Strategies (IGES), Hayama, Kanagawa, Japan 789 IPCC (2018) GLOBAL WARMING OF 1.5 °C - Summary for Policymakers Intergovernmental Panel on Climate Change, 790 IUCN (2015) Nexus trade-offs and strategies for addressing the water, agriculture and energy security nexus in Africa 791 IWA/IUCN/ICA 792 Jaafar H, King-Okumu C, Haj-Hassan M, Abdallah C, El-Korek N, Ahmad F (2016) Water resources within the Upper Orontes 793 and Litani Basins: a balance, demand and supply analysis amid the Syrian refugees crisis. IIED, London 794 Jaafar HH (2017) Water Resources Within the Orontes and Litani Basins - Water Balance, Demand and Supply Analysis IIED, 795 London 796 Jaafar HH, Ahmad FA (2019 in press) Time series trends of Landsat-based ET using automated calibration in METRIC and 797 SEBAL: The Bekaa Valley, Lebanon Remote Sensing of Environment 798 doi:https://doi.org/10.1016/j.rse.2018.12.033. 799 Jaafar HH, Woertz E (2016) Agriculture as a funding source of ISIS: A GIS and remote sensing analysis Food Policy 64:14-25 800 Jaafar HH, Zurayk R, King C, Ahmad F, Al-Outa R (2015) Impact of the Syrian conflict on irrigated agriculture in the Orontes 801 Basin International Journal of Water Resources Development 31:436-449 802 Jarray H, Zammouri M, Ouessar M, Zerrim A, Yahyaoui H (2017) GIS based DRASTIC model for groundwater vulnerability 803 assessment: Case study of the shallow mio-plio-quaternary aquifer (Southeastern Tunisia) Water Resources 804 44:595-603 805 Jarso I, Tari D, King-Okumu C (2017) Recommendations to the County Government of Isiolo for preparation of a strategic 806 plan on water, energy and climate change. IIED, London 807 Jillo B et al. (2016) Cracking the climate-water-energy challenge in the drylands of Kenya. IIED, London 808 Jomaa I, M. SS, R. J (2015) Sharp expansion of extensive groundwater irrigation, semi-arid environment at the northern 809 Bekaa Valley, Lebanon. Natural Resources 6:381-390 810 Kaur S, Aggarwal R, Lal R (2016) Assessment and Mitigation of Greenhouse Gas Emissions from Groundwater Irrigation 811 Irrigation and Drainage 65:762-770 812 King-Okumu C, Harou J, Orindi V (2018) Mitigating and adapting to 'arid' climate extremes: Locally prioritized ecosystem-813 based adaptations at the water-energy nexus. In: Leal W (ed) Handbook of Climate Change Resilience Springer, 814 Nairobi. 815 King-Okumu C, Jillo B, Kinyanjui J, Jarso I (2017) Devolving water governance in the Kenyan Arid Lands: from top-down 816 drought and flood emergency response to locally driven water resource development planning International 817 Journal of Water Resources Development 818 King C, Jaafar H (2015) Rapid Assessment of the Water-Energy-Food-Climate Nexus in 6 Selected Basins of North Africa and 819 West Asia undergoing Transitions and Scarcity Threats International Journal of Water Resources Development 820 31:343-359 821 King C, Salem B (2012) A socio-ecological investigation of options to manage groundwater degradation in the western 822 desert, Egypt Ambio 41:490-503 823 King C, Salem B (2013) Assessing the cost of groundwater degradation in the urbanizing desert area of Wadi El Natrun. In: 824 The Economy of Green Cities. Springer, pp 295-311 825 KNBS (2018) Chapter 18: Environmental Economic Accounts: Energy Accounts. In: Statistics KNBo (ed) Economic Survey, 826 2018. Nairobi, pp 282-827 Michael A, Khepar S, Sondhi S (2008) Water Wells & Pumps. Tata McGraw-Hill Education: McGraw Hill Education (India) Pvt 828 Ltd., 829 Michalscheck M, Petersen G, Gadain H (2016) Impacts of rising water demands in the Juba and Shabelle river basins on 830 water availability in south Somalia Hydrological Sciences Journal 61:1877-1889 831 MoE-EU-UNDP (2014) Lebanon Environmental Assessment of the Syrian Conflict and Priority Interventions 832 MoE (2016) Egypt's third national communication to the UNFCCC. Ministry of Environment, Egypt, Cairo 833 MoE/UNDP/GEF (2015a) National Greenhouse Gas Inventory Report and Mitigation Analysis for the Agriculture Sector in 834 Lebanon. Beirut, Lebanon 835 MoE/UNDP/GEF (2015b) National Greenhouse Gas Inventory Report and Mitigation Analysis for the Land Use, Land-Use 836 Change and Forestry Sector in Lebanon. Beirut, Lebanon

837 MoE/UNDP/GEF (2016) Lebanon's third national communication to the UNFCCC. Beirut, Lebanon. Ministry of Environment, 838 Lebanon, Beirut 839 MoE/UNDP/GEF (2017) Lebanon's Second Biennial Update Report to the UNFCCC. Ministry of Environment, Beirut, 840 Lebanon 841 MoEW/UNDP (2014) Assessment of Groundwater Resources of Lebanon. MoEW/UNDP, Beirut 842 Molle F, Gaafar I, El-Agha DE, Rap E (2016) Irrigation Efficiency and the Nile Delta Water Balance - Technical Report -843 December 2016. IWMI, Cairo 844 Molle F, Gaafar I, El-Agha DE, Rap E (2018) The Nile delta's water and salt balances and implications for management 845 Agricultural Water Management 197:110-121 846 Mutiga JK, Mavengano ST, Zhongbo S, Woldai T, Becht R (2010) Water Allocation as a Planning Tool to Minimise Water Use 847 Conflicts in the Upper Ewaso Ng'iro North Basin, Kenya Water Resources Management 24:3939-3959 848 MWRI (2010) Draft strategy for water resource management until 2050 (in Arabic). . MWRI, Cairo: 849 Ndirangu W (2017) Catalysing low cost green technologies for sustainable water service delivery in Kenya - Feasibility Study 850 Report. CTCN, UNEP DTU & WSTF, Nairobi 851 Nelson G, Robertson R (2008) Personal communication. Planning Commission (India) Government of India, New Delhi 852 Nelson G, Robertson R, Msangi S, Zhu T, Liao X, Jawajar P (2009) Greenhouse gas mitigation. Issues for Indian agriculture. 853 International Food Policy Research Institute, Environment and Production Technology Division, Washington, DC. OECD/IEA (2017) Energy Access Outlook. OECD International Energy Agency Paris 854 855 Oord A, Collenteur R, Tolk L (2014) Hydrogeological Assessment of the Merti Aquifer, Kenya. 856 Ouessar M, Bruggeman A, Abdelli F, Mohtar RH, Gabriels D, Cornelis WM (2009) Modelling water-harvesting systems in the 857 arid south of Tunisia using SWAT Hydrology and earth system sciences 13:2003 -2021 858 Ouessar M et al. (2004) An Integrated Approach for Impact Assessment of Water Harvesting Techniques in Dry Areas: The 859 case of Oued Oum Zessar Watershed (Tunisia) Environmental Monitoring and Assessment 99:127-140 860 RoL (2018a) Capital Investment Programme Report, CEDRE April 2018. Goverment of Lebanon, Beirut 861 RoL (2018b) Lebanon VNR Agenda 2030 and SDGs Main Messages Beiru 862 Saadé-Sbeih M, Haj Asaad A, Shamali O, Zwahlen F, Jaubert R (2018) Groundwater balance politics: Aquifer 863 overexploitation in the Orontes River Basin Water Alternatives 11:663-683 864 Salem B (2014) Omayed Biosphere Reserve, Egypt. In: Schaaf T, Cardenas MR, Lee C (eds) Innovative ways for sustainable 865 use of drylands - final report of the SUMAMAD project. United Nations Educational, Scientific and Cultural 866 Organization, Paris, pp 60-79 867 Salim MG (2012) Selection of groundwater sites in Egypt, using geographic information systems, for desalination by solar 868 energy in order to reduce greenhouse gases Journal of Advanced Research 3:11-19 869 Shah T (2009) Climate change and groundwater: India's opportunities for mitigation and adaptation Environmental 870 Research Letters 4:035005. 