

Article (refereed)

Kuliskova, P.; Horky, P.; Slavik, O.; Jones, J.I.. 2009 Factors influencing movement behaviour and home range size in ide *Leuciscus idus*. *Journal of Fish Biology*, 74 (6). 1269-1279. 10.1111/j.1095-8649.2009.02198.x

© 2009 The Fisheries Society of the British Isles

This version available at <http://nora.nerc.ac.uk/5708/>

NERC has developed NORA to enable users to access research outputs wholly or partially funded by NERC. Copyright and other rights for material on this site are retained by the authors and/or other rights owners. Users should read the terms and conditions of use of this material at <http://nora.nerc.ac.uk/policies.html#access>

This document is the author's final manuscript version of the journal article, incorporating any revisions agreed during the peer review process. Some differences between this and the publisher's version remain. You are advised to consult the publisher's version if you wish to cite from this article.

<http://www.blackwell-synergy.com>

Contact CEH NORA team at
nora@ceh.ac.uk

Editorial Manager(tm) for Journal of Fish Biology
Manuscript Draft

Manuscript Number: MS 08-244R1

Title: Factors influencing movement behaviour and home range size in ide *Leuciscus idus*

Short Title: Behaviour of ide *Leuciscus idus*

Article Type: Regular paper

Keywords: cyprinids; home range; migration; insectivores; lowland river.

Corresponding Author: Petra Kuliskova, MSc.

Corresponding Author's Institution: TGM - Water Research Institute

First Author: Petra Kuliskova, MSc.

Order of Authors: Petra Kuliskova, MSc.; Ondrej Slavik, PhD; Pavel Horky, PhD; Iwan J Jones, Dr

Abstract: 17 individuals of ide *Leuciscus idus* were radio-tracked weekly from September 2003 to September 2004 in the Elbe River, Czech Republic to examine migration patterns and the influence of environmental factors on the diurnal behaviour. Of the 10 environmental factors measured, ide were significantly influenced by turbidity, which increased diurnal movement and the home range size of the species. The peak of longitudinal movement occurred in the spring, indicating pre-spawning migration. Migrating fish moved downstream and later returned upstream to the vicinity of their original locations, displaying a homing behaviour.

1 **Factors influencing movement behaviour and home range size in ide**

2 *Leuciscus idus*

3
4 P. KULÍŠKOVÁ*¹, P. HORKÝ¹, O. SLAVÍK^{1Φ}, J. I. JONES[‡]

5 ¹ TGM – Water Research Institute, Podbabská 30, 160 62 Prague 6, Czech Republic

6 *Department of Ecology, Charles University, Faculty of Science, Viničná 7, 128 44 Prague 2,
7 Czech Republic

8 ‡ CEH Wallingford, Crowmarsh Gifford, Wallingford, UK

9 Φ University of South Bohemia České Budějovice, Research Institute of Fish Culture and
10 Hydrobiology, 389 25 Vodňany, Czech Republic

11
12 Running title: Behaviour of ide

13
14 17 individuals of ide *Leuciscus idus* were radio-tracked weekly from September 2003 to
15 September 2004 in the Elbe River, Czech Republic to examine migration patterns and the
16 influence of environmental factors on the diurnal behaviour. Of the 10 environmental factors
17 measured, ide were significantly influenced by turbidity, which increased diurnal movement
18 and the home range size of the species. The peak of longitudinal movement occurred in the
19 spring, indicating pre-spawning migration. Migrating fish moved downstream and later
20 returned upstream to the vicinity of their original locations, displaying a homing behaviour.

21
22 Key words: cyprinids, home range, migration, insectivores, lowland river

23
24 *Author to whom correspondence should be addressed. Tel.: +420 220 197 402; email:
25 petra_kuliskova@vuv.cz;

INTRODUCTION

27

28

29 It has been suggested that the daily activity rhythms of fish are the result of a complex trade-
30 off between growth and survival, which takes into account diel fluctuation in food
31 availability, food capture efficiency and predation risk (Metcalf *et al.*, 1999). Usually such
32 diel activity patterns are closely related to variations in the physical environment. For
33 example, a temperature-dependent shift in diel activity is supposed to be a consequence of
34 higher predation risk in cold water (Webb, 1978; Fraser *et al.*, 1995). Changes in diel activity
35 patterns have been related to the light intensity and duration corresponding with the daytime
36 (Harvey & Nakamoto, 1999), season (David & Closs, 2003) or the moon phase (Horký *et al.*,
37 2006). However, other environmental conditions, such as the influence of water turbidity
38 (Benfield & Minello, 1996; Sweka & Hartman, 2003) and water flow conditions, may have an
39 effect (Harvey & Nakamoto, 1999; Slavík *et al.*, 2007).

40 Ide, *Leuciscus idus* (L.), is a species of benthopelagic, riverine cyprinid inhabiting
41 deeper, slower flowing reaches of lowland middle-sized rivers east of the Rhine basin in
42 Europe to Siberia (Maitland & Campbell, 1992). It achieves a maximum size of 53 cm total
43 length (L_T), a body mass of 2.0 kg, and a recorded age of 14 years. Ide are visually oriented
44 feeders and predominantly consumes insects, although coarse fish or plant material might be
45 occasionally consumed (Cala, 1970). Ide spawn in spring and belonging to the phyto-
46 lithophilic spawning group (Balon, 1975). Their migratory pattern has been described as
47 potamodromous with a prevailing upstream migration (Cala, 1970; Müller, 1986). According
48 to Winter & Fredrich (2003), who observed migrations of ide in the middle reaches of the
49 River Elbe in Germany and the River Vecht in the Netherlands, ide is a flexible species
50 capable of adapting its movement pattern to different conditions of river systems. To
51 investigate within catchment variation in the migratory patterns of ide we tracked the

52 movements of individuals in a low stream order, upstream section of the Elbe River, closer to
53 the source than previous investigations (Winter & Fredrich, 2003). To assess which
54 environmental factors influence the migration and diurnal behaviour of ide, 17 specimens
55 were radio-tracked weekly from September 2003 to September 2004 in the Elbe River, Czech
56 Republic.

