

1 **Population abundance of recovering humpback whales (*Megaptera***  
2 ***novaeangliae*) and other baleen whales in the Scotia Arc, South Atlantic**

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21 **ABSTRACT**

22 Following the cessation of whaling, South Atlantic populations of humpback (*Megaptera*  
23 *novaeangliae*) and some other baleen whale species are recovering, but there has been  
24 limited monitoring of their recovery in the Scotia Arc, a former whaling epicentre and a  
25 hotspot for Antarctic krill (*Euphausia superba*). To inform the management of krill fisheries,

1 up to date assessment of whale biomass and prey consumption is essential. Using a model-  
2 based approach, we provide the first estimates of whale abundance and krill consumption for  
3 South Georgia and the South Sandwich Islands, and total abundance of humpback whales  
4 across their southwest Atlantic feeding grounds, using data collected in 2019. Humpback  
5 whale abundance was estimated at 24,543 (CV = 0.26; 95% CI 14,863 – 40,528), similar to that  
6 measured in Brazil on the main wintering ground for this population. The abundance of baleen  
7 whales in South Georgia and the South Sandwich Islands, including those not identified to  
8 species level, was estimated at 43,824 (CV = 0.15; 95% CI 33,509 – 59,077). Based on the  
9 proportion of humpback whales identified during the surveys (83%), the majority of these are  
10 likely to be humpback whales. Annual krill consumption by baleen whales was estimated to  
11 be in the range 4.8 – 7.2 million tons, representing 7 - 10% of the estimated krill biomass in  
12 the region. However, there is a need to better understand feeding rates in baleen whales, and  
13 further research into this field should be a priority in order to improve the accuracy and  
14 precision of prey consumption rate estimation.

15

## 16 **KEY WORDS**

17 HUMPBACK WHALE; BALEEN WHALE; ABUNDANCE; KRILL CONSUMPTION; SOUTH GEORGIA;  
18 SOUTH SANDWICH ISLANDS

19

## 20 **1. INTRODUCTION**

21 Baleen whales are important predators of krill, but were decimated by modern whaling during  
22 the 20<sup>th</sup> century (Rocha et al. 2014). In the South Atlantic, whaling was initially very intense in  
23 the vicinity of sub-Antarctic South Georgia island, and latterly spread across the Scotia Arc via  
24 offshore factory ships working with catcher vessels. Species were successively targeted, with

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1 humpback whales (*Megaptera novaeangliae*) hunted initially, followed by blue  
2 (*Balaenoptera musculus*), fin (*B. physalus*), and sei whales (*B. borealis*). There was only  
3 limited exploitation of southern right whales (*Eubalaena australis*) in the 20<sup>th</sup> century, as this  
4 species was already very rare following heavy exploitation in the early 19<sup>th</sup> century (IWC  
5 2001). Since whaling ended, the abundance of some cetacean species has been increasing.  
6 The southwest Atlantic population (referred to by IWC (1998) as Breeding Stock A) of  
7 humpback whales is thought to have recovered to a level similar to that in the 19<sup>th</sup> century,  
8 before its exploitation on an industrial scale (Bortolotto et al. 2016, Ward et al. 2011, Zerbini  
9 et al. 2019). In the light of the increasing size of this humpback whale population, and of other  
10 species in the region such as the blue whale (Branch et al. 2004, Calderan et al. 2020), fin  
11 whale (Viquerat & Herr 2017) and southern right whales (Best et al. 2001, Groch et al. 2005,  
12 Crespo et al. 2019), re-assessment of whale abundance and krill consumption rates by whales  
13 on their high latitude South Atlantic feeding grounds is urgently needed for inclusion in krill  
14 fishery management models (e.g. Trathan et al. in press).

15 Antarctic krill (*Euphausia superba*) is a critical species in Southern Ocean food webs and  
16 constitutes the main prey of a wide range of taxa, including species of fish, squid, seabirds,  
17 pinnipeds and cetaceans (Hill et al. 2006, Trathan & Hill 2016). The Antarctic Peninsula and  
18 Scotia Arc region has been identified as the area with the highest concentration of krill within  
19 the Southern Ocean (Atkinson et al. 2008), supporting high levels of predator biomass and  
20 other biodiversity, particularly in the northern Scotia Arc (e.g. Hogg et al. 2011). Southern  
21 Ocean fishery management is required to be sufficiently dynamic to be able to adapt quickly  
22 to changes in both prey and predator ecology, to spatially and temporally regulate harvesting  
23 at sustainable levels in order to minimise any negative impacts on the Antarctic ecosystem  
24 from such direct anthropogenic activities (Trathan et al. 2021). Within the Scotia Arc

1 spatiotemporal overlap with the summer feeding season of baleen whales has been limited  
2 to the South Orkney stratum, because the krill fishery has been restricted to winter months  
3 in the South Georgia stratum and there has been no krill fishing effort in the South Sandwich  
4 Islands stratum (CCAMLR 2021). However, both whales and the fishery exploit the same krill  
5 populations (Trathan et al. 2021). Catch limits need to be set with reference to updated  
6 estimates of both krill and predator biomass (Constable 2001 & 2011, Constable et al. 2000,  
7 Kawaguchi & Nicol 2020, Krafft et al. 2021). While krill are regularly monitored at a small  
8 number of locations in the Southern Ocean (e.g. Fielding et al. 2014, Reiss et al. 2008, Krafft  
9 et al. 2018), regular monitoring of predator biomass is also crucial for sustainable fishery  
10 management, particularly in areas targeted by krill fisheries. This is particularly important for  
11 baleen whales, as they are significant krill consumers and many species are increasing in  
12 abundance following their protection from commercial whaling (e.g. Branch et al. 2004,  
13 Leaper & Miller 2011, Bortolotto et al 2016, Calderan et al. 2020).

14 Since commercial whaling ended, very few large-scale surveys designed to assess whale  
15 recovery in the polar South Atlantic (Scotia Arc) region have been conducted. Three  
16 circumpolar surveys of the Southern Ocean were carried out between 1978/79 and 2003/04  
17 (International Whaling Commission's IDCR/SOWER), yielding circumpolar and regional  
18 abundance estimates for blue and humpback whales (Branch 2007 & 2011). However, these  
19 surveys were designed primarily to measure minke whale abundance (Matsuoka et al. 2003).

20 Minke whale occurrence is typically concentrated at the ice edge (Kasamatsu et al. 1998; Herr  
21 et al. 2019), so in the South Atlantic surveys were conducted in the southern Scotia Arc, while  
22 South Georgia in the northern Scotia Arc is the location where the majority of historical  
23 whaling effort took place. Modern whaling catches of >176,000 whales were made at South  
24 Georgia (Allison 2016), indicating that the northern Scotia Arc was a key feeding ground for

1 many baleen whale species. The Commission for the Conservation of Antarctic Marine Living  
2 Resources (CCAMLR) 2000 synoptic krill survey surveyed both whales and krill across the  
3 Scotia Arc, providing information on both whale abundance (Hedley et al. 2001) and krill  
4 consumption by whales (Reilly et al. 2004). However surveys of South Georgia waters were  
5 limited by poor weather conditions, and abundance estimates from this historically important  
6 region were very small ( $N = 116$  baleen whales, Hedley et al. 2001; no uncertainty estimates  
7 provided).

8 In 2019, this synoptic krill survey was (in large part) repeated (Krafft et al. 2021), and included  
9 cetacean observers collecting line-transect data over an area similar to the 2000 survey. For  
10 the first time, the remote South Sandwich Islands were also surveyed, an area known from  
11 satellite tracking to be an important feeding site for humpback whales (Zerbini et al. 2006 &  
12 2011, Horton et al. 2020) but never corroborated by direct surveys. Here we report for the  
13 first time estimates of total baleen whale abundance and krill consumption in the South  
14 Georgia and South Sandwich Islands (northern and eastern Scotia Arc) and provide a new  
15 feeding ground estimate of abundance for humpback whales in the Scotia Arc, spanning the  
16 known feeding range of the southwest Atlantic breeding stock. These are essential data to  
17 contribute to the CCAMLR krill risk assessment framework (CCAMLR 2019, paragraph 5.17),  
18 which combines localised estimates of krill consumption needs of various predators with  
19 estimates of krill biomass at the same spatio-temporal scale, to test the likely impacts of a  
20 range of different krill fishing scenarios.