871 Shah T, Durga N, Rai GP, Verma S, Rathod R (2017) Promoting Solar Power as a Remunerave Crop Economic and Polical 872 Weekly 52:14-19 873 Shouman ER, El Shenawy ET, Badr MA (2016) Economics analysis of diesel and solar water pumping with case study water 874 pumping for irrigation in Egypt International Journal of Applied Engineering Research 11:950-954 875 Swiderska K, King-Okumu C, Islam MM (2018) Ecosystem-based adaptation: a handbook for eba in mountain, dryland and 876 coastal ecosystems. IIED, London 877 UN E, FAO I (2014) System of Environmental-Economic Accounting 2012: Central Framework. 878 UNWater (2018) Sustainable Development Goal 6 Synthesis Report on Water and Sanitation Draft – 2nd May 2018 United 879 Nations, New York 880 Velis M, Conti KI, Biermann F (2017) Groundwater and human development: synergies and trade-offs within the context of 881 the sustainable development goals Sustainability Science 12:1007-1017 882 Verma S, Durga N, Shah T (2018) Solar Irrigaon Pumps and India's Energy-Irrigaon Nexus IWMI-TATA Water Policy Research 883 Highlight-02:6 884 Villholth KG, Lopez-Gunn E, Conti K, Garrido A, Gun JVD (eds) (2017) Advances in Groundwater Governance. 1st Edition. 885 CRC Press. 886 Wakeford JJ (2017) The Water-Energy-Food Nexus in a Climate-Vulnerable, Frontier Economy: The Case of Kenya - Report 887 Prepared for the United Kingdom Department for International Development. 888 WOCAT (2013) Water harvesting - Guidelines to good practice. CDE, Berne 889 Woldemariyam F, Ayenew T (2016a) Application of hydrochemical and isotopic techniques to understand groundwater 890 recharge and flow systems in the Dawa River basin, southern Ethiopia Environmental Earth Sciences 75 891 doi:10.1007/s12665-016-5777-0 892 Woldemariyam F, Ayenew T (2016b) Identification of hydrogeochemical processes in groundwater of Dawa River basin, 893 southern Ethiopia Environmental Monitoring and Assessment 188 doi:10.1007/s10661-016-5480-3 894 WRMA (2013a) Final Report - Surface and Groundwater Assessment and Planning in Respect to the Isiolo County Mid Term 895 ASAL Program Study Volume 1 Main Report. Earth Water Ltd, 896 WRMA (2013b) Part G: Ewaso Ng'iro North Catchment Area. In: National Water Master Plan 2030. Nairobi, p 112 897 WRMA (2014) Ewaso Ng'iro North Catchment Area (ENNCA) Catchment Management Strategy 2015-2022. Water 898 Resources Management Authority, Republic of Kenya, Nairobi, Kenya

- WRMA (2016a) Surface water and groundwater resources assessment in Wajir County for decision making: final report.
 Geekan Kenya Ltd for the Kenyan Water Resource Management Authority (WRMA), Nairobi
- 901 WRMA (2016b) Water Resources Assessment for Decision Making in Garissa County FINAL REPORT JUNE 2016. MTAP,
 902 Nairobi
- 903 WWDR (2014) The United Nations World Water Development Report 2014: Water and Energy Volume 1. World Water
 904 Development Report, UNESCO, UNWater, Paris
- Yang J et al. (2018) Quantifying the Sustainability of Water Availability for the Water-Food-Energy-Ecosystem Nexus in the
 Niger River Basin Earth's Future 6:1292-1310 doi:10.1029/2018EF000923





Figure 2: Registered water level variation in the Grés de Trias aquifer at Wadi Hjar

Source: WADIS-MAR project