

FIXED ON NITROGEN

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Abstract

Critical Loads for the Management of Nitrogen Acidification and Eutrophication Dr. Chris Evans

Critical loads have been extensively used, particularly in Europe, as a tool for managing nitrogen emissions. Critical loads define the acceptable pollutant loading to an ecosystem. If the critical load is exceeded, long-term ecosystem damage is expected. Nitrogen deposition is relevant to: i) critical loads for acidity, which define the acceptable combined loading of sulphur and nitrogen deposition, beyond which acidification damage occurs in terrestrial or freshwater ecosystems; and ii) critical loads for nitrogen as a nutrient, which define the specific nitrogen loading beyond which biodiversity loss will occur due to eutrophication. Critical loads vary according to the sensitivity of the ecosystem. They have been used at large (e.g. European) scales to define the most cost-effective strategy for emissions reductions (i.e. that which will lead to the greatest reduction in critical load exceedance for the amount spent), and are also now being used at smaller scales, e.g. to protect and manage individual sites of conservation importance. Although they have proven effective as a policy tool, critical loads nonetheless have a number of limitations. Most importantly, they predict damage to the ecosystem at long-term steady state. Since many ecosystems have the capacity to accumulate nitrogen over very long periods, the lag between critical load exceedance and observable ecosystem damage (or indeed deposition reductions and ecosystem recovery) may be long. New approaches, which overcome some of these limitations using dynamic models, and more detailed vegetation models, are described.



Critical Loads for the Management of Nitrogen Acidification and Eutrophication

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Overview

- 1. The critical loads concept
- 2. Critical loads for N and acidity
- 3. Critical loads for N as a nutrient
- 4. Dynamic models
- 5. Critical loads in Alberta
- 6. Conclusions



1. The Critical Loads concept



Eagle Mountains, Czech Republic, 2005



Regulating long-range pollutant emissions

Option 1: Best-available technology

Option 2: Effects-based approach

<u>**Critical Load</u>** = the **highest annual input** of the pollutant that, **at steady-state**, does not cause unacceptable **ecological** [*or human health*] effects</u>

<u>Critical Limit</u> = the highest steady-state concentration of the pollutant that does not cause unacceptable ecological [*or human health*] effects







Setting critical loads

1. Define an indicator of change for the receptor of interest:

Ecosystem structure Sensitive indicator species Nitrate leaching Soil acidification

2. Define a dose-response function

3. Define a damage threshold for the required level of ecosystem protection



CLs assume a damage threshold exists – if dose vs damage is linear, we have more of a problem...



2. Critical Loads for acidification





















Critical Loads and ecosystem damage 1) Nitrogen and acidity





Critical Loads and ecosystem damage 1) Nitrogen and acidity





Critical Loads and ecosystem damage 1) Nitrogen and acidity





Calculating critical loads for acidity UK Methods

- Skokloster classes
 - Heathland and grassland
 - Basically estimates of long-term buffering provided by weathering in different soils
 - 5 sensitivity classes
- Simple mass balance (SMB)
 Forests



Simple mass balance (SMB) model

- Based on a critical limit for UK forests this is Ca:Al = 1
- Balances acid inputs and outputs to derive a critical load that ensures the critical limit is not exceeded
- And the equations are...

$$\begin{split} & \mathsf{CL}_{max}(\mathsf{S}) = \mathsf{BC}_{dep} - \mathsf{CI}_{dep} + \mathsf{BC}_w - \mathsf{BC}_u + (1.5 \times \mathsf{Ca}_{le}/(\mathsf{Ca}:\mathsf{AI})_{crit}) + \mathsf{Q}^{2/3}(1.5 \times \mathsf{Ca}_{le}/((\mathsf{Ca}:\mathsf{AI})_{crit} \times \mathsf{K}_{\mathsf{Gibb}}) \\ & \mathsf{CL}_{min}(\mathsf{N}) = \mathsf{N}_i + \mathsf{N}_{de} + \mathsf{N}_u \\ & \mathsf{CL}_{max}(\mathsf{N}) = \mathsf{CL}_{max}(\mathsf{S}) + \mathsf{CL}_{min}(\mathsf{N}) \end{split}$$



The critical load function





UK 5th percentile Critical Loads for Acidity





European Critical Loads for Acidity





Critical Load Exceedance, UK





Critical Load exceedance in UK surface waters: What needs to be done to reduce exceedance?



