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1 Impacts of Nitrogen deposition on vascular plants in Britain: An analysis of
2 two national observation networks

3 **Running head:** Nitrogen impacts on vascular plants.

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18

1 **Abstract**

2 Large areas of Great Britain currently have nitrogen (N) deposition at rates which exceed the
3 thresholds above which there is risk of damage to sensitive components of the ecosystem (critical
4 loads). Previous studies have focussed primarily on the relationship of species richness to
5 nitrogen, whereas here we look at individual species. We used data from two national observation
6 networks over Great Britain to examine the response of individual vascular plant species to N in
7 acid grasslands, calcareous grasslands and heathlands. Presence absence records of individual
8 species, along with mean Ellenberg N scores, within 10km hectads were modelled against N
9 deposition whilst at the same time controlling for the effects of climate, land use and sulphur
10 deposition using generalised additive models. Ellenberg N showed a significant increase with
11 increasing N deposition in almost all habitats across both surveys indicating increased fertility.
12 Many individual species showed strong relationships with N deposition and clear negative trends
13 in species prevalence to increasing nitrogen were found in all habitats. A number of these species
14 were either habitat dominants or possessed traits known to be influential in controlling ecosystem
15 function. Many community dominants showing significant negative relationships with N
16 deposition highlight a potentially significant loss of function. Some species that showed negative
17 relationships to N showed signs of decline at low levels, far below the current critical load levels.
18 Some species also showed continuous changes as N deposition levels rose above the current
19 critical load values. This work contributes to the growing evidence base suggesting species level
20 impacts at low N deposition values.

21 **Keywords:** Air quality, critical loads, Ellenberg N, nitrogen deposition, pollution, vascular plants

22

1. Introduction

Atmospheric nitrogen (N) deposition poses a serious threat to sensitive semi-natural habitats in Great Britain (Hall et al., 2006a; RoTAP, 2011.). Estimated levels of nitrogen (N) deposition across Great Britain and Northern Ireland are shown in Fig. 1. Large areas of the country exceed the critical loads for nutrient N and critical levels for ammonia, and are predicted to continue to do so in 2020 despite reductions in emissions of reactive N gases (Hall et al., 2006b). The critical load is defined as the level below which significant harmful effects on specified sensitive elements of the environment do not occur (Nilsson and Grennfelt 1988). Therefore, by definition, as N levels surpass the critical load values there is possibility of a significant harmful effect on the environment. There is a growing body of evidence from both experiments and surveillance studies that suggest that increases in nitrogen have a negative effect on species richness and composition (e.g. Mountford et al., 1993; Wilson et al., 1995, Stevens et al., 2004, Maskell et al., 2010, Stevens et al., 2010a, Dupre et al., 2010, Smart et al., 2004, Van den Berg et al., 2010).

A negative relationship between species richness and N leads us to hypothesise that some individual vascular plant species show a decline in frequency of occurrence with increased N and are perhaps not found at all in some areas due to high N deposition levels. However, as some species lose out, decline and potentially become eliminated from the ecosystem, with increased levels of N, some species may flourish. Changes in mean indicator scores based on summarising changes across taxa can reveal strong signals of pollutant impacts. However, plant species respond individualistically so that assessment at an individual species level is required to understand how assemblage-level changes comprise different magnitudes and directions of change among the species present. Expressing habitat change at the species level also allows identification of potential indicators of air pollution and aids interpretation of critical load exceedance scenarios.

1 There have previously been a number of geographically large-scale studies attempting to
2 determine the impact of total inorganic N deposition on vegetation in Great Britain. For example,
3 Stevens et al., 2004 used a large scale spatial survey of acid grasslands to demonstrate declines in
4 species richness, when high N deposition was compared with areas of low N deposition. In the
5 RoTAP report (RoTAP, 2011.) results of eleven surveys conducted in Great Britain and Northern
6 Ireland are compared. Species richness declined with increasing N deposition in seven of the
7 eleven surveys. Further to this, species richness of vascular plants has been found to decrease in
8 response to N deposition in a number of experimental studies (e.g. Bobbink et al., 1998; Clark
9 and Tilman, 2008; Mountford et al., 1993).

10 The Countryside Survey has also been used to study the impact of N deposition on semi-
11 natural vegetation communities. Maskell et al., 2010 used data from the 1998 survey to show
12 declines in species richness in both acid grassland and heathland along the gradient of N
13 deposition covering Great Britain. Trait data was used to propose mechanisms for species loss.
14 These appeared to vary by habitat type; in acid grassland and heathland acidification appeared to
15 be the driving process rather than eutrophication (Maskell et al., 2010). In addition to this, large-
16 scale analyses of UK and European data have also detected parallel signals of eutrophication
17 (Dupre et al 2010; Smart et al 2004; RoTAP 2011). As N deposition increases, plant species and
18 traits associated with higher fertility have become more abundant and stress-tolerant species have
19 declined.

20 In some studies covering a range of habitats across Europe, Ellenberg values (Ellenberg et
21 al., 1991; Hill et al., 1999) have been used as indicators of N deposition effects (Diekmann and
22 Falkengren-Grerup, 2002; Jones et al., 2004). Ellenberg values describe the realised niche of a
23 plant in relation to its tolerance of certain environmental conditions on a scale of one to nine.
24 The most relevant to N deposition is the N score (nitrogen or nutrients). The Ellenberg N score
25 indicates overall fertility or productivity (Hill and Carey, 1997).

1 However, none of these previous studies have focussed on relationships between
2 individual species and N, whilst at the same time accounting for other confounding factors such
3 as climate. As the majority of air pollution policy is focussed on setting appropriate emissions
4 ceilings and the consideration of critical load exceedance, evidence from analysing individual
5 species responses to nitrogen needs to be considered so that critical loads are set at levels to
6 protect the risk of species loss.

7 The main aim of this paper is to analyse vegetation surveillance datasets on a national
8 scale, in terms of individual species responses and Ellenberg N scores, and collate evidence of
9 possible N deposition impacts/relationships. This adds the growing evidence base and provokes
10 thought on potential implications for current conservation commitments and biodiversity targets.
11 These include the aim to maintain or improve the status of wild flora and fauna and their
12 ecosystems and habitats in response to the Convention for Biological Diversity (CBD) Articles 8
13 and 9.

14 **2. Methods**

15 **2.1 Data**

16 Data were taken from two national vegetation datasets to look at relationships with N
17 deposition for three semi-natural habitats: calcareous grassland; acid grassland; and heathland.
18 The distributions of these three habitats over Great Britain are shown in Figure 8. The databases
19 used were: The Vascular Plant Database, which recorded the presence of all vascular plant
20 species growing within 10km² hectads in the wild between 1987 and 1999 (Preston et al., 2002);
21 and The Botanical Society of the British Isles (BSBI) Local Change Survey data which records
22 all vascular plant species growing in the wild in 811 2 x 2 km tetrads located within a regular grid
23 of 10km² hectads across Great Britain. The BSBI survey was conducted in both 1987-88 and
24 2003-04 (Braithwaite et al., 2006), but for this analysis only data from 2003-04 was used. For the
25 purposes of this assessment, any recorded subspecies were aggregated to the species level, with

1 hybrids and alien species excluded, and pooled over time as the subsequent analysis uses only the
2 spatial distribution of the data to inform the model.

3 Vascular plants were individually assigned to each of the target habitat types using
4 published preference data on habitat affinity and altitudinal ranges (Hill et al., 2004). Only those
5 species selected using the criteria above were subsequently used to generate summary variables.
6 Thus generalist species and those restricted to other non-target habitats were excluded from the
7 calculation of mean Ellenberg N values for each recording unit (ie. tetrad or hectad).

8 Hectads (10 x 10 km squares) and Tetrads (2 x 2km square) were processed to determine
9 upland and lowland strata and to locate focal Broad Habitats. Broad Habitat occurrence was
10 derived from the Land Cover Map 2000 25 m raster coverage. The upland mask is composed of
11 the upland delineation for England based on English Natural Areas, a Welsh upland mask
12 provided by Countryside Council for Wales (CCW), and the combined upland and montane
13 masks from Scottish Natural Heritage (SNH).

14 To understand fully the potential effects of N on vascular plants we investigated
15 relationships between atmospheric N deposition and: individual species' responses
16 (presence/absence in each recording unit based on current distribution); and Ellenberg N scores
17 (Hill et al., 1999; Hill et al., 2007). This allowed us to identify species that are especially
18 vulnerable to N deposition and those which are not found in areas of high deposition, whilst
19 Ellenberg N scores allow us to assess the extent to which these different species patterns reflected
20 prior knowledge of their association with different levels of fertility. Thus we calculated average
21 Ellenberg N scores based on the presence of species within each habitat. N scores are on a scale
22 of one to nine, where one represents the most infertile habitats and nine represents fertile habitats.

