

Article (refereed) - postprint

O'Hare, Matthew T.; Aguiar, Francisca C.; Asaeda, Takashi; Bakker, Elisabeth S.; Chambers, Patricia A.; Clayton, John S.; Elger, Arnaud; Ferreira, Teresa M.; Gross, Elisabeth M.; Gunn, Iain D.M.; Gurnell, Angela M.; Hellsten, Seppo; Hofstra, Deborah E.; Li, Wei; Mohr, Silvia; Puijalon, Sara; Szoszkiewicz, Krzysztof; Willby, Nigel L.; Wood, Kevin A. 2018. **Plants in aquatic ecosystems: current trends and future directions**. *Hydrobiologia*, 812 (1). 1-11. <https://doi.org/10.1007/s10750-017-3190-7>

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The final publication is available at Springer via <https://doi.org/10.1007/s10750-017-3190-7>

Contact CEH NORA team at
noraceh@ceh.ac.uk

1 **Plants in aquatic ecosystems: current trends and future directions**

2

3 Authors: Matthew T. O'Hare¹, Francisca C. Aguiar², Takashi Asaeda³, Elisabeth S. Bakker⁴, Patricia A.
4 Chambers⁵, John S. Clayton⁶, Arnaud Elger⁷, Teresa M. Ferreira², Elisabeth M. Gross⁸, Iain D. M.
5 Gunn¹, Angela M. Gurnell⁹, Seppo Hellsten¹⁰, Deborah E. Hofstra⁶, Wei Li¹¹, Silvia Mohr¹², Sara
6 Puijalón¹³, Krzysztof Szoszkiewicz¹⁴, Nigel J. Willby¹⁵, Kevin A. Wood¹⁶.

7 1. Centre for Ecology & Hydrology, Bush Estate, Penicuik, Scotland, EH27 0QB, UK

8 2. Forest Research Centre, School of Agronomy, University of Lisbon, Tapada da Ajuda 1349-017
9 Lisbon, Portugal

10 3. Department of Environmental Science, Saitama University, 255 Shimo-okubo, Sakura, Saitama,
11 338-8570, Japan

12 4. Department of Aquatic Ecology, Netherlands Institute of Ecology (NIOO-KNAW),
13 Droevendaalsesteeg 10, 6708 PB Wageningen, The Netherlands

14 5. Environment and Climate Change Canada, 867 Lakeshore Road, Burlington, Ontario, Canada L7R
15 4A6

16 6. National Institute of Water & Atmospheric Research, Gate 10 Silverdale Road, Hillcrest, Hamilton,
17 New Zealand

18 7. EcoLab, Université de Toulouse, CNRS, INPT, UPS, Toulouse, France

19 8. Université de Lorraine, LIEC UMR 7360 CNRS, Rue Général Delestraint, Bâtiment IBISE, F-57070
20 Metz, Lorraine, France.

21 9. School of Geography, Queen Mary University of London, London E1 4NS, UK

22 10. Finnish Environment Institute SYKE, Freshwater Centre, Paavo Havaksen tie 3, FI-90570 Oulu,
23 Finland

24 11. Key Laboratory of Aquatic Botany and Watershed Ecology, Wuhan Botanical Garden, Chinese
25 Academy of Sciences, Wuhan, 430074, China

26 12. Umweltbundesamt, Schichauweg 58, 12307 Berlin, Germany

27 13. Université de Lyon, Université Claude Bernard Lyon 1, CNRS, ENTPE, UMR 5023 LEHNA, F-69622
28 Villeurbanne, France

29 14. Poznan University of Life Sciences, Faculty of Environmental Engineering and Spatial
30 Management, Wojska Polskiego 28, 60-637 Poznan, Poland

31 15. Biological & Environmental Science, Faculty of Natural Science, University of Stirling, Stirling, FK9 4LA,
32 UK

33 16. Wildfowl & Wetlands Trust, Slimbridge, Gloucestershire, GL2 7BT, UK

34

35 *Key words:* Angiosperms; Botany; Herbivory; Limnology; Macrophytes; Submerged aquatic
36 vegetation; Trends in research

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38 *Running title:* Plants in aquatic ecosystems

