1 Method Variables Affecting the Bioaccessibility of Arsenic in Soil

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3	Louise Meunier ¹ , Joanna Wragg ² , Iris Koch ¹ , and Kenneth J. Reimer ¹ *,
4	1. Environmental Sciences Group, Royal Military College, P.O. Box 17 000, Station
5	Forces, Kingston, Ontario K7K7B4, Canada
6	2. British Geological Survey, Keyworth, Nottingham NG12 5GG, United Kingdom
7	* Corresponding author phone: (613)541-6000, ext 6161; fax: (613)541-6596; e-mail:
8	reimer-k@rmc.ca.
9	
10	ABSTRACT
11	
12	Arsenic bioaccessibility tests are now being commonly used in risk assessment. However,
13	concerns remain about the reliability of such tests because the bioaccessibility of arsenic
14	from soil may be susceptible to soil composition (including iron concentration), as well as
15	method considerations such as varying liquid-to-solid ratios and the chosen buffer system.
16	In this study, arsenic-contaminated tailings and soils were tested to compare two
17	bioaccessibility methods: one that uses glycine as a buffer, and a second that is more
18	physiologically based. With the glycine-buffered method, arsenic and iron bioaccessibility
19	increased in the presence of a higher buffer concentration at higher liquid-to-solid ratios,
20	whereas the results of physiologically-based tests were unaffected by variations in these
21	parameters. In the glycine-buffered system, interactions between iron and glycine may
22	influence the concentration of arsenic in solution, which may not be consistent with human

1	gastrointestinal conditions. The choice of a physiologically-based method may be more
2	appropriate to achieve representative arsenic bioaccessibility values toward estimating
3	risks to human health.
4	
5	Keywords: bioaccessibility; arsenic; glycine; contaminated soil; tailings
6	
7	INTRODUCTION
8	
9	Traditional human health risk estimates are based on the total concentration of
10	contaminants in soils, ^[1, 2] but several regulatory bodies worldwide now recognize that this
11	may overestimate the actual risk. ^[3-5] Instead, using the bioavailable fraction, or the fraction
12	of a substance that can be absorbed and reach systemic circulation, ^[6] may provide a more
13	representative estimation. This is especially important when the bioavailable fraction is
14	potentially much smaller than the total concentration of a contaminant. Such is the case for
15	arsenic in soils, which is considered only one-fifth as bioavailable as soluble (aqueous)
16	inorganic arsenic. ^[7]
17	
18	Arsenic bioavailability data may be obtained by measuring arsenic concentrations in the
19	blood and tissues of animals following their ingestion of arsenic-contaminated substrates.
20	Several in vivo ingestion studies have indicated that arsenic soil bioavailability depends on
21	soil type and contaminant source, and varies from 2 to 48%. ^[8-12] Because bioavailability

22 varies greatly, site and soil-specific measurements are required.

2	Although the use of in vivo models may provide representative bioavailability data for
3	estimating risks, the costs and ethics of using these models has led to the development of
4	laboratory-based in vitro extraction tests that simulate human in vivo conditions. Results
5	from these tests provide an estimation of bioaccessibility, which is defined as the fraction
6	of a substance that is soluble in the gastrointestinal environment and is available for
7	absorption. ^[13] Several methods have been developed to assess the bioaccessibility of soil
8	contaminants but the results vary between methods, and concerns remain regarding the
9	reliability of such measurements.
10	
11	Ruby et al. ^[11, 14] published one of the first studies examining a physiologically-based
12	extraction test (PBET). This test was based on rabbit gastrointestinal conditions and is
13	similar to paediatric conditions. It takes into consideration gastric and intestinal chemistry,
14	extraction pH, soil mass and fluid volume, stomach mixing and emptying rate, and small
15	intestinal transit time. This test is considered to be a valid predictor of oral lead
16	bioaccessibility. ^[15, 16] Rodriguez et al. ^[17] indicated that the arsenic bioaccessibility data
17	obtained by such an in vitro gastric method is strongly correlated with in vivo
18	bioavailability results, although the test slightly underestimated bioavailability. Several
19	additional studies have used variations of this physiologically-based method to measure the
20	bioaccessibility of metals, or to simplify the method. ^[17-19]
21	

1	Parallel method development has been carried out by various agencies including the
2	Solubility/Bioavailability Research Consortium (SBRC), which has established a simple in
3	vitro test using glycine as a buffer. ^[20] The SBRC calls for 0.4 M glycine, which is a much
4	higher concentration than the < 1 mM concentrations of glycine typically found in
5	biological fluids. ^[21] Although not physiologically representative, the gastric phase of the
6	SBRC method – also entitled the Relative Bioaccessibility Leaching Procedure – has been
7	validated against in vivo lead bioavailability data. ^[22] The USEPA recommends this method
8	to determine the relative bioavailability of lead for the purposes of risk assessment in
9	Region 8 (Colorado Area). ^[23] The in vivo database for arsenic is not as extensive as that
10	for lead, ^[20] and a glycine-buffered test has yet to be validated for estimating arsenic
11	bioavailability. ^[24]