23 **2.2 Driver variables**

24 Ecological drivers known to affect species occurrence were included in the model so that
25 the effects could be partialled out before looking into the relationship with total inorganic N

1 deposition. Additional variables included were: Intensity of land-use in each grid square;
2 minimum January temperature; maximum July temperature; total annual rainfall; and change in
3 sulphur deposition. Total annual rainfall was preferred to summer or winter rainfall due to the
4 high correlation between the seasons (>0.8) allowing the inclusion of only one term. For
5 completeness, total annual rainfall was therefore favoured. Intensity of land use was measured as
6 the proportion of arable plus improved grassland in the grid square as calculated from Land
7 Cover Map 2000. Climate data was based on long term averages from the Met Office covering
8 the period from 1980 to 2005. This long term average provided good representation of the spatial
9 variability in temperature, which was needed to capture the spatial differences in species
10 occupancy we modelled. Sulphur (S) deposition was the difference in model estimates for 1971
11 and 2005 as based on the FRAME model (Fine Resolution Ammonia Exchange) (Fournier et al.,
12 2000). For N deposition we used estimates at the 5 km² scale provided by CEH Edinburgh
13 (NEGTAP, 2001; Smith et al., 2000), calculated as the mean of the estimates for 1996, '97 and
14 '98 from the CBED model for deposition to moorland. We may perhaps expect plant response to
15 be a function of cumulative N deposition, however, data on this was not available, so current N
16 deposition was used as a proxy for cumulative deposition. If cumulative data were based on back-
17 casting from current deposition patterns (as in Dupre et al., 2010), we would expect to see very
18 similar results. Whenever the resolution of the botanical records was greater than that of the
19 driver variable data, the driver data were averaged over the size of the botanical recording unit.

20 **2.3 Modelling of response variables**

21 We used a generalised additive model (GAM) (Hastie and Tibshirani, 1990) based
22 approach to model both species presence/absence and Ellenberg N scores. This allowed the
23 effects of other covariates of interest to be modelled and accounted for before looking at the
24 relationship between the response variable and N. The flexible, non-linear, smooth terms used in
25 the GAM framework allow for the possibility that a species may be unaffected beneath a certain

1 level of N but showing signs of change beyond this threshold, thus conforming to the critical
2 loads definition.

3 Any spatial structure in the model residuals was accounted for in the GAM by including
4 an additional two-dimensional smooth term, i.e. a planar surface, in the model. This term is
5 defined as an interaction between the two coordinate axes. Therefore any spatial dependence
6 between hectads, missing spatial covariates or spatially structured recording effort is accounted
7 for as much as is possible. This, rather broad scale way of handling residual spatial structure was
8 assessed by repeating a subset of models in a Bayesian framework allowing for a complex,
9 localised, fine scale dependence structure to be specified. Results from such analyses confirmed
10 that the broader scale dependence method of the GAMs performed equally as well. Furthermore,
11 for the BSBI tetrad data a random effect term for hectad was also included to account for
12 differing levels of variability to data from within the same hectad to data from differing hectads.

13 Finally, to maintain an element of consistency across the analyses, all models were started
14 with the same set of terms and driver variables, which were included in the model in the same
15 order.

16 **2.4 Model Inference**

17 For both the hectad and tetrad models, regression parameters for any fixed linear terms
18 and parameters for each of the smooth terms were estimated, together with standard errors, by
19 either approximating the true likelihood or using quasi likelihood methods. Having fitted the
20 model with all terms, individual terms were selected based on F tests, the AIC score (Akaike,
21 1969) and the generalised cross validation score. Non significant terms were only removed if the
22 AIC and GCV scores improved upon their removal. The covariate corresponding to N was
23 always included in the model whether it was significant or not, as it was this relationship that
24 ultimately we were interested in. The final model was checked by examining plots of residuals
25 and quantile-quantile plots of the observed and predicted values.

1 To formally assess the impact of N, the p value from the F test associated with the N
2 covariate was returned along with a graphical plot of the estimated smooth term in the GAM for
3 N deposition. Both the p value and the plots together provide a full assessment of the modelled
4 relationship, including significance, direction and change points, that N has in relation to the
5 vegetation response variable.

6 Because of the complex modelling framework and the number of driver variables used in
7 building the model, data sets with a small number of observations could not be analysed and were
8 therefore removed. Also species with less than 10% and greater than 90% of presences in the data
9 set were also removed as convergence in the model was difficult to achieve.

10 **3 Results**

11 Results are presented separately for the analysis of the individual species' presence /
12 absence and the calculated Ellenberg N score. Furthermore, each of the three habitats: calcareous
13 grassland, acid grassland and heathland are listed individually for the species response analysis.
14 The number of hectads within each of the three habitats for the two recording schemes is shown
15 in Table 1. Histograms showing the ranges of N deposition covered within each habitat for which
16 data were analysed are given in Appendix A.

17 It is important to note that the low sample sizes for analysis are not solely because of too
18 few records in either recording scheme. It is the combination of sampling units coinciding with
19 grid cells with identifiable and hence larger areas of the focal habitat. i.e. the presence records for
20 species in cells where Land Cover Map did not pick up the habitat of interest were not included.

21 **3.1 Individual species response**

22 Removing the very rare and very common species from the analysis (those with an overall
23 presence of less than 10% and greater than 90% respectively) the presence/absence of the
24 remaining species occurring in the sampling units according to Table 1, was modelled as outlined

1 above. P values obtained from testing the null hypothesis of no association between
2 presence/absence and N deposition, having accounted for other driver variables, are given in
3 Appendix B for all species analysed. For species where a significant association with N
4 deposition was found, the relationships are plotted as probability of presence against increasing
5 total inorganic N deposition ($\text{kg N ha}^{-1} \text{ yr}^{-1}$). The critical load for each habitat, as defined by the
6 range midpoint of the published values by Bobbink and Hettelingh, 2011, is also included in
7 these plots.

8 **3.1.1 Acid grassland**

9 In the Vascular Plant Database there were sufficient data to test the probability of
10 presence with N deposition for eleven species (between 10 and 90% of hectads occupied) for
11 lowland acid grassland. Five of these species showed a significant relationship with N
12 deposition: *Cerastium arvense*, *Cerastium semidecandrum*, *Trifolium arvense*, *Vicia lathyroides*
13 and *Viola canina*. All five species showed similar trends (Fig. 4), with the probability of
14 presence declining rapidly with increasing deposition, including at levels below the critical load
15 of $15 \text{ kg N ha}^{-1} \text{ yr}^{-1}$, and then tailing off. Only one species was investigated for upland acid
16 grassland but did not show a significant response to N deposition.

17 In the BSBI data for lowland acid grasslands there was sufficient data on only one species
18 in the data set (Appendix B), *Senecio sylvaticus*, which did not show a significant relationship
19 with N deposition. For upland acid grassland again only one species had sufficient data for
20 analysis: *Agrostis vinealis*. *A. vinealis* showed a significant hump-backed response to N
21 deposition, increasing to $25 \text{ kg N ha}^{-1} \text{ yr}^{-1}$ then declining.

22 **3.1.2 Calcareous grassland**

23 In the Vascular Plant Database the presence / absence of 40 species in lowland calcareous
24 grassland were modelled. Of these, 17 showed significant relationships with N deposition
25 (Appendix B). Nine species showed negative relationships (Fig. 2a) with N deposition: *Allium*

1 *vineale*, *Anacamptis pyramidalis*, *Carlina vulgaris*, *Cynoglossum officinale*, *Echium vulgare*,
2 *Geranium columbinum*, *Ononis repens*, *Rosa micrantha* and *Spiranthes spiralis*. All of these
3 species showed changes in their probability of presence even at low levels of N deposition well
4 below the critical load of 20 kg N ha⁻¹ yr⁻¹. Four species showed positive relationships (Fig. 2b)
5 with N deposition: *Carex spicata*, *Knautia arvensis*, *Lathyrus nissolia* and *Stachys officinalis*. All
6 of which appeared to increase to around 25 kg N ha⁻¹ yr⁻¹ and then stabilise. There were also four
7 species (*Centaureum erythraea*, *Epipactis helleborine*, *Linum bienne* and *Pastinaca sativa*) that
8 showed hump-backed distributions (Fig. 2c) with the peak occurring around the critical load
9 level. For upland calcareous grasslands there were only sufficient data to analyse seven species.
10 Two species showed significant relationships with N deposition (Fig. 2d). *Alchemilla*
11 *xanthochlora* showed a significant increase with N deposition, whereas *Melica nutans* showed a
12 generally declining probability of presence with increasing N deposition. The slight increase at
13 low levels of deposition is due to the wide variation in the data at this point.

