



Article (refereed) - postprint

Jin, Shuqin; Zhang, Bin; Wu, Bi; Han, Dongmei; Hu, Yu; Ren, Chenchen; Zhang, Chuanzhen; Wei, Xun; Wu, Yan; Mol, Arthur P.J.; Reis, Stefan; Gu, Baojing; Chen, Jie. 2021. **Decoupling livestock and crop production at the household level in China**. *Nature Sustainability*, 4 (1). 48-55. <u>https://doi.org/10.1038/s41893-020-00596-0</u>

© The Author(s), under exclusive licence to Springer Nature Limited 2020

For use in accordance with Nature Research's Terms of Reuse of archived manuscripts

This version is available at https://nora.nerc.ac.uk/id/eprint/528400/

Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at https://nora.nerc.ac.uk/policies.html#access.

This document is the authors' final manuscript version of the journal article, incorporating any revisions agreed during the peer review process. There may be differences between this and the publisher's version. You are advised to consult the publisher's version if you wish to cite from this article.

The definitive version is available at https://www.nature.com/

Contact UKCEH NORA team at noraceh@ceh.ac.uk

The NERC and UKCEH trademarks and logos ('the Trademarks') are registered trademarks of NERC and UKCEH in the UK and other countries, and may not be used without the prior written consent of the Trademark owner.

1	Decoupling livestock and crop production at the household level in China
2	
3	Shuqin Jin ^{1,2,#} , Bin Zhang ^{1,#} , Bi Wu ¹ , Dongmei Han ³ , Yu Hu ¹ , Chenchen Ren ⁴ , Chuanzhen
4	Zhang ⁵ , Xun Wei ⁶ , Yan Wu ⁷ , Arthur P.J. Mol ² , Stefan Reis ^{8,9} , Baojing Gu ^{4,5,*} , Jie Chen ^{1,**}
5	
6	¹ Research Center for Rural Economy, Ministry of Agriculture and Rural Affairs of China,
7	Beijing 100810, PR China
8	² Environmental Policy Group, Wageningen University, Wageningen 6700 EW, the
9	Netherlands
10	³ School of Economy, Hebei University, Baoding 071000, PR China
11	⁴ Department of Land Management, Zhejiang University, Hangzhou 310058, PR China
12	⁵ College of Environmental and Resource Sciences, Zhejiang University, Hangzhou 310058,
13	PR China
14	⁶ China Rural Technology Development Center, Ministry of Science and Technology, Beijing
15	100045, PR China
16	⁷ School of Economics, Nanjing Audit University, Nanjing 211815, PR China
17	⁸ UK Centre for Ecology & Hydrology, Bush Estate, Penicuik, Midlothian, EH26 0QB, UK
18	⁹ University of Exeter Medical School, European Centre for Environment and Health,
19	Knowledge Spa, Truro, TR1 3HD, UK
20	
21	[#] These authors contributed equally to this work.
22	
23	*Corresponding Author at College of Environmental and Resource Sciences, Zhejiang
24 25	University, 866 Yuhangtang Road, Hangzhou 310058, PR China.
26	**Co-corresponding Author at Research Center for Rural Economy, Ministry of Agriculture
27	and Rural Affairs, 56 Zhuanta Hutong, Beijing100810, PR China.
28 29	E-mail addresses: bigu@ziu.edu.cn (B. Gu), chenije21st@sina.com (J. Chen)
30	
31	Animal manure used to be the major source of additional nutrients and crucial for
32	maintaining soil fertility and crop yield in traditional farming systems. However, it is
33	increasingly not recycled nowadays, wasting vital resources and damaging the
34	environment. By using long-term (1986-2017) data from a rural household survey
35	(>20,000 households) across China, here we show that the share of rural households with

both crop planting and livestock raising (CPLR) has sharply declined from 71% in 1986 to only 12% in 2017. Compared to households with only crop planting, the CPLR households apply less synthetic fertilizer and more manure per cropland area. However, manure production in one third of CPLR households has exceeded the nutrient requirement of crop growth on their croplands. Rebuilding the links between livestock and croplands at regional scale thus provides vital opportunities for the sustainable intensification of agriculture in China.

43

Feeding an increasingly affluent global population with less pollution is one of the major global 44 challenges integral to the attainment of the United Nations' Sustainable Development Goals 45 (SDGs)¹. Animal products have contributed to over half of the protein supply in developed 46 countries, and have shown a sharp increase in developing countries². To produce these animal 47 products, an increasing number of feedlots are built, which concentrate large numbers of 48 animals. At the same time, rearing livestock on a farm backyard has transitioned to industrial 49 livestock farming ^{3, 4}, and a decoupling of livestock and croplands on household level is 50 occurring (Fig. 1). The decoupling has substantially reduced the manure recycling rate and had 51 detrimental effects on the environment ^{5, 6}. Recycling manure to cropland is not only a challenge 52 for developing countries such as China, but also for developed countries such as United States 53 (US) and regions in Europe ^{7, 8}. Therefore, understanding why such a decoupling occurs and 54 how it affects the sustainability of crop-livestock coupled systems is crucial for rebuilding the 55 linkage between croplands and livestock production. 56