57

58

MATERIALS AND METHODS

59

60 STUDY AREA

61 The study was carried out on the upper part of the River Elbe, Czech Republic. The
62 river rises at 1383 m above sea level. It has a total length of 1091 km with a catchment area of
63 148,268 km². The Czech portion of the river is 368 km long and has a catchment area of
64 51,394 km². The primary river stretch studied was about 40 km long, from the weir at Střekov
65 (distance from the source 320 km; 50°38' N; 14°03' E) to the frontier with Germany (Fig. 1).
66 During spawning migrations, the stretch studied was extended as far as Meissen, Germany
67 (distance from the source 410 km; 51° 81' N; 13° 28' E) as fish were followed. The river
68 width in the area studied was 100 to 150 m, and the riverbanks have little aquatic vegetation
69 and are reinforced with rocks and concrete. The water was up to 6 m deep and no submergent
70 vegetation or floating plants were recorded. Across the whole study period, the average flow
71 was 293 m³ s⁻¹, with the maximum in winter (748 m³ s⁻¹) and the minimum in early autumn
72 (79 m³ s⁻¹).

73

74 FISH CAPTURE AND TAGGING

75 Fish were sampled by electrofishing (650 V, 4 A, pulsed D.C.) and seventeen
76 individuals were radio tagged. All fish were caught and released at 5 km long river stretch

77 (Fig. 1). The individuals were measured to the nearest mm (mean standard length 378 mm L_S ,
78 ranging from 285 to 450 mm) and weighed to the nearest g (mean fish body mass 755 g,
79 ranging from 450 to 1240 g). Fish were anaesthetized with 2-phenoxy-ethanol (0.2 ml l^{-1}).
80 Radio transmitters (MCFT 3B, 11 g in air, 14 x 43 mm, with an operational life estimated to
81 be 399 days; MCFT 3EM, 8.9 g in air, 11 x 49 mm, with an operational life estimated to be
82 278 days; Lotek Engineering, Inc., Canada) were implanted into the body cavity through a
83 midventral incision that was closed by three separate stitches, using a sterile, braided,
84 absorbable suture (Ethicon Coated Vicryl). The mass of the transmitter never exceeded 2 % of
85 the fish body mass in the air (Winter, 1983). Fish were held until they had recovered their
86 equilibrium and showed spontaneous swimming activity (*c.* 5 min. after surgery), then
87 released close to the site of capture. The transmitters had external antennae and their potential
88 range was approximately 300 m depending on the gain of receiver and tracking conditions.

89

90 SAMPLING PROCEDURES

91 All fish were tracked from a boat weekly during the period from 11 September 2003 to
92 21 September 2004. Once all the fish were positioned, one individual was randomly chosen
93 for a 24h tracking cycle. Fish positions were determined once in each three-hour period over a
94 diel cycle (0600 – 0859, 0900 – 1159, 1200 – 1459, 1500 – 1759, 1800 – 2059, 2100 – 2359,
95 2400 – 0259, 0300 – 0559 hours) using a GPS receiver. The interval between measurements
96 varied slightly depending on the tracking conditions (3 hours \pm 20 min.). The fish were
97 located using landmarks and positioned with the help of a GPS (GPS map 76S, Garmin Ltd.,
98 USA) using two radio receivers (Lotek SRX_400 receiver firmware versions W5 and W31)
99 and three-element Yagi antennas equipped with a compass. The fish direction was determined
100 by the double lateral extinction technique (bearing on the bisecting line of the two extinction
101 axes; Winter *et al.*, 1978). A computer program was developed to obtain fish position

102 coordinates and plot them on the map using the biangulation method proposed by White &
103 Garrot (1990).

104

105 HABITAT MEASUREMENT

106 Water temperature ($^{\circ}\text{C}$), dissolved oxygen concentration (mg l^{-1}), conductivity (μS), pH, and
107 turbidity (NTU) were measured by microprocessors (Oxi 196; pH/Cond 340i/SET; TURB
108 355 T; WTW, Germany). Light intensity (Ev) was measured by a SECONIC Super Zoom
109 Master L-68 (Seconic, Tokyo, Japan) at the expected locations of individuals during each
110 positioning. Measurements of the atmospheric pressure and the moon phase were conducted
111 with help of the Remote Weather station BAR 928 H (Huger Electronics, Germany). The Elbe
112 River Authority measured water flow daily at a gauging station located within the study
113 stretch.

114

115 DATA ANALYSES

116 Short term movements were defined as the distance (m) between the fish positions determined
117 in two subsequent three-hour intervals over a 24 hour cycle and are henceforth referred to as
118 “diurnal movements”. Although the fish were located every time, in several cases the signal
119 was so weak that triangulation could not be precise: These occasions were excluded from
120 further analysis. Longer term movements were determined from the difference (m) between
121 the locations of a fish in two successive week intervals and, henceforth, are referred to as
122 “longitudinal movements”. Data on fish movements were analyzed using Map Source Version
123 5.3 (Garmin Ltd., USA). The sizes of home ranges were determined using the Minimum
124 Convex Polygon method (Mohr, 1947). Fish that were used for home range analyses were
125 suggested to occupy home range; i.e. fish could not move during two subsequent weeks in the
126 longitudinal direction more than its usual extent of diurnal movements across the twenty-four-

127 hour cycle. Furthermore, fish that moved for the whole twenty-four-hour cycle in only one
128 direction (upstream or downstream) was suggested to exhibit a mobile or emigration phase of
129 a home range shift and was subsequently excluded from the analyses. Data concerning light
130 intensity were first entered into the analysis as the absolute values of illumination ($1 \text{ Ev} \cong 5$
131 lx ; $y = 0.6211e^{0.6943x}$, where $y = \text{lx}$, $x = \text{Ev}$), referred to as ‘intensity of illumination’.
132 Furthermore, three intervals with different light intensity were determined across the twenty-
133 four-hour cycle: twilight (light intensity ranged between 1 – 6 Ev), day (above 6 Ev), and
134 night (below 1 Ev); in further analysis, these categories will be referred to as ‘light intervals’.