13 In addition to the inherent variability introduced by any chosen method and their related 14 parameters (e.g. extraction pH, liquid-to-solid ratios), the bioaccessibility of arsenic may be affected by arsenic speciation^[25] and soil physicochemical properties. Specifically, 15 16 higher soil iron oxide content and lower soil pH both tend to yield lower bioaccessibility.^[26-28] In some cases, the bioaccessibility is negatively correlated with 17 arsenic concentration in soil,^[25, 29] but other studies have found the bioaccessibility of 18 arsenic to be independent of soil arsenic concentration.^[19] The negative dependence of 19 20 bioaccessibility on soil arsenic concentration may be an indication of solubility saturation, and Richardson et al.^[30] recommend carrying out tests at various liquid-to-solid ratios to 21 rule out this possibility. In some cases (where glycine was used as a buffer), the arsenic 22

1	bioaccessibility increased with an increased ratio, ^[31] whereas other studies (that did not use
2	glycine) have shown arsenic bioaccessibility to be independent of liquid-to-solid ratio. ^[25]
3	Thus the choice of buffer in estimating arsenic bioaccessibility may influence the results.
4	
5	In this study, arsenic-contaminated tailings, soils and reference materials representing a
6	range of iron to arsenic molar ratios were subjected to bioaccessibility tests using both a
7	physiologically-based test (Method P) and a glycine-buffered test (Method G) under
8	simulated gastric and gastric followed by intestinal conditions. The results were examined
9	to determine the effects of extraction pH, liquid-to-solid ratio, buffer choice and
10	concentration, and arsenic and iron concentrations on the bioaccessibility of arsenic. The
11	objective was to evaluate the variability introduced by the methods and sample
12	characteristics in measuring the bioaccessibility of arsenic.
13	
14	MATERIALS AND METHODS
15	
16	Sample Description and Preparation
17	
18	The test materials consisted of soils and tailings samples, reference materials, and solid
19	matrices. The soils and tailings samples were taken from arsenic-contaminated abandoned
20	gold mine sites in Nova Scotia, Canada, ^[32] and are described by Meunier et al. ^[25] The
21	standard reference materials selected were NIST2710 ^[33] and NIST2711 ^[34] , and a reference
22	soil from the British Geological Survey (BGS 102) was also tested. ^[35] Two more samples

1	consisted primarily of arsenic minerals: a mine tailings sample rich in scorodite
2	(FeAsO ₄ •2H ₂ O) collected at Lower Seal Harbour (LSH), Nova Scotia, Canada, ^[32] and a
3	sample of arsenopyrite (FeAsS) from Niñas de Panasqueira, Portugal (Ward's Natural
4	Science). Ironite® (Mineral Supplement 1-0-0), a commercial fertilizer that contains high
5	concentrations of arsenic and iron from mine waste was also tested. All tailings samples
6	were dried and sieved to a $<150 \ \mu m$ particle size. The arsenopyrite and LSH tailings
7	samples were ground and homogenized, using a mortar and pestle, to a ${<}150~\mu\text{m}$ particle
8	size. Ironite® was sieved to $<150 \ \mu m$. Reference materials and soil were used as received
9	(the particle size for NIST2710 and NIST2711 is <74 μ m, and the BGS102 particle size is
10	<50µm).
11	
12	Sample Elemental Composition

14 The certified arsenic and iron concentrations of the standard reference materials NIST2710 and 2711 were used.^[33, 34] All other samples were analyzed for major and trace elements by 15 16 aqua regia digestion. Analytical results for the Nova Scotia samples are described in detail by Parsons et al.^[32] Digestions of the remaining samples were performed according to 17 Method 200.7^[36] and analyzed by the Analytical Services Unit, Queen's university, 18 19 Kingston, ON using inductively coupled plasma-optical emission spectrometry (Varian AX-Vista Pro). The results for the total arsenic concentration in BGS102 (94 mg•kg⁻¹) is 20 comparable to unpublished data for this sample $(102 \text{ mg} \cdot \text{kg}^{-1})$.^[37] 21 22

1 Bioaccessibility Extractions

2

3	For each sample, the bioaccessibility of arsenic and iron was measured using two different
4	methods and two liquid-to-solid ratios (i.e. 100:1 and 2000:1). The glycine-buffered
5	Method G was based on a SBRC method modified from Kelley et al. ^[20] The
6	physiologically-based Method P was modified from Ruby et al. ^[11] and Rodriguez et al., ^[17]
7	and is similar to tests carried out by Cave et al. ^[38] and Koch et al. ^[39] These in vitro
8	experiments consisted of a gastric phase (phase 1, denoted P1), and a gastric followed by
9	an intestinal phase (phase 2, denoted P2).
10	
11	Experiments were carried out in batches of 8 to 12 samples, along with two duplicate pairs,
12	one blank and one reference material (NIST2710 or NIST2711) for each phase. All
13	chemicals and reagents used in the bioaccessibility tests were analytical grade or better,
14	and all solutions were prepared with distilled deionised water (DDW, Barnstead E-pure
15	reverse osmosis/ion exchange apparatus Water Purification System, minimum resistance
16	18 MΩ•cm).
17	
18	Gastric solution compositions are described in detail by Koch et al. ^[39] . Briefly, the gastric
19	solution of Method G contained 0.4 M glycine (Sigma) in DDW acidified with
20	hydrochloric acid (HCl Fisher) to a pH of 1.5 ± 0.05 (Acumet Excel XL15). For Method G,
21	an additional series of tests was performed at the 2000:1 liquid-to-solid ratio using a more

22 dilute (0.02 M) glycine solution. In Method P, the gastric solution includes porcine pepsin

(Sigma-Aldrich), sodium citrate (Caledon), malic acid (Sigma-Aldrich), glacial acetic acid
 (Fisher), and sodium chloride (Fluka) in DDW (concentrations can be found in Koch et
 al.^[39] The solution was similarly acidified to a pH of 1.80 ± 0.05.