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42 **Abstract**

43 Aquatic plants fulfil a wide range of ecological roles, and make a substantial contribution to the
44 structure, function and service provision of aquatic ecosystems. Given their well-documented
45 importance in aquatic ecosystems, research into aquatic plants continues to blossom. The 14th
46 International Symposium on Aquatic Plants, held in Edinburgh in September 2015, brought together
47 120 delegates from 28 countries and six continents. This special issue of *Hydrobiologia* includes a
48 select number of papers on aspects of aquatic plants, covering a wide range of species, systems and
49 issues. In this paper we present an overview of current trends and future directions in aquatic plant
50 research in the early 21st century. Our understanding of aquatic plant biology, the range of scientific
51 issues being addressed and the range of techniques available to researchers have all arguably never
52 been greater; however, substantial challenges exist to the conservation and management of both
53 aquatic plants and the ecosystems in which they are found. The range of countries and continents
54 represented by conference delegates and authors of papers in the special issue illustrate the global
55 relevance of aquatic plant research in the early 21st century but also the many challenges that this
56 burgeoning scientific discipline must address.

57

58 **Introduction**

59 In the early 21st century, researchers recognize the fundamental importance of plants that grow in
60 and around water to the structure, functioning and service provision of aquatic ecosystems
61 (Chambers et al., 2008). Aquatic plants interact with and influence the hydrological,
62 geomorphological and physico-chemical environments, and interact with a wide range of other
63 organisms, from microbes to vertebrates, for example, by providing habitat and food (Brix, 1997;
64 Engelhardt & Ritchie, 2001; Wood et al., 2017a). The current interest contrasts with the views of
65 earlier limnologists a century ago who considered aquatic plants to be largely unimportant in aquatic
66 ecosystems; for example, Shelford (1918) argued that "*One could probably remove all the larger*
67 *plants and substitute glass structures of the same form and surface texture without greatly affecting*
68 *the immediate food relations*". Over the past century the study of aquatic plants has expanded
69 considerably, because of the increased recognition of their importance in fundamental system
70 processes. Specialist journals have been established, such as Aquatic Botany (Den Hartog, 1975) and
71 Journal of Aquatic Plant Management, as well as conferences devoted to aquatic plant research.

72 As a consequence of the growth of aquatic plant research over recent decades, our views on many
73 key topics in aquatic botany have shifted (Vermaat & Gross, 2016; Phillips et al., 2016), and so this
74 introduction to the special issue on plants in aquatic systems presents an overview of current trends
75 and future directions in aquatic plant research in the early 21st century. It is a time of newly emerging
76 fields and the advancement of long-established research areas. The research is set against a
77 background of rapid environmental change that has been on-going for at least the last two centuries.
78 The pace of change is unremitting with demands on water resources set to increase globally
79 (Dudgeon et al., 2006; Vörösmarty et al., 2010). In the future the response of aquatic plant
80 dominated systems (e.g., shallow lakes and seagrass beds) to global temperature increases and
81 climatic extremes may well become a focus of research efforts. The in-depth understanding aquatic
82 botanists possess can only contribute positively to our understanding of how climate change will

83 perturb aquatic systems. Trends in aquatic plant research reflect the environmental pressures on
84 freshwater systems, legislative drivers, technical advances and developments in the wider fields of
85 ecology and environmental management.

86 Both national and international legislative drivers have had a clear impact on the direction of aquatic
87 plant research. In Europe, the implementation of the European Union (EU) Water Framework
88 Directive (WFD) (European Commission, 2000) led to a massive surge in research on monitoring
89 methods, their inter-calibration and the analysis of the resulting large multi-site datasets (Hering et
90 al., 2010). As the WFD implementation moves into its second phase, we now see a shift in focus to
91 restoration projects. We have learnt much during the implementation of this directive and it is likely
92 that we will see knowledge transfer from European scientists to colleagues in countries across the
93 globe. We see many countries in Asia and Africa now adopting reference based systems for
94 freshwater assessment (e.g., Kennedy et al., 2016).