14 In the BSBI local change survey data there were sufficient records for analysis of 17
15 species within lowland calcareous grasslands. Eight showed a significant relationship with N
16 deposition (Fig. 3a): *Bromopsis erecta*, *Campanula glomerata*, *Carex spicata*, *Centaurea*
17 *scabiosa*, *Daucus carota*, *Ononis repens*, *Sanguisorba minor* and *Viola odorata*. Of these, a
18 number of species show changes of a small magnitude or complex responses but there are clear,
19 steep declines for a number of species including *B. erecta*, *C. glomerata*, *C. spicata* and *O.*
20 *repens*. There were 12 species analysed in upland calcareous grasslands of which three species
21 showed significant correlations with N deposition: *Persicaria vivipara*, *Rubus saxatilis* and
22 *Thalictrum alpinum*. All of these species showed hump-backed relationships with N deposition
23 (Fig. 3b).

24 **3.1.3 Heathland**

25 Within the Vascular Plant Database there was sufficient data for analysis of four species
26 in lowland heathland. Two species showed a significant relationship with N deposition (Fig. 5a).

1 *Platanthera bifolia* had a positive relationship with N deposition, with probability of presence
2 increasing steadily as N deposition increased. *Viola canina* had a negative relationship with N
3 deposition declining considerably in its probability of presence between 10 and 25 kg N ha⁻¹ yr⁻¹.
4 In upland heathlands it was possible to analyse ten species (Appendix B). Two species showed
5 significant relationships with N deposition. *Arctostaphylos uva-ursi* had a negative relationship
6 with N deposition declining in probability of presence up to 15 kg N ha⁻¹ yr⁻¹ and then remaining
7 at very low levels (Fig. 5b). *Trientalis europaea* showed a hump-backed distribution peaking at
8 approximately 18 kg N ha⁻¹ yr⁻¹.

9 In the BSBI Local Change survey data there were insufficient data to perform analysis on
10 any of the 16 potential lowland heathland species in the data set. However, in upland heathland
11 there were seven species for which analysis could be performed (Appendix B). Of these, three
12 showed significant relationships with N deposition: *A. vinealis*, *Listera cordata* and *Vaccinium*
13 *vitis-idaea* (Fig. 6). *A. vinealis* and *L. cordata* showed similar hump-backed responses with peak
14 probability of presence at around 20 kg N ha⁻¹ yr⁻¹, whereas *V. vitis-idaea* showed a clear decline
15 in response to N deposition.

16 **3.2 Ellenberg N**

17 Plots of modelled relationships between Ellenberg N score and total inorganic N
18 deposition for all habitats are shown in Fig. 7a for the Vascular Plant database and 7b for the
19 BSBI Local change survey data.

20 Using data from the Vascular Plant Database, average Ellenberg N scores showed a clear
21 increase with increasing N deposition (indicating an increase in fertility) for lowland acid
22 grasslands (p=0.0005). For upland acid grasslands, however, there was no significant (p=0.32)
23 relationship between Ellenberg N and N deposition. This result was echoed in the analysis of the
24 BSBI Local Change survey data, showing a clear increase in the lowlands (p=0.01), but no
25 significant result in the upland acid grasslands (p=0.42).

1 There were significant results for both lowland ($p=0.004$) and upland ($p=0.0062$)
2 calcareous grasslands in the analysis of Ellenberg N scores from the Vascular Plant Database
3 with both communities showing a significant increase in mean Ellenberg N with increasing N
4 deposition. Results from the BSBI Local Change survey data in upland calcareous grasslands
5 showed no significant relationship between N deposition and mean Ellenberg N score ($p=0.72$),
6 whereas lowland calcareous grassland showed a significant ($p=0.0074$) increase. Unlike the acid
7 grasslands analysis however this relationship is non-linear.

8 Analysis of the vascular plant database for both upland and lowland heathlands showed a
9 significant ($p=0.0059$ & $p=0.0549$ respectively) increase in Ellenberg N score. The analysis of
10 BSBI Local Change survey data for heathlands showed mixed results between upland and
11 lowland communities with no significant response for the analysis of upland heathland ($p=0.98$)
12 but a clear, significant ($p=0.0015$) increase in Ellenberg N with increasing deposition for lowland
13 heathland.

14 **4 Discussion**

15 **4.1 Individual Species responses**

16 **4.1.1 Acid Grassland**

17 For lowland acid grassland, consistent trends were found in the Vascular Plant Database
18 with five species displaying significant responses showing a very similar negative relationship,
19 declining in occupancy to low levels before stabilising c.25 kg N ha⁻¹ yr⁻¹. With the exception of
20 *Viola canina*, which the BSBI Local Change survey results (Braithwaite, Ellis and Preston, 2006)
21 suggested was declining due to its vulnerability to eutrophication, all are small, low-growing or
22 prostrate species most typical of dry acid grassland in the south and east of England. All have
23 low Ellenberg N values and thus are typical of infertile environments. Unfortunately, there were
24 no species with suitable coverage (between 10 and 90% occupancy) in the BSBI Local Change
25 survey data to test the effects of N deposition and to enable us to confirm whether there were
26 consistent trends among species within tetrads that contained lowland acid grassland. It is

1 difficult to draw conclusions for upland acid grassland as only *Agrostis vinealis* displayed a
2 significant (hump-backed) response reaching an optimum in occupancy at ca. 25 kg N ha⁻¹ yr⁻¹ in
3 the BSBI Local Change survey data.

4 **4.1.2 Calcareous Grassland**

5 Results for the lowland calcareous grassland analysis using the Vascular Plant data
6 showed that around half the species with significant responses showed negative trends in
7 occupancy in relation to increasing N deposition. The BSBI Local Change survey data also
8 showed a high proportion of negative trends, though this was generally for different species than
9 noted in the Vascular plant data. Although the two datasets had numerous species in common,
10 only one species, the legume *Ononis repens*, showed a significant negative response in both. For
11 most other cases when a particular species was recorded in both datasets, a significant result in
12 one was paired with a non-significant result in the other. This could be because although a
13 particular species was recorded in the two datasets, it may not correspond to records from the
14 same locations. This could have a large effect in a spatial modelling approach as adopted here.

15 It was notable that most species displaying clear negative trends (in either dataset) tended
16 to be specialists of grazed nutrient-poor calcareous grassland in Great Britain and generally
17 showed declines in the BSBI Local Change survey results (Braithwaite, Ellis and Preston, 2006)
18 e.g. *Bromopsis erecta*, *Campanula glomerata*, *Carlina vulgaris*, *Spiranthes spiralis*. These were
19 typically species with low Ellenberg N values. Species displaying more complex (hump-backed)
20 or positive trends in the Vascular Plant Database dataset are generally more typical of taller and
21 slightly more mesotrophic swards (e.g. *Epipactis helleborine*, *Knautia arvensis*, *Lathyrus*
22 *nissolia*, *Stachys officinalis*). Increased biomass production following increased N deposition
23 would therefore favour the latter species and may partly account for the reported declines of some
24 of the rarer species dependent on shorter swards. *Carex spicata*, which also showed a positive
25 relationship to N deposition, was found to have increased in the BSBI Local Change Survey

1 results (Braithwaite, Ellis and Preston, 2006), which suggested that increases were potentially a
2 result of increased availability in gaps of drought-prone grasslands and verges.

3 In upland calcareous grassland, *Melica nutans* was the only species to show a clear
4 negative relationship with N, whereas *Alchemilla xanthochlora* had greatest probability of
5 occurrence at the highest N deposition levels. *Melica nutans* was shown to decline in the BSBI
6 Local Change survey results (Braithwaite, Ellis and Preston, 2006), possibly due to competition.
7 *Alchemilla xanthochlora* was also shown to decline in the BSBI Local Change survey results
8 (Braithwaite, Ellis and Preston, 2006), possibly a result of eutrophication. These results found by
9 Braithwaite, Ellis and Preston, 2006 were obtained from pooling data over multiple habitats,
10 whereas the relationship we find is based solely on upland calcareous grasslands. It is typically a
11 species of unimproved hay meadows yet is capable of surviving in tall-herb communities on
12 roadsides. Three species displayed a hump-backed relationship to N deposition in the BSBI Local
13 Change survey data, with an optimum in occupancy between 10-20 kg N ha⁻¹ yr⁻¹ (*Persicaria*
14 *vivipara*, *Rubus saxatilis*, *Thalictrum alpinum*). Again these results are difficult to interpret but
15 may represent the localised distribution of these species to a narrow altitudinal and geographical
16 range in the uplands, and therefore a highly restricted N deposition gradient, or a potential
17 unmodelled interaction between N deposition and another covariate in the model.