21

## 22 **2. METHODS**

### 23 **2.1. Cetacean survey field methods**

## Scotia Arc whale abundance

1 Three of the vessels that participated in the 2019 CCAMLR synoptic krill survey, the *RRS*  
2 *Discovery* (DIS), *FV Cabo de Hornos* (CDH) and *RV Kronprins Haakon* (KPH) contributed data to  
3 this study, although in the case of CDH and KPH only humpback whale data were contributed.  
4 Within the Scotia Arc, these surveys covered three main areas, defined by the Food and  
5 Agriculture Organisation of the United Nations (FAO) as Subarea 48.2 South Orkney Islands  
6 and Scotia Sea (SO), Subarea 48.3 South Georgia (SG) and Subarea 48.4 South Sandwich  
7 Islands (SS) (Figure 1). As in the 2000 survey, all vessels surveyed parallel transects across  
8 these strata, with DIS surveying the South Georgia and South Sandwich regions, CPH surveying  
9 South Georgia and the South Orkney Islands/Scotia Sea, and KPH surveying transects in all  
10 three strata.

11 Line transect survey protocols were standardized as far as possible prior to survey, but  
12 differed in practice, as numbers of observers differed across vessels. Observation watches  
13 were carried out through daylight hours during periods with visibility at least 1 km in Beaufort  
14 sea state <6; data collected outside these limits were removed from these analyses. Line  
15 transect effort was suspended whenever the vessel stopped for scientific sampling on station  
16 (trawl and net hauls, oceanographic sampling, etc.). On transect data were collected on vessel  
17 position, speed and heading, wind speed and direction, sea state, weather conditions and  
18 visibility every 15 minutes, or whenever environmental conditions changed. On DIS, a team  
19 of four observers maintained effort between 05:30 and 20:00 (local time). Two observers  
20 were on watch at any one time, with a third acting as data recorder; the observers rotated  
21 every hour, such that the same pair of observers always watched together. The observation  
22 platform located above the bridge was approximately 19 m above sea level (ASL) and had an  
23 unobstructed view of 180° forward. Distances were estimated using reticle binoculars and  
24 angles from the vessel's heading were measured with an angle board. On CDH, observations

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1 were carried out by a single observer during daylight hours from a platform 10.5 m ASL;  
2 observations were only made from the starboard side, since visibility there was significantly  
3 better than on the port side. The primary search area was from 350° - 45° relative to the ship's  
4 heading, with a secondary search area from 45° to 90°. Distances to whales were estimated  
5 by eye, and random checks on accuracy were made by using a stopwatch and comparing the  
6 distances using the ship's speed and assuming no animal movement. Observations were  
7 carried out mainly from inside the bridge or outside in front of the bridge when weather  
8 conditions allowed. On KPH, observer effort was carried out by two dedicated observers. The  
9 observation platform was approximately 25 m ASL and, due to an obstructed view to  
10 starboard, dedicated observations were usually made within the forward port quarter (270°  
11 - 360°, relative to the ship's heading). However, observers also scanned outside of this sector,  
12 and occasionally recorded observations within that part of the forward starboard quarter that  
13 was not obstructed.

14

### 15 **2.2. Estimation of abundance**

16 Line transect cetacean survey data collected by all three survey vessels were used to estimate  
17 the abundance of humpback whales across the Scotia Arc, which is thought to represent the  
18 summer feeding ground range of southwest Atlantic humpback whales (Zerbini et al. 2019).  
19 The abundance of all baleen whales was also estimated for the two CCAMLR Statistical  
20 subareas 48.3 (South Georgia; SG) and 48.4 (South Sandwich; SS), using survey data from DIS  
21 alone, to provide total whale biomass estimates for these areas.

22 The CCAMLR 2019 survey was designed primarily to estimate krill biomass and to investigate  
23 krill ecology and not as a line transect survey for cetaceans. A design-based approach to  
24 abundance estimation, for which there should be an equal probability of coverage throughout

1 the study area (Buckland et al. 2001, Hiby & Hammond 1989), was therefore not considered  
2 appropriate and a model-based method (Hedley et al. 1999, Hedley & Buckland 2004) was  
3 adopted. A two-stage process was followed in which detection functions were first fitted to  
4 perpendicular distance estimates and generalised additive models (GAM) were then applied,  
5 using an offset derived from the detection function to account for effective search area (Miller  
6 et al. 2013). Abundance was estimated by summing the product of predicted density and cell  
7 area over a grid of resolution 0.14° latitude and longitude (median cell area 140 km<sup>2</sup>).

8 Detection functions were fitted with the *R* package *Distance* (Miller et al. 2019). A multiple  
9 covariate distance sampling (MCDS) framework was implemented with group size, sea state,  
10 visibility and observer team as potential covariates, applied to count data from individual  
11 vessels and to data pooled across all three vessels. For the two vessels KPH and CDH, sightings  
12 made outside their respective 90° primary fields of view were excluded. Fitting detection  
13 functions to data pooled across the three vessels proved challenging due to small sample sizes  
14 within the study area for CDH and KPH. We therefore included sightings from transect lines  
15 in areas adjacent to the Scotia Sea study area in the detection function, which increased  
16 sample sizes and allowed good fits to be obtained. Half-normal and hazard-rate key functions  
17 with no adjustment terms were tested and the best fitting models were selected by  
18 minimising the Akaike Information Criterion (AIC).

19 Density surface models and associated variances were fitted to count data using the *R*  
20 package *dsm*, which also allowed for uncertainty in the detection function to be propagated  
21 through to the abundance estimates (Miller et al. 2013, Bravington et al. 2021). Data were  
22 partitioned into segments with mean length 10.6 km (range 0.9 – 35 km); the maximum length  
23 of segments was constrained by regular changes in observer teams. Coordinates were  
24 projected to UTM zone 21S so that *x* and *y* coordinates represented similar spatial scales. The

1 covariates included in the models are listed in Table 1, although only those found to be  
2 contributing significantly to model fit were retained. Environmental covariates were checked  
3 for correlation by constructing a Pearson correlation matrix and, of the variables retained in  
4 models, the maximum Pearson correlation value was 0.23. Negative binomial and Tweedie  
5 (Dunn & Smyth 2008) distributions were tested, all variables were included and by setting  
6 *Select = True*, covariates not contributing to model fit were automatically penalised out  
7 (Marra & Wood 2011). Plots of smooth functions were examined and, where the distribution  
8 of data points was markedly skewed, explanatory variables were log transformed to achieve  
9 a more even distribution. Model residuals were investigated using *rqgam.check*. The best  
10 fitting models were selected on the basis of maximising deviance explained and minimising  
11 the coefficient of variation attributed to the GAM, and the resulting predictions were checked  
12 by comparison with plots of effort and sightings made across the study area.

13 Spatial models were fitted to humpback whale density estimates pooled across all vessels for  
14 the entire study area. The *dsm* package assumes observations are carried out on both sides  
15 of the track-line, i.e. through 180° to generate an offset accounting for the effective area  
16 searched that takes the general form:

$$17 \quad \textit{Offset} = 2 * \textit{ESW} * l$$

18 where *ESW* is the effective strip width derived from the detection function and *l* is the line  
19 segment length. Therefore, to account for effort having been applied through 90° on KPH and  
20 CDH, segment lengths for those vessels were halved. Data from DIS only were used to model  
21 the abundance of all baleen whales in SG and SS, including unidentified baleen whales, but  
22 excluding sightings recorded as unidentified large whales, a category that could include sperm  
23 whales.

24

### 1 **2.3. Krill consumption by whales**

2 Lunge feeding by baleen whales is energetically costly and therefore its efficiency in terms of  
3 energy gained over cost is dependent on prey density (Goldbogen et al. 2011). However, given  
4 sufficiently dense aggregations of prey it is an extremely efficient method of feeding  
5 (Goldbogen et al. 2019, Potvin et al. 2021), permitting rapid weight gain. Lockyer (1981)  
6 concluded from seasonal catch data that most humpback whales arrived in South Georgia  
7 before the end of December with a body weight of around 15 t and weight increase appeared  
8 to slacken off in February with an overall weight gain of 15 t, effectively doubling the animal's  
9 weight in a period of only three months. Estimation of krill consumption by baleen whales is  
10 based on whale abundance estimates, estimated daily prey consumption rates and either the  
11 length of stay of whales in feeding areas or the proportion of their annual prey requirement  
12 consumed there. Reviews of methods for estimating prey consumption rates (Reilly et al.  
13 2004, Leaper & Lavigne 2007) point out some of the difficulties associated with the estimation  
14 of this parameter for species in which there is marked seasonal asymmetry in feeding rates  
15 and for which direct measurements of consumption are unavailable. Humpback whales  
16 undertake long migrations: in the case of the southwest Atlantic population their winter  
17 breeding habitat in the coastal waters of Brazil is located over 4,000 km from feeding habitats  
18 around the South Sandwich Islands and South Georgia (Clapham & Mead 1999, Zerbini et al.  
19 2011a). While humpback whales may feed opportunistically during migration (Danilewicz et  
20 al. 2009, Owen et al. 2017) the bulk of their annual energy requirements must be supplied by  
21 prey consumption during the period they forage within southern feeding grounds. There are  
22 essentially two approaches to the estimation of consumption rates, a supply approach,  
23 historically based on feeding rates derived from weights of the stomach contents of whales  
24 killed commercially or for research purposes, and a demand approach based on studies of