4

5 Test procedures are identical for both methods. In the 100:1 liquid-to-solid ratio tests, two 6 subsamples were tested for each sample (representing P1 and P2). Measured amounts of 7 gastric solution (20 ml) were poured into 50-ml polyethylene specimen containers and 8 heated to 37 °C. A previously measured quantity of solid sample (0.2 g) was added to this 9 prepared solution and the test containers were secured in a temperature-controlled flatbed 10 rotation incubator (New Brunswick Scientific Innova 4230) at 37 °C and 150 rpm under 11 aerobic conditions. The pH was measured after 30 minutes (and acidified as required), and 12 again at the end of the one-hour gastric phase. Samples were stirred over a hot plate (37 13 °C) while pH measurement took place, and the time taken to measure pH (less than six 14 minutes per sample) was included in the overall incubation time. At the end of the gastric 15 stage of the test, samples used in the gastric phase only (P1) were removed from the 16 sample set.

17

The remaining samples were retained for the gastric followed by intestinal phase of the test (denoted P2). At the beginning of P2, the solutions were modified to simulate intestinal conditions by raising the pH to 7.0 ± 0.2 using a saturated NaOH (Sigma-Aldrich) solution for Method G,^[20] and a 10 M solution of Na₂CO₃ (Fluka) for Method P. Porcine bile (Sigma) and porcine pancreatin (Sigma-Aldrich) was also added to all intestinal solutions.

Test containers were returned to the incubator. The pH was measured after two hours (and
 adjusted as required), and again at the end of the four-hour intestinal stage of the test. The
 time taken to measure pH (less than 12 minutes during the intestinal stage of the test) was
 included in the overall incubation time.

5

6 In the 2000:1 liquid-to-solid ratios tests, a single subsample was prepared for each solid 7 sample (to be used for both P1 and P2). Measured amounts of gastric solution (300 ml) 8 were poured into 1-L polyethylene specimen containers and heated to 37 °C. A previously 9 measured quantity of solid sample (0.15 g) was added to this prepared solution. Test 10 containers were heated and shaken, and the pH was measured and adjusted as described 11 above. At the end of the gastric portion of the test, a solution aliquot (representing the P1 12 extract) was removed by syringe and filter (0.45 µm, PVDF membrane, Milipore), and 13 replaced with fresh gastric solution. The intestinal stage of the test was conducted as 14 described above. 15 16 At the end of each experiment, extracts were transferred to centrifuge tubes, centrifuged at

3800 rpm (2970•g) for 20 minutes, and the supernatant was filtered (0.45 μm, PVDF
membrane, Milipore).

19

All collected extracts from Method G and Method P were stored frozen (to -18 °C) and reserved for analysis. Thawed aliquots of each extract were diluted with 2% nitric acid (HNO₃ Fisher) solution and analyzed for total arsenic and iron concentrations by

1	inductively coupled plasma – mass spectrometry (Thermo Electron Corporation X-Series ^{II})
2	in collision cell mode as described by Smith et al. ^[40] Bioaccessibility results are expressed
3	as a concentration or as a percentage. The quantity of a substance (mg•kg ⁻¹ dry weight)
4	extracted during a bioaccessibility test is defined as the bioaccessible concentration.
5	Dividing this quantity by the total concentration in a given sample and multiplying by 100
6	gives the percent bioaccessibility.
7	
8	Quality Assurance and Quality Control
9	
10	Quality control tests and results are summarized in Table 1. Tests include blanks,
11	duplicates and a standard reference material (SRM), NIST2710, used in bioaccessibility
12	extractions, as well as matrix spikes and a calibration check solution for ICP-MS analysis.
13	The ICP-MS instrument detection limit was based on three standard deviations of eight
14	replicate measurements of a low concentration solution (1 and 10 μ g•L ⁻¹ for arsenic and
15	iron respectively); 88 of the 101 blanks returned numbers below the arsenic and iron
16	instrument detection limit (1 μ g•L ⁻¹ for arsenic; 10 μ g•L ⁻¹ for iron). All blanks were below
17	the limit of quantification (3 μ g•L ⁻¹ for arsenic; 20 μ g•L ⁻¹ for iron). All SRM arsenic and
18	iron results were within laboratory control limits (Table 1). Given the low variability in
19	SRM iron bioaccessibility between all methods and all liquid-to-solid ratios, these results
20	were combined and are shown as a single entry in Table 1. The average of the relative
21	percent difference (RPD) between extraction duplicates was less than 17%. However, RPD
22	values as high as 76% were recorded for duplicates associated with one sample (MG6),

and may be attributed to heterogeneity of this particular sample for reasons unknown at
 this time.