18 **4.1.3 Heathland**

19 In the Vascular Plant Database, only two species displayed significant trends for lowland
20 heath, *Viola canina* decreasing and *Platanthera bifolia* increasing significantly in probability of
21 presence with increasing N deposition. *V. canina* is a species with a low Ellenberg N score and its
22 low stature means it does not compete well with competitive species. The latter result is very
23 surprising as *P. bifolia* is an uncommon plant of unimproved pastures and wet heathlands. It was
24 shown to be declining in the BSBI Local Change survey results (Braithwaite, Ellis and Preston,
25 2006), which they suggest was due to a long term response to management of moorland for
26 grouse. Having applied the habitat masks, there were no species in lowland heathland recorded in

1 the BSBI Local Change survey data that had sufficient coverage (between 10 and 90%
2 occupancy) to allow testing of the spatial effects of N deposition.

3 The results for upland heathland showed two sub-shrubs *Arctostaphylos uva-ursi*
4 (Vascular Plant Database) and *Vaccinium vitis-idaea* (BSBI Local Change survey data)
5 displaying clear negative trends in occupancy with increasing N deposition. Both of these
6 species have low Ellenberg N values. Two species associated with more humid upland
7 heathlands, *Trientalis europaeus* and *Listera cordata*, displayed distinct optima at around 20 kg
8 N ha⁻¹ yr⁻¹, possibly reflecting their restricted distributions along the N deposition gradient. Both
9 are elusive species. An alternative hypothesis for *Listera cordata* might be its dependence on
10 aerial deposition of nutrients for growth in nutrient-poor habitats. This species typically roots
11 within *Sphagnum* tussocks and therefore attains most of its nutrients from decaying matter or
12 rainwater.

13 **4.2 Ellenberg N**

14 There was a general trend for higher Ellenberg N scores in areas of high N deposition
15 across many habitats indicating higher fertility rates. The most prominent changes occurred in
16 lowland acid grasslands, where Ellenberg N showed this relationship in both data sets. However,
17 the magnitude of these differences across the N deposition gradient were generally quite small -
18 differences of approximately one unit on the Ellenberg N scale between deposition of 5 and 30 kg
19 N ha⁻¹ yr⁻¹. It is somewhat surprising that lowland grasslands appear to show greater differences
20 than upland ones. Due to their tendency to have soils with a lower nutrient status, we might
21 expect upland grasslands to be more sensitive. Although this is dependent on the species present
22 being capable of exploiting any excess N. For upland acid grasslands the lack of a convincing
23 Ellenberg N score response to N deposition is not entirely surprising as this has also been found
24 in research scale spatial surveys which have suggested acidification from N deposition may be

1 the main driver of reductions in species richness rather than eutrophication (Maskell et al., 2010;
2 Stevens et al., 2010).

3 The two datasets generally show consistent results between the habitats with only slight
4 differences in the shape of the response, whilst the general pattern and modelled scores are
5 broadly aligned.

6 **4.3 Ecosystem effects**

7 Whilst deletions of rare species and additions of generalist species constitute changes in
8 conservation value, impacts on ecosystem properties can also result. These impacts can often be
9 understood in terms of the mass effect of dominant species in controlling productivity and litter
10 quality, coupled with the functional diversity effect of subordinate species (Grime 1998; Smith &
11 Knapp 2003; Laughlin 2011). Because the scale of our observational data was occupancy of large
12 grid cells (10x10km and 2x2km) common generalists and many ecosystem dominants were not
13 examined since they are either not specialised to those focal habitats vulnerable to nitrogen
14 deposition or were likely to be present in many grid squares and so insensitive to the nitrogen
15 deposition gradient. However, a number of the species for which we found significant
16 relationships are either habitat dominants or possess traits known to be influential in controlling
17 ecosystem function. For example three leguminous N-fixers showed negative responses to the
18 nitrogen gradient. This could indicate a loss of capacity to fix N and sequester carbon, yet two of
19 the three species, *Vicia lathyroides* and *Trifolium arvense*, are not common and, where they do
20 occur, occupy subordinate positions in the dominance hierarchies of each grassland type.
21 Community dominants include *Tricophorum cespitosum* (upland heath), *Vaccinium vitis-idaea*
22 (upland heath) and *Bromopsis erecta* (lowland calcareous grassland). Their negative relationships
23 with N deposition imply a potentially more significant loss of function. This situation is more
24 likely for populations at their range edge.

1 Most of the species for which significant effects were found are of restricted distribution
2 in Britain and at low biomass even within their favoured habitat. On this basis, their changing
3 abundance may have had less direct impact on ecosystem function than correlated increases in
4 biomass of species that may have responded positively to increased nutrient supply. The impact
5 of increasing species is likely to be significant both in terms of their role in competitive
6 suppression of shorter, stress-tolerators and in driving influential changes in ecosystem
7 productivity and nutrient cycling. The scale of the study makes it difficult to speculate further on
8 ecosystem impacts since losses or gains from survey squares are likely to exhibit large lag effects
9 relative to influential changes in biomass. This also means that biomass-linked changes in
10 function may well have either significantly pre-dated loss from the grid cell or, for additions,
11 should follow subsequently if colonisation precedes biomass increase. Because of the subordinate
12 position of many of these species in their favoured habitat, it may be that ecosystem impacts have
13 been or will be relatively minor and that observed species changes are most significant in terms
14 of loss of the cultural ecosystem service linked to the conservation of wild species diversity.
15 However, recent work has helped elucidate the supporting role of a diverse species pool in
16 maintaining ecosystem function (Isebell et al 2011). We cannot therefore, rule out the possibility
17 of reduced resilience following species deletions among the ecosystems studied here.

18

19 **4.4 Conclusions**

20 This study set out to evaluate the response of individual vascular plant species to N
21 deposition and assess how the shape and direction of any relationships relate to current critical
22 load levels. After accounting for a number of additional explanatory variables, we have found
23 evidence that a number of species show clear responses to N even at low levels. Although the
24 models allow for the possibility of no change up to a threshold after which we see a decline, there
25 are few instances where this is indeed the case. Where we do see significant relationships to N

1 deposition, there are many changes that occur well below the critical load, with Ellenberg N
2 values increasing from the lowest N levels and some species showing strong responses at low
3 deposition. This highlights the possibility that current European empirical critical loads for N do
4 not protect all species and this gives concern in respect of meeting nature, conservation and
5 biodiversity targets of the UK and Europe. Ongoing effects are also found above critical load
6 values indicating there may be benefits from reductions in deposition however small and at any
7 level.

8 Within all three habitats, there are a number of species with low Ellenberg N values, which may
9 have been expected to respond negatively to N deposition but do not show a significant
10 relationship with N deposition here. This may be for a number of reasons as this analysis is based
11 on data using large recording units, there may be interactions with climate not detected and
12 presence absence data may be too coarse and insensitive to change. Abundance of individual
13 species would be more sensitive to change but targeted surveys are needed to collect such data.
14 The main benefits of using a large scale study with comprehensive nationwide coverage, as done
15 here, are that we have good coverage over the N deposition gradient with which to build our
16 models, any inference drawn is not spatially restricted and there is no chance that we only model
17 an unusual niche of a given species.

18 However, using large scale data and large recording units meant that the assignment of
19 species to a habitat was at a coarse scale and may not be as accurate as smaller scale data and
20 smaller recording units such as quadrats, which contain species data from known habitat types. It
21 is also important to note that the analysis presented did not extend to species which were too rare
22 or too abundant. We were therefore only working with the middle range of species and
23 conclusions are based on modelling these species. It may be, however, that the rare or abundant
24 species show more sensitive changes or differing relationships to those presented here. This may
25 be an avenue of further research, as would be the investigation of other drivers variables and the
26 possibility of including interaction terms to look at 2 dimensional species responses.

1 To ensure that relationships seen here did not reflect a potential correlation between N
2 deposition and habitat area, the theory being that the larger the habitat area the more likely we are
3 to find a record, some selected species were re-modelled with habitat area included as an
4 additional covariate in the model. However, no differences to the original analysis were found
5 and the plotted relationships remained unchanged.