1 energy budgets. Recent work on energy budgets has focussed on the costs of anthropogenic  
2 disturbance (e.g. Pirotta et al. 2021, Braithewaite et al. 2015) for which relative measures are  
3 meaningful, but in order to estimate absolute biomass of prey consumed, a series of  
4 parameters have to be estimated and assumptions made, leading Reilly et al. (2004) to  
5 conclude the more direct supply approach to be more reliable. However, assumptions  
6 regarding digestion rates, the length of time ingested food remains in the stomach and the  
7 frequency with which the stomach is filled introduce unquantifiable uncertainty in that  
8 approach (Leaper & Lavigne, 2007). Few quantitative data are available on the weights of  
9 mysticete stomach contents. Kawamura (1980) reviewed qualitative data available from  
10 commercial whaling on the prey species of whales and provided data on the weights of fin  
11 whale stomach contents from catches in the Bering Sea. Tamura et al. (1997) estimated daily  
12 consumption rates from minke whales lethally sampled in Antarctic regions and Reilly et al.  
13 (2004) extrapolated from these estimates to larger species using an allometric model  
14 approximated by an exponential function. Allometric scaling from the smallest baleen whale  
15 species does not take into account differences in feeding rates between species arising from  
16 niche separation (Friedlaender et al. 2006). Given the uncertainties in both approaches, we  
17 opted for the energetic demand method.

18 Population demography also needs to be taken into account in consumption rate estimation  
19 as energetic demands and therefore prey consumption rates vary according to age, sex and  
20 reproductive status; recent insights into this intraspecific variation have been gained by  
21 quantitative assessment of total body lipid from historic whaling data (Irvine et al. 2017). The  
22 population demography of humpback whales using the Scotia Arc (i.e. sex-specific  
23 proportions of adults and juveniles) is not directly known, so was inferred from the available  
24 literature. In their review of humpback whale life history parameters, Zerbini et al. (2010)

1 considered the range of parameters compatible with maximum rates of population growth.  
2 A population rate of increase ( $r$ ) in the range 7% - 11% (Zerbini et al. 2010) resulted in the  
3 proportion of breeding age females being in the range 34% - 44%. However Zerbini et al.  
4 (2010) aimed to measure maximum possible growth rates in a population that was far from  
5 recovery, and the southwest Atlantic population is thought now to be close to recovery, with  
6 median predicted population growth between 2015 and 2019 at  $r = 2.94\%$  (Zerbini et al.  
7 2019). In the western North Atlantic, lower population growth rates ( $r \sim 3\%$ , Stevick et al.  
8 2003) are associated with proportions of breeding females at the lower end of this range  
9 (averaged annual proportion of mature animals in the Gulf of Maine feeding ground is 68.7%,  
10 Robbins 2007). Therefore, for the purposes of this study we took a conservative approach and  
11 assumed the Scotia Arc humpback population is likely similar to the lower range value from  
12 Zerbini et al. (2019) of 34% breeding age females and a sex ratio of 1:1 (i.e. 68% total mature  
13 animals, similar to the mature proportion estimated for the Gulf of Maine feeding ground).  
14 Mean body mass of humpback whales is based on Trites & Pauly (1998). For other baleen  
15 whale species we have used body mass estimates compiled by Reilly et al. (2004) which  
16 assume a mean body mass for each species; we did not adjust for population demography as  
17 the proportion of adult and juvenile components are not known for these species in the South  
18 Atlantic.

19 The basal metabolic rate (BMR) is the energy required by an organism to maintain  
20 homeostasis. Kleiber (1975) identified an allometric relationship between BMR and body  
21 mass in animals, calculated as  $BMR = 70M^{0.75}$  where BMR is measured in kilocalories per day  
22 and  $M$  is body mass in kilograms. All whales require additional energy for activities that  
23 include locomotion and prey capture, in females of reproductive age for foetal growth in  
24 pregnancy and for subsequent lactation, mature males may incur additional energetic costs

1 associated with reproduction and calves require energy for growth. Further energetic costs  
2 may arise from stressors such as anthropogenic disturbance, attacks by predators or changes  
3 in the distribution of prey.

4 When BMR is expressed in kiloJoules per day, Kleiber's equation becomes  $BMR = 293.1M^{0.75}$   
5 and a multiplication factor  $>1$  may then be applied to estimate the average daily metabolic  
6 rate, usually referred to as the field metabolic rate (FMR) of the form  $FMR = \beta(293.1M^{0.75})$   
7 (Leaper & Lavigne 2007), where  $\beta$  may be expected to vary according to age, sex and  
8 reproductive status. In their review of methods for the estimation of prey consumption by  
9 whales, Leaper & Lavigne (2007) found most researchers had assumed  $\beta$  to be in the range 2-  
10 5 but cautioned that such values were not necessarily derived from actual data and found no  
11 evidence for an upper bound of  $\beta > 4$ . A direct approach to the measurement of FMR has been  
12 applied to a wide range of terrestrial mammals, some pinniped and odontocete species using  
13 a doubly labelled water (DLW) method (Lifson & McClintock 1966, Nagy 1983, Speakman  
14 1997, Nagy et al. 1999). However, the DLW method requires intervention in the form of taking  
15 blood samples and injecting labelled water that, in the case of cetaceans, has restricted its  
16 application to captive animals. Nagy (2005) reviewed the relationship between FMR and body  
17 mass using data derived from DLW experiments and developed the following formula for  
18 terrestrial mammals:

$$FMR=4.82M^{0.734}$$

19  
20 Evolutionary adaptation to life in low temperature aquatic environments may, however,  
21 result in elevated metabolic rates compared with terrestrial mammals (e.g. Costa & Williams  
22 2000). When applied to pinnipeds and bottlenose dolphins (*Tursiops truncatus*) the FMR  
23 values estimated by the DLW method were up to six times higher than those estimated by  
24 Nagy's equation for terrestrial mammals (Costa & Williams 2000). In a study of harbour

1 porpoise (*Phocoena phocoena*) energetics, FMR estimates obtained by the DLW method were  
2 correlated with respiration rates in captive animals, which were then compared with  
3 respiration rates of free-ranging porpoises to obtain FMR estimates up to two times those  
4 predicted by Nagy's equation (Rojano-Donāte et al. 2018). However, differences between  
5 small odontocetes and baleen whales in body size, behavioural activity and feeding strategies  
6 suggests caution in extrapolation of the high FMR values estimated for the smaller species.  
7 Simple allometric scaling does not take into account that much of the body mass of large  
8 whales consists of blubber, which has a lower metabolic rate than other tissues, such as  
9 muscle (Costa & Williams 2000) which may therefore tend to lower BMR and therefore FMR.  
10 On the other hand, the energetic implications associated with conversion of ingested food to  
11 blubber and the reverse process in fasting periods to mobilise stored energy may tend to raise  
12 metabolic rates, as for example the efficiency in converting stored energy to milk energy in  
13 lactating mammals has been estimated at 84% (Ofstedal 2000). Using a state-dependent  
14 foraging model, Wiedenmann et al. (2011) estimated blue whales expend between 2.07 and  
15 3.1 times the daily estimate of BMR (with a mean value of 2.29).  
16 Comparing Nagy's equation for FMR with Kleiber's BMR equation in humpback whales gives  
17 a  $\beta$  value of 2.25. We therefore take a conservative approach and assume that in males and  
18 non-breeding females  $\beta$  is in the range 2-2.5. Given a birth rate in the range 0.36 – 0.42  
19 (Zerbini et al. 2010) we assume breeding age females are either pregnant or lactating in two  
20 out of every three years. Pregnancy requires additional energy to grow and support the foetus  
21 and placenta, and lactation is one of the most energetically expensive physiological processes  
22 in mammals (Hanwell & Peaker 1977). In a study of grey whale (*Eschrichtius robustus*)  
23 energetics, the estimated cost of pregnancy was an additional 18% and of lactation 28% of  
24 annual energy requirements (Villegas-Amtmann et al. 2015). Analysis of historical whaling

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1 records from Western Australia (Irvine et al. 2017) found pregnant female humpback whales  
2 had energy stores 26.2% higher than non-pregnant females. Using the data given above for  
3 the increased energy store of pregnant humpback whales and the proportional increase of  
4 lactation over pregnancy in grey whales, we estimate the mean  $\beta$  for females over a three  
5 year period to be in the range 2.8-3.5. Immature animals need additional energy for growth  
6 and in the period from weaning at one year of age to reaching sexual maturity at the age of  
7 five, their average annual gain in body mass is approximately 7% (Chittleborough 1965,  
8 Clapham 2000). Innes et al. (1987) using allometric functions of body mass found feeding rates  
9 of growing juvenile marine mammals to be approximately double that of adults, although this  
10 was largely based on pinnipeds with baleen whales represented by a single study of captive  
11 grey whale calves (Wahrenbrock et al. 1974). We therefore assumed  $\beta$  to be in the range 3-4  
12 for juvenile humpback whales.

