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Peaucelle, Marc; Bacour, Cédric; Ciais, Philippe; Vuichard, Nicolas; Kuppel, Sylvain; Peñuelas, Josep; Belelli Marchesini, Luca; Blanken, Peter D.; Buchmann, Nina; Chen, Jiquan; Delpierre, Nicolas; Desai, Ankur R.; Dufrene, Eric; Gianelle, Damiano; Gimeno-Colera, Cristina; Gruening, Carsten; Helfter, Carole; Hörtnagl, Lukas; Ibrom, Andreas; Joffre, Richard; Kato, Tomomichi; Kolb, Thomas E.; Law, Beverly; Lindroth, Anders; Mammarella, Ivan; Merbold, Lutz; Minerbi, Stefano; Montagnani, Leonardo; Šigut, Ladislav; Sutton, Mark; Varlagin, Andrej; Vesala, Timo; Wohlfahrt, Georg; Wolf, Sebastian; Yakir, Dan; Viovy, Nicolas. 2019. **Covariations between plant functional traits emerge from constraining parameterization of a terrestrial biosphere model**. *Global Ecology and Biogeography*, 28 (9). 1351-1365, which has been published in final form at <u>https://doi.org/10.1111/geb.12937</u>

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1 Co-variations between plant functional traits emerge from constraining 2 parameterization of a terrestrial biosphere model.

- 3 **Running title:** Functional traits variability inferred from data assimilation
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56 Acknowledgments

This work was performed using HPC resources from GENCI-TGCC (Grant 2017-57 58 A0030106328). The authors would like to acknowledge the financial support from the 59 European Research Council Synergy grant ERC-SyG-2013-610028 IMBALANCE-P. The study was supported by the National Centre of Excellence (272041), ICOS-Finland (281255) 60 and Academy professor project (284701) funded by the Academy of Finland. N.B. 61 62 acknowledges various funding sources for the Swiss FluxNet, particularly from the SNF (grants: 20FI21 148992, 20FI20 173691). L.M. was supported by the Swiss National Science 63 Foundation under the 40FA40_154245/1 grant agreement. This work used eddy covariance 64 data acquired and shared by the FLUXNET community, including these networks: 65 AmeriFlux, AfriFlux, AsiaFlux, CarboAfrica, CarboEuropeIP, CarboItaly, CarboMont, 66 ChinaFlux, Fluxnet-Canada, GreenGrass, ICOS, KoFlux, LBA, NECC, OzFlux-TERN, 67 TCOS-Siberia, and USCCC. The ERA-Interim reanalysis data are provided by ECMWF and 68 69 processed by LSCE. The FLUXNET eddy covariance data processing and harmonization was carried out by the European Fluxes Database Cluster, AmeriFlux Management Project, and 70 Fluxdata project of FLUXNET, with the support of CDIAC and ICOS Ecosystem Thematic 71 72 Center, and the OzFlux, ChinaFlux and AsiaFlux offices. L.B.M. acknowledges the support of 73 the RUDN University program 5-100. L L.S. was supported by the Ministry of Education, Youth and Sports of CR within the National Sustainability Program I (NPU I), grant number 74 75 LO1415.

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79 Co-variations between plant functional traits emerge from constraining 80 parameterization of a terrestrial biosphere model.

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Key words: Plant functional traits, ORCHIDEE, terrestrial model, optimization, data
 assimilation, plant acclimation.

- 84
- 85 Abstract
- 86 Aim

Mechanisms of plant trait adaptation and acclimation are still poorly understood and 87 consequently lack a consistent representation in terrestrial biosphere models (TBMs). Despite 88 the increasing availability of geo-referenced trait observations, current databases are still 89 insufficient to cover all vegetation types and environmental conditions. In parallel, the 90 growing number of continuous eddy-covariance observations of energy and CO₂ fluxes has 91 92 enabled modelers to optimize TBMs with these data. Past attempts to optimize TBMs 93 parameters mostly focused on model performance overlooking the ecological properties of ecosystems. The aim of this study is to assess the ecological consistency of optimized trait-94 related parameters while improving the model performances for gross primary productivity 95 (GPP) at sites. 96

- 97 Location
- 98 World

99 Time period

100 1992-2012

101 Major taxa studied

102 Trees and C3 grasses

103 Methods

We optimized parameters of the ORCHIDEE model against 371 site-years of GPP estimates
from the FLUXNET network and we looked at global co-variation among parameters and
with climate.

107 **Results**

The optimized parameter values are shown to be consistent with leaf-scale traits, in particular well-known trade-offs observed at the leaf level, echoing the leaf economic spectrum theory. Results show a marked sensitivity of trait-related parameters to local bio-climatic variables and reproduce observed relationships between traits and climate.

112 Main conclusions

Our approach validates some biological processes implemented in the model and enables us to study ecological properties of vegetation at the canopy level, as well as some traits that are difficult to observe experimentally. This study stresses the need for 1) implementing explicit trade-offs and acclimation processes in TBMs, 2) improving the representation of processes to avoid model-specific parameterization as well as 3) performing systematic traits measurements at FLUXNET sites in order to gather information on plant ecophysiology and plant diversity, together with micro-meteorological conditions.

120

121 Introduction

Terrestrial biosphere models (TBMs) describe the different processes controlling exchanges of energy and trace gases between the atmosphere and the biosphere. Process-based TBMs are useful tools for understanding the dynamics of ecosystems under changing environment, for present-day to future conditions.

In most TBMs, the worldwide vegetation is divided into plant functional types (PFTs) based on general characteristics of the photosynthetic pathways, phenology, structure and physiology. Different PFTs usually share the same equations but use different parameter values to describe generic processes (photosynthesis, respiration), while biome-specific equations may be used for phenology and allocation. Therefore, for a given PFT, only the differences in climate and soil properties can determine spatial and temporal gradients in ecosystem state variables.