135

136 STATISTICAL ANALYSES

137 Associations between the variables were tested using the Linear Mixed Model (LMM).
138 Separate models were applied for the following dependent variables: diurnal movements
139 (LMM I), home range size (LMM II) and longitudinal movements (LMM III). All of the data
140 were square root transformed to achieve normality before analyses. To account for the
141 repeated measurements of the same individuals across the period of observation, analyses
142 were performed using mixed model analysis with individual fish and date nested within
143 individual fish (LMM I, II) and individual fish and date nested within individual fish (LMM
144 III) as a random factors, using PROC MIXED (SAS, version 9.1).

145 PROC MIXED is the way to cope with repeated-measures experiments with people or
146 animals as subjects, where subjects are declared random because they are selected from the
147 larger population to which you want to generalize (SAS Institute Inc., 2004). For the LMM I
148 model, fixed effect used were the classes ‘moon phase’ (8 levels), ‘season’ (spring, summer,
149 autumn, winter), and ‘light interval’ (day, night, twilight), and the continuous variables were
150 ‘turbidity’ (range 5.5 – 44 NTU), ‘fish mass’ (1293–3946 g), ‘water temperature’ (0–24 °C),
151 ‘water flow’ ($79\text{--}748 \text{ m}^3 \text{ s}^{-1}$), ‘atmospheric pressure’ (992–1033 hPa), ‘conductivity’ (332–

152 425 μS), ‘light intensity’ (0 – 15.1 Ev), and ‘dissolved oxygen’ (5.5–12.9 mg l^{-1}). For the
153 LMM II - III models, fixed effects used were the same as for the LMM I model except for the
154 ‘light interval’ (day, night, twilight) and ‘light intensity’ (0 – 15.1 Ev) that were excluded
155 from the analyses. The significance of each fixed effect (including interactions) in the
156 analyses was assessed by the F-test, upon sequential dropping of the least significant effect,
157 starting with a full model. Fixed effects and their interactions that were not statistically
158 significant are not discussed further. In unbalanced designs with more than one effect, the
159 arithmetic mean for a group may not accurately reflect the response for that group, since it
160 does not take other effects into account. Therefore, the least-squares-means (LSMEANs) were
161 used. LSMEANs (further referred to as ‘adjusted means’) are, in effect, within-group means
162 appropriately adjusted for the other effects in the model. Adjusted means (Adj P) were
163 computed for each class; differences between classes were tested by the *t*-test. For multiple
164 comparisons, we used the Tukey-Kramer adjustment. Associations between the dependent
165 variable and other continuous variables were estimated by fitting a random coefficient model
166 using PROC MIXED as described by Tao *et al.* (2002). With this random coefficient model,
167 we calculated predicted values for the dependent variable and plotted them against the
168 continuous variable with predicted regression lines. The degrees of freedom were calculated
169 using the Kenward-Roger method (Kenward & Roger, 1997).

170

171

172

RESULTS

173

174 Final LMM models contained the fixed factors ‘turbidity’ for diurnal movements (LMM I),
175 ‘turbidity’ and ‘season’ for home range area size (LMM II), and ‘season’ for longitudinal
176 movements (LMM III). Details of the models are shown in Table I. Descriptive data of the

177 extent of diurnal movements, longitudinal movements, total distance migrated during
178 spawning and home range per individual are provided in Table II. The other environmental
179 variables tested (water temperature, dissolved oxygen, conductivity, pH, atmospheric
180 pressure, moon phase and light intensity) were not found to have a significant effect.

181

182 DIURNAL MOVEMENTS AND HOME RANGE SIZE OF IDE

183 During the whole study fish did not remain at one exact position, however they occupied
184 defined home ranges between which they relocated (e.g. during spring migration). Mean
185 home range size was $19,495.8 \pm 13,890.9 \text{ m}^2$ (Table II), but both diurnal movement and home
186 range size appeared to vary in a consistent manner. Repeated measurements indicated that
187 both diurnal movement [Fig. 2(a)] and the home range size [Fig. 2(b)] of ide increased with
188 increasing turbidity. The relationship between flow and turbidity was not statistically
189 significant; increased turbidity was a consequence of both surface run-off and phytoplankton
190 growth. Home range size was significantly smaller (Tukey-Kramer Adj. $P < 0.05$) during
191 winter than other seasons [Fig. 3(a)].

192 Final GLMM I model indicated the influence of the light interval nested within season on
193 the diurnal movements of ide (Table I); however, differences among classes were
194 insignificant, and hence the character of dependence was not possible to determine (Tukey-
195 Kramer Adj. $P > 0.05$).

196

197 LONGITUDINAL MOVEMENTS OF IDE

198 Longitudinal movements of the ide were significantly larger (Tukey-Kramer Adj. $P < 0.01$) in
199 the spring, with non-significant differences among other seasons [Fig. 3(b)]. Almost all
200 individuals, with one exception, displayed downstream spring migrations, most of them
201 remaining within Czech part of Elbe River (40 km long; Fig. 1). Six individuals moved to

202 near the confluence of the River Bílina at Ústí nad Labem, including one individual that
203 moved 19 km upstream to reach this spot (the only upstream migrating individual). Four
204 individuals moved downstream to near the town of Malé Březno and a further two to near the
205 border with Germany. Five individuals undertook longer migrations (68 – 100 km) to reach
206 spawning sites near Dresden and Meissen in the German part of River Elbe (Fig. 1). The
207 individuals that undertook longest migration started their run earliest, at the end of February.
208 All final destination of migration were shallower riffles with a gravel substrate. Later in the
209 season, ide displayed homing behaviour and returned to within 0.5 – 2 km of the starting
210 position.

211

212 DISCUSSION

213

214 Behaviour of visually oriented animals is known to be affected by visibility, as they rely on
215 visual cues for orientation and feeding. In aquatic ecosystems, the visibility is determined not
216 only by the light intensity but also by the water turbidity (Benfield & Minello, 1996).
217 Turbidity imposes a considerable environmental constraint with a potential to affect whole
218 fish communities (Colby *et al.*, 1972; Diehl, 1988). It may shape the habitat choice patterns
219 (Miner & Stein, 1996), social interactions (Valdimarsson & Metcalfe, 2001) or reproductive
220 behaviour of the fish, in terms of reduced sexual selection (Järvenpää & Lindström, 2004;
221 Heubel & Schlupp, 2006). Increased turbidity influences visually-oriented fish by decreasing
222 their visual range (Utne-Palm, 2001), typically affecting foraging efficiency by reducing the
223 distance at which a predator detects prey (Benfield & Minello, 1996; Sweka & Hartman,
224 2003). Benfield & Minello (1996) evaluated the influence of turbidity on predation rates of
225 gulf killifish, *Fundulus grandis* Baird and Girard, and that significantly fewer prey were
226 consumed in tanks containing turbid water. Brook trout, *Salvelinus fontinalis* L., become

227 more active in higher turbidity, thus increasing the chance of encountering potential prey by
228 enlarging the total volume of water searched (Sweka & Hartman, 2001a). Hence, we suggest
229 that ide extend their diurnal movements and home range size as a result of the reduced
230 foraging success in turbid water.