3

4 RESULTS AND DISCUSSION

5

6 The 19 samples used in this study cover a wide range of arsenic concentrations (from 94 to 420 000 mg•kg⁻¹) and iron to arsenic molar ratios (from 1.2 to 2600). Of these, the 13 7 8 Nova Scotia tailings samples (identified by the prefixes CAR-, GD-, MG-, and NB-) were 9 included in a previous study investigating the effects of soil composition and mineralogy on arsenic bioaccessibility from a suite of 29 tailings and soil samples.^[25] These 13 10 11 samples represent the wide range of arsenic concentrations and physical characteristics 12 encountered throughout the Nova Scotia gold mining districts. They were selected for 13 further experiments in the present study to compare the results previously obtained using 14 Method P (at the 100:1 and 2000:1 liquid-to-solid ratios) with the arsenic bioaccessibility 15 values obtained by Method G. The results of both methods are presented in Table 2 16 (arsenic bioaccessibility), and Table 3 (iron bioaccessibility). The percent arsenic 17 bioaccessibility ranges from near zero to 79%. The range of iron bioaccessibility results is 18 much smaller, varying from near zero to a maximum of 19%. 19

As previously indicated, physicochemical soil properties and arsenic speciation may result in large variations within methods. Furthermore, variations between methods may be attributed to a number of parameters including extraction pH, liquid-to-solid ratio, method,

1	and buffer concentration. First, the results are examined with respect to the extraction pH
2	to determine if the choice of buffer affects pH stability, how the bioaccessibility pH
3	conditions affect the bioaccessibility. The discussion then addresses the effects of liquid-
4	to-solid ratio, buffer concentration, and iron on the bioaccessibility.
5	
6	Effect of Buffer on the Stability of Method pH
7	
8	The ease with which pH is measured and maintained throughout a bioaccessibility test can
9	be an important factor in designing a test that is simple and economical, while providing a
10	representative measure of the soluble fraction of a contaminant. ^[41] Buffers that are
11	effective at gastric (pH < 2) and intestinal (pH \sim 7) pH are desirable. Glycine (used in
12	Method G, pK _a s 2.3 and 9.6), as well as the combination of citrate, malate and acetate
13	(used in Method P, pK _a s 3-3.4, 5-6.4) meet this requirement, since the solutions are
14	diprotic with pK _a s near the appropriate pHs.
15	
16	During the gastric stage of the test, pH measurements (before adjustment) varied by ± 0.07
17	for Method G (using the 0.4 M glycine concentration), which was a smaller variation than
18	that observed for Method P (± 0.15). Under intestinal conditions, greater variations in pH
19	measurements were recorded, and Method G (0.4 M glycine) once again demonstrated
20	more stability (± 0.17) than Method P (± 0.87). However, when the dilute 0.02M glycine
21	concentration was used, wider variations were recorded under gastric and intestinal
22	conditions (± 0.16 and ± 1.1 respectively). Therefore, at its higher 0.4 M concentration, the

1 glycine buffer provides a more effective buffering capacity than both Method G at the 2 dilute 0.02 M glycine and Method P. This suggests that the higher buffer concentration 3 provides useful pH control, but the rest of the study examines other associated effects. The 4 next section outlines how the pH in each buffer mixture affects the bioaccessibility results. 5 6 Effect of Gastric and Intestinal Extraction Conditions on the Bioaccessibility of Arsenic 7 8 Bioaccessibility tests are intended to represent a worst-case scenario for the solubility of a contaminant in a simulated gastro-intestinal tract.^[41] These tests thereby provide more 9 10 conservative (protective) adjustments in human health risk assessments. Since the bioaccessibility of arsenic may be influenced by the extraction pH,^[11] bioaccessibility tests 11 12 were carried out under both gastric (acidic pH) and intestinal (neutral pH) conditions. In 13 more than 70 percent of the results (all methods tested), the bioaccessibility of arsenic is 14 greater in P2 compared to P1 (Table 2). Inspection of the results in Table 2 reveals that 15 several tailings samples (NB11A, NB6B, GD2, LSH Tailings, CAR 1, and GD1) returned 16 a higher arsenic bioaccessibility result in P2 for all 5 treatments. The Ironite® and 17 NIST2710 sample exhibited the reverse trend of having a higher P1 arsenic 18 bioaccessibility in all cases. For the lower 100:1 liquid-to-solid ratios, variations between 19 P1 and P2 were relatively small (paired *t*-tests, \log_e transformed data, p > 0.52), but these 20 variations were statistically significant when the higher 2000:1 liquid-to-solid ratio was 21 used (paired *t*-tests, \log_{e} transformed data, p < 0.011) for both Methods G and P (Table 2).

Variations between P1 and P2 were not as apparent in the iron bioaccessibility results

1 (Table 3). In four of the five treatments, statistical analysis reveals that the extraction pH 2 did not significantly affect the bioaccessibility of iron (paired *t*-tests, \log_e transformed data, 3 p > 0.064). The only exception is for Method G at the 100:1 liquid-to-solid ratio, where the 4 P1 results were greater than the P2 iron bioaccessibility for all but one sample (paired *t*-5 tests, \log_e transformed data, p = 0.009).