95 The global financial crash in 2008 has exacerbated the difficulty in obtaining research funding in
96 many countries, and immediate output in terms of results reigns over the long-term understanding
97 of complex interactions and processes (Krugmann, 2012). In Europe we have also seen a reduction in
98 core funding for national research organizations and university researchers who work on aquatic
99 plant management issues and there are concerns that there will be a slow erosion of the research
100 base. The United Kingdom's decision in 2016 to leave the EU will likely have implications for site-level
101 conservation of aquatic plants under the EU Habitats Directive (Council of the European
102 Communities, 1992), although it is not yet clear what will replace the EU Directives in UK law. In the
103 USA, the Department of Energy has been planning to increase hydropower output by retro-fitting
104 turbines to pre-existing dams that are currently only used for flood control or water supply. While
105 the election in the USA of President Trump in 2016, who is a climate change sceptic and pro-fossil
106 fuel advocate, makes the implementation of this policy much less certain, it is worth noting that it did
107 have substantial cross-party support. If this work is undertaken it could reduce the USA's carbon

108 production and reduce its requirement to buy in fossil fuels from abroad, but careful assessment of
109 downstream impacts on aquatic plants and other taxa will need to be undertaken. In China the
110 current five-year plan, which has significant green policies, has energized the environmental sector
111 and led to substantial efforts to exchange knowledge with western countries. We hope this exchange
112 will lead to greater international collaboration between aquatic botanists in the future. In developing
113 countries there is a need too for the services of aquatic botanists where rapid population expansion
114 and the intensification of resource use have increased demands on water supplies and other natural
115 resources. A striking example is the numerous hydropower plants constructed in South America that
116 have caused profound changes in aquatic ecosystems, including macrophyte community composition
117 and patterns of colonization (e.g., Martins et al., 2013). Yet at the same time as these enormous
118 ecological changes, many developing countries also face reduced research funding and weakened
119 environmental legislation, which limits conservation efforts (Azevedo-Santos et al., 2017). The
120 conference attracted delegates from many developing countries and we would strongly encourage
121 their future participation.

122 While global financial trends and legislative drivers have affected the direction of research, technical
123 advances in survey and analytical methodologies have also been influential. Some established
124 techniques have become increasingly used in aquatic botany, for example, molecular biology and
125 stable isotope analysis. Recent reductions in the cost of stable isotope analysis have facilitated their
126 use. Developments in ecological modelling and computational biology have allowed aquatic plants to
127 be incorporated into models that can predict interactions between macrophytes and other
128 organisms (e.g., Wood et al., 2014; Stillman et al., 2015). The continued development of remote
129 sensing, drone technology and the software to interpret aerial photography, now allows new types
130 of spatial analysis. Moreover, the potential for drones to carry Light Detection and Ranging (LIDAR)
131 equipment could facilitate aquatic plant-sediment interaction studies. The rise of ‘citizen science’
132 represents greater public participation in scientific research and has the potential to aid data
133 collection (McKinley et al., 2017). Similarly, the emergence of R (the free statistical software

134 environment) has encouraged the development and sharing of new analytical techniques (R Core
135 Development Team, 2016).

136 Aquatic botanists work from an especially strong position where the physiology of the plants is well
137 described and there is a deep knowledge of the plants' roles in system function. Aquatic plants have
138 many advantages over other aquatic biota as study organisms: they are sessile, they can be
139 accurately mapped, rapidly surveyed and cultured easily in the laboratory, and they are increasingly
140 being used by a wide variety of researchers. Although, historically, there was an assumption that
141 publishing aquatic botany studies in high impact journals was challenging, there is anecdotal
142 evidence that this is no longer the case.

143 Against this background of environmental and societal change, aquatic botanists met recently to take
144 stock of their discipline at the 14th International Symposium on Aquatic Plants, held in Edinburgh in
145 September 2015. The symposium series originally began as an aquatic weeds meeting but over time
146 the focus of the symposia changed as research and management interests altered. As our
147 understanding and appreciation of the different roles that macrophytes play has increased, so too
148 have the breadth of topics addressed at the symposia. The conference continues to attract delegates
149 involved in the practical management of aquatic systems and those working directly in research. The
150 synopsis which follows is based primarily on the conference output. The 14th International
151 Symposium was attended by 120 delegates from 28 countries and six continents, and featured 79
152 oral presentations in addition to over 30 poster presentations. Although the 2015 symposium and
153 the 13 preceding symposia were held in Europe, henceforth, every second symposium will be held
154 outside Europe to reflect the global nature of the subject and the attendees. Global regions often
155 diverge in approaches and attitudes towards macrophytes, for instance, weed management with
156 herbicides is well accepted in the United States yet largely prohibited in Europe. Therefore, truly
157 international conferences are vital in order to provide opportunities for global debates on such key
158 issues. The next conference will take place in February 2018 in New Zealand and it will be jointly held