231 In riverine systems, increased turbidity is usually associated with increased flow
232 during hydrologic events (Sahoo *et al.*, 2006). However, low discharge may have the opposite
233 effect: increased water residence time during low water flow may allow the buildup of
234 phytoplankton biomass (Lane *et al.*, 2007). Here there was no significant relationship between
235 discharge and turbidity, suggesting both potential sources and that the behaviour of ide was
236 influenced by the water turbidity per se. Home range size varied consistently with season and
237 turbidity. The influence of turbidity is likely to be due to its effect on visibility. Reduced diurnal
238 movement in winter may be due to lower food availability and/or temperature related
239 metabolism.

240 The winter season is a period of reduced activity in cyprinid fish (Bauer & Schlott,
241 2004). They tend to remain in areas with the most appropriate conditions for wintering, as
242 was shown for example in bream *Abramis brama* L. migrating into lentic refugia (Molls &
243 Neumann, 1994). A restricted home range may be a direct consequence of a reduced
244 metabolic rate linked to low temperatures as well as a result of efficient energy conservation
245 or the use of locally restricted refuge during harsh conditions (Brown & Mackay, 1995;
246 Hiscock *et al.*, 2002).

247 Many freshwater fish species, including cyprinids, undertake long distance migrations
248 during the breeding season (Baras & Cherry 1990; Lucas 2000). Previous reports of ide
249 indicated an upstream pre-spawning migration followed by downstream movement after
250 spawning (Cala, 1970; Müller, 1986), including studies in the middle reaches of the River
251 Elbe (Winter & Fredrich, 2003). In contrast, we observed that ide undertook similar long

252 distance migrations in spring but in the opposite direction, i.e. downstream during spring and
253 returning upstream towards formerly occupied areas later in the season. These findings are
254 partly consistent with Cala (1970) from Kävlingeån in Sweden, where ide also displayed large
255 downstream migration in spring. However, the latter case is more complicated as the fish
256 migrated downstream to coastal waters in the spring, where they remained for the consecutive
257 summer, only returning to the river in autumn (Cala *op. cit.*). Such inconsistency in the
258 direction of migration may indicate that the movements of fish are shaped by multiple factors
259 that vary even within river systems. An obvious constraint is the presence of lateral
260 obstructions that hamper fish migration (Lucas & Frear, 1997; Horký *et al.*, 2007), although
261 in our study no fish were observed to move to the vicinity of the weir at Strekov during spring
262 migration. The location of suitable spawning areas (Pollux *et al.*, 2006) and channel
263 morphology (Lau *et al.*, 2006) may also be essential.

264 Although ide are declining in numbers, classified as vulnerable by IUCN Red List
265 criteria (2001) and protected as an endangered species (Lusk *et al.*, 2004), few references
266 regarding its behaviour exist (Cala 1970; Winter & Fredrich 2003). Our data demonstrate that
267 the turbidity may substantially influence the movement patterns of this species. As turbidity is
268 influenced by both eutrophication and changes in land-use (Duchrow & Everhart, 1971), and
269 increased turbidity has a negative effect on foraging success and growth of fish (Sweka &
270 Hartman 2001a; Sweka & Hartman 2001b), eutrophication of the river catchment could be an
271 important negative influence on ide abundance and distribution.

272 Our findings further indicate that their migratory behaviour is shaped by multiple factors that
273 vary even within river systems. Whilst encouragingly this may indicate a degree of plasticity
274 in the species, more work is needed to understand the factors influencing these migrations and
275 hence direct conservation efforts to improve the breeding success of the remaining
276 populations.

277

278 **Acknowledgements:**

279 The authors wish to thank A. Slavíková, A. Petrušek and three anonymous referees for
280 helpful comments on earlier versions of the manuscript and technical staff at the Water
281 Research Institute in Prague for assistance with the fieldwork. The study was financially
282 supported by grant from the Ministry of Environment of the Czech Republic (SP/2e7/229/07).

283

284

References

285

286 Balon, E. K. (1975). Reproductive guilds of fishes: A proposal and definition. *Journal of*
287 *Fisheries Research Board of Canada* **32**, 821–864.

288 Baras, E. & Cherry, B. (1990). Seasonal activities of feale barbel *Barbus barbus* L. in the
289 River Ourthe (Southern Belgium) as revealed by radio-tracking. *Aquatic Living*
290 *Resources* **3**, 283-294.

291 Bauer, C. & Schlott, G. (2004). Overwintering of farmed common carp (*Cyprinus carpio* L.)
292 in the ponds of a central European aquaculture facility – measurement of activity by
293 radio telemetry. *Aquaculture* **241**, 301–317.

294 Benfield, M. C. & Minello, T. J. (1996). Relative effects of turbidity and light intensity on
295 reactive distance and feeding of an estuarine fish. *Environmental Biology of Fishes* **46**,
296 211-216.

297 Brown, R. S. & Mackay, W. C. (1995). Fall and winter movements of and habitat use by
298 cutthroat trout in Ram River, Alberta. *Transactions of the American Fisheries Society*
299 **124**, 873-885.

300 Cala, P. (1970). On the ecology of the ide *Idus idus* (L.) in the River Kävlingeån, south
301 Sweden. *Report of the Institute of Freshwater Research, Drottningholm* **50**, 45-99.