6

7 For this type of experiment, we have determined that the acceptable within-laboratory 8 repeatability is $\pm 30\%$, which is consistent with the uncertainty estimated for total arsenic concentration results from accredited laboratories^[42] for such environmental samples. 9 10 Therefore, two results that are within 30% of each other are no more distinguishable than 11 two replicate analyses. The 30% uncertainty was used as a benchmark for comparison of 12 P1 and P2 values in Tables 2 and 3, where the higher result of either P1 or P2 is indicated 13 in bold only where the relative percent difference (RPD) between the two results is greater 14 than the 30%. The RPD varied by more than 30% in 53 of the 95 arsenic bioaccessibility 15 measurements, and in 62 of the 90 iron bioaccessibility measurements.

16

17 The predominance of higher P2 results for arsenic, illustrates that carrying out

18 bioaccessibility measurements in one phase only, as is typically done when the gastric

19 phase (P1) is assumed to represent worst-case conditions, would not necessarily provide

- 20 the most conservative estimates, especially in the case of arsenic-contaminated mine
- 21 tailings. In the next sections, the higher value of P1 or P2 (typically chosen as a worst-case

scenario in risk assessments) for each sample was used in the comparisons between
 methods.

3

4 Effect of Liquid-to-Solid Ratios

5

6 Previous arsenic bioaccessibility results reveal that, for the 13 Nova Scotia mine tailings 7 samples included in the present study, Method P was insensitive to liquid-to-solid ratios ranging from 100:1 to 5000:1.^[25] To test the effects of this variable, bioaccessibility 8 9 extractions were carried out on an additional six samples (Table 2). There were no 10 significant differences between the two liquid-to-solid ratios tested by Method P (paired t-11 tests, \log_e transformed data, p = 0.74 for the entire set of 19 samples, and p = 0.37 when 12 comparing only the six additional samples). This finding is consistent with other results from physiologically-based bioaccessibility tests.^[43, 44] The results are illustrated Figure 1, 13 14 where similarities between the two liquid-to-solid ratios are apparent for Method P, with 15 exceptions noted only for samples BGS102 and NIST2710 (higher arsenic bioaccessibility 16 at the higher liquid-to-solid ratio) and GD5 (higher arsenic bioaccessibility at the lower 17 liquid-to-solid ratio). In such a case, the bioaccessibility results obtained at the 100:1 liquid-to-solid ratio with Method P would be used for risk assessment purposes.^[30] 18 19 Conversely, large variations were observed in the case of Method G (Figure 1). Compared 20 with the 100:1 liquid-to-solid ratio, the arsenic bioaccessibility results were significantly 21 higher when the higher 2000:1H (0.4 M glycine) ratio was used (paired *t*-tests, log_e 22 transformed data, p < 0.0010). In such a case, the values obtained using the higher liquid-

1	to-solid ratio should be retained. ^[30] However at this 2000:1 H liquid-to-solid ratio, a
2	greater amount of buffer is available, and we hypothesize that the buffer concentration may
3	affect arsenic bioaccessibility.
4	
5	Effect of Buffer Concentration
6	
7	Further experiments were conducted to determine whether the bioaccessibility of arsenic
8	was influenced by the liquid-to-solid ratio, or by the amount of glycine present. An
9	additional series of tests was therefore performed using 0.02 M glycine at the 2000:1
10	liquid-to-solid ratio. At this dilute glycine concentration, the ratio of glycine to solid
11	sample is the same as in the 100:1 (0.4 M glycine) liquid-to-solid ratio test. In other words,
12	for this experiment, the amount of glycine used in the 100:1 ratio was diluted to a 2000:1
13	ratio. This test is denoted 2000:1L to differentiate it from the higher 0.4 M glycine
14	concentration test identified as 2000:1H.
15	
16	The results for Method G at 2000:1H (0.4 M glycine) generally give higher
17	bioaccessibility results than the other two method G treatments, for which the results are
18	similar, as shown in Figure 1. This trend is observed for a large majority of samples (15 of
19	the 19 samples), and suggests that a higher amount of glycine increases arsenic
20	bioaccessibility, even though statistical analysis did not reveal significant differences (1-
21	way ANOVA with Bonferroni adjustment for the three Method G treatments, $p = 0.17$; and
22	1-way ANOVA of all five treatments with Bonferroni adjustment, $p = 0.45$). A larger data