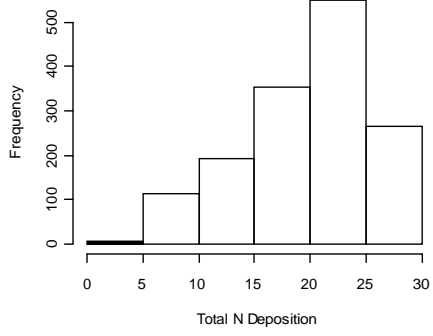
6 Our work has sought to separate out the effect of N deposition from other variables
7 known to affect the responses of interest. To this end we included in the model all extra variables
8 we thought ecologically meaningful and for which data were available. We also included a
9 generic spatial surface and used a flexible class of models. All of which mean we have made less
10 assumptions, are less open to bias and have accounted for more possible confounding issues than
11 any other similar study. We therefore see this as a key contribution to a growing body of
12 scientific evidence suggesting species level impacts in all three habitats investigated and potential
13 effects at low deposition values beneath current critical load levels.

14 **Appendix: A**

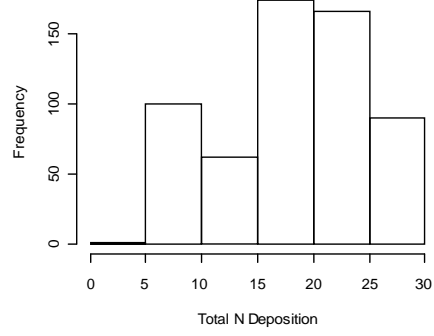
15 Histograms showing the distribution of N deposition in the Vascular plant database between 0 –

16 $30 \text{ kg N ha}^{-1} \text{ yr}^{-1}$

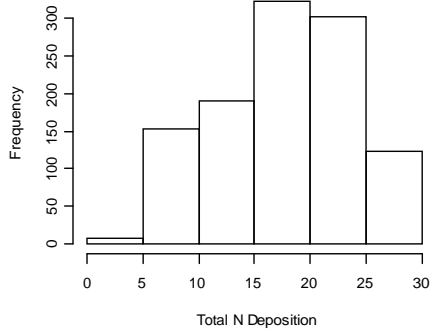
Lowland Calcareous Grassland



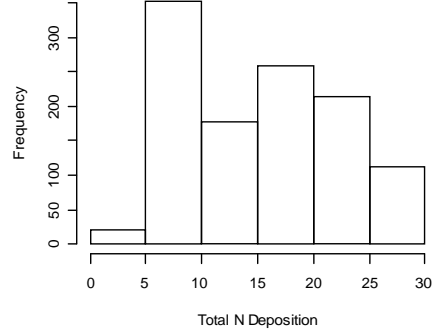
Upland Calcareous Grassland



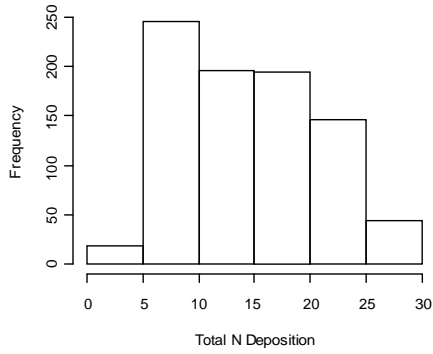
Lowland Acid Grassland



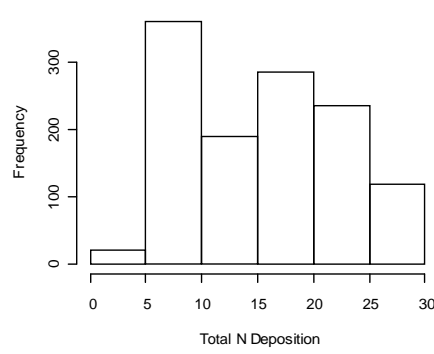
Upland Acid Grassland



Lowland Heathland



Upland Heathland



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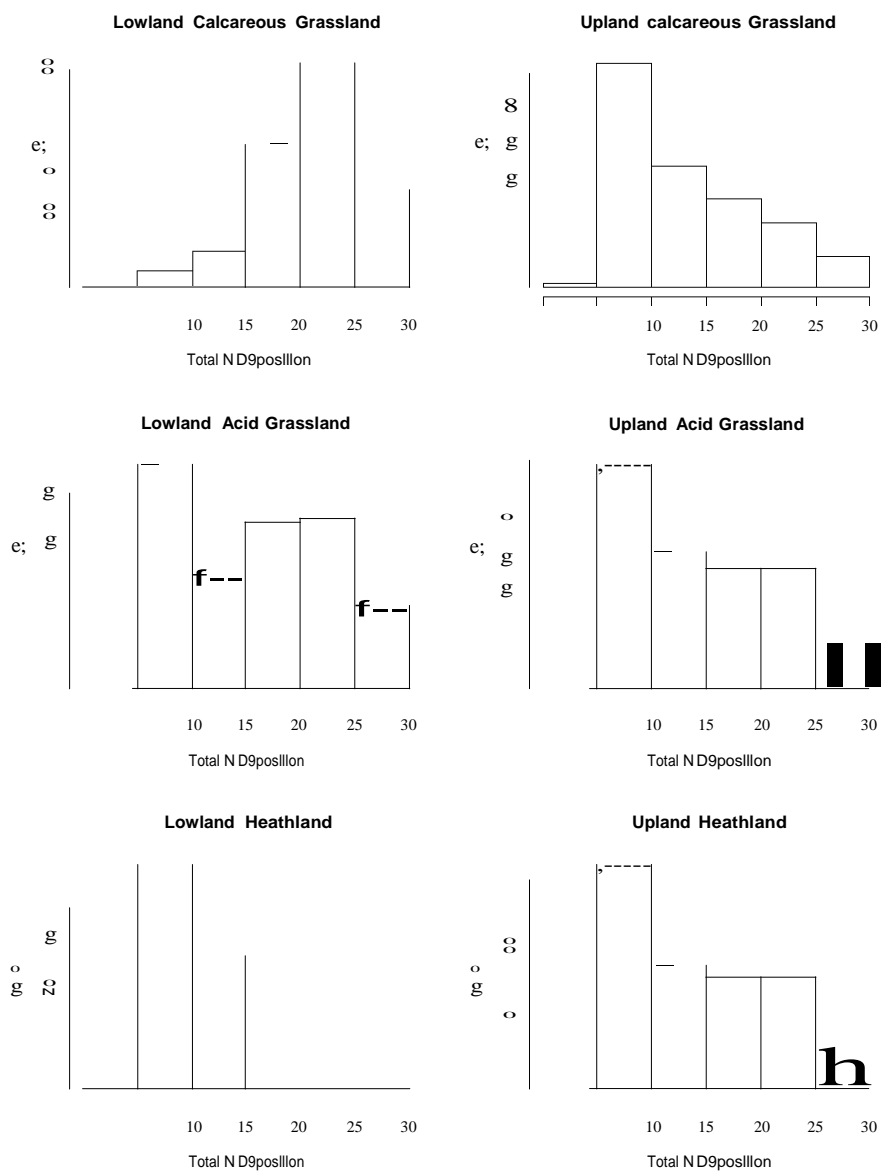
4

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1 Histograms showing the distribution of N deposition in the BSBI local change survey data
 2 between 0 – 30 kg N ha⁻¹ yr⁻¹.



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1 **Appendix B:**

2 Full table of results showing the p values testing the null hypothesis that there is no relationship
 3 between species occurrence and N deposition. Significant results are highlighted.