The prescribed values of PFT-specific parameters are derived from discrete observations 133 obtained at varying spatial scales (organs, individuals, ecosystems; Reich et al., 2007; Kattge 134 et al., 2009) and in specific environmental conditions, despite the modulation of real world 135 plant traits by climate (Wright et al., 2005; Ordoñez et al., 2009; van Ommen Kloeke et al., 136 137 2012; Maire et al., 2015) and soil properties (Fisher et al., 2012). In addition, some TBM parameters relate to traits that are difficult to measure experimentally (e.g. root turnovers or 138 139 carbon allocation), or are model-specific. These parameters can hardly be directly optimized 140 from observations and their adjustment to varying environmental conditions can only be determined by labor intensive multi-factorial ecosystem manipulation experiments (Luo et al., 141 142 2017). This rigid determination of parameter values, combined with the use of single PFT to 143 cover a range of different species (Peaucelle et al., 2016), hinders a realistic representation of the past, present and future ecosystem dynamics both at the local or regional scale, and their 144

response to global drivers such as climate, elevated CO_2 and nutrient fertilization (Hartig *et al.*, 2012, Atkin *et al.*, 2015; Kroner & Way, 2016; Reich *et al.*, 2016).

To overcome the rigidity of the PFT representation, various approaches have been proposed 147 to provide continuous distributions of plant functional traits related to model parameters. 148 These approaches range from extrapolating trait observations across spatial gradients 149 (Verheijen et al., 2015), to estimating optimal trait values according to ecological theories and 150 plant-centered approaches (Reu et al., 2011; Pavlick et al., 2013; Prentice et al., 2014). The 151 152 drawback of these different approaches is that they require both spatial and temporal observations for model calibration and/or validation. Despite the increasing number of geo-153 referenced trait observations (Kattge et al., 2011), current databases are insufficient to cover 154 all vegetation types and environmental conditions for projections at the ecosystem level 155 (Musavi et al., 2015, 2016). Moreover, trait observations should be co-located with process 156 and meteorology data to understand linkages between traits and ecosystem function (Law et 157 al., 2008), which is rare in existing databases although increasingly addressed for some 158 biomes (Bjorkman et al., 2018). Long-term monitoring of functional traits is needed to assess 159 the adjustments to climate. As such information is still lacking, approaches have been 160 161 developed that confound the spatial and temporal dimensions of trait variability.

Another modeling strategy consists in optimizing TBMs against observed variables sensitive 162 163 to ecosystem-level parameters in order to overcome these limitations. This approach assumes 164 that the model structure is unbiased, so that realistic parameters values can be estimated when model simulations best match observations. Because biometric variables are sparse and often 165 depend on processes not represented in models (Thum et al., 2017), energy and trace gas flux 166 167 measurements are more appealing to optimize TBM parameters. Eddy-covariance data provide near-continuous observations of CO₂, latent heat and sensible heat fluxes, and are 168 therefore well suited for better constraining photosynthesis, respiration, transpiration and 169

carbon phenology model parameters. Eddy-covariance measurements have been extensively
used to improve specific performances of TBMs, i.e. their ability to reproduce specific
observed ecosystem behaviors (Carvalhais *et al.*, 2010; Kuppel *et al.*, 2012; Santaren *et al.*,
2014; Schürmann *et al.*, 2016). However, such model calibrations are disconnected, by
construction, from ecological theory or trait-based relationships, and do not exploit the full
potential of continuous flux observations across the globe, which provide both spatial and
temporal information.

In this study, we aim at assessing the consistency of model trait parameters optimized against 177 eddy-covariance flux tower measurements of growth primary productivity (GPP) using the 178 state-of-the-art ORCHIDEE land surface model (Krinner et al., 2005). In addition to classical 179 optimization analyses (i.e. looking for the optimal parameter sets resulting in the highest 180 model improvement), we focus here on the variability of optimized parameter values and on 181 inter-traits correlations or trait-environment correlations. By doing so, we address the 182 183 following research questions: 1) Are the parameters retrieved by optimizing the model against flux tower records consistent with known relationships between traits (i.e. trade-offs) ? or 2) 184 between traits and environmental variables? and 3) What new relationship can be identified 185 with this approach? 186

187

188 Methods

189 *The ORCHIDEE model*

The land surface model ORCHIDEE (v1.9.6, without nitrogen cycle) computes biosphere-190 atmosphere exchanges, consistently with water and carbon storage using ordinary differential 191 equations (Krinner et al., 2005) (Figure 1). Given meteorological forcing, plant and soil 192 193 conditions, the model simulates photosynthesis, all components of the surface energy budget and hydrological processes at a half-hourly time step, while the dynamics of carbon storage 194 are calculated daily. In ORCHIDEE, the land surface is discretized into 12 plant functional 195 types (PFT) and bare soil (Table S1.1, Appendix S1). All PFTs share the same equations, but 196 use different parameter values, except for phenology (budburst/senescence), which is PFT-197 specific (Botta et al., 2000). 198

199

200 Eddy-covariance GPP

We used half-hourly flux observations from eddy-covariance sites within the FLUXNET 201 network (https://fluxnet.fluxdata.org). The sites were selected on the basis of spatial 202 homogeneity and the dominance of a vegetation type that could easily be matched to one of 203 the PFTs in ORCHIDEE, excluding crops and C₄ grasses. The vegetation type information at 204 each site was obtained from http://fluxnet.ornl.gov. The list of analyzed FLUXNET sites (98 205 sites, 371 site-year) and the corresponding PFTs is given in Appendix S2. The following 206 analyses rely on GPP derived from net ecosystem exchange (NEE; reference with variable 207 USTAR threshold) after accounting for ecosystem respiration calculated using the method of 208 Reichstein et al., (2005) provided in the FLUXNET dataset. Years with less than 80% of 209 available half hourly observations were discarded. 210

211

Because ORCHIDEE needs continuous half-hourly meteorological forcing, we gap-filled time series of weather variables using the interpolation algorithm developed by Vuichard & Papale (2015). Linear interpolation was applied between available observations when the gapduration in the meteorological data was less than six hours. Otherwise, the variables were interpolated and bias corrected using the ERA-interim reanalysis (~80km, Dee *et al.*, 2011). Snow and rain were identified according to air temperature (threshold for snow being 0°C).