1	set may reveal statistically significant differences, which are only apparent in pair-wise
2	comparisons of the results for individual samples. In paired <i>t</i> -tests, results for Method G at
3	0.4 M glycine and the 2000:1H ratio are significantly different from the results obtained in
4	the other four treatments (paired <i>t</i> -tests, \log_e transformed data, $p < 0.0010$). The higher
5	bioaccessibility may result from glycine stabilizing cationic species in solution, including
6	iron (as described by Castillo and Ramirez), ^[45] and arsenic associated with this iron may
7	be released (i.e. when iron remains in solution, so does arsenic). Therefore, in the Method
8	G experiments, it is likely the amount of glycine present and the influence of a glycine-iron
9	complex that affects the arsenic bioaccessibility, rather than variations in the liquid-to-solid
10	ratios. In the following section, the relationship between iron and arsenic is further
11	examined by comparing bioaccessibility results and iron to arsenic concentrations.
12	
13	Effect of Iron Concentration on the Bioaccessibility of Arsenic
14	
15	The arsenic bioaccessibility results presented in Figure 1 are shown in decreasing order of
16	iron to arsenic molar ratio for each sample, and the results appear generally lower as this
17	ratio decreases, although this trend is not statistically significant (linear regression, $r < r$
18	0.23, $p > 0.19$). However, for all five treatments, the percent arsenic bioaccessibility is
19	negatively correlated with the total arsenic concentration (linear regression, $r > 0.50$, $p < 0.50$
20	0.033) and with the total iron concentration (linear regression, $r > 0.49$, $p < 0.035$). This is
21	consistent with previous findings for similar tailings samples. ^[25, 29] As demonstrated in the
22	previous sections, this negative relationship is not associated with solution saturation.

Previous findings have established that the highest arsenic concentrations in the tailings
 samples are associated with sparingly soluble arsenic minerals,^[25] and the decreased
 bioaccessibility is likely associated with an increased proportion of these minerals.

4

5 When the percent arsenic and iron bioaccessibilities are compared, the only method that 6 differs slightly from the others is Method G at the higher 0.4 M glycine and 2000:1H ratio, 7 for which the results show a weak positive correlation (linear regression, r = 0.22) (Figure 8 2). While this slope (0.13) is not significantly different from zero (p = 0.36), it is higher 9 than that for the other four treatments (slopes < 0.073, r < 0.32, p > 0.51). Thus, when a 10 higher liquid-to-solid ratio is used with a 0.4 M glycine concentration, higher bioaccessible 11 arsenic concentrations appear to be associated to the greatest extent with higher 12 bioaccessible iron concentrations. This result supports the hypothesis that extraction in the 13 presence of greater amounts of glycine is conducive to solubilising more iron (and 14 associated arsenic) than amounts that would be extracted under more physiologically 15 representative conditions.

16

17 CONCLUSION

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For the series of soils and tailings studied, the bioaccessibility of arsenic estimated by
Method P is not significantly affected by varying solid-to-liquid ratios and changes in pH,
whereas differences in bioaccessibility results were observed for the glycine-buffered
Method G. The bioaccessible arsenic concentration was also higher when a greater amount

1	of glycine was available in solution, which may be influenced by the interaction between
2	glycine and iron. The consequence is that using a glycine buffer in arsenic bioaccessibility
3	extractions may unduly influence the results by introducing methodological artefacts that
4	do not mimic the human gastro-intestinal conditions. At this time, neither Method P nor
5	Method G has been accepted by any regulatory agency as a validated method for arsenic
6	(against in vivo bioavailability data), ^[46] although some comparisons between Method G
7	and modified Method P arsenic bioaccessibility results with bioavailability results are
8	available ^[47, 48 and other references therein, 49] . In vivo hamster arsenic data ^[50] are also available for
9	one of the samples in the present study (Ironite®). For this sample, results from all
10	methods tested (28-32% for Method P; 25-39% for Method G) compared equally well with
11	arsenic bioavailability values $(31 \pm 6\%)$ and did not vary significantly between methods.
12	For this reason, it is not possible to unambiguously distinguish between methods based on
13	in vivo data; clearly more bioavailability data are required to compare and validate
14	bioaccessibility methods.
15	
16	The present study demonstrates that the practicality of a method and the potential for

17 artefacts resulting from that same method must be balanced. Method G has been accepted 18 as a useful method to estimate lead bioaccessibility, and hence bioavailability, but its 19 application to other elements and soil samples may require careful consideration and study. 20 The choice of a bioaccessibility method that is both physiologically based and robust with 21 respect to changes in method variables, such as Method P, may provide more