It will only be possible to remove critical load exceedance in these areas by reducing N deposition



Time lags between exceedance and damage

- For S deposition, exceedance of critical loads may lead to relatively rapid damage
- Delays occur due to:
 - Base cation buffering
 - S adsorption (mainly in unglaciated soils)
 - S reduction (mainly in wetlands)
- For N deposition, lags between critical load exceedance and damage may be much longer.
- Delays are primarily due to soil N immobilisation



Predicted steady state

Significance of lags in N leaching

Now



Curtis et al., Environmental Pollution (2005)

- Critical loads models generally predict a much higher level of steadystate N leaching than is currently observed
- Lag times appear to be long
- But if NO₃ leaching does reach predicted levels, future acidification could be as bad, or worse, than the 1970s-80s.



Nitrogen sources and sinks at Llyn Llagi, Wales

1. Present day





Nitrogen sources and sinks at Llyn Llagi, Wales

2. Future (1)





Nitrogen sources and sinks at Llyn Llagi, Wales

2. Future (2)





Nitrogen sources and sinks at Llyn Llagi, Wales

2. Future (3)





Nitrogen sources and sinks at Llyn Llagi, Wales

3. Steady State

But not all sites like this – some parts of Europe leaching most or all of incoming N already

In a managed forest, N uptake may reduce N leaching (but, N deposition may be higher)





3. Critical Loads for Nitrogen as a Nutrient





Nitrogen as a nutrient

- Nitrogen is a major nutrient required by all plants, and the limiting nutrient in most northern ecosystems
- Many natural habitats are characterised by slowgrowing species adapted for low-N conditions.
- With increased N deposition, these species are out-competed by faster-growing species more able to exploit increased N availability
- The results is a loss of biodiversity, or of characteristic plant species.



Critical Loads and ecosystem damage 3) Nitrogen and biodiversity





Critical Loads and ecosystem damage 3) Nitrogen and biodiversity





Evidence that N deposition is causing eutrophication of UK ecosystems



Countryside Survey, changes between 1990 and 1998

Plant Atlas, changes between 1930-69 and 1987-99





Critical Loads for Nitrogen as a nutrient Empirical critical loads (Used for other UK ecosystems)

- Based on experimental/field evidence of thresholds for change in species composition, plant vitality or soil processes
- Focused on communities likely to be sensitive to N deposition, of conservation value and with a reasonably wide distribution
- European ranges defined at a workshop in Berne, 2002
- Reliant on a large amount of scientific data, and a certain amount of expert judgement
- Countries decide which communities to protect, and where within the range to set their critical loads



Berne empirical critical loads, and their application in the UK

(a) Ecosystem (with corresponding EUNIS class, where used)	(b) 2001 UK mapping value	(c) Critical load range in 1996 Mapping Manual	(d) Critical load range from Berne workshop	(e) Revised UK mapping value
Grasslands Dry acid and neutral closed grassland (E1.7) Calcareous grassland (E1.26) Montane grassland Hay meadows (E2.2) Montane hay meadows (E2.3) Arctic/sub-alpine grass Moist/wet oligotrophic grass (E3.5) Molinia meadows (E3.51) Nardus stricta swards (E3.52) Moss/lichen mountain summits (E4.2) Inland dune pioneer grass (E1.94) Inland dune silicaceous grass (E1.95)	25 25 ⁽¹⁾ 12	20-30 # 15-35 # 10-15 (#)	10-20 # 15-25 ## 20-30 (#) 10-20 (#) 10-15 (#) 10-20 # 15-25 (#) 10-20 # 5-10 # 10-20 (#) 10-20 (#)	15 20 - - - 15 - - - - - - - - - - - - - - -
Heathland/moorland Lowland dry heaths (F4.2) Lowland Erica wet heaths (F4.11) Upland Calluna wet heaths (F4.11) Arctic/alpine heaths (F2) Tundra (F1)	17 15 7.5	15-20 ## 17-22 # 10-20 (#) 5-15 (#)	10-20 ## 10-25 # 10-20 (#) 5-15 (#) 5-10 #	12 15 15
Coastal habitats Coastal stable dune grasslands (B1.4) Shifting coastal dunes (B1.3) Coastal dune heaths (B1.5) Moist-wet dune slacks (B1.8) Dune slack pools (C1.16) Salt marshes (A2.64 & A2.65)		20-30 #	10-20 # 10-20 # 10-20 (#) 10-25 (#) 10-20 (#) 30-40 (#)	15 15 - - -
Softwater oligotophic lakes Permanent oligotrophic lakes (C1.1) Bogs mires and fens		5-10 ##	5-10 ##	-
Ombrotrophic and raised bogs (D1) Poor fens (D2.2) Rich fens (D4.1) Montane rich fens (D4.2)	10	5-10 #	5-10 ## 10-20 # 15-25 (#) 15-25 (#)	10 15 -