4 **Vascular Plant Database**

Habitat	Species	P Value	Change
Upland Acid Grassland			
	<i>Viola lutea</i>	0.4131	
Upland Calcareous Grassland			
	<i>Alchemilla xanthochlora</i>	0.0406	Positive
	<i>Galium boreale</i>	0.2170	
	<i>Melica nutans</i>	0.0016	Negative
	<i>Persicaria vivipara</i>	0.3801	
	<i>Rubus saxatilis</i>	0.4809	
	<i>Thalictrum alpinum</i>	0.9106	
	<i>Viola lutea</i>	0.3051	
Upland Heath			
	<i>Arctostaphylos uva-ursi</i>	0.0001	Negative
	<i>Cornus suecica</i>	0.3167	
	<i>Cryptogramma crispa</i>	0.5060	
	<i>Diphasiastrum complanatum</i>	0.9520	
	<i>Empetrum nigrum</i>	0.2198	
	<i>Lycopodium clavatum</i>	0.3648	
	<i>Rubus chamaemorus</i>	0.7752	
	<i>Trichophorum cespitosum</i>	0.6715	
	<i>Trientalis europaea</i>	0.0237	Hump-Back
	<i>Vaccinium vitis-idaea</i>	0.0919	
Lowland Acid Grassland			
	<i>Anthriscus caucalis</i>	0.7252	
	<i>Aphanes australis</i>	0.2999	
	<i>Cerastium arvense</i>	0.0212	Negative
	<i>Cerastium semidecandrum</i>	0.0022	Negative
	<i>Myosotis ramosissima</i>	0.0990	

<i>Ornithopus perpusillus</i>	0.6780	
<i>Scleranthus annuus</i>	0.2995	
<i>Senecio sylvaticus</i>	0.9415	
<i>Trifolium arvense</i>	0.0344	Negative
<i>Vicia lathyroides</i>	0.0457	Negative
<i>Viola canina</i>	0.0000	Negative

Lowland Calcareous Grassland

<i>Allium vineale</i>	0.0365	Negative
<i>Anacamptis pyramidalis</i>	0.0507	
<i>Blackstonia perfoliata</i>	0.0571	
<i>Brachypodium pinnatum</i>	0.2820	
<i>Bromopsis erecta</i>	0.8306	
<i>Carex spicata</i>	0.0478	Positive
<i>Carlina vulgaris</i>	0.0226	Negative
<i>Catapodium rigidum</i>	0.7550	
<i>Centaureum erythraea</i>	0.0087	Hump-Back
<i>Centaurea scabiosa</i>	0.8592	
<i>Cirsium acaule</i>	0.5282	
<i>Cirsium eriophorum</i>	0.0905	
<i>Clinopodium vulgare</i>	0.2215	
<i>Cynoglossum officinale</i>	0.0000	Negative
<i>Daucus carota</i>	0.2341	
<i>Echium vulgare</i>	0.0070	Negative
<i>Epipactis helleborine</i>	0.0280	Hump-Back
<i>Filago vulgaris</i>	0.1019	
<i>Geranium columbinum</i>	0.0176	Negative
<i>Geranium sanguineum</i>	0.1388	
<i>Hypericum hirsutum</i>	0.1314	
<i>Inula conyzae</i>	0.0880	
<i>Knautia arvensis</i>	0.0014	Positive
<i>Lathyrus nissolia</i>	0.0030	Positive
<i>Linum bienne</i>	0.0306	Hump-Back
<i>Ononis repens</i>	0.0109	Negative

<i>Ononis spinosa</i>	0.6074	
<i>Ophrys apifera</i>	0.1344	
<i>Orchis morio</i>	0.2136	
<i>Origanum vulgare</i>	0.2222	
<i>Pastinaca sativa</i>	0.0154	Hump-Back
<i>Picris hieracioides</i>	0.4054	
<i>Poa angustifolia</i>	0.9579	
<i>Rosa micrantha</i>	0.0105	Negative
<i>Rosa rubiginosa</i>	0.3157	
<i>Sanguisorba minor</i>	0.1927	
<i>Sherardia arvensis</i>	0.3197	
<i>Spiranthes spiralis</i>	0.0074	Negative
<i>Stachys officinalis</i>	0.0186	Positive
<i>Viola odorata</i>	0.3596	

Lowland Heath

<i>Platanthera bifolia</i>	0.0081	Positive
<i>Radiola linoides</i>	0.0527	
<i>Scleranthus annuus</i>	0.6258	
<i>Viola canina</i>	0.0008	Negative

1

2 BSBI Local Change Survey Data

Habitat	Species	P Value	Change
Upland Acid Grassland			
	<i>Agrostis vinealis</i>	0.0060	Hump-Back
Upland Calcareous Grassland			
	<i>Alchemilla alpina</i>	0.2052	
	<i>Galium boreale</i>	0.7008	
	<i>Persicaria vivipara</i>	0.0001	Hump-Back
	<i>Rubus saxatilis</i>	0.0321	Hump-Back
	<i>Saxifraga aizoides</i>	0.2127	
	<i>Saxifraga oppositifolia</i>	0.1236	
	<i>Selaginella selaginoides</i>	0.0605	

	<i>Thalictrum alpinum</i>	0.0231	Hump-Back
Upland Heath			
	<i>Agrostis vinealis</i>	0.0206	Hump-Back
	<i>Arctostaphylos uva-ursi</i>	0.1448	
	<i>Listera cordata</i>	0.0001	Hump-Back
	<i>Lycopodium clavatum</i>	0.0880	
	<i>Rubus chamaemorus</i>	0.0707	
	<i>Trichophorum cespitosum</i>	0.0701	
	<i>Vaccinium vitis-idaea</i>	0.0074	Negative
Lowland Acid Grassland			
	<i>Senecio sylvaticus</i>	0.1240	
Lowland Calcareous Grassland			
	<i>Brachypodium pinnatum</i>	0.4144	
	<i>Bromopsis erecta</i>	0.0052	Negative
	<i>Campanula glomerata</i>	0.0000	Negative
	<i>Carex spicata</i>	0.0183	Negative
	<i>Catapodium rigidum</i>	0.1609	
	<i>Centaurea scabiosa</i>	0.1047	
	<i>Centaureum erythraea</i>	0.0000	Negative
	<i>Crepis capillaris</i>	0.2595	
	<i>Daucus carota</i>	0.0218	Negative
	<i>Knautia arvensis</i>	0.1527	
	<i>Ononis repens</i>	0.0021	Negative
	<i>Origanum vulgare</i>	0.5387	
	<i>Pastinaca sativa</i>	0.9980	
	<i>Salvia verbenaca</i>	0.4535	
	<i>Sanguisorba minor</i>	0.0000	Negative
	<i>Stachys officinalis</i>	0.0962	
	<i>Viola odorata</i>	0.0251	Hump-Back

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4 BSBI and Biological Records Centre for providing access to their database and to all the
5 botanical recorders who have collected these data over many years. We are also grateful to CEH
6 Edinburgh for providing N and S deposition data.

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1 **Table 1:** The number of sampling units (hectads or tetrads) across Great Britain containing each of the three studied
2 habitats from each of the two surveys.

	Habitat	Lowland	Upland
Vascular Plant Database	Acid grassland	1143	1203
	Calcareous grassland	1546	662
	Heathland	855	1284
BSBI Local Change survey data	Acid grasslands	164	158
	Calcareous grasslands	377	209
	Heathland	43	231