1	representative arsenic bioaccessibility results for the purpose of estimating risks to human
2	health.
3	
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11	
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Medium	Measurement	Method	Experiment	N	Range	Average
Blank ^b	Arsenic concentration	Method P (P1, P2)	All L:S	19	< 3.0 µg•L ⁻¹	_
Blank	Iron concentration	Method P (P1, P2)	All L:S	22	$< 20 \ \mu g \cdot L^{-1}$	_
Blank	Arsenic concentration	Method G (P1, P2)	All L:S	30	$< 2.0 \ \mu g \cdot L^{-1}$	_
Blank	Iron concentration	Method G (P1, P2)	All L:S	30	$< 20 \ \mu g \cdot L^{-1}$	_
Spike recovery ^c	Arsenic recovery (%)	Method P (P1, P2)	L:S = 100:1	6	93-116%	98%
Calibration check ^d	Arsenic recovery (%)	Instrument	_	36	85-101%	93%
Calibration check	Iron recovery (%)	Instrument	_	36	95-119%	103%
SRM NIST2710	Bioaccessible arsenic (%)	Method P (P1)	Control limits ^e (L:S = $100:1$)	23	24-57%	40%
SRM NIST2710	Bioaccessible arsenic (%)	Method P (P2)	Control limits $(L:S = 100:1)$	37	23–49%	36%
SRM NIST2710	Bioaccessible arsenic (%)	Method G (P1)	Control limits $(L:S = 100:1)$	37	29-70%	40%
SRM NIST2710	Bioaccessible arsenic (%)	Method G (P2)	Control limits $(L:S = 100:1)$	29	4.0-38%	21%
SRM NIST2710	Bioaccessible iron (%)	All methods	Control limits (All L:S)	19	0.29–15%	4.5%
SRM NIST2710	Bioaccessible arsenic (%)	Method P (P1)	L:S = 100:1	5	23-45%	39%
SRM NIST2710	Bioaccessible arsenic (%)	Method P (P2)	L:S = 100:1	5	31-38%	38%
SRM NIST2710	Bioaccessible arsenic (%)	Method G (P1)	L:S = 100:1	5	42-56%	49%
SRM NIST2710	Bioaccessible arsenic (%)	Method G (P2)	L:S = 100:1	5	16-37%	25%
SRM NIST2710	Bioaccessible iron (%)	All methods	All L:S	10	0.36-12%	4.8%
Duplicate pairs	RPD, bioaccessible arsenic (%)	Method P (P1, P2)	All L:S	38	0.22-20%	7.2%
Duplicate pairs	RPD, bioaccessible iron (%)	Method P (P1, P2)	All L:S	30	0.70–59%	17%
Duplicate pairs	RPD, bioaccessible arsenic (%)	Method G (P1, P2)	All L:S	52	0.23-47%	11%
Duplicate pairs	RPD, bioaccessible iron (%)	Method G (P1, P2)	All L:S	52	0.42–76%	12%

Table 1. Results of quality control tests^a.

a. Abbreviations: N = number of replicates; L:S = liquid-to-solid ratio; P1 = gastric phase; P2 = gastric + intestinal phase; Method P = physiologically-based; Method G = glycine-buffered; SRM = standard reference material; RPD = relative percent difference, calculated as $100 \cdot$ (result of primary sample – result of duplicate sample) • average⁻¹. All other abbreviations are described in the text.

b. Instrument detection limit is $1 \mu g \cdot L^{-1}$ for arsenic and $10 \mu g \cdot L^{-1}$ for iron. Results presented in this table represent limits of quantification.

c. Spike recovery of 100 ppb•L⁻¹ potassium arsenate (KH₂AsO₄, Fluka reagent grade) in blank PBET solution added immediately before analysis.

d. Calibration check solutions used in the ICP-MS analysis include both 50ppb and 750ppb multi-element solutions (PlasmaCAL), prepared from a different source solution than the ICP-MS calibration solutions.

e. All control limits for SRM are average ± three standard deviations for all laboratory results recorded (between 2001 and 2009 for arsenic; between 2006 and 2008 for iron), excluding the results of the present study.

			Method P ^b				(0.4	(\mathbf{M})	Method G^{c}		(0.02M)	
	Molar	Total	L:S=	100:1	L:S=2	2000:1	L:S=	100:1	L:S=20	000:1H	L:S=2	000:1L
Sample	Ratio Fe:As	Arsenic mg•kg ⁻¹	<i>P1</i>	P2	P1	P2	<i>P1</i>	P2	<i>P1</i>	P2	<i>P1</i>	P2
BGS102	2600	94	2.3	2.8	12	9.5	2.8	2.1	8.4	9.7	3.5	5.2
NIST2711	360	110	47	37	45	38	40	34	45	51	45	44
MG6	250	320	2.5	1.4	3.9	3.5	2.6	0.33	10	14	3.3	5.0
NB7	79	460	31	34	25	29	33	26	43	58	35	39
NIST2710	72	630	33	23	47	36	42	17	79	35	48	39
NB6A	54	740	16	18	12	14	21	21	37	30	23	26
Ironite®	37	4300	28	4.7	32	30	25	0.39	39	30	39	3.7
NB11A	18	5800	14	16	6	13	12	18	10	24	7.4	15
NB6B	11	7200	3.9	5.6	2.8	4.5	3.3	4.5	4.1	9.7	2.2	4.8
NB12	9.4	8200	13	29	12	26	16	15	24	54	17	33
GD5	5.3	7200	47	49	32	37	44	26	58	61	48	43
MG3	3.3	24 000	11	12	10	7.2	16	8.4	31	35	21	22
MG4	3.3	21 000	2.0	2.7	0.99	1.8	4.1	3.3	9.0	13	2.1	3.8
GD2	2.5	19 000	2.4	4.1	0.82	2.4	7.8	16	10	13	2.9	7.1
CAR2	1.6	310 000	0.62	0.27	0.10	0.19	0.48	0.52	0.44	0.54	0.44	0.41
LSH Tailings	1.4	200 000	1.0	1.6	0.62	0.84	0.97	2.6	1.1	3.7	0.83	2.8
CAR1	1.3	$77\ 000$	2.1	5.0	0.50	1.6	2.4	3.3	2.4	9.8	2.1	8.8
GD1	1.3	210 000	0.13	0.32	0.05	0.42	0.13	2.2	0.27	2.8	0.22	0.93
Arsenopyrite	1.2	420 000	0.17	0.22	0.24	0.31	0.23	0.16	0.24	0.36	0.34	0.36

Table 2. Percent arsenic bioaccessibility^{ab}.