- Boreal forest
 10-20 kg N/ha/yr, 'quite reliable'
- Onset of NO₃ leaching, N mineralisation
 - forest surveys, fertilisation experiments
- N/P and N/Mg imbalances in trees
 - forest surveys, fertilisation experiments
- Ground vegetation change
 - fertilisation experiments (e.g. displacement of Vaccinium myrtillus by Deschampsia flexuosa at > 5 kg N/ha/yr in N. Sweden)



2) Tundra

5-10 kg N/ha/yr, 'quite reliable'

- Vegetation change
 - One set of fertilisation experiments receiving 10 kg N/ha/yr, Svalbard, showing changes in species composition of moss layer, decrease in lichens.



- 3) Alpine grasslands 10-15 kg N/ha/yr, '*expert judgement*'
- Vegetation change
 - One experiment in Switzerland showing biomass increase after 4 years addition of 20 kg N/ha/yr
- Extrapolation from (better studied) lowland grasslands



4) Blanket bogs

5-10 kg N/ha/yr, 'reliable'

- Increased N in peat and peat water
 - Experiments, field surveys
- Changes in moss growth and N content
 - Experiments, field surveys
- Increases in vascular plants over mosses
 - Experiments, field surveys



UK 5th percentile nutrient N critical loads



- Cl_{nut}N for most of UK in the range 10-20 kg N/ha/yr
- Lower values for high mountain ecosystems



European 5th percentile nutrient N critical loads





Exceedance of 5th percentile nutrient N critical loads





Critical load exceedances across Europe 2010 forecast





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4. Dynamic Models





Dynamic Models

- Critical loads are essentially models of steadystate chemistry
- Dynamic models predict the *time* at which damage (or recovery) will occur
- Much current work in Europe is focused on modelling, in particular:
 - Setting 'Target Loads' the target deposition required to achieve acceptable chemical status by a given target date
 - Modelling biodiversity impacts by relating vegetation status to soil chemical status



Target loads for acidity, N European surface waters





Modelled lags in N leaching (Llyn Llagi again)

MAGIC calibrated present day





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Modelled lags in N leaching (Llyn Llagi again)

MAGIC predicted 2100





MAGIC modelling of lags in N leaching (Llyn Llagi again)

MAGIC predicted 2100

Predicted Steady State



Dynamic models suggest that many ecosystems are a long way (centuries?) from the steady state NO_3 leaching levels indicated by the steady state mass balance



Predicting biodiversity change with dynamic models: MAGIC-GBMOVE

UK Countryside Survey: 16,691 vegetation survey plots. Species recorded, Ellenberg values for fertility (Eb N), acidity (Eb R) and moisture (Eb F) calculated

Subset of sites to relate Ellenberg values to abiotic conditions (soil pH, moisture, C/N ratio)

C:N ratio

Ellenberg fertility

GBMOVE: Empirical relationships derived to predict probability of occurrence as a function of nitrogen, acidity and other environmental drivers







Predicting biodiversity change with dynamic models: MAGIC-GBMOVE



Note that GBMOVE does not assume a threshold

MAGIC simulation



GBMOVE predictions





Calculating critical loads with dynamic models: 1. Netherlands





Calculating critical loads with dynamic models: 1. Netherlands



Comparison of critical loads estimated by the method of van Dobben et al. (2006) with those estimated by the Steady State Mass Balance