13 The energetic content of krill varies with age, sex and reproductive status (Färber-Lorda et al.  
14 2009). Schaafsma et al. (2018) reviewed the energetic value of zooplankton species in the  
15 Southern Ocean, quoting values of  $5.45 \text{ kJg}^{-1}$  wet weight for female and  $3.83 \text{ kJg}^{-1}$  for male  
16 *Euphausia superba* Dana sampled in South Georgia (Clarke 1984). Assuming a 1:1 sex ratio of  
17 krill we use a mean value of  $4.64 \text{ kJg}^{-1}$  and divide the estimated annual energy requirement  
18 of whales by this value to obtain the weight of krill needed to be assimilated to meet the  
19 energy demand. The assimilation efficiency of prey by baleen whales was estimated to be  
20 80% (Lockyer 1981, Kenny et al. 1997) resulting in a multiplication factor of 1.25 for the  
21 quantity of prey needed to be consumed. A further assumption needs to be made of the  
22 proportion of overall annual consumption of prey by humpback whales during the summer  
23 period spent in high latitude habitats, for which we took the value of 85% suggested by  
24 Lockyer (1981).

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### **3. RESULTS**

Effort tracks and humpback whale sightings from all three vessels in the study area are shown in Figure 2, revealing concentrations of humpback whale sightings in two main areas, one to the east and northeast of the South Sandwich Islands, the other to the north and northwest of South Georgia. For each vessel and each survey region, effort expressed as total line length is summarised in Table 2, and the number of humpback whale sighting events and counts are summarised in Table 3, which excludes sightings truncated at perpendicular distances >2.5 km and outside the 90° primary fields of view for CDH and KPH. Sightings of all mysticete species recorded by DIS are summarised in Table 4. These sightings and effort data are available from the British Antarctic Survey Polar Data Centre (DOI: XXX).

#### **3.1. Abundance of humpback whales in the Scotia Arc**

The detection function selected was a hazard rate function with truncation at 2.5 km incorporating sea state and observer team as covariates (Figure 3); a Tweedie distribution (Dunn & Smyth 2008) was selected for the GAM, based on goodness of fit. The variables retained were tabulated in the model output summary with smooths plotted in Figure 4, and explained 41.4% of the total deviance (Table 5). Phytoplankton density emerged as a highly significant explanatory variable, both as chl-*a* values concurrent with the survey and those for the previous year. These two variables were not significantly correlated, although the values for April 2018 were correlated with adjacent months (data not shown), so the variable CHL\_Apr2018 was interpreted as representing a generic value for phytoplankton density in the previous autumn. The predicted abundance of humpback whales across the Scotia Sea was 24,543 (CV = 0.26, 95% CI 14,863 – 40,528) (Table 6). The resulting density surface model

1 (Figure 5) showed two main areas of high density, one to the north-west of South Georgia  
2 and the other mainly to the north and north-east of the South Sandwich Islands.

3

### 4 **3.2. Abundance of baleen whales in South Georgia and the South Sandwich Islands**

5 For combined baleen whales, the best fitting detection function also included observer team  
6 and sea state as covariates (Figure 6). The GAM retained fewer variables compared with the  
7 humpback whale model (Table 7, Figure 7) and deviance explained was lower than for  
8 humpback whales at 21%. The point estimate of baleen whale abundance (Table 8) was  
9 43,824 (CV = 0.15, 95% CI 33,509 – 59,077); the density surface model is shown in Figure 8.

10

### 11 **3.3. Estimation of krill consumption**

#### 12 *3.3.1. Humpback whales*

13 Table 9 lists the FMR for individual whales and annual energy requirements (AER) for each  
14 age and sex class of the humpback whale population. For the estimated humpback whale  
15 population (n = 24,543) this results in an annual consumption of 3.5-4.5 million tonnes (Mt)  
16 of krill. Assuming 85% is consumed in high latitude feeding grounds, the total wet weight of  
17 krill consumed in the study area would be 3-3.8 Mt.

18

#### 19 *3.3.2. All baleen whales*

20 Estimating the consumption by all baleen whales was challenging because 27% of the total  
21 count of whales input to the abundance model were not identified to species level. We  
22 assume that the species composition of unidentified sightings and the resulting abundance  
23 estimates were in proportion to that of the identified sightings, although there may be bias  
24 in the unidentified category associated with differences between species in size, behaviour or

1 prominence of diagnostic identification features. Table 10 lists the FMR for individual whales  
2 of mean body mass and the AER for the estimated abundance of each baleen whale species.  
3 Assuming  $\beta$  to be in the range 2-3, 85% of annual consumption would amount to 4.8-7.2 Mt  
4 of krill consumed.

5

#### 6 **4. DISCUSSION**

7 Here we provide the first estimates of whale abundance and krill consumption for the waters  
8 surrounding South Georgia and the South Sandwich Islands (north and east Scotia Sea),  
9 regions that were historically whaled and that currently support a regulated commercial krill  
10 fishery (Trathan et al. 2021). Total abundance of baleen whales across this region is estimated  
11 to be 43,824 (CV = 0.15) and based on the large proportion of humpbacks identified during  
12 the surveys (83%, Table 3), the majority of these whales are likely to be that species. South  
13 Georgia was an epicentre of commercial whaling in the early 20<sup>th</sup> century (Hart 2001), and  
14 over 173,000 whales were killed within a day's sailing of its shores (Allison 2016). The fishery  
15 closed in the 1960s, and whales were rarely seen for the next three decades. Humpback and  
16 southern right whales re-appeared at South Georgia in the 1990s (Moore et al. 1999) but their  
17 occurrence was patchy. For example, the SOWER survey of South Georgia waters in 2000  
18 provided a regional estimate of only 116 baleen whales (Hedley et al. 2001; no estimate of  
19 uncertainty provided). Reports from visitors to South Georgia showed an increase in sightings  
20 up to the 2001 - 2005 period, but reduced sightings in the latter half of that decade  
21 (Richardson et al. 2012). Since 2013, sightings of humpback whales in the summer months of  
22 January and February have increased rapidly however (Jackson et al. 2020). Our 2019  
23 estimate of baleen whale abundance in South Georgia waters is 31,676 (CV=0.16), suggesting

1 a dramatic increase in whales returning to this historically important habitat over the last 20  
2 years.

3

#### 4 **4.1. Humpback whale recovery in the Scotia Arc**

5 The large estimate of humpback whale abundance in the Scotia Arc presented here confirms  
6 the strong recovery of the southwest Atlantic population reported from their winter breeding  
7 habitats in Brazil (Zerbini et al. 2019). The feeding ground range of the southwest Atlantic  
8 humpback population is estimated to span 20-50°W and south from 40°S to the ice edge in  
9 the east, and between 50-70°W from 40°S to the Falkland Islands (Zerbini et al. 2011b).  
10 Humpbacks are now rarely seen in the latter location, although they were occasionally caught  
11 there (Frans & Augé 2016). Modelled recovery of the southwest Atlantic population predicts  
12 breeding ground abundance to be over 24,000 in 2019 (median N = 24,925, probability  
13 intervals 22,369 – 27,007, Zerbini et al. 2019). This estimate is closely concordant with our  
14 abundance estimate of 24,543 (CV = 0.26%) humpback whales present across the Scotia Arc  
15 in the same year. The CV of 26% compares favourably with earlier design- and model-based  
16 abundance estimates in the southwest Atlantic region (Branch 2011, Hedley et al. 2001),  
17 especially given the circumstances of differences between vessels, constraints imposed by  
18 team size, the very large area covered, frequent poor viewing conditions and a survey not  
19 primarily designed as a cetacean line transect survey. Larger team sizes would have been  
20 ideal, such that a dual platform approach could be used to estimate the probability of  
21 detection on the track-line (Buckland et al. 2004). Here, we do not have data to correct for  
22 the proportion of animals potentially missed on the track-line (i.e. we assumed  $g(0) = 1$ ), so  
23 our estimates may be negatively biased and therefore represent conservative estimates of  
24 abundance. Humpback whale abundance in the Scotia Sea was previously estimated at 2,493

1 (CV=0.55) in 2001 (Hedley et al. 2001, Reilly et al. 2004); our new surveys indicate an order of  
2 magnitude increase in the abundance of this species in the last two decades, and given the  
3 survey constraints we consider this a conservative estimate.

