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1. Introduction

Plutonium (Pu) originates in the environment primarily as a consequence of nuclear weapons testing, 520 atmospheric tests were conducted worldwide between 1945 and 1980.^{1,2} However, only 10% of these experiments were conducted in the southern hemisphere, resulting in significantly less fallout in the tropics than in the mid-latitudes of the northern hemisphere. This makes the analysis of ultra-trace Pu isotopes in tropical soils challenging.^{3,4} Due to their long retention time and minimal spatial variability, Pu isotopes have been recently utilised as an alternative fallout radionuclide tracer for determining soil erosion rates. Due to the much longer half-lives of ²³⁹Pu and ²⁴⁰Pu (24 110 and 6561 years, respectively),

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Optimisation of plutonium separations using TEVA cartridges and ICP-MS/MS analysis for applicability to large-scale studies in tropical soils⁺

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The analysis of plutonium (Pu) in soil samples can inform the understanding of soil erosion processes globally. However, there are specific challenges associated for analysis in tropical soils and so an optimal analytical methodology ensuring best sensitivity is critical. This method aimed to demonstrate the feasibility of sample preparation and analysis of Pu isotopes in African soils, considering the environmental and cost implications applicable to low-resource laboratories. The separation procedure builds upon previous work using TEVA columns, further demonstrating their usefulness for the reduction of uranium (U) interference in ICP-MS analysis with enhanced selectivity for Pu. Here several steps were optimised to enhance Pu recovery, reducing method blank concentration, and improving the separation efficiency through the determination of the elution profiles of U and Pu. The elimination of the complexing agent in the eluent, increased the spike recovery by improving matrix tolerance of the plasma, and simplified the separation procedure, improving throughput by 20%. The subsequent method was validated through the analysis of Certified Reference Material IAEA-384, where high accuracy and improved precision of measurement were demonstrated (measured value 114 \pm 12 versus certified value 108 \pm 13 Bq kg⁻¹). Optimisation of the column separation, along with the analysis of the samples using O₂ gas in ICP-MS/MS mode to mass shift Pu isotopes away from interfering molecular U ions provided a simple, robust, and cost-effective method with low achievable method detection limits of 0.18 pg kg⁻¹ ²³⁹⁺²⁴⁰Pu, applicable to the detection of ultra-trace fallout Pu in African soils.

> approximately 99% of the original activity remains in soils, allowing for their stable and long-term use as a tracer compared to ¹³⁷Cs, which has a half-life of only 30 years.^{5,6} In addition, more than six times as many atoms of ²³⁹⁺²⁴⁰Pu were initially dispersed compared to ¹³⁷Cs, despite the latter's significantly higher activity in the environment. This combination of long half-life and higher atom content makes mass spectrometry techniques better suited to Pu isotopes, whereas radiometric decay counting techniques are more appropriate for the higher specific activity ¹³⁷Cs. Consequently, recent developments in mass spectrometry techniques have the potential to increase the sensitivity of quantification of Pu isotopes and subsequently the availability of analytical methods applicable to tropical soils.7,8 This raises the potential of using Pu as a soil erosion tracer in the tropics, where the risk of soil degradation is increasing due to extreme weather patterns.3

> Radiometric and mass spectroscopy techniques, such as alpha spectrometry, accelerator mass spectrometry (AMS), and inductively coupled plasma mass spectrometry (ICP-MS), can be used to analyse Pu in a variety of samples.^{7–10} In recent years, ICP-MS has gained popularity due to its low detection limits, short analytical time, high sample throughput, relatively simple



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operation, and lower instrument cost.5,11-14 Despite the benefits of ICP-MS, the technique is severely constrained by polyatomic interferences, most notably uranium hydrides (UH⁺). In addition, ²³⁸U tailing interferes with the detection of mass-to-charge ratio (m/z) 239 due to its concentration being several orders of magnitude greater than that of ²³⁹Pu in soils.¹⁵ This has placed a requirement for a high level of enrichment and elemental separation prior to analysis for the accurate determination of trace Pu isotopes.⁷ The Pu isotopes can be separated from the interfering isotopes in the matrix and pre-concentrated to assure maximum sensitivity, using techniques such as coprecipitation, ion exchange chromatography, and extraction chromatography.¹⁶⁻¹⁹ Utilising the selective TEVA column resin (Eichrom Technologies) to effectively remove U and preconcentrate Pu is common for determination by ICP-MS.^{16,20} Nygren et al. (2003)¹⁹ reported that TEVA resin exhibited the maximum separation yield of Pu from soils and sediments compared to other methods. This finding is corroborated by the scientific literature and can be attributed to both the relatively low U content of the resin and the large variation in acid dependency of k' which is greater than 3 orders of magnitude between U and Pu with nitric acid concentration between 2 and 4 M.²¹ Therefore, this paper concentrates on improving the analysis of ultra-trace Pu isotopes in east African soils by further optimising the chemistry of TEVA columns for the latest generations of mass spectrometric instrumentation, applicable to analysis of large-scale studies in tropical soils.