219

220 *Data assimilation procedure*

The parameters of ORCHIDEE were optimized with the ORCHIDAS package developed by: Kuppel *et al.*, (2012); Bacour *et al.*, (2015); MacBean *et al.*, (2015) and Peylin *et al.*, (2016); (https://orchidas.lsce.ipsl.fr/; Figure 1). Gaussian distributions of parameter and observation errors being assumed, a gradient-based approach was used to minimize the Bayesian cost function *J* (Tarantola, 2005):

$$\mathbf{J}(\mathbf{x}) = \frac{1}{2} \left[\left(\mathbf{y} - H(\mathbf{x}) \right)^{\mathrm{T}} \mathbf{R}^{-1} \left(\mathbf{y} - H(\mathbf{x}) \right) + (\mathbf{x} - \mathbf{x}_{\mathbf{b}})^{\mathrm{T}} \mathbf{B}^{-1} (\mathbf{x} - \mathbf{x}_{\mathbf{b}}) \right]$$
(1)

This function quantifies the difference between observations (y) and simulations (H(x)) (here GPP), and between *a priori* (x_b) and optimized parameters (x). The *B* and *R* matrices are the prior error covariance matrices for parameters and observations, respectively (including in the latter case eddy-covariance measurement and model errors).

Both *R* and *B* were taken as diagonal, as discussed in Kuppel *et al.* (2012). The J(x) function was iteratively minimized with the L-BFGS-B algorithm (Byrd *et al.*, 1995), which notably allows bounding the range of variation of the parameters to optimize. After model calibration (i.e. minimizing *J*), the posterior error covariance matrix (*A*), providing the full statistical distribution of the optimized parameters was estimated by: $A = [H^{T} R^{-1}H + B^{-1}] - 1$ (2) where *H* is the Jacobian of model at the minimum of *J* (Tarantola, 2005). The covariances of errors between parameters contained in the non-diagonal terms of *A* inform about the ability of observations given the structure of *H* to solve for parameters individually, or in combination. High error covariance between two parameters relates to the equifinality problem, whereby different values of these parameters result in model outputs equally matching the observations (relative to *R*).

241

242 *Optimized parameters*

We restricted our exercise to the parameters involved in the assimilation of CO₂ following 243 previous sensitivity analyses from Kuppel (2012). We analyzed 14 parameters controlling 244 long-term and inter-annual GPP variability (Table 1). The key equations involving each 245 optimized parameter as well as their effect on the simulated GPP are described in Table S1.2 246 (Appendix S1). The parameters were related to photosynthetic capacity, phenology, carbon 247 allocation and the water budget. Photosynthetic capacity parameters were the maximal rate of 248 249 carboxylation limited by CO_2 (*Vcmax*), the ratio between the maximal rate of carboxylation limited by light and Vcmax (Vi/Vc), the optimal temperature of photosynthesis (Topt) and the 250 slope of the Ball-Berry model for stomatal conductance (gslope). Parameters driving 251 phenology were the specific leaf area (SLA), leaf longevity (Lage), summer maximal leaf area 252 index (LAImax) and the temperature for leaf senescence (Csenes). Allocation parameters were 253 the minimal fraction of LAImax for the use of carbohydrate reserves (Klai) and the period 254 after budburst during which the use of carbohydrates is allowed (*tauleaf*) for the formation of 255 new leaves. Finally, two parameters involved in the water status of the plant were the 256 exponential factor describing the root profile and length (Kroot) and the minimal threshold at 257 which photosynthesis becomes limited by minimum water potential (Wlim). In addition, two 258

scaling factor *Kbm* (initial biomass of leaves for evergreen species) and *bbdate* (spring
burdburst date) were added in the optimization to allow adjusting the seasonal timing of GPP.

The range in variation of the three parameters corresponding to observable traits (SLA, Vcmax 261 and Lage) was set from the TRY database for each PFT (Niinemets et al., 1999; Deng et al., 262 2004; Meir et al., 2007; Kattge et al., 2009, 2011; Domingues et al., 2010; Cernusak et al., 263 2011; Azevedo & Marenco, 2012; van de Weg et al., 2012; Nascimento & Marenco, 2013). 264 Species from the TRY database were assigned to corresponding PFTs based on available 265 metadata about plant structure, leaf phenology and climate information extracted from 266 species' latitude and longitude coordinates. We chose as a reference range the 2.5 - 97.5 267 percentile of the trait distributions from TRY. The variation ranges for the other parameters 268 were fixed based on expert judgment (Kuppel et al., 2014). 269

270

271 Simulations and assimilation set-up

At each flux tower site, we assumed that the eddy-covariance flux footprint was entirely 272 composed by a single PFT (Appendix S2). The model was forced by local meteorological 273 observations (see Meteorological data section) and soil texture from the harmonized 274 worldwide soil database (Nachtergaele et al., 2012) to define the residual and saturation water 275 contents, and the saturated hydraulic conductivity in the soil model (Ducoudré et al., 1993; 276 277 Krinner et al., 2005) based on Van Genuchten (1980). Initial soil carbon pools in equilibrium with local climate were obtained with an analytical spin-up procedure (Lardy *et al.*, 2011; Xia 278 et al., 2012). Initial biomass was simulated until reaching equilibrium (generally after a ~300 279 280 year-long simulations using the studied year meteorological data and constant CO₂ set to level of the year), thus different from the real stand age observed at each site. 281

282

We optimized GPP averaged over 15 days using moving windows to avoid noise from high 283 frequency variations in the parameter optimization that could induce convergence issues 284 (Bacour et al., 2015). As far as test data from eddy-covariance measurements are concerned, 285 high frequency variations in fluxes include also variation in the boundary layer that are 286 unrelated to the fluxes at the surface (Ibrom et al., 2006). Santaren et al. (2007) estimated that 287 for parameters related to photosynthesis and phenology, optimization based on half-hourly 288 observations did not improve the results. For each site, the optimizations were conducted 289 year-by-year to account for trait variability over time (Wu et al., 2013). 290

Following MacBean *et al.* (2015), each calibration (site-year) used ten replicates representing different starting parameter sets with values randomly picked within their allowed variation range (Table S1.3). Only the best calibration out of these ten replicates was retained for analyses. This procedure increases the chances of finding the global minimum of J as Santaren *et al.* (2014) showed that the gradient-based algorithm was sensitive to initial conditions with a non-linear and complex model such as ORCHIDEE.