4

#### 5 **4.2. Baleen whale abundance**

6 The CV of the abundance estimate for all baleen whales (15%) was lower than that for  
7 humpback whales alone, probably due to the larger sample size. When apportioned to species  
8 in the same proportions as for counts of identified species, the point estimate of humpback  
9 whale abundance derived from the all baleen whale category (36,444) was higher than the  
10 estimate obtained from identified humpback whale sightings (24,543), although it fell within  
11 the 95% CI of 14,863 – 40,528. Such an increase is to be expected given the assumed  
12 predominance of humpback whales in the unidentified category and suggests that the point  
13 estimate for humpback whales is conservative. Pooling data from all baleen whale species  
14 into a single category potentially raises problems for modelling if there are differences in  
15 detectability and habitat partitioning, leading to variation in the influence of environmental  
16 parameters or variation in prey species. In this case, species identity could not be used as a  
17 covariate because 27% of all baleen whale sightings were unidentified to species level.

18 Since the global moratorium on commercial whaling came into effect in 1986, populations of  
19 all baleen whale species (Allison 2016) have not been hunted, and are showing evidence of  
20 recovery in the southwest Atlantic (Calderan et al. 2020, Weir et al. in press, Crespo et al.  
21 2019, Zerbini et al. 2019, Viquerat & Herr, 2017). Here, we estimate that over 45,000 baleen  
22 whales used South Georgia and South Sandwich Islands waters in summer 2019, a signal that  
23 whales are again becoming a significant component of the ecosystems of these waters. The  
24 majority of the whales sighted in both areas were humpback whales. Humpback whales were

## Scotia Arc whale abundance

1 the first species to be exploited during the modern whaling period (Zerbini et al. 2011b), and  
2 therefore have had more time to recover than later exploited species such as the blue, fin and  
3 sei whales. Studies on wintering grounds show that humpback whales can have population  
4 growth rates up to 12% (Wedekin et al. 2017, Noad et al. 2019), which are also higher than  
5 those known for right whales (~7%, Cooke et al. 2001) or anticipated for fin and blue whales  
6 (Vikingsson et al. 2009, Williams et al 2011). Satellite tracking of humpback whale movements  
7 on their Southern Ocean feeding grounds in the east Antarctic shows association with the  
8 marginal ice zone and with areas recently uncovered by retreating ice (Andrews-Goff et al.  
9 2018, Riekkola et al. 2019), likely associated with elevated krill abundance. Similarly, in the  
10 western Antarctic Peninsula region, tracked humpback whales showed a preference for  
11 habitats proximal to sea ice, while avoiding bays in which elevated levels of sea ice persisted  
12 in summer (Friedlaender et al. 2021). In the Scotia Arc, the main feeding ground of humpback  
13 whales is closer to the polar front than the ice edge (Zerbini et al. 2006, Horton et al. 2020).  
14 This is likely a function of the density and abundance of krill present in the northern Scotia  
15 Arc, advected to South Georgia (Murphy et al. 2007).

16 Baleen whale abundance at South Georgia (~31,000 whales) was roughly double that seen at  
17 the South Sandwich Islands, although the South Georgia stratum was less than 10% larger in  
18 area than the South Sandwich Islands stratum. This abundance is likely due to the unique  
19 oceanography of the South Georgia shelf (Young et al. 2014) which makes South Georgia a  
20 high density area for krill, and therefore attracts krill predators (Trathan et al. 2007). South  
21 Georgia lies between two major fronts, the Polar Front to the north, and the southern  
22 Antarctic Circumpolar Current front to the south. The latter loops around three sides of the  
23 island (Figure 1) and brings Antarctic krill into the shelf waters in large quantities. The South  
24 Sandwich Islands are more remote and less well studied, but also support significant

1 populations of krill predators (e.g. Stryker et al. 2020) as they lie along the southern boundary  
2 of the Antarctic Circumpolar Current and receive krill from the Antarctic Peninsula and  
3 Weddell Sea. Whaling was limited at the South Sandwich Islands, with 1,134 whales estimated  
4 to have been killed in local waters (Allison 2016). This was initially due to the remoteness of  
5 the location and poor weather conditions, but the less restricted factory ships did not  
6 subsequently hunt there either, suggesting that this area either was not a historical hotspot  
7 for whales, or that they were so heavily exploited across the region that by the time factory  
8 ship whaling began there were insufficient numbers to make hunting worthwhile at the South  
9 Sandwich Islands (Hart, 2006).

10

#### 11 **4.3. Krill consumption by baleen whales**

12 In this study, we assume that all baleen whales using the Scotia Arc region are feeding on  
13 Antarctic krill. However, some baleen whale species in the Scotia Arc may also feed on other  
14 prey species. While all species feed predominantly on Antarctic krill south of 60°S, fin whales  
15 may also feed on *Thysanoessa macrura* (Herr et al. 2016) and right whales may feed on  
16 copepods (Valenzuela et al. 2019, Jackson et al. 2020), although south of 50°S they are  
17 anticipated to feed predominantly on krill (Tormosov et al. 1998). Feeding on other species  
18 was therefore expected to be negligible. A further assumption made in order to estimate krill  
19 consumption is that 85% of the annual consumption by whales takes place within their high  
20 latitude feeding habitats. The proportion of annual energy requirements ingested during the  
21 summer season at high latitudes is unknown and this is a topic that deserves further research  
22 in order to better understand the annual energy budgets of migratory baleen whales (Owen  
23 et al. 2017). Migration is not synchronous, as whales leave their winter breeding habitat over  
24 a period of two months and this may have ecological consequences as numbers in the feeding

1 grounds gradually increase and subsequently gradually decrease across the feeding seasons.  
2 The sex ratio of humpback whales is biased towards males in breeding habitats (Barlow et al.,  
3 2011), suggesting females may remain in feeding habitats during years in which they are not  
4 breeding. Acoustic monitoring has confirmed that some humpback whales remain in high  
5 latitude habitats year round (Van Opzeeland et al. 2013, Schall et al. 2020) as do some blue  
6 whales (Širović et al. 2009) and similarly, humpback, blue, fin and minke whales have been  
7 acoustically detected in high latitude Arctic habitats year round (e.g. Moore et al. 2012). This  
8 suggests the simple paradigm of migration between summer feeding habitats at high latitudes  
9 and winter breeding habitats at low latitudes does not apply to all individuals of a population  
10 in each year, and that for some individuals a higher proportion of their annual energy needs  
11 may be met in high latitude habitats. We might also expect the abundance of whales in high  
12 latitude feeding habitats to be greater than in lower latitude breeding areas. This variation in  
13 migratory behaviour raises issues for management and calls for further research to better  
14 understand demographic aspects of migration phenology and the winter ecology of baleen  
15 whales at high latitudes.

16 The estimate of 4.8 - 7.2 million tons of krill consumed by all baleen whales in South Georgia  
17 and the South Sandwich Islands represents 7 - 10% of the total available krill biomass  
18 estimated during the 2019 season in the CCAMLR 2000 survey area, extending from the  
19 Antarctic Peninsula to the South Sandwich Islands (Krafft et al. 2021). In the 2019/2020 fishing  
20 season, the commercial krill fishery took 115,319 tonnes of krill from South Georgia and no  
21 krill from the South Sandwich Islands (CCAMLR 2021). Spatiotemporal overlap between the  
22 krill fishery and the summer feeding season of baleen whales has been limited to the South  
23 Orkney (48.2) stratum, in areas where we predict relatively low densities of humpback  
24 whales. In the South Georgia (48.3) stratum, fishing has been restricted to the winter months

1 of June to September and there has been no fishing effort in the South Sandwich (48.4)  
2 stratum (Table 11). Although a range of krill consumption values based on the 95% CI of  
3 abundance estimates could be provided, that would not be equivalent to a 95% CI in  
4 consumption rates due to the unknown variance of energy budget estimates, which were not  
5 statistically based. Such an approach was therefore considered potentially misleading and  
6 instead, a range of consumption estimates were calculated for both humpback and all baleen  
7 whales, based on point estimates of whale abundance and likely energy requirement rates  
8 relative to body weight. There is clearly room for improvement in better understanding  
9 energy budgets of migratory whales and further research into this field should be a high  
10 priority in order to improve the precision of prey consumption rates. Here, we have aimed at  
11 providing a range of estimates, recognising the uncertainty inherent in methods of calculating  
12 consumption rates.