As a result of recent developments in reaction cell technology, tandem ICP-MS (ICP-MS/MS), also known as triple quadrupole ICP-MS (ICP-QQQ-MS), provides an alternative method of analysis to determine Pu. This method is increasingly utilised for the determination of Pu isotopes due to its enhanced abundance sensitivity, which effectively eliminates the interference of the ²³⁸U peak tailing on the measurement of ²³⁹Pu and ²⁴⁰Pu.^{7,8,14,22} In addition, the main interference of ²³⁸UH⁺ on ²³⁹Pu can be eliminated by utilising different reaction gases in the collision-reaction cell.^{4,7,8,10,15,23} The quadrupole mass filter positioned in front of the collision-reaction cell permits the pre-selection of species, thereby enhancing the reaction efficiency of the collision cell and prohibiting the formation of secondary polyatomic interference. In addition, the use of a second quadrupole reduces peak tailing, resulting in enhanced mass resolution. Gases, including ammonia (NH₃), carbon dioxide (CO_2) , and oxygen (O_2) , have been proposed for the elimination of UH⁺ interferences by ICP-MS/MS.^{4,7,15} The most frequently employed of these gases is NH₃, which effectively mass shifts the 238U1H interference away from 239Pu via the preferential reaction of U with the NH₃ gas. Using this method, detection limits of 0.16 fg g^{-1} for ²³⁹Pu have been achieved; however, despite the ability of NH3 to react with the U interference, the use of NH₃ gas poses several safety concerns, making its use undesirable.^{8,24} Using O₂, the lowest detection limits of 0.06 fg g^{-1} for ²³⁹Pu have been reported; this can be attributed to the formation of PuO2⁺, which effectively mass shifts the Pu isotopes to m/z 271 to avoid interference from 238 UH⁺ at m/z 239.⁴ The m/z ratio of PuO₂⁺ exceeds the mass range of many early ICP-MS/MS instruments; however, user

demands for analysing heavy elements in mass shift modes has led instrument manufacturers developing instruments with an extended m/z range up to 275.

The aim of this study was to accurately determine fallout Pu activity concentrations in tropical soils for the subsequent determination of soil erosion rates with an improved separation and analysis method for ultra-trace Pu determination. The two objectives to achieve this aim are: (1) adapting and optimising a separation method using TEVA cartridges for the removal of matrix interferences with pre-concentration of ultra-trace Pu isotopes to reduce waste and increase throughput; and (2) establishing a robust analytical method for the determination of ultra-trace level Pu isotopes with sufficient sensitivity for African soil samples using oxygen as a reaction gas for ICP-MS/MS.

2. Materials and methods

2.1 Reagents and materials

All reagents used were of analytical grade. Water used throughout had a resistivity of 18.2 MQ at 25 °C and was obtained from a Milli-Q gradient system (Millipore, MA, USA). Extra-pure nitric acid (70%) was obtained from Thermo Fisher Scientific. Ultra-pure HCl was obtained from ROMIL. A solution of 242 Pu (2.1 × 10⁻³ kBq/14.2 ng) of unknown origin was created from a stock solution (0.21 kBq) using a 100-fold dilution with a solution of 5% HNO3 and 2.5% HCl in water. This ²⁴²Pu solution was then used to spike soil samples prior to digestion. Samples were filtered using 0.45 µm hydrophilic PTFE syringe filters (Thermo Fisher Scientific, UK). Oxidation of Pu species was achieved using >97% purity NaNO₂ (Sigma-Aldrich). The elution of Pu from the columns used a solution of 0.05 M ammonium oxalate, which was created by dissolving diammonium oxalate monohydrate (Supelco, Sigma-Aldrich) in water. Quality control was achieved using certified reference material obtained from the International Atomic Energy Agency (IAEA). The certified reference material IAEA-384 (radionuclides in Fangataufa Lagoon sediment with a certified value of 108 \pm 13 Bq kg⁻¹ ²³⁹⁺²⁴⁰Pu) was selected, as this reference soil has been widely used within the literature for the determination of Pu isotopes and no suitable soil CRM was available. To account for differences in the sediment CRM and the analysed soils, an in-house reference material was created using soils collected in the UK. Column separations of Pu isotopes were performed using TEVA pre-packed columns (2 ml, 100-150 µm) from Eichrom Technologies. Elemental standard Ir (Spex CertiPrep) was used as an internal standard. Calibration of Pu concentrations via ICP-MS used a U elemental standard solution (Fisher chemical). Silicate sand, which was subjected to the same dissolution procedure as the CRM and samples, was used as the method blank.