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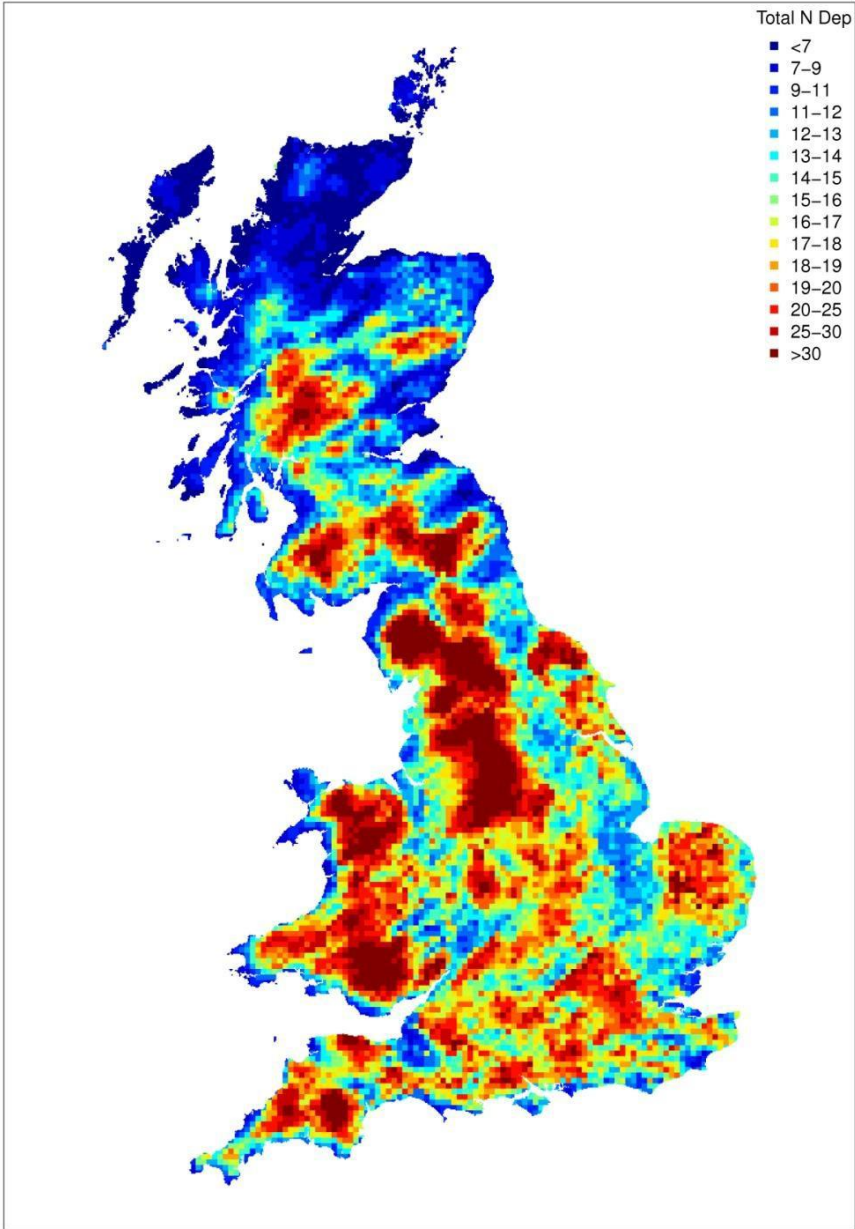
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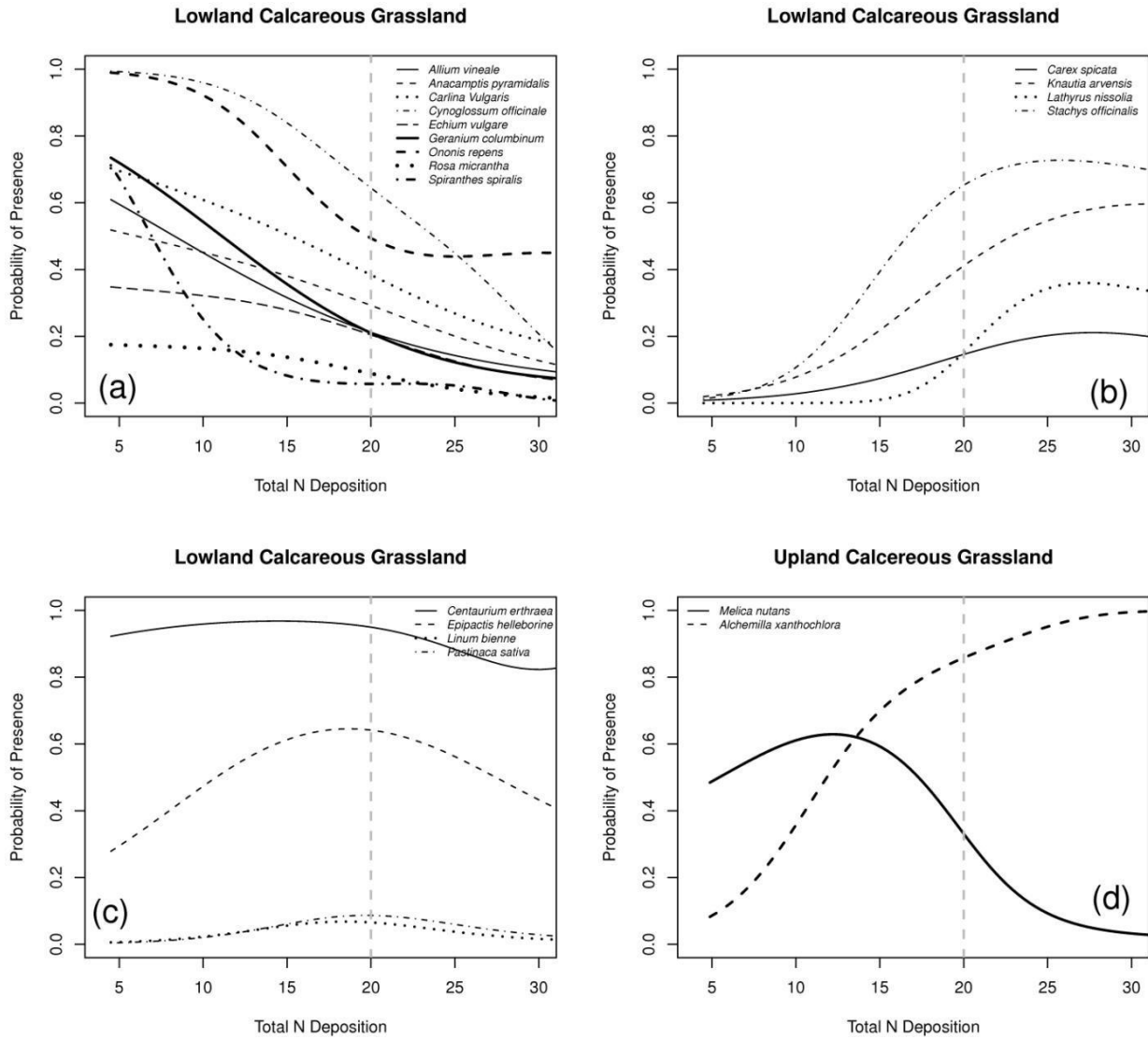
2 **Figure 1:** Total inorganic N deposition ($\text{kg N ha}^{-1} \text{ yr}^{-1}$) to heathland and rough grazing land in Great Britain. CBED
3 deposition data for 1996 – 98.



4

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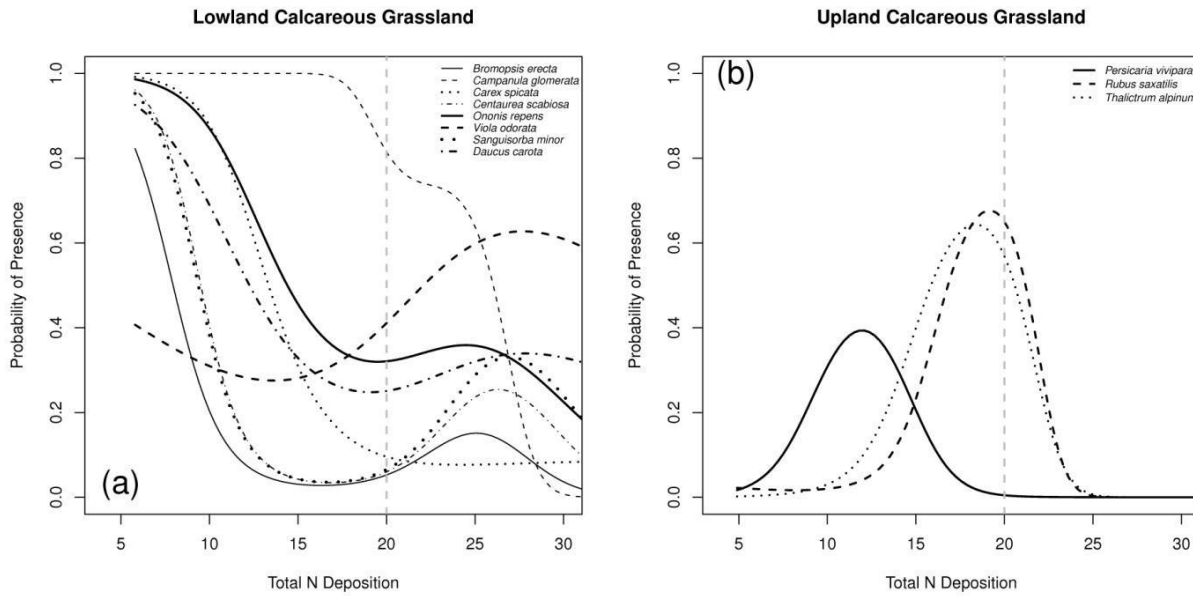
1 **Figure 2:** Change in probability of presence for species showing statistically significant relationships to N
 2 deposition in calcareous grasslands in the Vascular Plant database against increasing total inorganic N
 3 deposition ($\text{kg N ha}^{-1} \text{ yr}^{-1}$). The dashed grey line indicates the midpoint of the critical load range for calcareous
 4 grasslands as defined by Bobbink and Hettelingh, 2011.