- 302 Colby, P. J., Spangler, G. R., Hurley, D. A. & McCombie, A. M. (1972). Effects of
303 eutrophication on salmonid communities in oligotrophic lakes. *Journal of Fisheries*
304 *Research Board of Canada* **29**, 975-983.
- 305 David, B. O. & Closs, G. P. (2003). Seasonal variation in diel activity and microhabitat use of
306 an endemic New Zealand stream-dwelling galaxiid fish. *Freshwater Biology* **48**(10),
307 1765-1781.
- 308 Diehl, S. (1988). Foraging efficiency of three freshwater fishes: effects of structural
309 complexity and light. *Oikos* **53**, 207-214.
- 310 Duchrow, R. M. & Everhart, W. H. (1971). Turbidity measurement. *Transactions of the*
311 *American Fisheries Society* **100**, 682-690.
- 312 Fraser, N. H. C., Heggenes, J., Metcalfe, N. B. & Thorpe, J. E. (1995). Low summer
313 temperatures cause juvenile Atlantic salmon to become nocturnal. *Canadian Journal*
314 *of Zoology* **73**, 446-451.
- 315 Harvey, B. C. & Nakamoto, J. R. (1999). Diel and seasonal movements by adult Sacramento
316 pikeminnow (*Ptylocheilus grandis*) in the Eel River, north-western California.
317 *Ecology of Freshwater Fish* **8**, 209-215.
- 318 Heubel, K. U. & Schlupp, I. (2006). Turbidity affects association behaviour in male *Poecilia*
319 *latipinna*. *Journal of Fish Biology* **68**(2), 555-568. doi:10.1111/j.0022-
320 1112.2006.00941.x
- 321 Hiscock, M. J., Scruton, D. A., Brown, J. A. & Clarke, K. D. (2002). Winter movement of
322 radio-tagged juvenile atlantic salmon in Northeast Brook, Newfoundland.
323 *Transactions of the American Fisheries Society* **131**, 577-581.
- 324 Horký, P., Slavík, O., Bartoš, L., Kolářová, J. & Randák, T. (2006). The effect of the moon
325 phase and seasonality on the behaviour of pikeperch in the Elbe River. *Folia*
326 *Zoologica*, **55**(4), 411-417.

- 327 Horký, P., Slavík, O., Bartoš, L., Kolářová, J. & Randák, T. (2007). Behavioural pattern in
328 cyprinid fish below a weir as detected by radio telemetry. *Journal of Applied*
329 *Ichthyology* **1**, 1-5. doi: 10.1111/j.1439-0426.2007.00848.x
- 330 Järvenpää, M. & Lindström, K. (2004). Water turbidity by algal blooms causes mating system
331 breakdown in shallow-water fish, the sand goby *Pomatoschistus minutus*. *Proc. R.*
332 *Soc. Lond. B* **271**, 2361 - 2365.
- 333 Kenward, M. G. & Roger, J. H. (1997). Small sample inference for fixed effects from
334 restricted maximum likelihood. *Biometrics* **53**, 983-997.
- 335 Lane, R. R., Day, J. W. J., Marx, B. D., Reyes, E., Hyfield, E. & Day, J. N. (2007). The
336 effects of riverine discharge on temperature, salinity, suspended sediment and
337 chlorophyll *a* in a Mississippi delta estuary measured using a flow-through system.
338 *Estuarine, Coastal and Shelf Science* **74**, 145-154.
- 339 Lau, J. K., Lauer, T. E. & Weinman, M. L. (2006). Impacts of channelization on stream
340 habitats and associated fish assemblages in East Central Indiana. *American Midland*
341 *Naturalist* **156**, 319-330.
- 342 Lucas, M. C. & Frear, P. A. (1997). Effects of a flow-gauging weir on the migratory
343 behaviour of adult barbel, a riverine cyprinid. *Journal of Fish Biology* **50**, 382-396.
- 344 Lucas, M. C. (2000). The influence of environmental factors on movements of lowland-river
345 fish in the Yorkshire Ouse system. *The Science of the Total Environment* **251/252**,
346 223-232.
- 347 Lusk, S., Hanel, L. & Lusková, V. (2004). Red List of the ichthyofauna of the Czech
348 Republic: Development and present status. *Folia Zoologica* **53**(2), 215-226.
- 349 Maitland, P. S. & Campbell, R. N. (1992). *Freshwater fishes of the British Isles*. London:
350 Harper Collins.

351 Metcalfe, N. B., Fraser, N. H. C. & Burns, M. D. (1999). Food availability and the nocturnal
352 vs. diurnal foraging trade-off in juvenile salmon. *Journal of Animal Ecology* **68**, 371-
353 381.

354 Miner, G. J. & Stein, R. A. (1996). Detection of predators and habitat choice by small
355 bluegills: effects of turbidity and alternative prey. *Transactions of the American*
356 *Fisheries Society* **125**, 97-103.

357 Mohr, C. O. (1947). Table of equivalent populations of North American mammals. *American*
358 *Midland Naturalist* **37**, 223-249.

359 Molls, F. & Neumann, D. (1994). Fish abundance and fish migration in gravel-pit lakes
360 connected with the River Rhine. *Water Science & Technology* **29**, 307-309.

361 Müller, K. (1986). Seasonal anadromous migration of the pike (*Esox lucius*, L.) in coastal
362 areas of the northern Bothnian Sea. *Archiv für Hydrobiologie* **107**, 315-330.

363 Pollux, B. J. A., Korosi, A., Verberk, W. C. E. P., Pollux, P. M. J. & van der Velde, G.
364 (2006). Reproduction growth, and migration of fishes in a regulated lowland tributary:
365 potential recruitment to the River Meuse. *Hydrobiologia* **565**, 105-120. doi:
366 10.1007/s10750-005-1908-4

367 Sahoo, G. B., Ray, C. & De Carlo, E. H. (2006). Use of neural network to predict flash flood
368 and attendant water qualities of a mountainous stream on Oahu, Hawaii. *Journal of*
369 *Hydrology* **327**, 525-538.