57

The demand for and the economic returns from livestock products have been found important 58 factors affecting the growth of livestock industries ³. However, these changes alone may not 59 immediately lead to the decoupling of croplands and livestock, and the transition of livestock 60 production from smallholder to industrial farming. In China today, small and medium scale 61 livestock farms still play an important role for the supply of animal products ³. The underlying 62 63 reasons why smallholder farmers, traditionally the major form of livestock production in China, give up livestock production is still not well understood. Traditionally small scale livestock 64 farming was matched with a corresponding amount of cropland cultivation at household level, 65 fostering within-household manure recycling 4, 9. Whether the decoupling of livestock 66 production and croplands on household level will result in lower manure recycling rates has not 67 been robustly quantified to date. 68

China is the world largest market for animal products and also the largest consumer of synthetic 70 fertilizers applied to croplands, accounting for about one third of global total nitrogen fertilizer 71 consumption². Overall synthetic fertilizer use efficiency (fertilizer nutrient harvested in crops 72 divided by total fertilizer use) is lower than 50% and the average manure recycling ratio is lower 73 than 40% in China, indicating that over half of fertilizer and manure nutrients are lost to the 74 environment¹⁰. As a consequence, agriculture has become the dominant source of air and water 75 pollution ¹¹. Reducing these nutrient losses has become a grand challenge for China in the 76 context of achieving the SDGs¹². Within contrast to large-scale farming in Europe (> 30 ha) or 77 the US (>150 ha), the average cropland size is only around 0.5 ha per rural household in China 78 ¹³. Application patterns of synthetic fertilizers and manure in large-scale farming are 79 substantially different from those on smallholder farms ^{14, 15}. How these differences affect the 80 coupling/decoupling of livestock and croplands is not well understood and requires further 81 82 studies. In this paper, we contribute to a better understanding by basing our research on longterm data (1986-2017) from a rural household panel survey across China (>20,000 households, 83 84 Extended Data Fig. 1 to Fig. 3). We address the following key questions: (1) to what extent the decoupling between livestock and cropland production occurs at household level; (2) whether 85 such decoupling leads to manure nutrient loss and increased use of synthetic fertilizers; and (3) 86 what are key reasons for such a decoupling to occur and pathways to future recoupling between 87 livestock and croplands. 88

89

90 **Results**

In this paper, we divided all surveyed households into four key groups: (I) combined crop planting and livestock raising (CPLR), (II) only crop planting, (III) only livestock raising, and (IV) no crop or livestock. These category-IV households generally engage in agricultural activities through labor rental such as work in large farms operated by other households or in non-agricultural sectors. We maintain these households in the survey since these farmers still live in the villages and reflect the changes in rural China.

97

Household share The overall share of CPLR households declined sharply from 71% in 1986 to 12% in 2017, while households with only crop planting increased from 26% to 57% during the same time period (Fig. 2a). It suggests that on-farm decoupling between livestock and cropland at household level did occur in China to a large extent. Meanwhile, the share of rural households no longer participating in agricultural production increased substantially from 3% to 31% between 1986 and 2017.

Households with only livestock raising accounted for around 1% during the study period (Fig. 105 2a). Under the Household Contract Responsibility System (HCRS), each rural household is 106 allocated some cropland area. As a result, there are few households with only livestock raising, 107 but no crop planting. However, this decoupling does not indicate the disappearance of livestock 108 production in China; on the contrary, more professional and centralized livestock farms are 109 emerging (Extended Data Fig. 4). The majority of livestock production in 1986 originated from 110 rural households. By 2010, in contrast, approximately half of the livestock production 111 originated from rural households, i.e. CPLR and livestock-only households³. The remaining 112 50% of livestock production is derived from industrial-scale livestock farms, who make up less 113 than 1% of total livestock farms in China (Extended Data Fig. 4). These are not normally part 114 of the survey, but industry statistics specifically for industrial farms are used to reflect the 115 changes of these farms. The survey of rural households utilized in this study complements these 116 analyses of industrial farms. 117

118

Animal stocking density In the context of decoupling, we found that livestock density (i.e., 119 pig number equivalent per cropland per household, see Methods) in CPLR households has 120 increased from 9 to 31 pigs per hectare (ha) cropland between 1986 and 2017 (Fig. 2d). The 121 largest livestock farm was found with over 5,000 pigs from one rural household. Increasing 122 animal stocking density resulted in a situation where the average manure amount produced at 123 CPLR households exceeded the nutrient requirement by their associated cropland (which is 15 124 pigs per ha cropland, equivalent to 75 kg N ha⁻¹) since approximately 2002. In 1986, 90% of 125 CPLR households raised less than 15 pigs per ha cropland, and by 2017 this value declined to 126 63% (Fig. 2b). It means that in 2017 in over one third of the CPLR households manure 127 production exceeded the nutrient carrying capacity of their cropland (total nitrogen required by 128 crops on these croplands). If this manure surplus is not transported to and applied on 129 130 neighboring cropland, it leads to nutrients being lost to the environment. And as surplus manure 131 is either discarded as waste, or applied as excess manure to cropland fields, the lack of uptake by crops leads to increased losses to the environment. 132

133

Mechanization and the decline of draft animal use We found an increasing trend of mechanization for both CPLR and crop-only households between 2004 and 2017 (Fig. 2c). However, the degree of mechanization is much lower for CPLR households compared to that of crop-only households. Accordingly, we noted that the share of households utilizing draft animals declined sharply from 1995 to 2017, which is consistent with the increase inmechanization (Fig. 2d).