159 with the Aquatic Plant Management Society of North America. The conference will also be supported
160 by our colleagues from China, where there has been an upsurge in research interest in aquatic plants
161 in recent years.

162 Traditionally, authors of conference presentations elaborated their contributions as full papers
163 published in a special issue of *Hydrobiologia* (e.g., Caffrey et al., 1996; Caffrey et al., 1999; Caffrey et
164 al., 2006; Pieterse et al., 2010; Ferreira et al., 2014). Thus, in this special issue of *Hydrobiologia*, we
165 present a number of studies of aquatic plants that comprise the peer-reviewed proceedings of the
166 14th International Symposium on Aquatic Plants. In the remainder of this paper, we present an
167 overview of current trends and future directions in aquatic plant research in the early 21st century.
168 We focus on the following key areas of study, each of which represented a key session during the
169 conference: (i) physical habitat interactions, (ii) riparian processes, (iii) ecological stoichiometry and
170 nutrient cycling, (iv) trophic interactions – focused on plant herbivore interactions, (v) community
171 responses to environmental change in space and time, (vi) aquatic plant monitoring, (vii)
172 ecotoxicology, (viii) restoration, (ix) the future of invasive species management and (x) fundamental
173 science.

174

175 **Overview of current trends and future directions in aquatic plant research**

176 *Physical habitat interactions and riparian processes*

177 The interactions between plants and water flow and sediments has been championed sporadically
178 for over forty years, but in the last decade work has accelerated as the importance of the
179 interactions for ecology, hydrology and fluvial geomorphology were fully realized. Plants influence
180 physical processes: transport of solutes, sediment deposition/resuspension, hydraulic conditions and
181 light transmittance (O'Hare, 2015; Klančnik et al., 2017). In turn the physical environment affects
182 macrophytes. Its effects are induced by mean velocity, turbulence and water level (O'Hare, 2015).

183 Macrophytes can be affected at scales, from individual plants to populations and communities. This is
184 exemplified by plant growth which is known to be influenced from the microscale, for example, cell
185 ultrastructure (Atapaththu et al., 2015), to macroscale, for example, biomechanical traits (Puijalon et
186 al., 2011; Schoelynck et al., 2014). Current developments in our understanding of these complex two-
187 way interactions between aquatic vegetation and physical factors are tightly linked to fluid dynamics
188 modelling (Marjoribanks et al., 2014; Verschoren et al., 2016).

189 While aquatic botanists have tended to focus on aquatic macrophytes, geographers have been
190 examining both instream and riparian vegetation. An especially exciting development is the
191 realization that vegetation fringing a river's edge has a substantial influence on fluvial
192 geomorphological processes. In effect, nearshore plants (emergent and submerged) help engineer
193 river form (Gurnell, 2014; Gurnell et al., 2016). This has significant practical implications as
194 alterations to hydrology and fluvial geomorphology are as widespread as nutrient pollution in
195 Europe, effecting approximately half of all water bodies (Kristensen, 2012). We speculate that this
196 reflects an unmeasured but global trend as evidenced by the contributions from Africa and Asia to
197 this session on impacts of flow disturbance and regulation. Regulation by hydropower dams
198 influences the colonization rates of aquatic and riparian vegetation, with synergic impacts when
199 rivers are subjected to sediment removal or impaired by storage reservoirs (Aguiar et al., 2016). Such
200 disturbances create ecosystems prone to alien plant invasions, and regulation alters the growth
201 trajectories, composition and complexity of native communities (Bunn & Arthington, 2002). During
202 the conference the concerning case of Podostemaceae in West-Africa (strictly aquatic angiosperms)
203 was highlighted, where six species are critically endangered and four species have become extinct
204 due to altered flows (personal communication). Such issues can be overcome: for example,
205 implementing environmental flows that inundate geomorphological structures and create slack
206 waters helped with the restoration of regulated rivers by enhancing recruitment and colonization
207 (Rivaes et al., 2015; Souter et al., 2014). While most research in this field focuses on rivers, data from
208 the UK and Denmark indicate artificial water-level fluctuations in lakes affects macrophytes (e.g.,