370 SAS Institute Inc. (2004). SAS/STAT® 9.1 User's Guide. Cary, NC: SAS Institute Inc.

371 Slavík, O., Horký, P., Bartoš, L., Kolářová, J. & Randák, T. (2007). Diurnal and seasonal
372 behaviour of adult and juvenile European catfish as determined by radio-telemetry in
373 the River Berounka, Czech Republic. *Journal of Fish Biology* **71**(1), 101-114. doi:
374 10.1111/j.1095-8649.2007.01471.x

375 Sweka, J. A. & Hartman, K. J. (2001a). Effects of turbidity on prey consumption and growth
376 in brook trout and implications for bioenergetics modeling. *Canadian Journal of*
377 *Fisheries and Aquatic Sciences* **58**, 386-393.

378 Sweka, J. A. & Hartman, K. J. (2001b). Influence of turbidity on brook trout reactive distance
379 and foraging success. *Transactions of the American Fisheries Society* **130**, 138-146.

380 Sweka, J. A. & Hartman, K. J. (2003). Reduction of reactive distance and foraging success in
381 Smallmouth bass, *Micropterus dolomieu*, exposed to elevated turbidity levels.
382 *Environmental Biology of Fishes* **67**(4), 341-347.

383 Tao, J., Little, R., Patetta, M., Truxillo, C. & Wolfinger, R. (2002). *Mixed the SAS System*
384 *Course Notes*. Cary, NC, USA: SAS Institute Inc.

385 Utne-Palm, A. C. (2001). Visual feeding of fish in a turbid environment: physical and
386 behavioural aspects. *Mar. Fresh. Behav. Physiol.* **35**, 111-128.

387 Valdimarsson, S. K. & Metcalfe, N. B. (2001). Is the level of aggression and dispersion in
388 territorial fish dependent on light intensity? *Animal Behaviour* **61**, 1143-1149.

389 Vinyard, G. L. & O'Brien, W. J. (1976). Effect of light and turbidity on the reactive distance
390 of Bluegill (*Lepomis macrochirus*). *Journal of Fisheries Research Board of Canada*
391 **33**, 2845-2849.

392 Webb, P. W. (1978). Temperature effects on acceleration of rainbow trout, *Salmo gairdneri*.
393 *Journal of the Fisheries Research Board of Canada* **35**(11), 1417-1422.

394 White, G. C. & Garrott, R. A. (1990). *Analysis of wildlife radio-tracking data*. New York:
395 Academic Press, 383 pp.

396 Winter, H. V. & Fredrich, F. (2003). Migratory behaviour of ide: a comparison between the
397 lowland rivers Elbe, Germany, and Vecht, The Netherlands. *Journal of Fish Biology*
398 **63**, 871-880. doi:10.1046/j.1095-8649.2003.00193.x

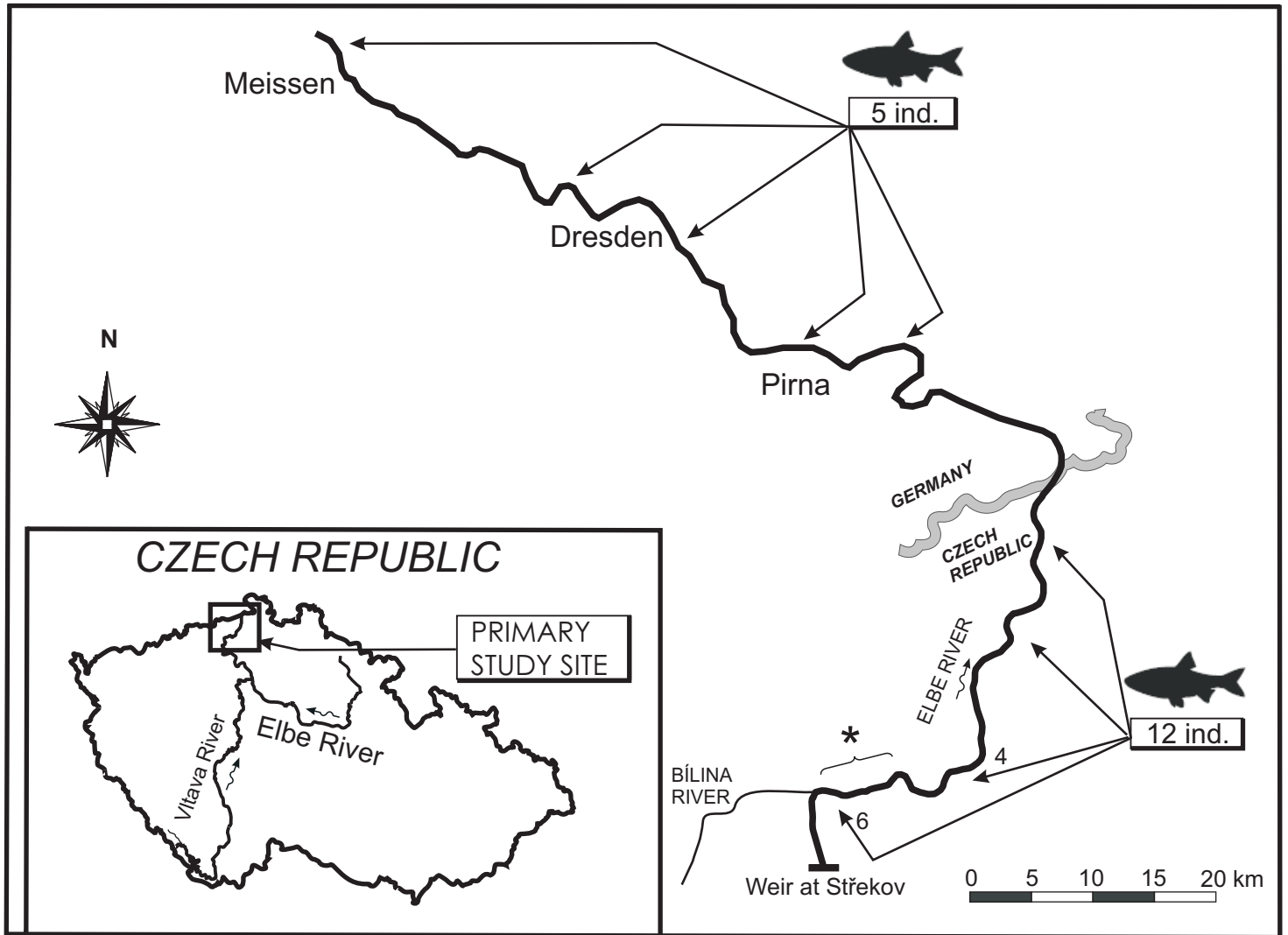
399 Winter, J. D. (1983). Underwater Biotelemetry. In *Fisheries Techniques* p. 371-395. (Nielsen,
400 L. A. & Johnsen, D., eds), pp. 371 – 395. Bethesda, MD: American Fisheries Society.
401 Winter, J. D., Kuechle, V. B., Sinff, D. B. & Tester, J. R. (1978). Equipment and methods for
402 radio tracking freshwater fish. In: *Univ. Minnessota: Agricultural Experiment Station,*
403 *Miscellaneous Reports 152.*