140

Manure, fertilizer use and farmland size Compared to households with only crop production, 141 CPLR households use lower amounts of synthetic fertilizers and more manure per ha cropland 142 (Fig. 3). It means CPLR households use manure nutrients instead of synthetic fertilizer input, 143 and the average substitution ratio (calculated based on the difference of synthetic nitrogen 144 fertilizer uses in CPLR and crop-only households) was 11% during the study period. 145 Nevertheless, with the increase of synthetic fertilizer use, the substitution ratio declined from 146 18% in 1995 to only 10% in 2017. The manure uses in crop only households also suggest that 147 part of the manure produced from CPLR or livestock only households is transported to and 148 applied on neighboring croplands. However, we found that the proportion of manure use in 149 CPLR households is indeed higher than that in households with only crop production, and 150 higher animal stocking density corresponded with a higher manure recycle ratio (Fig. 3c & d). 151 It suggests that the decoupling of livestock and croplands reduces the use of manure, may 152 leading to environmental pollution and decline of cropland soil fertility. 153

154

However, we found that the synthetic nitrogen fertilizer uses significantly increase with animal 155 stocking density (Fig. 3b). It implies that CPLR households with higher animal stocking density 156 use not only more manure than CPLR household with low animal stocking density, but also 157 more synthetic fertilizers. Farmland size decreases sharply with the increase of animal stocking 158 density (Fig. 4a). This indicates that farmers with smaller farmland size raise more livestock, 159 substituting income from crop yield with income from animal products. Meanwhile, small 160 farmland size leads to more synthetic fertilizers use per hectare ¹³, and thus nitrogen application 161 rate significantly increases with animal stocking density (Fig. 4b). This is surprising as these 162 farmers have abundant amounts of manure but still use a large amount of synthetic fertilizers. 163 This may occur due to the inconvenience for smallholders to store and apply manure on their 164 165 small farmland areas. It is easy for smallholders to use more fertilizers to increase yield other than relying on other inputs for instance through machinery use and advanced knowledge for 166 nutrient management. 167

168

169 **Spatial variation** The share of CPLR households declined for all the villages across China 170 between 1986 and 2017 (Fig. 5a & b). The largest decline was observed for the North China 171 Plain, which is the major crop production area in China, and along the East Coast. Western 172 China showed a comparatively smaller decline of the share of CPLR households, especially in 173 the hilly regions, such as in southwestern China. The share of households with draft animals 174 also declined sharply for most villages across China in this period and ratios of draft animal use 175 above 5% were only found in some hilly villages by 2017 (Fig. 5c & d).

176

Manure input has declined to less than 5% of total nutrient input to croplands in most crop-only households in 2017. Values above 5% are mainly found in middle and western China. While for most CPLR households manure use is still higher than 5% (Extended Data Fig. 5a & b), a substantial decline was also found in the North China Plain. In contrast, much higher degrees of mechanization were found in crop-only households compared to that of CPLR households, especially in the North China Plain where CPLR households have become an exception by 2017 (Extended Data Fig. 5c & d).

184

185 **Discussion**

A decoupling between livestock and croplands has been observed at household level in China. 186 This is illustrated well from two aspects: first, only 5% of rural households keep draft animals, 187 reducing manure production for recycling; second, only 12% of rural households still pursue 188 combined animal and crop production, while an increasing share of CPLR households now 189 produce manure in excess of the crop requirements of their croplands. The overload of manure 190 leads to a large amount of nutrient loss to the environment⁹. These findings indicate that manure 191 production has become more concentrated at household level, leading to a reduced potential for 192 on-farm recycling of manure to croplands. Previous studies deduced that less synthetic 193 fertilizers would be used if coupling of livestock and croplands prevailed ^{9, 16}, and our results 194 provided solid evidence to support such conclusions in principle (Fig. 3). CPLR households are 195 shown to use fewer synthetic fertilizers and more manure compared to crop-only households. 196 Although the result is based on nitrogen fertilizers, the same findings still hold for phosphorus 197 and potassium fertilizers when used them for the analysis (Extended Data Fig. 6). It reveals that 198 199 the conclusion of decoupling between livestock and croplands and their effects on manure use is robust, and more attention should be given to such a decoupling process. 200

201

However, as soon as the animal stocking density exceeds the carrying capacity of cropland associated with the farm, overuse of both synthetic fertilizers and manure is found. While it seems paradox that more synthetic fertilizers are used on farms with coupled livestock and croplands, these findings imply that the coupling between livestock and croplands on household

level is only functional when manure production is not exceeding the carrying capacity of 206 surrounding croplands. Once animal numbers exceed the threshold, surplus manure would need 207 to be transported far away to other croplands, leading to an increase in transportation costs and 208 result in a reduction of manure recycling ^{16, 17}. With the increase of livestock production, 209 rebuilding the linkage between livestock and cropland beyond the household level and at the 210 regional level is thus a crucial step towards nutrient recycling and thus sustainable 211 intensification. We indeed find the crop only households also use manure that is transferred 212 from their neighbors. However, matching livestock and croplands exactly on their distribution 213 would benefit the manure recycle, just like the CPLR households within which livestock and 214 croplands are tightly coupled. Thus, relocating livestock based on the distribution of croplands 215 will lead to a reduction in transportation cost of manure and increase the manure recycle ratio 216 ^{6,18}, promoting the recoupling between livestock and croplands on regional scale. 217