297

298 Analyses

We only retained calibrations for which the optimized model reproduced GPP observations with high precision. The rationale for this was that optimized parameters from model runs which agreed poorly with GPP observations provided little or no useable information. The filtering was performed using a two-step procedure.

First, the criterion for 'improved GPP simulation' was the relative site-year posterior RMSE
(RMSE_{re}) between observed and optimized GPP:

$$RMSE_{re} = \frac{RMSE}{mean(GPP_{obs})} \tag{4}$$

Whenever the value of $RMSE_{re}$ was higher than the all- $RMSE_{re}$ median plus one interquartile range (IQR), the site-year was removed from the analysis. We also discarded sites with 'inconsistent parameters values', i.e. with too large differences between the ten replicates at the same site reflecting convergence issues (equifinality) of the algorithm.

Secondly, for sites with at least two $RMSE_{re}$ below 10 % among the ten replicates, we estimated the coefficient of variation (CV) of parameters across the replicates. We retained only years for which the median CV was below the median of all CV plus one IQR of their distribution. This filtering provided optimized parameters from 371 site-years (over 516 initially considered) for 98 sites (over 116; Appendix S2) spanning seven PFTs located in boreal, temperate and tropical areas (Table S3.4; Appendix S3).

315

For each parameter, we calculated the uncertainty reduction (UR) as:

$$UR = 1 - \frac{\sigma_{post}}{\sigma_{prior}} \tag{5}$$

With σ_{post} and σ_{prior} being the *posterior* and *prior* parameter uncertainties (square root of the diagonal elements of *A* and *B*). We then separated in the analysis the well- from the poorlyconstrained parameters. Well-constrained parameters are defined as those with 1) *UR* higher than the median of *UR* distributions for all parameters and 2) a low correlation of error with other parameters (from the A matrix, Eq. 2). Note that a strong error correlation making two parameters poorly constrained individually is still an interesting result as it indicates a range of possible tradeoffs between these two parameters.

The optimized parameter values were regressed against the local background bio-climatic variables (Table 2) for each site, and against the soil relative water content (volume of water by volume of soil) simulated by ORCHIDEE. Bio-climatic variables were averaged over the whole year and over the length of the growing season (GSL). For temperate sites, the growing

season was defined as the period with daily temperature above 5°C and relative soil water 328 content above 0.2 (Violle et al., 2015). In some tropical regions, the growing season length is 329 potentially limited by water availability (wet/dry seasons), we thus kept the same definition as 330 for temperate ecosystems. For boreal sites, we adapted the definition of the growing season 331 such as weekly temperature must be above 0°C. Analyses were performed with the R.3.2 332 software (R Core Team, 2016) and standardized major axis (SMA) analyses were performed 333 with the 'Imodel2' package (Legendre, 2014). Because we sought to compare simulated 334 335 correlations with common ecological properties observed at the global scale, we analyzed different groups of PFTs: all PFTs together; deciduous versus evergreens; needleleaves versus 336 broadleaves; and C₃ grasses (Table S1.1). Regressions were performed both with and without 337 a logarithmic transformation of the data. 338

339 Results and comparison to existing literature

- 340 *Optimization performances*
- A full description of the optimization performances and parameter uncertainty reduction canbe found in Appendix S3.
- In all cases, the optimized GPP time series better agrees with observations than the prior ones, with the RMSE being reduced by 76.6 ± 13.0 % (Table S3.4; Appendix S3). The median posterior RMSE_{re} is 0.19 and the IQR is 0.11. The median CV over all parameters is 0.24 (IQR=0.13). After optimization, the parameter uncertainty (Eq. 5) is reduced by 30 % on average (Table S3.5; Appendix S3).
- The posterior error correlation matrix *A* (Eq.2) reveals a positive correlation between *Vcmax* and several other parameters including (Figure 2): *Topt* (r= 0.57 ± 0.05); *gslope* (r= -0.37 ± 0.04); *Kroot* (r= 0.24 ± 0.07) and *Vj/Vc* (r= -0.31 ± 0.04). There also exists a negative correlation between *Kroot* and *gslope* (r= -0.38 ± 0.08), between *Kroot* and *Wlim* (r= -0.30 ± 0.09) and between *LAImax* and *Klai* (r= -0.37 ± 0.16) (Figure 2).
- Jointly analyzing information from the uncertainty reduction (Appendix S3) and the crossparameter error correlation enables to distinguish between: 1) well constrained parameters (*Lage* and *SLA* for evergreens/ *Lage* and *Csenes* for deciduous); 2) well constrained parameters with a risk of equifinality (*gslope, Kroot, LAImax, Topt* and *Vcmax*); and 3) poorly constrained parameters (*Vj/Vc, Klai, Tauleaf* and *Wlim;* Table 1). In the following analyses, trait co-variations have to be interpreted in respect to confidence intervals (posterior error) in parameter estimates.

360 *Co-variation between parameters*

We analyzed cross-site correlations between optimized parameters in relation to expected trait
 relationships. The co-variation between all parameters is illustrated in Figure S4.2 (Appendix

S4). For more clarity and considering the large number of parameters, we only describe here the relationships involving four parameters related to phenology (*SLA*, *Lage*) and photosynthesis (*Vcmax, gslope*). All relationships are provided in Table S4.6 (Appendix S4).