- Differences occur because:
 - The SMART model approach allows greater 'acceptable' N leaching than the DMB
 - Estimated N immobilisation is higher
- Compared to empirical critical loads, van Dobben approach gives similar range but no correlation for individual habitat types



Calculating critical loads with dynamic models: 2. Sweden



The ForSAFE Model

ForSAFE modelled vegetation change





Calculating critical loads with dynamic models: 2. Sweden





Calculating critical loads with dynamic models: 2. Sweden

ForSAFE critical loads and exceedances for individual sites

Site	Time of vegetation response	Critical load deposition kg ha ⁻¹ yr ⁻¹	Present deposition kg ha ⁻¹ yr ⁻¹	Excess deposition kg ha ⁻¹ yr ⁻¹	Required deposition reduction %
Högbränna	1910	1.1	1.5	0.4	27
Brattfors	1890	0.9	2.0	1.1	55
Storulvsjön	1925	2.0	3.5	1.5	43
Högskogen	1928	4.8	7.9	3.2	40
Örlingen	1910	3.6	8.5	3.9	52
Edeby	1918	3.9	7.8	3.9	50
Blåbärskullen	1880	1.6	8.5	6.9	81
Höka	1920	4.0	8.9	4.9	55
Hensbacka	1922	7.4	18.0	10.6	59
Söstared	1868	2.1	20.0	17.9	89
Gynge	1870	2.8	8.3	5.5	66
Fagerhult	1915	3.7	7.5	3.8	51
Bullsäng	1870	2.1	15.0	12.9	86
Timrilt	1889	3.6	23.0	19.4	84
Vång	1910	7.8	17.0	9.2	54
Västra Torup	1866	2.4	27.0	24.6	91



5. Critical Loads in Alberta





Alberta vs Europe: N deposition levels

Wet Deposition of Nitrogen





1995-97 Total N deposition



Alberta vs Europe: Acidity Critical Loads

- Acidity critical loads applied to both
- Methods appear fundamentally similar:

Net Acidifying Potential:

NAP =
$$([SO_4^{2^-}] - [Ca^{2^+} + Mg^{2^+}])_{wet} + [NO_3^-]_{leached}$$

ForSust model: Steady state mass balance approach

- 95% protection level, similar chemical thresholds used
- Range of acidity critical loads (0.25 to 1.0 keq/ha/yr) similar to Europe, but with lower maximum values



Alberta vs Europe: Damage vs Recovery

- In Europe, critical loads are, or have been, exceeded across much of the area, so emphasis is on reduction of CL exceedance and modelling timescales of recovery
- In Alberta, critical loads haven't been exceeded anywhere, so emphasis is on avoiding damage

European Target Loads:

The target deposition required to achieve recovery by a specified date at a currently exceeded site: 'Have to do more'

Albertan Target Loads:

Somewhere between current deposition and the critical load (~90%).

'Factor of safety'



Alberta vs Europe: Eutrophication

- Critical loads for N as a nutrient have not yet been applied to Alberta
- Evidence from Europe is that ecosystems may be more sensitive to N deposition with regard to eutrophication than with regard to acidification
- One possibility is to adopt the critical loads for N as a nutrient developed in Europe
- But Albertan ecosystems and plant species differ significantly from those in Europe – need to ensure that sensitivity to N deposition is similar before applying European values.
- Ideally, a combination of experiments and linked soil-vegetation condition surveys are required to establish local species sensitivity to N deposition



Conclusions

- Critical loads aren't perfect!
 - They do not consider timescales of change
 - They simplify complex ecosystem processes by which deposition impacts on environmental quality into 1 (or 2) numbers
 - Chemical criteria and damage thresholds are not always well defined or verified
 - Long-term sinks, particularly for N, are uncertain
 - They assume a threshold that might not really exist
- Dynamic models can address some of these limitations, but are unlikely to entirely replace critical loads
- And whatever their failings, critical loads have proven to be a highly effective means of translating science into policy, and take significant credit for the success of negotations to reduce acidifying emissions in Europe



Critical loads have worked...



Callitriche hamulata (water starwort)