218

To quantify the underlying driving forces of decoupling, we estimated the changes of the shares 219 220 in CPLR farms, the use of draft animal and livestock production per land area using panel models (Table 1). Results suggest that mechanization, synthetic fertilizer use and the income 221 share derived from non-agricultural sector activities have significant adverse effects on both 222 draft animals raising and the CPLR share. Each 1% increase in machinery and fertilizer use 223 corresponds with a reduction of 0.1% and 0.05% of draft animal raising and CPLR shares, 224 respectively. The degree of mechanization in Chinese agriculture has increased 8-fold between 225 1978 and 2017, and overall mechanization had reached 65% of the farms by 2017¹⁹. This is 226 similar for synthetic fertilizer use, which has increased 7 times in the same period ¹⁹. The low 227 transportation and application cost of synthetic fertilizers (per amount of nutrient applied) 228 229 largely promote its use in contrast to manure use, despite the fact that manure production also increased substantially since 1978 and is available as a waste product ³. 230

231

With the increase in urbanization, over 200 million farmers have been attracted to seek for part-232 time employment in urban areas ²⁰. Compared to the high income potential from non-233 agricultural sectors or large industrial farming sectors, smallholder livestock raising and manure 234 recycling are less lucrative for farmers ¹³. Due to rise in part-time jobs, which mainly attract 235 young or middle aged farmers, the remaining farms with a primary focus in agriculture are 236 mainly operated by older men or women, who are more likely to reduce the labor-intensive 237 livestock raising and manure recycling activities ¹⁴. The labor shortage, combined with the 238 increase in mechanization and synthetic fertilizer application, lead to a substantial decline in 239

- 240 draft animal raising and CPLR shares (Table 1).
- 241

Farmland size has an inverted U-shaped relationship with livestock raising (Table 1). Both 242 increasing and decreasing farmland size can reduce the livestock raising potential with a turning 243 point at around 3.7 - 4.8 ha. While farmland size below the turning point will lead to increase 244 uptake of non-agricultural part-time jobs of farmers or change to livestock-only farmers, 245 farmland size above the turning point will promote the professional operation of large-scale 246 crop production units (Table 1). Based on this analysis, an optimal farmland size at about 4 ha 247 may be conducive to increase the recoupling of livestock and croplands at household level. 248 With the increase of current average farmland size (~ 0.5 ha) to such an optimal level, the 249 250 increase in share of CPLR households and a simultaneous reduction in the livestock density e.g. number of pigs per ha cropland, could successfully lower manure loading to croplands and 251 252 consequently reduce nutrient losses to the environment. Safeguarding that livestock stocking densities are set not to exceed the surrounding farmland's carrying capacity has been suggested 253 254 to be crucial for the development of green agriculture by the Chinese government, labelled as "Suitable Scale Farming" and "Cropland-based Livestock Farming". Our study provides a 255 preliminary quantification of the such suitable levels at national scale, and future research 256 would be required to refine the results on local scale (e.g. county) for a nation-wide 257 implementation of scale farming. 258

259

In the context of mechanization, synthetic fertilizer uses and urbanization, the trend towards 260 decoupling between livestock and croplands is not easy to reverse at household level. However, 261 262 the observed response of CPLR share to changes in farmland size and the manure use in croponly households provides two potential pathways. First, increase farmland size through policy 263 regulation such as the Land Transfer System (LTS) or reform of land tenure system ¹⁵. Such a 264 change could trigger increases in farmers' income with both extra income from livestock 265 production and increase in farm size ¹³. Under such a pathway, farm size could increase to 266 267 around 4 ha, 8 times larger than current average levels. Labor productivity would likely increase not only from efficiency gains of larger parcels of land managed, but also from extra livestock 268 269 raising and hence income generation opportunities for farmers. Second, fostering recoupling of livestock and croplands on regional scale ⁹. Although the crop-only households do not have 270 their own manure production activities, they can utilize manure produced by neighboring farms 271 on their croplands. With a continued increase of farmland size after the turning point, crop-only 272 273 management for farmers would provide a viable option for farmers to increase income, given

the scale effect of farming and specialized production ¹³. Therefore, despite a de-facto separation of crop and livestock production, rebuilding the linkage between livestock and croplands through cooperation between crop-only and livestock-only households on local and regional scales needs to be facilitated ²¹.

278

Even though livestock production from large industrial-scale feedlots increases dramatically, 279 smallholder farmers still need to play an important role in animal products supply in the long-280 term³. Crop production is still dominated by smallholder farms and their farmland size is much 281 lower than the optimal farmland size, i.e. 4 ha²². Matching livestock production and cropland 282 is crucial for the recoupling of livestock and cropland in the coming decades. To reduce the 283 284 water pollution from livestock production, Chinese governments have relocated pigs from South China to North China; however, this relocation may lead to both food insecurity and new 285 pollution since not considering the linkage between livestock and croplands ²³. This highlights 286 the importance to couple livestock and croplands on mitigation of livestock pollution. 287 Increasing farmland size as a starting point to reduce synthetic fertilizer use, then relocating 288 livestock based on the distribution of croplands and their farm size can close the nutrient cycle 289 within agriculture at local to regional level ⁶. To achieve this, multiple stakeholders need to be 290 involved, including cropland farmers, livestock farmers, governments, enterprises and scientists. 291 Governments play an important role due to the state-owned land tenure system and 292 environmental control criteria require policy regulation in China⁸. Although central 293 government has implemented measures such as subsidizing manure recycling to rebuild the 294 linkage between livestock and croplands, many of these measures still focus solely on large-295 scale farms, not including small-scale livestock farms that account for about half of meat 296 production in China ⁹. Furthermore, incentive mechanisms are needed to bring together 297 cropland and livestock farmers to facilitate collaboration on either self-organized manure 298 trading or commercial services offering such transfers. Scientists are needed to provide a 299 systematic underpinning for the design of a suitable coupling system, including determining 300 301 optimal farm size, distribution of feedlots and crop structures etc. Scientific research can further support government to determine appropriate environmental control criteria for cropland-302 303 livestock coupling systems.