We observed a negative correlation between SLA and Lage for all PFTs (r=-0.63; Table 3) as 366 well as for evergreens (r=-0.67) and broadleaves PFTs (r=-0.53), separately. The slope of the 367 emerging relationship between LMA (1/SLA) and Lage (1.91; 1.63-2.24 95% confidence 368 interval; p < 0.05) for all PFTs was close to the observed slope from field observations (1.71; 369 1.62-1.82; Wright et al. 2004). Results highlighted other co-variations between Lage and 370 Vcmax (r=-0.59 overall PFTs), gslope and Lage (r=-0.7 for broadleaves), LAImax and SLA 371 (r=0.6 for needleleaves), and between SLA and Vcmax (r=-0.55 for evergreens). Here again, 372 373 the slope between Lage and Vcmax emerging for broadleaves PFTs (-1.69) was close to observations (-1.13; Xu et al. 2017). 374

No relationships were reported between *gslope* and *Lage* or between *glsope* and *SLA*, but a trade-off between the stomatal conductance (*gs*) and *Lage* was observed experimentally (Reich *et al.*, 1992; Poorter & Bongers, 2006), as well as a positive correlation between *gs* and *SLA* (Poorter & Bongers, 2006). The optimizations showed opposite relationships between *gslope* and *SLA* depending on the PFT: a positive significant correlation was obtained for deciduous PFTs and a negative significant correlation for evergreens and grasses (Table 3).

The positive relationship between *SLA* and *LAImax* emerging from optimized parameters for coniferous PFTs was consistent with the positive correlation between *LAI* and *SLA* reported by Pierce *et al.* (1994) for coniferous forests. Finally, a negative correlation between *SLA* and *Vcmax* has been observed experimentally for two gymnosperms species (Niinemets *et al.*, 2007), confirming the negative relationships found in our study for needleleaves. Despite the equifinality risk between *gslope* and the soil water stress *Wlim* in Figure 2, the positive correlation observed for broadleaves (r=0.7) and evergreens (r=0.52) was comparable to observations from independent data compiled by Lin *et al.* (2015).

Other significant correlations from the optimized parameters (Table S4.6, Figure S4.2; Appendix S4) could not be verified against observations because of the correlation of errors observed in Figure 2 or because of the scarcity of ecological data preventing us to conclude about the true nature of those correlations, as for example between *gslope* and *Vcmax*.

394

395 Variation of trait-related parameters with climate

We analyzed correlations between parameters and climate variables (Table 4, Figure S5.4; Appendix S5). As for co-variations between parameters, we only described here those implying *SLA*, *Lage*, *Vcmax* and *gslope*. All relationships are listed in Table S5.7 and more detailed analysis are available in Appendix S5.

400 We found a strong negative correlation between leaf lifespan (Lage) and temperatures (MAT, *TMIN*; r=-0.78/-0.65; Figure 3a) for evergreen PFTs. This correlation was independently 401 reported at global scale (Wright et al., 2005; van Ommen Kloeke et al., 2012) and confirmed 402 by Reich et al. (2014) who showed higher needle longevity with cold temperatures for boreal 403 species. However, the observed positive correlation between Lage and MAT at the global 404 405 scale for deciduous PFTs (Wright et al., 2005; van Ommen Kloeke et al., 2012) was not found specifically for deciduous systems in our study. Nevertheless, a positive correlation was 406 observed for C3 grasses and broadleaves (including deciduous). We also found a strong 407 408 negative correlation between Lage and the mean annual precipitations (MAP) for evergreens PFTs (r = -0.65), consistent with field data (van Ommen Kloeke *et al.*, 2012). In addition, a 409

410 negative correlation between *Lage* and incident shortwave radiation (*SW*) for evergreens was
411 obtained, consistent with field observations (Poorter & Bongers, 2006).

412

Regarding SLA, we found opposite sensitivities to MAT for evergreen (r=0.65) and deciduous 413 forests (r=-0.55). This result is consistent with independent leaf-scale data showing a positive 414 correlation between SLA and MAT for evergreen species (Figure 3b) and a negative 415 correlation for deciduous ones (Wright et al. 2005). The model calibration also resulted in a 416 positive correlation between the relative precipitation (RELP; Table 2) and SLA for deciduous 417 trees (r = 0.60; Figure 3c). Regarding the positive correlations obtained between SLA with 418 419 Kroot or gslope (Table 3), it suggests that SLA is highly sensitive to water stress for deciduous trees. For evergreens, positive correlation between SLA and precipitation also emerges when 420 considering the length of the growing season (*MAPgs*, r = 0.57; Table 4); which is consistent 421 422 with trait data (Wright et al., 2005). For evergreens, SLA was positively correlated to SW (r=0.53), a relationship observed by Givnish et al. (2004) and Poorter & Bongers (2006). 423

424

425 In their meta-analysis of stomatal conductance parameters from observations of several PFTs, Lin et al. (2015) showed that the slope of the stomatal conductance is positively correlated to 426 427 the mean air temperature over the growing period and to soil moisture stress. Here, our results show the same correlation between gslope and soil moisture during the growing season 428 (r=0.71; Figure 3d) and relative precipitation (r=0.66) for deciduous or broadleaved PFTs. On 429 the contrary, we find that *gslope* is negatively correlated with mean annual precipitation for 430 C_3 grasses (r=-0.59), and with shortwave radiation for broadleaved PFTs (r=-0.63). Medlyn et 431 al. (2011) suggested that gslope is proportional to the photosynthesis compensation point for 432 CO₂, and consequently to growth temperatures of the plant (Bernacchi et al., 2001). This 433

assumption is supported by the data from Lin *et al.* (2015). In our study, the relationship
between *gslope* and temperature was not supported.

436

Finally, *Vcmax* is mostly sensitive to temperature and light for broadleaved PFTs, with a negative correlation observed with *MAT* (r=-0.52) and *SW* (r=-0.54). This result contradicts previous observations by Ali *et al.* (2015), who suggested a positive correlation between *Vcmax* and seasonal temperature and light variations.

441

442 **Discussion**

443 Uncertainties and shortcomings of the approach

This section provides an overview of possible shortcomings of our approach that may explain some residual mismatch between model and observations. Several factors can impact the optimized value of the parameters, potentially aliasing the observed relationships: 1) flux measurements errors and errors in ecosystem respiration estimates used to derive gap-filled GPP; 2) optimization protocol/setup errors; and 3) model systematic errors deriving from absent or poorly represented processes in the model.