1 a. Results are presented in order of decreasing molar iron to arsenic ratio. Numbers in bold indicate the

2 higher of either P1 or P2 result where the relative percent difference, calculated as $100 \cdot (P1 - P2) \cdot$

3 average⁻¹ is greater than the acceptable laboratory repeatability (30%).

4 b. Abbreviations: L:S = liquid-to-solid ratio; P1 = gastric phase; P2 = gastric + intestinal phase;

5 Method P = physiologically-based; Method G = glycine-buffered.

6 c. For Method G, the first two extractions used a 0.4 M glycine concentration, and a third extraction was

7 performed using 0.02 M glycine.

			Method ^c			(0)	1 11)	Me	ethod G^{d}	(0.02M)		
	Molar Patio	Total Incr	L:S=	100:1	<i>L:S=2</i>	2000:1	L:S=	100:1	L:S=2	2000:1H	L:S=2	2000:1L
Sample	Kallo Fe:As	mg•kg ⁻¹	<i>P1</i>	P2	<i>P1</i>	P2	<i>P1</i>	P2	<i>P1</i>	P2	<i>P1</i>	P2
BGS102	2600	180 000	0.50	0.39	1.2	1.2	0.92	0.03	1.2	0.41	1.4	0.07
NIST2711	360	29 000	1.9	2.4	2.3	2.1	2.5	0.16	3.0	1.3	2.8	0.82
MG6	250	60 000	0.83	0.01	0.82	1.4	1.9	0.13	2.8	1.7	0.86	0.28
NB7	79	27 000	1.5	0.57	1.3	1.8	4.1	0.38	3.3	3.7	3.5	1.8
NIST2710	72	34 000	0.36	4.0	4.2	3.7	6.1	0.84	12	6.0	6.3	4.4
NB6A	54	30 000	0.74	0.45	0.57	0.87	2.7	0.63	3.1	1.6	2.4	0.28
Ironite®	37	120 000	5.2	15	14	12	14	0.06	18	12	19	0.50
NB11A	18	92 000	0.99	1.2	0.44	0.96	2.2	0.22	1.3	1.9	1.3	0.83
NB6B	11	60 000	0.87	0.69	0.60	0.87	3.1	0.25	1.9	1.5	1.7	0.79
NB12	9.4	64 000	n/a	n/a	2.8	4.8	9.1	2.2	8.2	9.7	8.3	6.3
GD5	5.3	29 000	n/a	n/a	5.5	6.4	13	1.1	16	13	16	5.2
MG3	3.3	59 000	n/a	n/a	2.2	7.6	8.6	0.43	14	10	10	1.5
MG4	3.3	53 000	1.1	1.1	1.1	1.1	4.8	0.81	7.2	5.9	3.4	0.92
GD2	2.5	36 000	0.89	2.2	0.92	1.2	6.6	1.1	7.3	4.0	2.6	2.3
CR2	1.6	360 000	0.23	0.23	0.09	0.10	0.38	0.10	0.32	0.19	0.35	0.05
LSH Tailings	1.4	210 000	0.67	0.89	0.42	0.35	0.94	0.35	0.91	1.6	0.85	0.66
CR1	1.3	75 000	n/a	n/a	0.41	0.83	2.6	1.5	2.5	6.0	2.5	0.81
GD1	1.3	200 000	n/a	n/a	0.16	0.37	0.28	0.47	0.43	0.69	0.42	0.44
Arsenopyrite	1.2	380 000	0.18	0.22	0.25	0.29	0.32	0.09	0.26	0.37	0.48	0.27

1 2 **Table 3.** Percent iron bioaccessibility^{ab}.

3 a. Results are presented in order of decreasing molar iron to arsenic ratio. Numbers in bold indicate the

4 higher of either P1 or P2 result where the relative percent difference, calculated as $100 \cdot (P1 - P2) \cdot$

5 average⁻¹ is greater than the acceptable laboratory repeatability (30%).

6 b. Abbreviations: L:S = liquid-to-solid ratio; P1 = gastric phase; P2 = gastric + intestinal phase;

7 Method P = physiologically-based; Method G = glycine-buffered.

8 c. Items marked n/a were not analyzed for iron concentrations.

9 d. For Method G, the first two extractions used a 0.4 M glycine concentration, and a third extraction was

10 performed using 0.02 M glycine.



method in decreasing order of iron to arsenic molar ratio. Results on the right side of the dotted line are associated with the vertical axis
shown on the right.



Figure 2: Bioaccessible arsenic and iron concentrations (mg•kg⁻¹) for all samples under both Method P and
Method G at the two liquid-to-solid ratios (100:1 and 2000:1). For Method G at the 2000:1, a 0.4 M glycine
concentrations was used in the 100:1 and 2000:1H extraction, and a 0.02 M glycine concentration was used
in the 2000:1L extraction.