1 FIG. 1 Map showing the location of the study site with highlighted distances of spring
2 downstream migrations and number of individuals (ind.) migrating within Czech and
3 German part of River Elbe. Bracket with asterisk indicate the river stretch where the fish
4 were caught and released after tagging. Arrow (\rightsquigarrow) indicates the direction of river
5 flow.

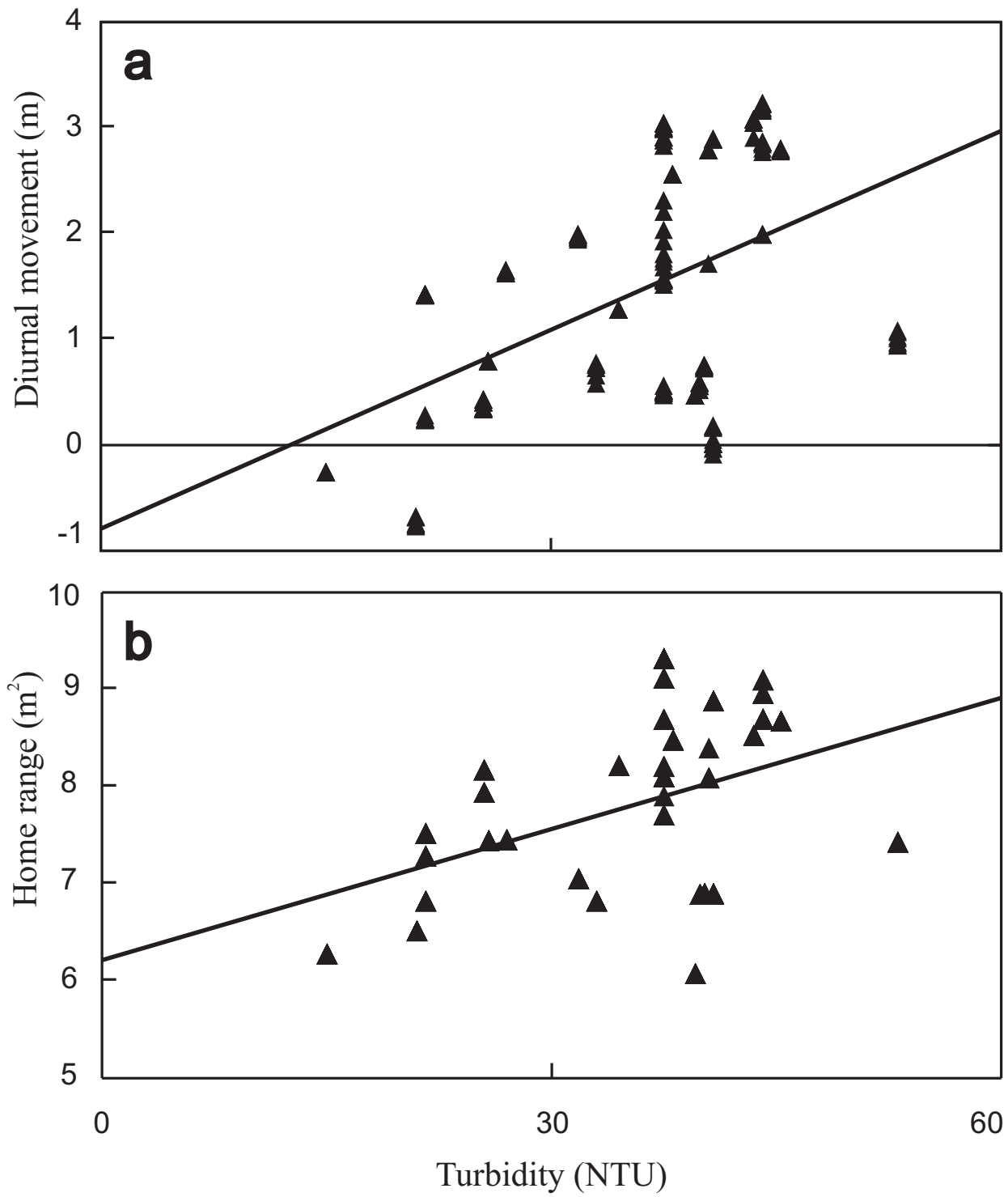
6
7 FIG. 2 Relationship between diurnal movements (a) and home range size (b) of ide and
8 turbidity. Predicted values are from square root transformed data. The curves were fitted
9 by: $y = 0.0188x - 0.0395$, ($r^2 = 0.38$) for diurnal movements and $y = 0.0189x + 7.4581$,
10 ($r^2 = 0.34$) for home range size.

11
12 FIG. 3 Home range size (a) and longitudinal movements (b) of ide across seasons. Asterisks
13 indicate significant differences between groups (* $P < 0.05$; ** $P < 0.01$). Values are
14 adjusted means \pm S.E. of square root transformed data.