First, we restricted our analysis to GPP. This flux is not directly measured but estimated from 450 NEE measured using the eddy-covariance method with an estimate of ecosystem respiration 451 determined using empirical models (Reichstein et al., 2005), and thus can be biased by 452 several factors (see Appendix S3 for a list of these factors). We chose GPP over a 453 454 combination of NEE and latent heat or evapotranspiration fluxes, which has often been used to optimize ORCHIDEE (Kuppel et al., 2012; Bacour et al., 2015; Peylin et al., 2016), 455 because it implies the optimization of more parameters related to soil, respiration and energy 456 457 budget, and therefore increases the risk of equifinality. To reduce the uncertainties, it is

458 necessary to lower the correlation of errors between parameters by assimilating 459 complementary biophysical variables. For example, assimilating both GPP and LAI estimates 460 at the site level could improve the evaluation of parameters such as *SLA* or *Lage*, and 461 consequently improve the estimation of photosynthesis parameters.

Second, the Bayesian framework is based on the assumption that the model/observation errors 462 are random and that the model structure is "true". Any bias of model structure is expected to 463 be aliased onto the estimated parameters (MacBean et al., 2016) and might therefore impact 464 465 the retrieved correlations. For instance, missing processes would be compensated during the optimization by adjusting parameters (e.g. light attenuation, vertical distribution of LAI, 466 diffuse light, horizontal light distribution in the stand) to non optimal values. Also, while 467 traits are usually measured at the leaf level, our approach rather focuses on traits at the canopy 468 level (given the structure of ORCHIDEE and the assumed exponential attenuation of light and 469 LAI from top to bottom of canopy (Krinner et al. 2005; Table S1.2), and the assimilation of 470 471 GPP data). As an additional test, we conducted the above analyses using multi-year instead of single-year observations in order to add more constraints on parameters (see Figure S4.3 & 472 S5.5). The same relationships were found as with single-year observations, thus strengthening 473 474 our conclusions, showing that spatial correlations are observed even when taking into account a possible temporal variability of traits. 475

Finally, a wrong representation of species and the lack of representation of representation of traits variability within a community in ORCHIDEE can affect simulated processes, which will ultimately impact the estimated parameter values (see Appendix S3 for a discussion on initial site conditions). Especially, the C_3 grass PFT represents diverse grasslands, with different species, ecophysiology (Adams *et al.*, 2016) and management practices (Merbold *et al.*, 2014). This results in an increased variability and a high range of estimated plant functional traits (Figure S3.1). A refinement of the PFT definition may improve the robustness of optimizations (for instance by separating natural or semi-managed biomes), or
by distinguishing genera or major species (Peaucelle *et al.*, 2016).

In order to decrease the impact of uncertainty in PFT composition and reduce the correlation errors between parameters, the use of concomitant observations of traits and carbon fluxes at the FLUXNET sites would enable a) to constrain known parameters and b) to validate optimized traits. However, functional trait observations at FLUXNET sites as well as a precise description of species composition are not yet systematic (Musavi *et al.*, 2015, 2016).

490

491 *Ecological consistency of trait relationships*

The optimization of model parameters managed to reproduce many known ecological properties. The optimized parameters consistently matched the well-known relationships resulting from the leaf economic spectrum theory (LES, Reich *et al.*, 1999; Wright *et al.*, 2004). Particularly our results align with the trait theory that long lived canopies are metabolically less active and are consistent with the LES empirical evidence that plants invest either in structure or photosynthesis (Liu *et al.*, 2010; Reich, 2014).

Our results also reproduced several observed trait-climate relationships at the PFT level. 498 Globally, evergreen PFT parameters showed a strong dependency on mean annual 499 temperature and radiation, while parameters for deciduous PFTs exhibited a strong sensitivity 500 to precipitation and soil moisture over the growing season (Figure S5.4). As postulated by 501 Reich (2014), climate exerts a control on the average leaf characteristics at the community 502 level. The observed relationships obtained at the PFT level might reflect, not only differences 503 in plant response to climate, but also differences in plant community composition (Shi et al., 504 2015). These results suggest that both the development of acclimation processes and trait-505

based approaches are needed in TBMs if we seek to capture the effect of biogeography on
ecosystem characteristics (Lu *et al.*, 2017; Fisher *et al.*, 2018).

508

Finally, while the results clearly highlight that photosynthesis and phenological mechanisms implemented in ORCHIDEE are robust enough to reproduce known behaviors of several vegetation species, belowground processes still appear poorly represented, which resulted in weakly constrained parameters and trait co-variations inconsistent with literature. These discrepancies are primarily due to a lack in eco-physiological knowledge reflecting the actual difficulty to study belowground ecological processes. The rooting system uses model-specific parameters (*Kroot*) that are hardly comparable to measured functional traits.

516

517 Concluding remarks and recommendations

The approach presented in this study is a new and effective way to validate the processes implemented in TBMs, to better define vegetation response to climate (Liang *et al.*, others, 2018), and could help improving existing data assimilation frameworks (Kaminski *et al.*, 2013; LeBauer *et al.*, 2013; Arsenault *et al.*, 2018) by bringing ecological constraints. The availability of continuous observations from eddy-covariance flux measurements gives a unique opportunity to resolve the different components of the short and long-term variability of traits through this approach.

525 Our results show that optimized leaf-related parameters align with plant trait theory, and 526 highlight the need to implement acclimation processes and trait-based approaches in models 527 instead of using constant parameters to reduce uncertainties in spatio-temporal patterns of the 528 modeled carbon fluxes. A first step would be to assess the behavior of the model at the global 529 scale when trait-climate relationships characterized in this study are explicitly implemented. 530 In parallel, relationships highlighted in this study may help to develop or validate new 531 methods to simulate plant acclimation. Used in a prognostic way, this approach could enable 532 to study correlations at the canopy scale and to assess the behavior of trait-related parameters 533 that are difficult to observe experimentally.