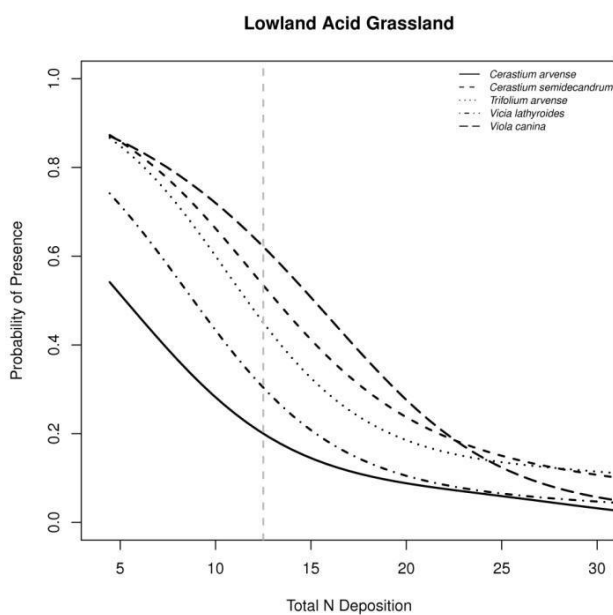
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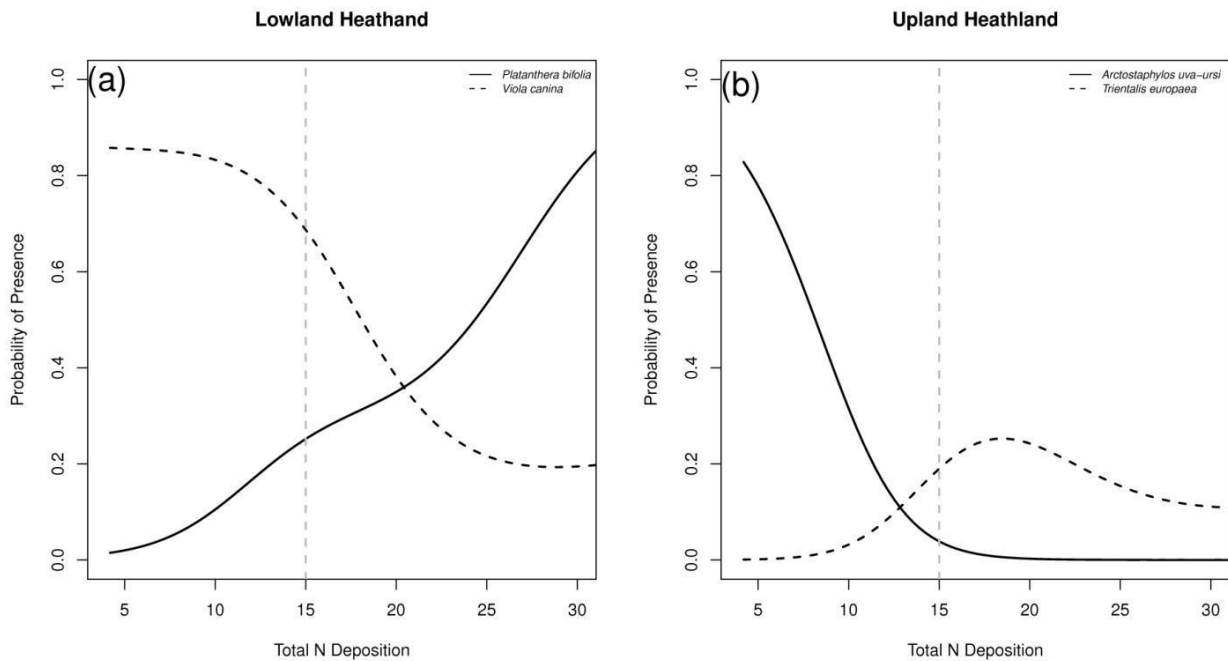
1 **Figure 3:** Change in probability of presence for species showing statistically significant relationships to N
 2 deposition in calcareous grasslands in the BSBI local change survey data against increasing total inorganic N
 3 deposition ($\text{kg N ha}^{-1} \text{ yr}^{-1}$). The dashed grey line indicates the midpoint of the critical load range for calcareous
 4 grasslands as defined by Bobbink and Hettelingh, 2011.



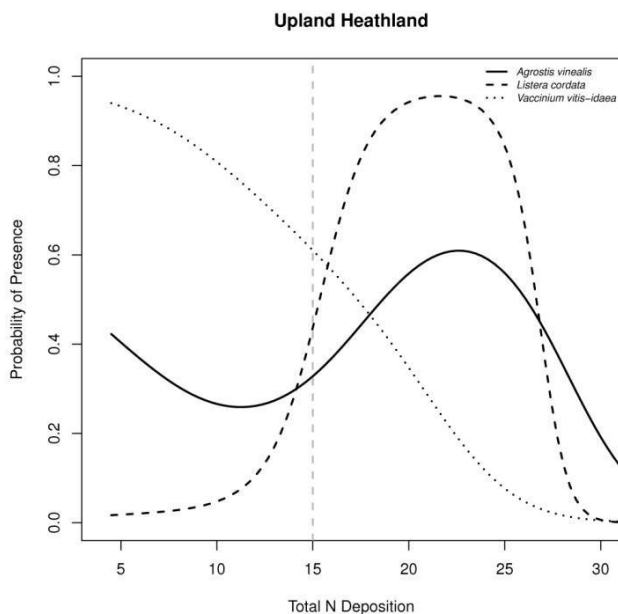
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 6 **Figure 4:** Change in probability of presence for species showing statistically significant relationships to N deposition
 7 in lowland acid grasslands in the Vascular Plant database against increasing total inorganic N deposition (kg N ha^{-1}
 8 yr^{-1}). The dashed grey line indicates the midpoint of the critical load range for acid grasslands as defined by Bobbink
 9 and Hettelingh, 2011.



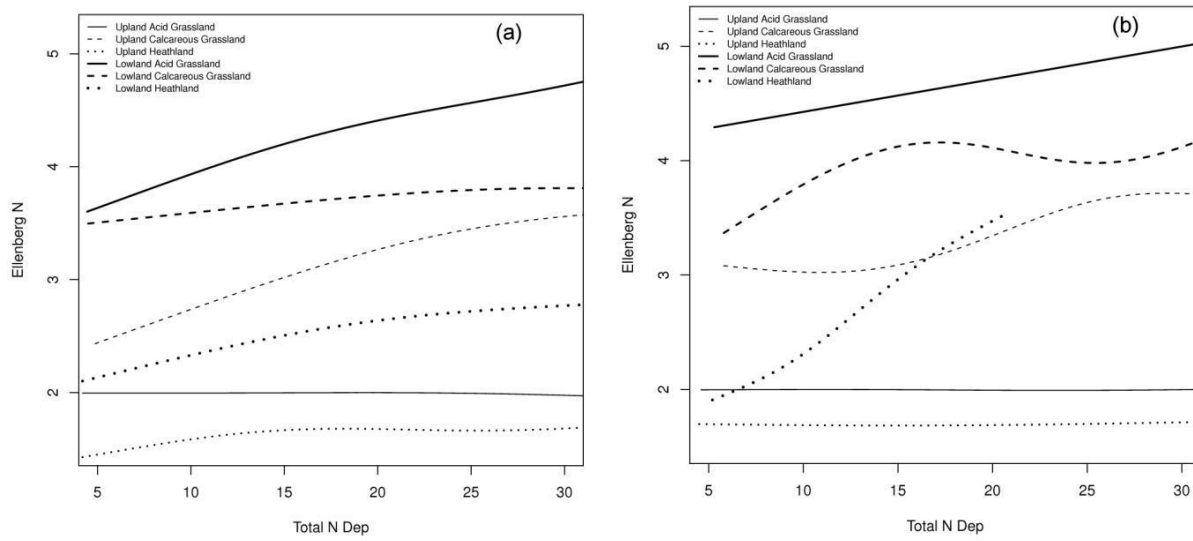
1 **Figure 5:** Change in probability of presence for species showing statistically significant relationships to N deposition
 2 in heathland in the Vascular Plant database against increasing total inorganic N deposition ($\text{kg N ha}^{-1} \text{yr}^{-1}$). The
 3 dashed grey line indicates the midpoint of the critical load range for heathlands as defined by Bobbink and
 4 Hettelingh, 2011.



5
 6 **Figure 6:** Change in probability of presence for species showing statistically significant relationships to N deposition
 7 in upland heathland in the BSBI local change survey data against increasing total inorganic N deposition (kg N ha^{-1}
 8 yr^{-1}). The dashed grey line indicates the midpoint of the critical load range for heathlands as defined by Bobbink and
 9 Hettelingh, 2011.



1 **Figure 7:** Modelled response of Ellenberg N scores against total inorganic N deposition ($\text{kg N ha}^{-1} \text{ yr.}^{-1}$) using data
 2 from (a) the Vascular Plant database and (b) the BSBI local change survey.



3
 4 **Figure 8:** Distribution of three semi-natural habitats: calcareous grassland; acid grassland; and heathland over Great
 5 Britain according to Land Cover Map 2000. Shown and mapped at 1km^2 resolution when habitat was dominant type
 6 in the 1km square.