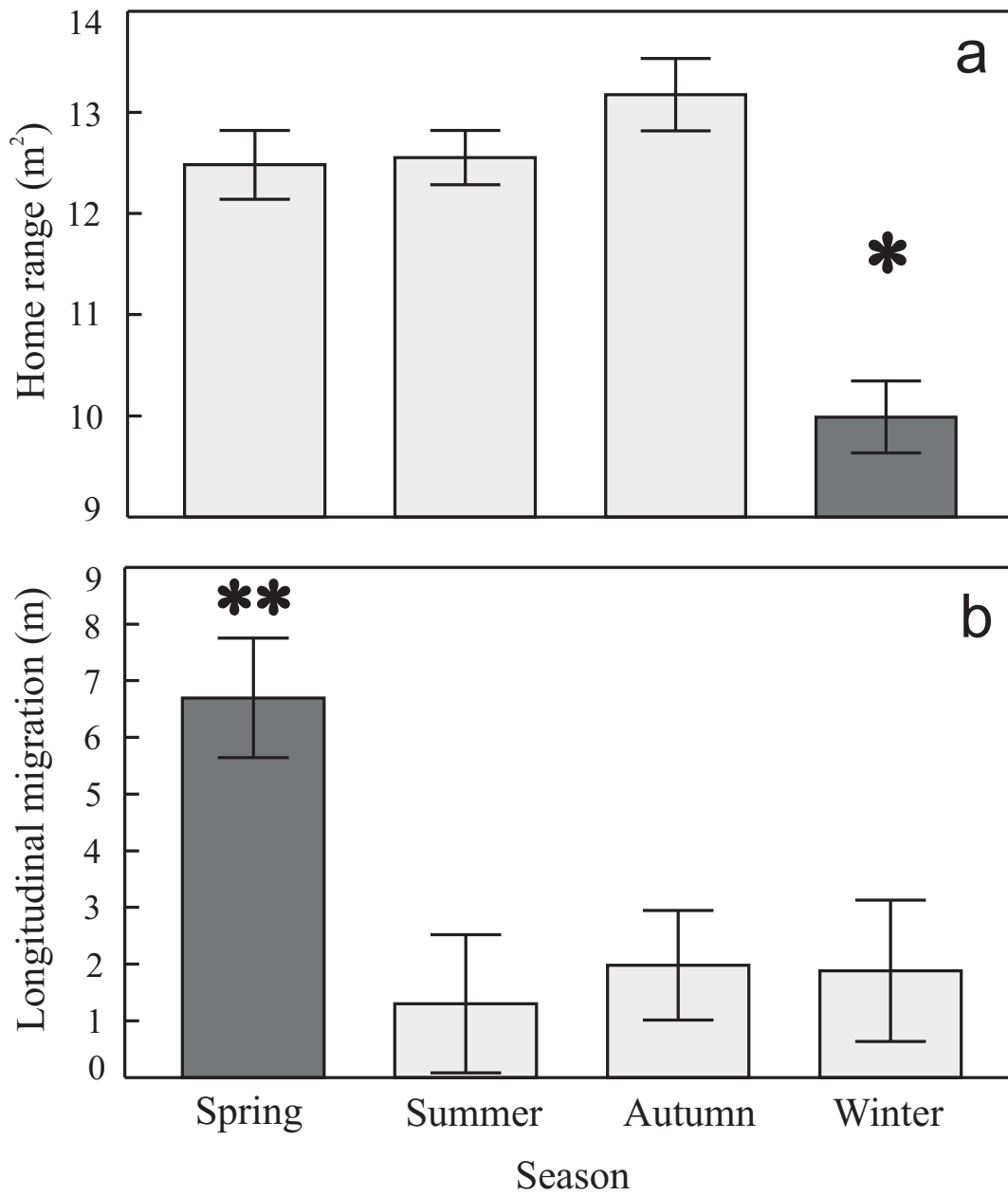
Figure



Figure



Figure



1 TABLE I. Type 3 tests of fixed effects for diurnal movements, home range area size, and
 2 longitudinal movements.

3

Effect	Num DF	Den DF	F	P<
LMM I (<i>diurnal movement</i>)				
turbidity	1	256	7.82	0.0056
light interval(season)	11	256	3.20	0.0004
LMM II (<i>home range size</i>)				
turbidity	1	255	68.37	0.0001
season	3	255	56.12	0.0001
LMM III (<i>longitudinal movement</i>)				
season	3	255	5.79	0.0008

4

5 TABLE II. Tagged individuals of ide with mean (\pm S.D.) of their recorded diurnal movements,
 6 longitudinal movements (LM), total distance migrated during spawning (TD) and
 7 home range (HR).

Individual code	DM (m)	LM (m)	TD (m)	HR (m²)
12	50.1 \pm 100.5	1 824.8 \pm 5 168.2	13 649	29 817.6 \pm 23 924.05
14	24.8 \pm 68.2	4 908.7 \pm 13 711.2	99 729	19 506.4 \pm 14 325.2
16	39.4 \pm 89.7	6 372.6 \pm 6 803.6	13 284	23 465.8 \pm 12 698.7
17	15.6 \pm 24.2	135.6 \pm 361.8	3 882	7 400.4 \pm 4 441.02
18	45.7 \pm 95.8	5 510.2 \pm 10 119.3	54 160	11 342.1 \pm 8 432.6
24	32.8 \pm 84.3	1 447.8 \pm 3 193.4	18 633	36 157.5 \pm 22 001.1
28	77.8 \pm 120.1	6 238.5 \pm 13 801.3	85 505	24 158.2 \pm 17 325.8
40	21.1 \pm 54.3	1 710.7 \pm 1 909.6	2 908	8 352.1 \pm 5 368.3
43	65.7 \pm 118.6	2 811.3 \pm 5 122.2	18 284	25 931.0 \pm 18 963.2
44	34.3 \pm 40.5	5 517.6 \pm 12 024.6	37 406	13 700.2 \pm 7021.4
45	38.1 \pm 81.6	982.5 \pm 3 421.7	4 145	27 536.5 \pm 19871.8
46	20.4 \pm 53.8	1 628.1 \pm 4 236.9	9 380	23 658.8 \pm 21 879.6
47	43.2 \pm 94.4	2 874.3 \pm 3 465.5	26 089	17 502.3 \pm 13 468.2
48	37.3 \pm 71.1	1 382.4 \pm 2 678.6	4 332	19 875.4 \pm 13 489.7
49	19.2 \pm 52.9	1 348.2 \pm 1 825.1	3 546	26 849.7 \pm 21 487.2
50	28.4 \pm 65.3	2 949.5 \pm 8 236.4	67 946	13 672.3 \pm 9 237.4
51	30.2 \pm 41.5	2 131.4 \pm 5 628.5	13 680	2 502.7 \pm 2 030.6

8

This piece of the submission is being sent via mail.