Several known ecological properties, observed at the site/leaf scale, emerged from model-data 534 assimilation. However, quantitative comparisons with observations were possible only for two 535 of them, SLA and Lage, which are also the two most studied traits. This is mainly because 536 TBMs use model-specific parameters that cannot be directly compared to standard trait 537 observations, but also because concomitant observations of functional traits, both in space and 538 time, are scarce in the literature. A recommendation to the TBM community would be to 539 540 make use of parameters (and processes) that can be related directly to observations in order to unit vegetation model and functional traits (for instance the use of the Specific Root Length 541 for belowground processes). 542

We argue that co-located systematic and standardized trait observations (starting with key 543 traits related to phenology -SLA, LAI-, photosynthesis -Vcmax, Jmax, Topt-, water transport -544 gs- and allocation -Carbon:Nitrogen ratio, shoot/root-; (Law et al., 2008) along with 545 biometric data are needed at the FLUXNET sites or within other environmental observation 546 networks such as ICOS (Integrated Carbon Observation System) or NEON (National 547 Ecological Observatory Network) if we seek to distinguish temporal and spatial components 548 549 of trait variability across biomes and climates. The creation of a FLUXNET trait database could improve our comprehension of trait acclimation and help us to disentangle the 550 differences observed at regional and local scales, to improve the up-scaling of processes from 551 552 the leaf to the canopy/ecosystem level and to properly calibrate/validate ecosystem models.

Supporting information 553 Appendix S1: Description of PFTs, model parameters and equations 554 Table S1.1: List of plant functional types. 555 Table S1.2: List of main equations involving optimized parameters. 556 Table S1.3: Default parameter value and range allowed by the optimization algorithm. 557 Appendix S2: List of FLUXNET sites used for the analyses (xlsx file). 558 559 **Appendix S3**: Optimization performances Figure S3.1: Distribution of optimized parameter values. 560 Table S3.4: Mean *a priori* and *a posteriori* RMSE between observations and simulations. 561 562 Table S3.5: Mean parameter uncertainty reduction between *prior* and *posterior* simulations. **Appendix S4**: Relationships between traits 563 564 Figure S4.2: Correlation matrices between traits optimized against site-year GPP. Figure S4.3: Correlation matrices between traits optimized against site GPP. 565 Table S4.6: Extended Table 3 with all relationships 566

- 567 Appendix S5: Relationships between traits and climate
- 568Figure S5.4: Correlation matrices between traits and environmental variables optimized569against site-year GPP
- 570Figure S5.5: Correlation matrices between traits and environmental variables optimized571against site GPP
- 572 Table S5.7: Extended Table 4 with all relationships
- 573

574 Data accessibility

- 575 All FluxNet data can be downloaded at: https://fluxnet.fluxdata.org
- 576 Information about the ORCHIDEE model, source code and contact: http://orchidee.ipsl.fr/
- 577 Information about the data assimilation system ORCHIDAS: https://orchidas.lsce.ipsl.fr/
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Table 1: Description of the 14 optimized parameters and associated processes. All the parameters are common to each PFT. Kbm and bbdate are scaling factor added in the model to improve the optimization of the seasonal cycle of the GPP but are not analyzed in the study. (See Table S1.2; Appendix 1 for the detailed equations involving each parameter).

Parameter	Description	Processes involved
SLA	Specific leaf area ($m^2 gC^{-1}$)	Photosynthesis, Phenology, Allocation
Lage	Leaf lifespan (days)	Photosynthesis, Phenology
Vcmax	Maximal carboxylation rate limited by CO_2 (µmol m ⁻² s ⁻¹)	Photosynthesis
Vj/Vc	Ratio between the maximal carboxylation rate limited by light and Vcmax	Photosynthesis
Topt	Optimal temperature of the photosynthesis (°C)	Photosynthesis
gslope	Slope of the Ball-Berry relationship for the stomatal conductance	Photosynthesis, energy budget
LAImax	Maximal leaf area index	Photosynthesis, Phenology, Allocation
Klai	Minimal fraction of LAImax for the use of carbohydrate reserves	Allocation
bbdate	Budburst date (day of the year)	Phenology
tauleaf	Period after budburst during which the use of carbohydrates is allowed	Allocation
Csenes	Temperature for leaf senescence (used only for deciduous)	Phenology
Kbm	Multiplicative factor for the initial leaf biomass (used only for evergreens)	Phenology, Allocation
Kroot	Exponential factor describing the root profile and depth	Water budget, Photosynthesis
Wlim	Minimal threshold at which the photosynthesis becomes limited by water availability	Photosynthesis

Variable	Description	Units
LAT	Latitude	°N
MAT	Mean annual temperature	Celsius
TMAX	Mean temperature of the warmest month of the year	Celsius
TMIN	Mean temperature of the coldest month of the year	Celsius
TVAR	Temperature difference between TMAX and TMIN	Celsius
DTR	Yearly average of diurnal temperature range	Celsius
MAP	Mean annual precipitation	mm y ⁻¹
REH	Mean annual relative humidity	%
SW	Mean annual downward shortwaves radiation (visible and near IR)	$W m^{-2}$
PDRY	The driest quarter of the year is determined (to the nearest week), and the total precipitation over this period is calculated.	mm y ⁻¹
RELP	PDRY divided by MAP	fraction
SHUM	Yearly averaged soil humidity	fraction
GSL	MATgs, DTRgs, SWgs, MAPgs, REHgs and SHUMgs are the same variables averaged during the growing season of the plant	-

Table 2: Description of bio-climatic variables calculated at each site and for each year.

Table 3 : Relationships between trait-related parameters. For some relationships, values are logtransformed (x). For each relationship is the number of sites, as well as the correlation coefficient (r, blue when negative; red when positive). Only relationships with an absolute and significant (pvalue<0.05) correlation coefficient higher than 0.5 are listed for the different groups of PFT : all, broadleaves (bro ; TroEB, TemEB, TDB, BDB), needleleaves (need ; TEN, BEN), evergreens (ever ; TroEB, TemEB, TEN, BEN), deciduous (dec ; TDB, BDB) and C3 grasses (gra). Note that evergreens include needleleaves and that broadleaves include deciduous. The type of relationship is given for each trait: 0= verified with ecological observations; 1=partially verified on similar data or 3= different from observations. When available, the reference for verification is given. Well constrained parameters are in bold, parameters with a risk of equifinality are normal, poorly constrained parameters are in italics. Refer to Table 1 for the description of each parameter.