304

The trend of decoupling between livestock and cropland farming with the increase of industrial livestock farms is not a problem faced only by China, but also found in other regions around the world with economic growth and increasing urbanization ⁵. Similar decoupling processes

have already been observed in developed regions in Europe and the US decades ago ^{6, 16}. These 308 challenges have been partly overcome and manure recycling has accounted for about half of 309 the nutrient input to cropland in these regions ²⁴. Yet, excess nutrient pollution from manure 310 still contributes to substantial damage and costs to the environment and human health there ²⁵. 311 However, the situation in emerging economies such as China and India is more serious due to 312 the degree and rapid nature of agricultural growth, resulting in substantial livestock-related 313 environmental pollution pressure³. Recoupling livestock and cropland is thus an urgent and 314 complex challenge to address. In developing countries (e.g., countries in Africa) decoupling 315 between livestock and croplands is not a major challenge yet due to the costs and availability 316 of synthetic fertilizers and low urbanization level ²⁴. If China could solve this challenge, it could 317 provide an example for these countries to avoid the decoupling process during their agricultural 318 transition. This is not only important for achieving SDGs in China, but also other parts of the 319 320 world in the context of globalization and international trade.

321

322 Methods

In the past, rural households normally had two kinds of livestock (Fig. 1). One was a draft 323 animal, such as ox, horses, donkeys, and mules. These animals are used for ploughing and for 324 short-distance transport. Draft animals were mainly fed with straw, which are digested, excreted 325 and the manure returned to croplands to provide important nutrient input before synthetic 326 fertilizers were commonly used. Other animals are reared for meat and other animal products, 327 such as pigs, chickens, ducks and others, and mainly grain-fed. Their manure is another 328 important organic fertilizer source. Before industrial farming took hold, smallholder households 329 330 normally engaged in both crop planting and livestock raising and thus could recycle the manure within their household farm operation. With economic growth and urbanization, linkage 331 between livestock and croplands is broken. Industrial livestock farming is blooming and 332 centralized on small piece of lands far from croplands. Draft animals were gradually replaced 333 by machines and manure substituted by synthetic fertilizers, and straw feed was also replaced 334 by forages or grains. Non-recycled manure and straw have meanwhile become key contributors 335 to environmental pollution, from eutrophication to air pollution due to ammonia (NH₃) emission 336 337 from manure and pollutants (e.g., fine particles) emission during straw burning in fields.

338

Household survey In this paper, the household data were obtained from Fixed Observation Rural Survey (FORS) that was established in 1984. At that time, the HCRS had just been established in China, leading to the emergence of smallholder farmers. Before the HCRS was introduced, collective farms were the major organization form for agricultural production at village level. Thus, the FORS at household level under the HCRS provide the longest and largest dataset of farming survey data, which provides detailed information for scientific research in China.

346

The FORS has formally started since 1986. The management office is in the Research Center 347 for Rural Economy, which belongs to the Ministry of Agriculture and Rural Affairs, China. The 348 system surveys cover more than 20,000 farming households and more than 300 villages in 31 349 provinces including autonomous regions and municipalities, except Hong Kong, Macao and 350 Taiwan (Extended Data Fig. 1 and Fig. 2). The sample of farmers was obtained by stratified 351 352 sampling methods combining classification sampling and random sampling. Once the sample households were confirmed in 1986, they remained unchanged for a long time and follow-up 353 investigation has been continued. Only when the chosen rural household moved to an urban 354 area permanently or all the household members are deceased, we shall amend the sample; 355 otherwise, they would stay in the FORS. Thus, some households no long participate in 356 agricultural activities are still included in our survey to reflect the changes of rural China. Due 357 to the FORS only surveys rural households, thus, the independent agricultural companies such 358 as industrial farming for crops and livestock are not included. 359

360

The survey method requires sample households to keep daily accounts (Extended Data Fig. 3), and the investigators regularly visit the households to summarize and collate data at the end of the year. The surveys include detailed information about the basic demographic characteristics, income and expenditure, and production and operation of farming households. Due to issues of data continuity and availability, this paper uses the number of draft animals and all livestock species to calculate pig equivalent numbers for comparison. The cropland area uses the cultivated land operated by farming households.

368

Coupling calculation According to key parameters of crop planting and livestock raising, farmers can be divided into four categories, including CPLR, only crop planting, only livestock raising, and no crop or livestock. The crop planting only household is defined as that the area of cultivated land is greater than zero, but without livestock raising. The livestock raising only household is defined as that the number of draft animals or livestock raised is greater than zero, but the cropland area is zero. The no crop or livestock household has neither livestock numbers, nor farmland area. The CPLR households are those who have both cropland cultivation and 376 livestock raising.