13 Climate change is driving changes in the ecology of a wide range of marine organisms  
14 (Poloczanska et al. 2016) including the distribution of Antarctic krill, which may be contracting  
15 southwards (Atkinson et al. 2019). Ecological relationships between sympatric krill predators,  
16 such as humpback and minke whales are likely to be affected by rapid warming in the  
17 Antarctic region (Friedlaender et al. 2021). Southern right whales have shifted their feeding  
18 distribution in the southeast Atlantic in recent decades which may be a response to changes  
19 in the distribution of prey (van den Berg et al. 2020) and it is expected that other whale species  
20 may similarly respond to changes in the distribution of prey. The increase in abundance of  
21 recovering whale populations may in itself drive ecological changes, on the one hand through  
22 increased competition for food resources that could lead to niche specialisation and on the  
23 other hand through increased primary productivity driven by faecal fertilisation of the ocean  
24 (Lavery et al. 2014, Roman & McCarthy 2010). Models of the impacts of climate change on

1 Southern Ocean ecosystems predict future declines in whale populations as a consequence  
2 of expected future reduced prey availability and increased interspecific competition (Tulloch  
3 et al. 2019), although the complexity of ecological relationships in the oceanic food web  
4 causes inevitable uncertainty. Current changes in whale and krill ecology highlights the need  
5 for further research into the processes involved in this predator-prey relationship, especially  
6 in developing a more robust, statistically based understanding of feeding rates and  
7 requirements, and the need for regular monitoring of abundance and distribution of these  
8 key species.

9

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3

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7 **TABLES**

8 Table 1 – Covariates included in the generalised additive models (GAMs).

Variable name	Description
(x,y)	Lon & lat coordinates transformed to UTM zone 21S
depth	Depth extracted from GEBCO 2019
sdDepth	Standard deviation of depth in 20km grid cells
SST	Monthly sea surface temperature extracted from NOAA Ocean Color
sstFronts	SST fronts extracted using package <i>grec</i> , Belkin & O’Reilly (2009)
CHL	Monthly chl- <i>a</i> extracted from NOAA Ocean Color
chlFronts	CHL fronts extracted using package <i>grec</i> , Belkin & O’Reilly (2009)
CHL_mmmyyyy	Monthly chl- <i>a</i> extracted from NOAA Ocean Color for 12 preceding months
SSS	Sea surface salinity
zooc	Reprocessed zooplankton concentration as carbon in sea water, Lehody et al (2015)
iceDist	Distance to sea ice at its maximum extent in September 2018
curl	FNMOE Wind and Ekman Transport Data. Curl of Surface Wind Stress

9  
 10 Table 2 - Realised effort (Km) for each vessel in each region (as defined by CCAMLR  
 11 Statistical Areas, Figure 1)

Vessel	South Georgia	South Orkneys	South Sandwich	All
DIS	2123	0	3031	5154
CDH	560	415	0	975
KPH	1276	1672	206	3154
All	3042	1043	3134	7219

12  
 13 Table 3 - Summary of humpback whale sightings by each vessel in each region. Note that for  
 14 CDH and KPH only sightings within their primary 90° field of view are included and for DIS  
 15 sightings were truncated at a perpendicular distance of 2.5 km.

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Vessel	South Georgia		South Orkneys		South Sandwich		All	
	Sightings	Count	Sightings	Count	Sightings	Count	Sightings	Count
DIS	94	169			77	145	171	314
CDH	12	19	8	21			20	40
KPH	12	20	22	29	13	15	47	64
All	118	208	30	50	90	160	238	418

1  
2 Table 4 - Numbers of mysticete sightings recorded by DIS, data truncated at a perpendicular  
3 distance of 2.5 km.

Species	South Georgia (SG)		South Sandwich (SS)		SG & SS	
	Sightings	Count	Sightings	Count	Sightings	Count
Antarctic blue whale	5	8			5	8
Fin whale	12	15	22	34	34	49
Sei whale			1	1	1	1
Antarctic minke whale			2	2	2	2
Humpback whale	94	169	77	145	171	314
Southern right whale	1	1	2	3	3	4
Unidentified baleen whale	49	69	31	40	80	109
All baleen whales	161	262	135	225	296	487

4  
5 Table 5 – Approximate significance of smooth terms in the humpback whale model, where  
6 edf is the estimated degrees of freedom, Ref.df constrains the maximum number of knots in  
7 the smooth function and p-value is an approximate indication of significance.

Variable term	edf	Ref.df	p-value
s(x,y)	6.97	29	<0.001
s(log(CHL))	2.07	9	<0.01
s(log(CHL_Apr2018))	2.44	9	<0.001
s(sstFronts)	0.32	9	0.15
s(iceDist)	0.13	9	0.066

8  
9 Table 6 – Abundance estimates for humpback whales across the Scotia Sea

Region (CCAMLR area)	Stratum area (km <sup>2</sup> )	Point estimate	95% CI	CV
South Georgia (48.3)	1,033,249	12,103	7,145 – 20,499	0.27
South Sandwich (48.4)	944,953	11,656	5,865 – 23,164	0.36
South Orkney (48.2)	856,650	785	208 – 2,960	0.76
Scotia Sea total	2,834,852	24,543	14,863 – 40,528	0.26

10  
11 Table 7 – Approximate significance of smooth terms in the model for all baleen whales in

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1 South Georgia and the South Sandwich Islands. Column identifiers as given in Table 5.

Variable term	edf	Ref.df	p-value
<b>s(x,y)</b>	12.16	16	<0.001
<b>s(log(CHL))</b>	2.48	9	0.08
<b>s(sstFronts)</b>	1	9	0.11

2

3 Table 8 – Abundance estimates for all baleen whales in South Georgia and the South

4 Sandwich Islands

Region (CCAMLR area)	Stratum area (km <sup>2</sup> )	Point estimate	95% CI	CV
<b>South Georgia (48.3)</b>	1,033,249	30,905	22,361 – 42,713	0.17
<b>South Sandwich (48.4)</b>	944,953	12,919	7,796 – 21,409	0.26
<b>Total</b>	1,978,202	43,824	33,509 – 59,077	0.15

5

6 Table 9 - Field metabolic rate (FMR) for individual whales and annual energy requirements

7 (AER) for each age and sex sector of the humpback whale population.

Sex	Age	Body weight (kg)	FMR (kJ/day)	% Population	Population AER (MJ * 10 <sup>6</sup> )
<b>Male</b>	Immature	12,314	3507 - 4676	16%	1,473 – 1,964
	Mature	28,323	4367 - 5458	34%	3,897 – 4,871
<b>Female</b>	Immature	15,645	4197 - 5596	16%	1,762 – 2,350
	Mature	32,493	6534 - 7986	34%	6,047 – 7,559
<b>Total</b>					13,179 – 16,743

8

9 Table 10 - Field metabolic rate (FMR) for individual whales and annual energy requirements

10 (AER) for the estimated abundance of baleen whale species.

Identified species	Proportion of identified species	$\bar{N}$	Mean body weight (kg)	Biomass (tons)	FMR (MJ/day)	AER (MJ * 10 <sup>6</sup> )
<b>Humpback whale</b>	83.16%	36,444	25,151	958,014	1,171 – 1,756	15,573 – 23,360
<b>Antarctic blue whale</b>	2.11%	925	84,328	81,319	2,901 – 4,351	979 - 1,469
<b>Fin whale</b>	12.89%	5,649	48,768	288,045	1,924 – 2,886	3,966 – 5,950
<b>Sei whale</b>	0.26%	114	17,780	2,143	903 – 1,354	37.5 – 56

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<b>Minke whale</b>	0.53%	232	6,565	1,583	428 - 641	36 - 54
<b>Right whale</b>	1.05%	460	55,880	26,943	2,130 – 3,196	358 - 537

1

2

3 Table 11. Monthly catch history of *Euphausia superba* (tonnes) in 2019 and 2020 for the strata SO,  
4 SG and SS (CCAMLR 2021).