Parameters		r	PFT	log	SMA slope	n sites	references	Туре
	SLA	-0.67	ever	Х	-1.39	49	(Reich <i>et al.</i> , 1999); (Wright <i>et al.</i> , 2004)	0
Lage		-0.53	bro	х	-3.47	37		0
		-0.63	All	х	-1.92	98		0
	Vcmax	-0.90	Bro	х	-1.69	37	(Xu <i>et al.</i> , 2017)	0
Lage		-0.65	Dec		-2.15	23		0
		-0.59	All	х	-3.13	98		0
gelone	Lage	-0.70	Bro	х	-0.74	37	(Reich et al., 1992)	1
gstope		-0.57	Grass		0.00	26	(Poorter & Bongers, 2006)	1
	SLA	-0.62	Ever		-534.01	49	(Poorter & Bongers, 2006)	3
gslope		0.52	Dec		418.99	23		1
		-0.51	Grass		-235.65	26		3
LAImax	SLA	0.60	Need		422.11	35	(Pierce et al., 1994)	1
CT A	Vcmax	-0.55	Ever	х	-1.28	49		1
SLA		-0.53	Need	х	-0.75	35	(Infinemets <i>et al.</i> , 2007)	1
	Wlim	0.70	Bro	х	1.61	37	(I.) (J. 2015)	3
gsiope		0.52	Ever	х	1.47	49	(Lin <i>et al.</i> , 2015)	3

Table 4: Relationships between trait-related parameters and climate variables. For some relationships, traits values are log-transformed (x). For each relationship is given the correlation coefficient (r). Only relationships with an absolute (and significant p-value<0.05) correlation coefficient higher than 0.5 are listed for the different groups of PFT : all, broadleaves (bro; TroEB, TemEB, TDB, BDB), needleleaves (need; TEN, BEN), evergreens (ever; TroEB, TemEB, TEN, BEN), deciduous (dec; TDB, BDB) and C3 grasses (gra). The type of relationship is given for each trait: 0= verified with ecological observations; 1=partially verified on similar data; 2= not verified or, 3=different from observations. When available, the reference for verification is given. Well constrained parameters are in bold, parameters with a risk of equifinality are normal. Refers to Table 1 and 2 for the description of each parameter and climate variables respectively.

Trait	Climat	r	PFT	log	SMA slope	Référence	Туре
	LAT	0.59	ever		24.90	(Reich et al., 2014)	0
		-0.56	bro		-13.44	-	2
	MAP	0.66	grass		1.14	-	2
		-0.65	need	х	-0.66	(van Ommen Kloeke et al., 2012)	0
	MAT	-0.78	ever	х	-16.95	(Wright et al., 2005; van Ommen Kloeke et al., 2012;	0
Гаде		-0.62	need	х	-17.93	Reich <i>et al.</i> , 2014)	0
Luge		0.54	grass	х	107.81	-	2
		0.53	bro	Х	30.14	-	2
	SW	-0.53	ever	х	-1.84	(Poorter & Bongers, 2006)	1
	511	0.52	bro	Х	3.85	-	2
	TMIN	-0.65	ever	Х	-30.99	(Wright <i>et al.</i> , 2005; van Ommen Kloeke <i>et al.</i> , 2012; Reich <i>et al.</i> , 2014)	
	MAP	0.54	need	х	0.37		0
	MAPgs	0.57	ever	х	0.47	(wfight <i>et al.</i> , 2005)	0
	MAT	0.65	ever	Х	12.16		0
	MATgs	-0.63	bro	х	-0.86	(Wright <i>et al.</i> , 2005)	0
SLA		-0.55	dec	х	-0.96		0
	RELP	0.60	dec	х	0.25	-	2
		0.59	bro		0.08	_	2
	SW	0.53	ever		0.00	(Givnish <i>et al.</i> , 2004; Poorter & Bongers, 2006; Reich <i>et al.</i> , 2014)	1
	MAP	-0.59	grass	х	-1.12	-	2
	PDRY	0.58	dec		0.02		1
	REH	0.64	dec		19.24		1
gelope	DELD	0.66	bro		42.67	(Lin et al., 2015)	1
gsiope	KELI	0.58	dec		29.05		1
	SHUMgs	0.71	dec		20.53		1
	SW	-0.63	bro		-0.10	-	2
	SWgs	-0.55	dec		-0.08	<u> </u>	2
	MAT	-0.52	bro		-4.77	(Ali <i>et al.</i> , 2015)	3
Vcmax	RELP	0.60	bro		511.72	-	2
	SW	-0.54	bro		-1.15	(Ali <i>et al.</i> , 2015)	3



Figure 1: Schematic representation of the modeling protocol followed in this study. For each FLUXNET site-year (blue), the model ORCHIDEE (green) was calibrated with the data assimilation system ORCHIDAS (red) in order to reproduce GPP observations. The ORCHIDAS system uses a gradient-based approach (L-BFGS-B) to reduce the cost function J(x). For each site-year, 14 parameters (listed in Table 1) were optimized 10 times with different initial values. The best calibration, i.e. leading to the minimum value of J(x), was retained. This procedure was repeated for each site-year, resulting in 371 sets of 14 independently optimized parameters. Finally, correlations between optimized parameters and climate were explored using standardized major axis regressions.



Figure 2: Error correlation between optimized parameters (derived from the A matrix) averaged over deciduous trees, evergreen trees, and C3 grass. The color scale gives the error correlation coefficient. For more clarity, the coefficient is indicated in % in each matrix cell. The description of each parameter is listed in Table 1.



Figure 3: Four examples of co-variations obtained between optimized parameters (Table 1) and environmental conditions (Table 2) of the sites for PFTs TroEB (black square), TEN (red square), TemEB (green triangle), TDB (blue square), BEN (cyan dots) and BDB (pink dots). Each point represents the mean optimized parameter (environmental variable) value for one site while the error bars represent the inter-annual variability (no bars means only one year of measurement). The red line represents the slope of the standardized major axis regression.