377

The coupling of livestock and croplands refers to recycling manure to croplands and the manure loading is within the carrying capacity of cropland at household level. For example, environmental legislation has been implemented to limit the animal stocking density to 2.5 cow units per ha cropland surrounding the feedlots in the Netherlands ²⁶. Otherwise, farmers have to pay for manure disposal. In this paper, "pig-farmland ratio" is established to reflect the degree of recoupling between livestock and croplands.

$$PFR = \frac{Pig \text{ equivalent (head)}}{Farmland (ha)}$$

Based on The Technical Guidelines for Measuring the Bearing Capacity of Soil Contaminated 385 by Livestock and Poultry Manure issued by Ministry of Agriculture and Rural Affairs in 2018, 386 we adopted the standard of PFR = 15 pigs per ha farmland to measure whether the livestock 387 raising is over the limit. One pig equivalent normally represents manure production with 5 kg 388 nitrogen ²⁷. When $PFR \le 15$, the recycling of manure to cropland is within the carrying capacity 389 (nutrient requirement by crops) and has no significant environmental pollution effects. However, 390 when PFR > 15, the manure load exceeds the carrying capacity, leading to environmental 391 pollution if manure is only applied on farm. All other livestock types are converted into pig 392 number equivalents according to the conversion standard of 100 pigs = 15 cows = 30 beef cattle393 (draft cattle) = 250 sheep = 2500 poultry.394

395

Panel model analysis The long-term survey allows us to estimate the relation between decoupling of livestock and croplands with machinery and synthetic fertilizer use, farmers' income source, farm size, while controlling for compounding factors such as year, location using panel model analysis. The panel model compiles data on both temporal and spatial scales (1986-2017, over 20,000 households), which can reduce the impact of time invariant-omitted variables and improve the effectiveness of estimates.

402

We estimated the following equation using data on households that still have cropland cultivation:

405 $Y_{it} = \alpha + \gamma \cdot \ln Input_{it} + \rho \cdot Income + \theta_1 \cdot farmsize_{it} + \theta_2 \cdot farmsize_{it}^2 + \sum_k \beta_k x_{kit} + \varepsilon_i$

- 406 where subscript *it* denotes households *i* in time t; Y is the draft animal and the decoupling of livestock and croplands for the households in model 1 and 2, respectively. It is dummy variable, 407 i.e. 1 and 0, referring to having or no livestock raising, respectively. Y is continuous variable 408 in model 3 referring to livestock density in CPLR households (Table 1); Input is synthetic 409 fertilizer and degree of mechanization for crop production; *Income* is the share of farmers' 410 income from non-agricultural sectors; *farmsize* is the farmland area in the household; x_k 's 411 are various control variables affecting the recoupling of livestock and croplands, including 412 413 dummy variable for region, etc.; γ , ρ , θ and β_k are estimated coefficients; and ε_i is the error term. The detailed description of panel model is listed in SI text. 414
- 415

416 Data availability

417 Data of the main findings can be found in supplementary information, and any further data that 418 support the findings of this study are collated from literature sources as cited or available from 419 the corresponding author upon reasonable request.

420

421 **References:**

422 1. Griggs, D., *et al.* Sustainable development goals for people and planet. *Nature* 495, 305423 307 (2013).

424 2. FAO. FAOSTAT: FAO Statistical Databases. (Rome, Italy, 2020).

- 3. Bai, Z., *et al.* China's livestock transition: Driving forces, impacts, and consequences. *Sci Adv* 4, r8534 (2018).
- 427 4. Oenema, O. Nitrogen budgets and losses in livestock systems. *International Congress*428 *Series* 1293, 262-271 (2006).
- 5. Sutton, M.A., *et al. Our Nutrient World: the challenge to produce more food and energy with less pollution.* (Centre for Ecology and Hydrology (CEH), 2013).

431 6. van Grinsven, H.J.M., *et al.* Reducing external costs of nitrogen pollution by relocation of
432 pig production between regions in the European Union. *Reg Environ Change* 18, 2403-2415
433 (2018).

- 434 7. Sutton, M.A., *et al.* Too much of a good thing. *Nature* **472**, 159-161 (2011).
- 435 8. Gu, B., Zhang, X., Bai, X., Fu, B. & Chen, D. Four steps to food security for swelling cities.
 436 *Nature* 566, 31-33 (2019).
- 437 9. Zhang, C., *et al.* Rebuilding the linkage between livestock and cropland to mitigate
 438 agricultural pollution in China. *Resour Conserv Recy* 144, 65-73 (2019).