5

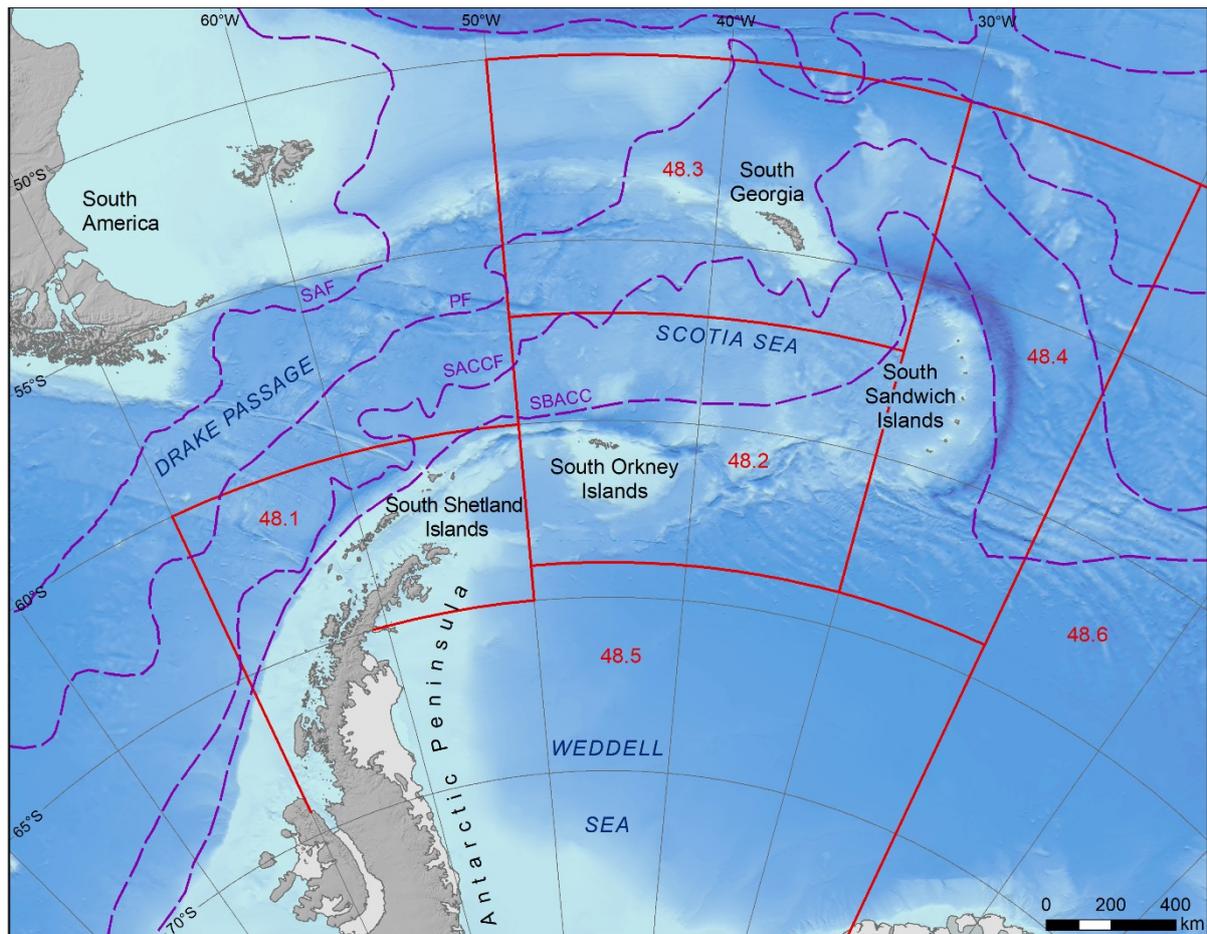
Stratum	2019			2020		
	SO (48.2)	SG (48.3)	SS (48.4)	SO (48.2)	SG (48.3)	SS (48.4)
<b>December</b>	21010	0	0	5460	0	0
<b>January</b>	37422	0	0	63065	0	0
<b>February</b>	45770	0	0	59169	0	0
<b>March</b>	56556	0	0	31376	0	0
<b>April</b>	1816	0	0	15125	0	0
<b>May</b>	0	0	0	10	0	0
<b>June</b>	0	751	0	229	18152	0
<b>July</b>	0	16693	0	0	35536	0
<b>August</b>	0	37317	0	0	43500	0
<b>September</b>	0	17037	0	0	18131	0

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7

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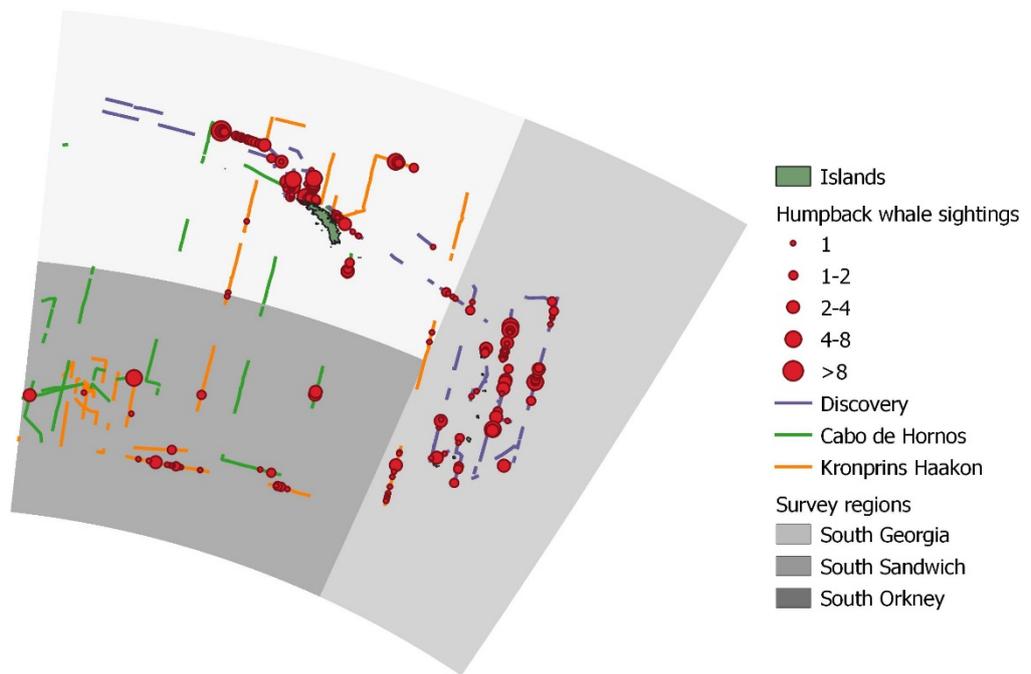
1 **FIGURES**



2

3 Figure 1. The study area showing the location of the three CCAMLR Statistical Areas strata  
4 48.2 (South Orkney Islands), 48.3 (South Georgia) and 48.4 (South Sandwich Islands). The  
5 Polar Front (PF), the Southern Antarctic Circumpolar Current Front (SACCF) and the  
6 Southern Boundary of the Antarctic Circumpolar Current (SBACC) are marked by dashed  
7 lines.

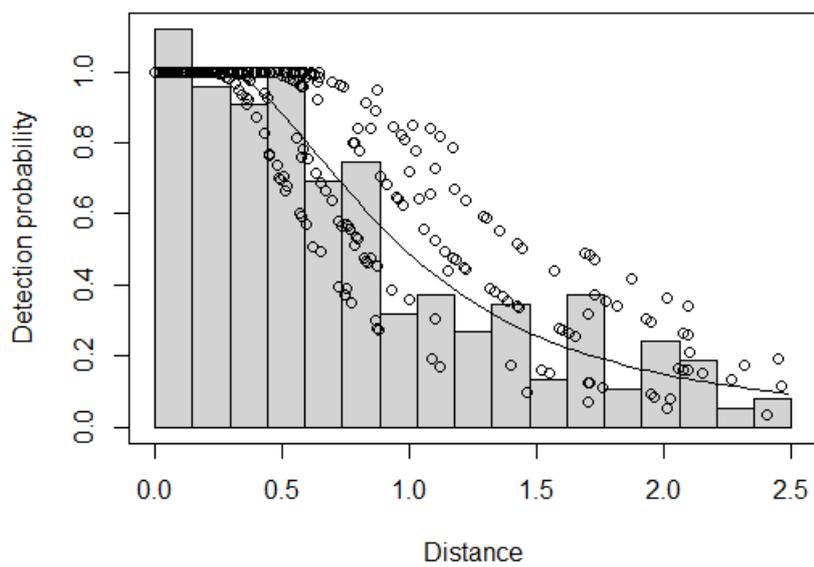
# Scotia Arc whale abundance



1

2 Figure 2. Transect lines and humpback whale sightings overlaid on survey region boundaries.

3 For orientation see Figure 1.



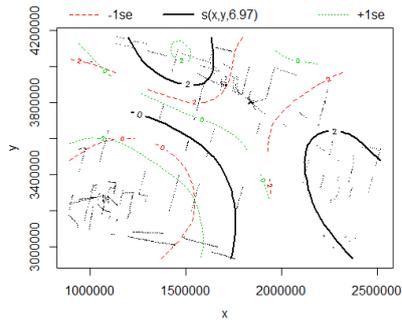
4

Scotia Arc whale abundance

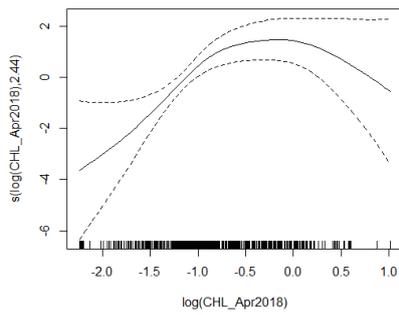
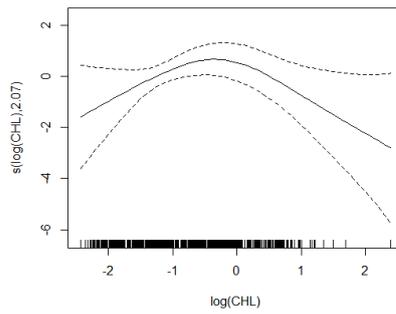
- 1 Figure 3. Perpendicular distances and fitted detection function for humpback whales, with
- 2 observer team and sea state as covariates.