209 Baastrup-Spohr et al., 2015; May & Spears, 2012; Smith et al., 1987), and that shoreweed (*Littorella*
210 *uniflora* (L.) Asch.) has potential as a model species in ecological studies of both lake productivity and
211 morphometry (e.g., Baastrup-Spohr et al., 2016; Robe & Griffiths, 2000).

212 In due course, this field of research has the potential to produce novel tools for management,
213 especially nature-based solutions to flooding, and fresh insights into the ecology of aquatic plants. A
214 research effort equivalent to that which elucidated the basic mechanisms of lake eutrophication
215 (Vollenweider, 1968) will likely be required to resolve these major research questions. With this
216 realization will come a far greater appreciation of the role of both instream and riparian vegetation in
217 engineering physical habitats. Further collaborative research between geographers and ecologists
218 will emerge.

219

220 *Ecological stoichiometry and nutrient cycling*

221 Ecological stoichiometry bridges ecology and ecosystem functions or processes at various levels,
222 from individuals to communities. Despite clear theories (Elser et al., 2000), elemental requirements
223 and the influence of environmental factors on nutrient uptake seem more complex for aquatic plant
224 systems. At a global scale, silica is a nutrient which is in surprisingly short supply in marine
225 environments requiring frequent inputs from freshwater systems. The role of macrophytes and other
226 primary producers in influencing silica delivery is gaining increasing interest and its accumulation in
227 macrophytes may be a functional trait that enables them to adapt to environmental conditions
228 (Schoelnyck & Struyf, 2016). At local scales, macrophytes strongly influence their physico-chemical
229 environment. Aquatic weed mats may constitute important hotspots for greenhouse gas emissions in
230 temperate shallow lakes, but wetland vegetation can also assist in nitrogen assimilation (Ribaudo et
231 al., 2017; Volkmann et al., 2016). Yet, the relation between environmental nutrient availability and
232 macrophyte nutrient content is often less clear. For example, research, presented during the
233 conference, showed that upland streams with proliferations of pond water-crowfoot (*Ranunculus*

234 *peltatus* L.) tend to have a low N:P ratio at overall very high nitrogen and phosphorus concentrations
235 (personal communication). Although intra-specific C:N:P stoichiometry of submerged macrophytes
236 correlates to sediment and water nutrient availability, inorganic carbon availability may also play a
237 strong role in their nitrogen-based metabolism (Hussner et al., 2016). Further research, presented
238 during the conference, found that macrophyte tissue nutrient concentrations appear more closely
239 related to plant growth form than to phylogeny (personal communication).

240

241 *Trophic interactions – focused on plant herbivore interactions*

242 Since the seminal paper by Lodge (1991) on herbivory of aquatic plants, researchers have been
243 devoting considerable attention to plant-herbivore interactions in aquatic ecosystems. Now, in the
244 early 21st century, it has now been demonstrated, unequivocally, that herbivores can provide strong
245 top-down regulation of macrophyte beds (Bakker et al., 2016; Wood et al., 2017a). These top-down
246 mechanisms can interact with recovery from stress; for example, recovery of macrophyte beds after
247 eutrophication attracts herbivorous water birds, but the colonization process can be hampered by
248 strong vertebrate herbivory. In contrast, smaller invertebrate grazers may assist recovery of
249 eutrophic systems. They stimulate submerged macrophyte growth and establishment by consuming
250 periphyton (instead of the tougher macrophytes) that would otherwise reduce light availability for
251 macrophytes (Bakker et al., 2016; Wood et al., 2017a).