- 439 10. Gu, B., Ju, X., Chang, S.X., Ge, Y. & Chang, J. Nitrogen use efficiencies in Chinese
- 440 agricultural systems and implications for food security and environmental protection. *Reg*441 *Environ Change* 17, 1217-1227 (2017).
- 442 11. Gu, B., Ju, X., Chang, J., Ge, Y. & Vitousek, P.M. Integrated reactive nitrogen budgets and
 443 future trends in China. *Proc Natl Acad Sci USA* 112, 8792-8797 (2015).
- 12. Ma, L., et al. Exploring Future Food Provision Scenarios for China. Environ Sci Technol
- 445 **53**, 1385-1393 (2018).
- 446 13. Wu, Y., et al. Policy distortions, farm size, and the overuse of agricultural chemicals in
- 447 China. Proc Natl Acad Sci USA 115, 7010-7015 (2018).
- 448 14. Ju, X., Gu, B., Wu, Y. & Galloway, J.N. Reducing China's fertilizer use by increasing farm
 449 size. *Global Environ Chang.* 41, 26-32 (2016).
- 450 15. Fan, L., *et al.* Decreasing farm number benefits the mitigation of agricultural non-point
 451 source pollution in China. *Environ Sci Pollut R* 26, 464-472 (2019).
- 452 16. Naylor, R. Losing the Links Between Livestock and Land. *Science* **310**, 1621-1622 (2005).
- 453 17. Willems, J., *et al.* Why Danish pig farms have far more land and pigs than Dutch farms?
- 454 Implications for feed supply, manure recycling and production costs. *Agr Syst* 144, 122-132
 455 (2016).
- 456 18. Garnier, J., et al. Reconnecting crop and cattle farming to reduce nitrogen losses to river
- 457 water of an intensive agricultural catchment (Seine basin, France): past, present and future.
 458 *Environ Sci Policy* 63, 76-90 (2016).
- 459 19. National Bureau of China. National data. (Beijing, 2019).
- 460 20. Bai, X., Shi, P. & Liu, Y. Society: Realizing China's urban dream. *Nature* 509, 158-160
 461 (2014).
- 462 21. Zheng, C., Liu, Y., Bluemling, B., Mol, A.P.J. & Chen, J. Environmental potentials of
- 463 policy instruments to mitigate nutrient emissions in Chinese livestock production. Sci Total
- 464 *Environ* **502**, 149-156 (2015).
- 465 22. Cui, Z., *et al.* Pursuing sustainable productivity with millions of smallholder farmers.
 466 *Nature* 555, 363-366 (2018).
- 467 23. Bai, Z., et al. China' s pig relocation in balance. Nature Sustainability 2, 888 (2019).
- 468 24. Zhang, X., *et al.* Managing nitrogen for sustainable development. *Nature* **528**, 51-59 (2015).
- 469 25. van Grinsven, H.J.M., et al. Costs and Benefits of Nitrogen for Europe and Implications
- 470 for Mitigation. *Environ Sci Technol* **47**, 3571-3579 (2013).
- 471 26. Oenema, O., et al. Nitrogen in current European policies. in The European Nitrogen

- 472 Assessment: Sources, Effects and Policy Perspectives (ed. M.A. Sutton, et al) (Cambridge
- 473 University Press, Cambridge, 2011).
- 474 27. Gu, B., *et al.* Toward a generic analytical framework for sustainable nitrogen management:
- 475 application for China. *Environ Sci Technol* **53**, 1109-1118 (2019).
- 476

477 Acknowledgements

- 478 This study was supported by the National Key Research and Development Project of China
- 479 (2016YFD0201304, 2018YFC0213300), National Natural Science Foundation of China
- 480 (41822701 and 41773068), National Social Fund of China (18ZD48) and the Fundamental
- 481 Research Funds for the Central Universities (2019XZZX004-11). S.R.'s contribution was
- 482 supported by the UK Natural Environment Research Council (NERC) National Capability
- 483 programme SUNRISE (NE/R000131/1).
- 484

485 **Competing interests**

- 486 The authors declare no competing interests.
- 487

488 Author contributions

- 489 S.J. and B.G. designed the study. B.W. prepared the data. B.Z., Y.H., C.R., C.Z. and B.G.
- analyzed the data and prepared the figures. B.G. wrote the paper and S.R. revised the paper. All
- 491 authors contributed to discussing the results and writing the manuscript.
- 492

493 Additional information

494 Supplementary information is available for this paper at

- 496 **Correspondence and requests for materials** should be addressed to B.G.
- 497

	Model 1	Model 2	Model 3
	Draft animal	CPLR	Animal land ⁻¹
Ln machinery and fertilizer use (\$ ha ⁻¹)	-0.107***	-0.048***	-0.096***
	(0.002)	(0.002)	(0.007)
Non-agricultural income share (%)	-0.002***	-0.005***	-0.005***
	(0.000)	(0.000)	(0.001)
Ln farm size (ha)	2.010^{***}	1.026***	-
	(0.172)	(0.017)	
Farm size ²	-0.272***	-0.108***	-
	(0.045)	(0.003)	
Year	Yes	Yes	Yes
Province	Yes	Yes	No
N	215,854	211,096	76,609
Pseudo/Adj. R ²	0.1962	0.2209	0.6742
Method	LBS	LBS	FEP

Table 1 Panel model analysis on the decoupling between livestock and cropland

499Robust standard errors (SEs) are in parentheses. *** p < 0.001; LBS, logit binary selection; FEP,500Fixed effect panel. Data year is from 2004 to 2017 due to availability. The detailed501interpretations of variables and models are in *SI Text* and summary statistics are listed in502Supplementary Table 1.