3

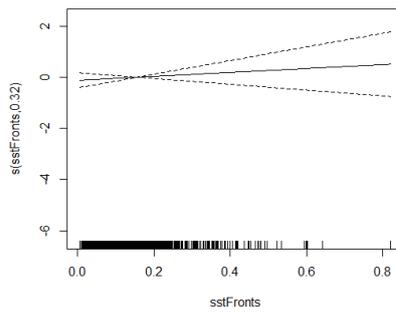
- 4 Figure 4. Plots of smooth functions for the humpback whale GAM.



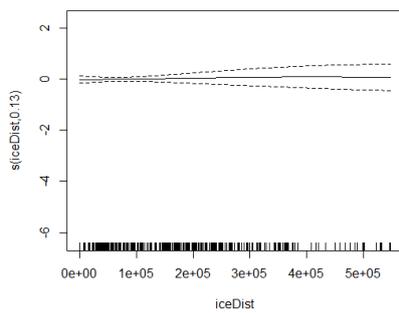
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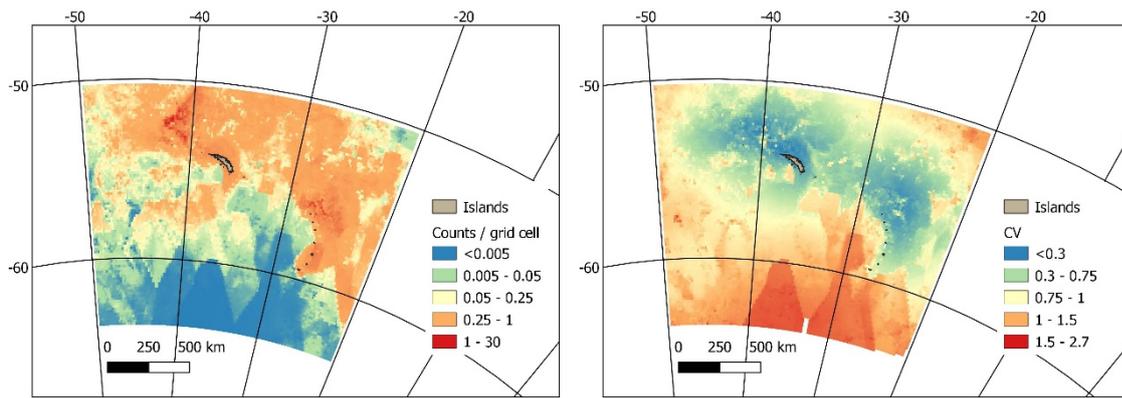
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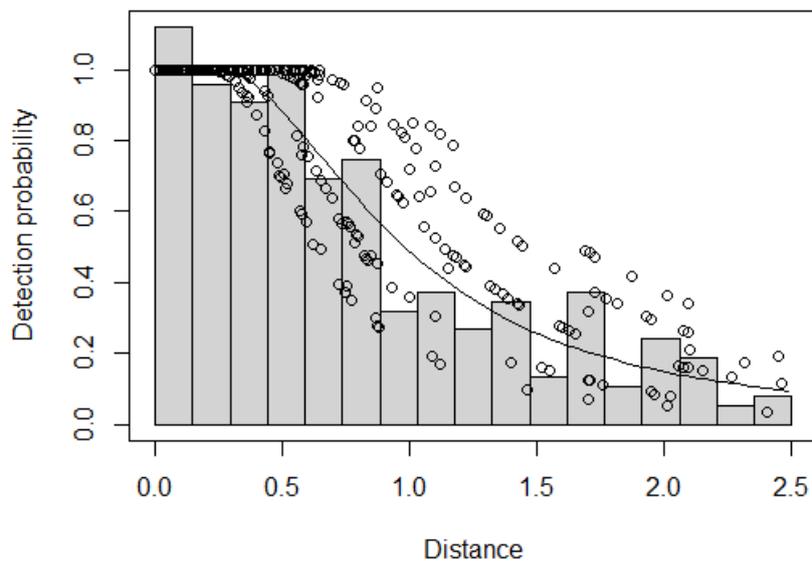
## Scotia Arc whale abundance



1

2 Figure 5. Predicted humpback whale density expressed as counts per grid cell (left) and CV of  
3 estimates (right). Latitude and longitude grid lines at 10° intervals.

4

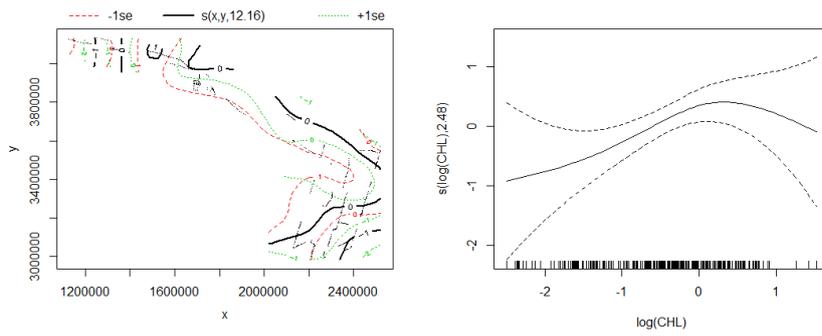


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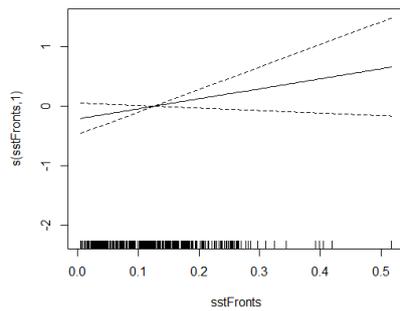
6 Figure 6. Perpendicular distances and fitted detection function for baleen whales, with  
7 observer team and sea state as covariates.

8

# Scotia Arc whale abundance

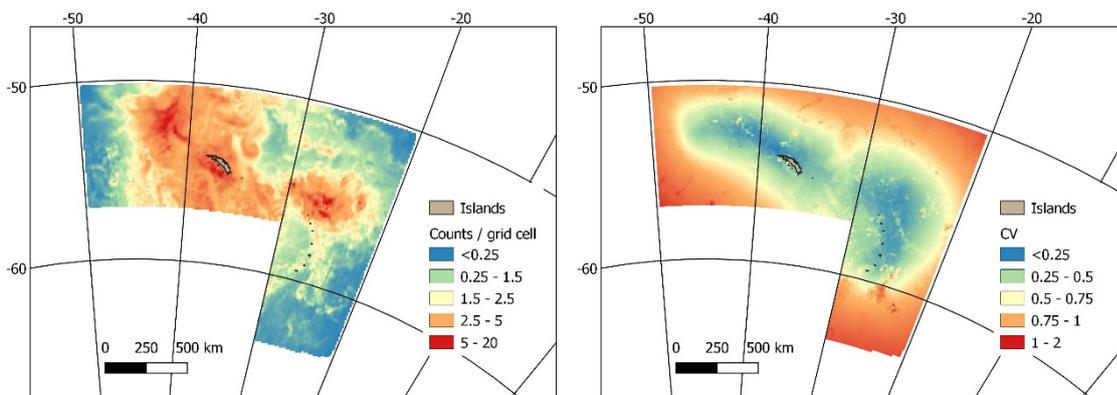


1



2

3 Figure 7. Plots of smooth functions for the baleen whale GAM.



4

5 Figure 8. Predicted baleen whale density expressed as counts per grid cell (left) and CV of  
 6 estimates (right). Latitude and longitude grid lines at 10° intervals.