252 Recognizing the importance of herbivory opens new research avenues by scaling up from
253 macrophyte beds to aquatic ecosystem functioning, as herbivores affect methane emission, carbon
254 cycling and regime shifts (Hidding et al., 2016). Furthermore, there is an urgent need to predict how
255 global change will alter trophic interactions as a result of exotic species invasions (Redekop et al.,
256 2017), temperature rises (Zhang et al., 2017) or changes in hydrological patterns (Wood et al.,
257 2017b). Finally, current and future conservation challenges lay in predicting and managing the
258 consequences of recovery of larger vertebrate herbivores, through re-introductions such as the

259 Eurasian beaver (*Castor fiber* L.) in Europe (e.g., re-wilding), as well as by strong local herbivore
260 population increases in species such as mute swans (*Cygnus olor* Gmelin).

261

262 *Community responses to environmental change in space and time*

263 The study of the responses of aquatic plant communities to environmental change in space and time
264 is both a mature field of research and one with critical new questions being asked. Current research
265 effort has seen a continued focus on the role of bottom-up regulation through environmental drivers
266 (e.g., Fernández-Aláez et al., 2017) and competitive processes between macrophyte species (e.g.,
267 Gérard & Triest, 2017; Nunes & Camargo, 2017) in shaping aquatic plant community composition.
268 Our understanding of how connectivity can influence floodplain macrophyte populations has now
269 matured to the point where scenario modelling is feasible, for example, on the Murray-Darling
270 system in Australia where species richness of floodplain plant communities can be predicted as a
271 function of channel connectivity in the watershed (Campbell et al., 2014). Furthermore, recent
272 studies of aquatic plant responses to floods in large floodplains have offered support for the flood
273 homogenization hypothesis (Thomaz et al., 2007). Floodplain inundation has received less attention
274 on smaller systems; however, comparative assessments of the importance of different aquatic
275 habitats to a Scottish regional flora confirmed the importance of riverine backwaters (Keruzoure et
276 al., 2013), a habitat that had been previously neglected. That study illustrated an increasing
277 awareness of spatial processes operating beyond individual sites, and the associated issue of scale-
278 dependent responses. Thus, for example, the effects of land use on macrophyte richness in lakes are
279 scale-dependent and are of greater importance at small spatial scales relative to the influence of
280 hydrological connectivity (O'Hare et al., 2012). Looking beyond the immediate is one of the most
281 powerful approaches of space and time analyses, and frequently produces insightful findings. Not
282 only do we see this in relation to hydrological connectivity but also in legacy signals, for example, the
283 lakes of northwest Europe are geologically young due to their glacial origins, with the signal of
284 glaciation still evident in the composition of their flora (Alahuhta et al., 2017).

285

286 *Aquatic plant monitoring*

287 Changes in the abundance or composition of an aquatic plant community are often obvious signals of
288 alteration in the ecological condition of a lake or stream. In fact, a recent review of assessment
289 methods used to implement the EU Water Framework Directive showed that the majority of
290 methods are based on macroscopic plants (28% of all methods), followed by benthic invertebrates
291 (26%) (Birk et al., 2012). Moreover, unlike many other biological indicators, macrophytes are equally
292 good at detecting eutrophication/organic pollution and hydrological/morphological changes (Birk et
293 al., 2012). Historically, surveys of abundance and composition were challenging in terms of both field
294 effort and taxonomic ability. As identified at this symposium, improved methods for mapping
295 abundance and composition of aquatic vegetation are now becoming available: high-resolution aerial
296 images of lake and rivers taken with unmanned aircraft systems permit identification, mapping and
297 abundance estimates of non-submerged species while near-infrared-sensitive DSLR cameras can be
298 used to map spatial distribution and depth of submerged species (e.g., Visser et al., 2015).

299 Research is continuing to show that community metrics (e.g., cover, diversity and richness) and
300 species frequency of occurrence are often related to water quality, lending support for the
301 development of macrophyte-based indices for classification of fresh waters and brackish water
302 ecosystems and seagrass beds (Spears et al., 2016). Although many macrophyte indices are based
303 only on hydrophytes due to their dependency on the quality of the aquatic environment, the
304 importance of helophytes has been demonstrated as indicators of the eutrophication process, for
305 example, in the bioassessment of lowland lakes (Kolada, 2016). Biochemical measurements may also
306 provide a new tool for bioassessment: for example, during the conference evidence was presented
307 that ^{15}N and C:N values from caged duck weed (*Spirodela* sp.) were found to relate to the proximity
308 and timing of sewage manure or fertilizer inputs into rivers in South Africa (personal
309 communication). Despite encouraging advances in both methods for mapping aquatic vegetation and

310 approaches for assessing water quality, physical factors such as hydrological modifications to water
311 courses or inter-annual variation in water levels can confound the relationship between macrophyte
312 occurrence and water quality, necessitating caution when deciding the status of a water body based
313 on limited (temporal or spatial) macrophyte data.