504 Figure Legends

Fig. 1 | Decoupling of livestock and cropland. The top section represents the traditional 505 situation - "Coupled livestock and cropland". Livestock raising provides manure and draft 506 animals for cropland, while cropland provides feed for livestock. Only small amounts of feed 507 and fertilizer are required from import, and pollutant emission are insignificant. The bottom 508 section represents the emerging situation – "Decoupled livestock and cropland". The recycling 509 between livestock and croplands is no longer intact, and large amounts of imported feed and 510 synthetic fertilizers are needed. Substantial amounts of pollutants are emitted to the 511 environment, leading to air and water pollution, biodiversity loss, soil acidification and global 512 warming. 513

514

515 Fig. 2 | Temporal changes of shares of households and draft animals and machinery use.

(a) four types of household shares; (b) livestock raising density in Crop planting and livestock
raising (CPLR) households; (c) machinery use in crop-only and CPLR households; (d) draft
animal share and animal stocking density in all households. NCL, no crop or livestock; Crop only crop planting; Livestock - only livestock raising; CPLR - crop planting and livestock
raising; <15, 15-30, 30-75 and >75 refer to livestock raising density with pig equivalent per ha
cropland in CPLR households. Error bars refer to standard errors (SEs).

- 522
- Fig. 3 | Temporal changes of fertilizer and manure use. (a) application rates of synthetic 523 nitrogen (N) fertilizer in crop-only and Crop planting and livestock raising (CPLR) households; 524 (b) application rates of synthetic N fertilizer in CPLR households with different livestock 525 526 density; (c) manure share of total fertilizer use in crop-only and CPLR households; (d) manure share in CPLR households with different livestock density; NCL - no crop or livestock; Crop -527 only crop planting; Livestock - only livestock raising; <15, 15-30, 30-75 and >75 refer to 528 livestock raising density with pig equivalent per ha cropland in CPLR households. Error bars 529 530 refer to standard errors (SEs).
- 531

Fig. 4 | **Farmland size, animal stocking density and synthetic nitrogen (N) fertilizer use.** (a) farmland size in households with different livestock density; (b) application rate of synthetic N fertilizer and farmland size. The blue bar and points represent the crop-only households while the green bars and points represent the Crop planting and livestock raising (CPLR) households with different animal stocking density. <15, 15-30, 30-75 and >75 refer to livestock raising density with pig equivalent per ha cropland in CPLR households. Error bars refer to standard

538	errors (SEs).	
-----	---------------	--

540	Fig. 5 Spatial variations of share of CPLR households and draft animals in all surveyed
541	villages across China. (a) Crop planting and livestock raising (CPLR) household share in 1986;
542	(b) CPLR household share in 2017; (c) draft animal share in 1986; (d) draft animal share in
543	2017. Base map is adopted from GADM data (https://gadm.org/).
544	
545	Extended Data Fig. 1 Locations of the selected villages of the Fixed Observation Rural
546	Survey (FORS). Base map is adopted from GADM data (https://gadm.org/).
547	
548	Extended Data Fig. 2 Locations of the selected counties of Fixed Observation Rural
549	Survey (FORS). Base map is adopted from GADM data (https://gadm.org/).
550	
551	Extended Data Fig. 3 The daily account of rural household on all their production and
552	consumption activities related to agriculture.
553	
554	Extended Data Fig. 4 The proportion of household and industrial livestock farming
555	systems in 2010s. (a) production proportion; (b) farm number proportion. Fixed Observation
556	Rural Survey (FORS) normally can cover household livestock farms, but not industrial farms
557	which are operated by independent companies. But due to the number of industrial farms is
558	less than 1% of total livestock farms in China, normally not survey but statistical counting of
559	industrial farms is used.
560	
561	Extended Data Fig. 5 Spatial variations of manure and machinery use in all surveyed
562	villages in 2017 across China. (a) manure share in crop-only households; (d) manure share in
563	Crop planting and livestock raising (CPLR) households; (c) machinery use in crop-only
564	households; (d) machinery use in CPLR households. Base map is adopted from GADM data
565	(https://gadm.org/).
566	
567	Extended Data Fig. 6 Temporal changes of fertilizer use in Crop only and CPLR
568	households. (a) application rates of synthetic phosphorus (P) fertilizer; (b) application rates of
569	synthetic P fertilizer in Crop planting and livestock raising (CPLR) households with different
570	livestock density; (c) application rates of synthetic potassium (K) fertilizer; (b) application
571	rates of synthetic K fertilizer in CPLR households with different livestock density. Crop - only 18

- 572 crop planting; Livestock only livestock raising; CPLR crop planting and livestock raising;
- 573 <15, 15-30, 30-75 and >75 refer to livestock raising density with pig equivalent per ha
- 574 cropland in CPLR households. Error bars refer to standard errors (SEs).



De-coupled livestock and cropland system









发生日期	金用 途	金额		
7月3日	All Rock 6004	360.00		
月8日	六(夏·唐·蒙米 700)	49.00		
月/2日	在得小果好雪生 150代.	1820,00		
2月27日	成保禁保养 好、144904.	18550.00		
3月16日	· · · · · · · · · · · · · · · · · · ·	490.00		
4月26日	前生 12 年7天- × 16000 4×1.06	1750.00		
月六日	山達 1日本事 小麦 28004.×118	7024.00		
9月20日	法住 果不好 200 K	68.00		
()月6日	· · · · · · · · · · · · · · · · · · ·	820,00		
11月 (-日)	·信息了富大果 1200年.	1560,00		
вн		E B		