314

315 *Ecotoxicology*

316 The banning of herbicides for use in aquatic systems across the EU resulted in a shift in research
317 away from studies on the efficacious use and impacts of pesticides in controlling aquatic plants. A
318 strong research focus remains, however, on the effects of pesticides and other pollutants derived
319 from terrestrial systems on aquatic plants (Coutris et al., 2011; King et al., 2016).

320 This was the first time an ecotoxicology session was held at the conference and it focused on linking
321 ecological studies with chemical risk assessment, with the overarching aims to make assessment
322 methods more realistic and to identify emerging plant-contaminant issues. The work presented in
323 the session indicated a continuing shift toward the use of more realistic test species. To refine risk
324 assessments, laboratory studies used more realistic exposure conditions than standard techniques;
325 an example was presented at the conference in which pesticide exposure pulses, typical of running
326 water bodies, caused less harm to gibbous duckweed (*Lemna gibba* L.) than standard exposure
327 conditions (personal communication). A higher tier approach, using mesocosms, proved effective
328 when investigating indirect effects of chemicals on plant populations and communities. On plant-
329 contaminant issues, the interaction between chemical contaminants and other stressors was evident;
330 for example, evidence presented at the conference showed that the stoichiometry (C:N:P) of
331 Eurasian water milfoil (*Myriophyllum spicatum* L.) was not only influenced by light and nutrients, but
332 also by herbicides and the metalloid arsenic (personal communication). Field monitoring and
333 biomarker assays revealed a significant relationship between the decline of dwarf eelgrass (*Zostera*

334 *noletii* Hornem.) in the Vaccarès lagoon in France and its exposure to chemical contaminants
335 including metals and pesticides (personal communication).

336

337 *Restoration*

338 Management of aquatic macrophytes is an essential part of freshwater restoration projects (Phillips
339 et al., 2016). Macrophyte restoration can have multiple benefits, for example, supporting
340 endangered waterfowl and fish species or limiting the spread of invasive species, such as Nuttall's
341 waterweed (*Elodea nuttallii* (Planch.) H. St. John), in Europe. To successfully restore macrophytes,
342 consideration of the following factors can be helpful: the genetic background of macrophyte
343 population used, native seed bank viability, control of herbivores and, in the case of eutrophic lakes,
344 the use of geo-engineering tools which reduce internal P loading, (Combroux et al., 2001; Guittonny-
345 Philippe et al., 2015; Hussner et al., 2017). Restoration science is still under development and new
346 data are desirable; monitoring using macrophyte growth forms can provide a cost-effective tool for
347 evaluating the effect of individual restoration projects while long-term records of macrophyte
348 dynamics can provide valuable information for assessment of broader, global scale change (Ecke et
349 al., 2016).

350 Throughout the history of this symposium the loss of lake macrophytes due to eutrophication has
351 been a core issue. Now, in the 21st century, research on the mechanisms of eutrophication continues
352 but with a somewhat different emphasis; we now see more work presented on systems that are in
353 recovery. Research has turned to drivers that influence the recovery trajectory; for example, trophic
354 interactions involving herbivores, which have been somewhat neglected in the past, and issues
355 associated with the role of invasive species.

356

357 *The future of invasive species management*

358 The spread of invasive species and decline in biodiversity is associated with accelerating
359 globalisation, human migration and increasing pressures on freshwater supplies; however, whilst
360 challenging, successful invasive species management has been demonstrated using combinations of
361 lake and aquatic plant-based approaches matched with appropriate management tools (Havel et al.,
362 2015). In some cases, regime shifts amongst aquatic flora, such as floating to submerged vegetation,
363 may follow from the use of classical biological control (Cuda et al., 2008; Bakker et al., 2016). Yet in
364 other cases invasive aquatic plants may not be considered the primary drivers of change, adding to
365 debate surrounding the anthropocentric interpretation of benefits (vs detriments) for many non-
366 native species in impacted habitats. Increasingly, there is a focus towards, arguably, bigger more
367 'threatening' issues such as climate change in the management of invasive species that could result
368 in greater impacts from existing nuisance aquatic plants at a global level. For example, alien aquatic
369 species can reduce the diversity of native seedbanks, thereby, jeopardising future restoration.
370 Targeted experimental work in both field and laboratory conditions is allowing researchers to
371 understand competitive interactions between native and invasive species (Gérard & Triest, 2017).
372 Continued research investment is required to manage the spread of invasive species. The
373 development of new knowledge and techniques will likely provide new opportunities in the future
374 for more effective invasive species management and aquatic restoration (e.g., Lozano & Brundu,
375 2017).

376

377 *Fundamental science*

378 Applied aspects dominate much of current aquatic plant research, such as aquatic plant populations'
379 restoration, monitoring and ecological quality assessment, and different forms of response of aquatic
380 plants to human disturbance or novel ways to control plant overgrowth. Nonetheless, fundamental
381 science is often the basis for management actions, and indeed many failures relate to the lack of
382 taxonomic resolution, the misunderstanding of species autecology and role in the ecosystem, or

383 undefined tolerance responses over the disturbance gradient. Fundamental science, thus, provides,
384 in large part, the key to successful plant management.

385 In spite of the development of genetic and cytoplasmic tools, morphological traits are still relevant as
386 well as the role of population traits, for example, for dispersal and survival. Many ecosystem
387 processes are also driven by vegetation, shaping succession of both plant and animal communities, in
388 the short- and long-terms, in which interspecific competition and environmental constraints
389 determine the end point. Understanding such processes is fundamental for biomanipulation,
390 ecosystem restoration and the proper management of both constructed and natural wetlands.

391

392 **Conclusions**

393 Both the conference presentations and this resulting special issue of *Hydrobiologia* reflect the broad
394 discipline that aquatic botany has become over the last century. Research interest in aquatic plants
395 range from the use of aquatic plants as model organisms, to the roles of aquatic plants within
396 ecosystems and to the conservation of aquatic plants themselves. Furthermore, the range of
397 countries and continents represented by conference delegates and authors of papers in this special
398 issue illustrate the global relevance of aquatic plant research in the early 21st century.

399 Currently, the International Symposia on Aquatic Plants are dominated by research on freshwater
400 taxa, and in particular those found in shallow lakes. However, greater integration of freshwater
401 macrophyte and marine seagrass research efforts, and their associated literatures, would benefit our
402 overall understanding of aquatic plant biology, management and conservation. Whilst aquatic plant
403 species may differ across ecotones, the processes that shape aquatic plant assemblages, such as
404 bottom-up and top-down control and competitive processes, will share common elements. For
405 example, recent research into herbivory on aquatic plants has synthesized information from
406 freshwater, brackish and marine ecosystems (e.g., Bakker et al., 2016; Wood et al., 2017a).

407 Our understanding of aquatic plants, the range of scientific issues being addressed and the range of
408 techniques available to researchers, have all arguably never been greater. This is to be welcomed, as
409 the challenges facing researchers and practitioners have also never been more pressing. Climate
410 change, rising human demand for resources including water, pollution of freshwater resources, the
411 spread of invasive non-native species, land-use changes and intensification, together with the
412 degradation, fragmentation and loss of aquatic habitats, all present huge challenges to the
413 conservation and management of both aquatic plants and the ecosystems in which they are found
414 (Dudgeon et al., 2006; Vörösmarty et al., 2010; Short et al., 2016). The 15th International Symposium
415 on Aquatic Plants, to be held in New Zealand in February 2018, will be an excellent opportunity to
416 assess our progress in meeting these challenges and to identify the areas in which we need to do
417 more.

418

419 **Acknowledgements**

420 We are grateful to André Padial, Baz Hughes, and two anonymous reviewers for their helpful
421 comments on earlier drafts of this manuscript.

422

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424

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