2.2 Sample collection

As part of this study, two bulk reference soils were collected from farmland within the UK and Zambia. The UK soil was collected from a cattle farm (Hoveringham, Nottinghamshire, England) during August 2020. Soil was collected from 3 locations across the farm, down to a depth of 15 cm. The soil was then dried in an oven overnight, disaggregated, sieved to <2 mm and finally milled \leq 53 µm using a planetary ball mill (Retsch GmbH, Germany), ready for dissolution. This UK reference soil was then homogenised and used throughout the method development as both a quality control sample and to optimise the separation method. The Zambian soil was created from the combination of agricultural soils collected from Kitwe, Zambia, by Hamilton *et al.* (2020).²⁵ The analysed soils were prepared in the same way as the UK reference soil and were used to verify the usefulness of the final method for use on African soils, where Pu levels are much lower than in the Northern Hemisphere.

To validate the method for the determination of soil redistribution rates, a soil core was collected within the Oroba valley, Nandi County, Kenya (Fig. 1). Within the valley, a site with no overall soil re-distribution, representative of a reference site, was collected. Here, a core was taken to a depth of 30 cm and broken down into 10 sections to determine the Pu inventory within the site. To collect the sample, a pit was dug, and a bulk density tin with a diameter of 5 cm and a height of 3 cm, was inserted into the pit wall. The sample was then taken from the tin and dried in an overnight oven, weighed, disaggregated, and then sieved to <2 mm. The resulting soil was then weighed again to determine the soil density, before finally being milled to \leq 53 µm using a planetary ball mill before dissolution. This sample was then analysed to determine the Pu inventory at the site and to demonstrate the methods usability at depth where levels of Pu are significantly lower.

2.3 Sample preparation

2.3.1 Dissolution. Prior to analysis, all samples are weighed into glass beakers and then ignited (ashed) at 550 °C for 12 hours to remove organic carbon. Plutonium does not form

volatile oxides and therefore would not be lost. The ashed sample was then weighed into a suitable PTFE beaker, and the volume of concentrated nitric acid required to leach plutonium from soils was investigated, along with the optimum soil mass (soil/acid ratio) to ensure optimum sensitivity while limiting environmental impact by reducing waste. The first step of the optimisation was to determine the optimal mass of sample to use, determined using the UK reference soil (Newark, UK). Six masses ranging from 1 g to 50 g of soil were investigated (1, 5, 10, 20, 35, 50 g). To determine the optimal mass, the soils were digested in 2 ml HNO₃ per 1 g soil which is the minimum volume of HNO3 suggested by Ketterer et al. (2004).26 The variability in measurement of ²³⁹⁺²⁴⁰Pu by ICP-MS/MS was determined for three replicate samples at each mass over three separate analytical runs. Subsequently, the optimal volume of acid to digest the soil was investigated to ensure the minimum volume of acid was used to reduce waste and therefore the environmental impact, whilst ensuring the maximum extraction of ²³⁹⁺²⁴⁰Pu from the soils. Volumes between 0.5 ml and 3 ml per gram of sample were investigated using the UK reference soil. To the leach mixture, 50 pg (70 μ l of 0.2 Bq ml⁻¹) of the ²⁴²Pu spike was added, and then the solution was covered by a watch glass. The soil solution was heated on a hotplate at 70 $^{\circ}$ C for 24 hours, cooled to room temperature, and then centrifuged at 3000 rev. per min for 15 minutes. The supernatant was then filtered through a 0.45 µm hydrophilic PTFE filter and collected in a PTFE beaker. A volume of water was added to the pellet in the centrifuge tube. The pellet was redistributed into the water through vigorous shaking and was subsequently centrifuged again. The supernatant was then filtered and added to the PTFE beaker, adjusting the concentration of HNO₃ to 8 M. For the Pu within the sample to be effectively separated using the TEVA column, it must first be in the IV oxidation state. To do this, 0.02 g NaNO₂ per 1 ml of solution was added, and the



Fig. 1 Sample location of soil core within the Oroba valley, Nandi County, Kenya, and sampling technique.



Fig. 2 Outline of column separation employed for the preparation of Pu isotopes in soil using a TEVA resin.

sample was placed back on the hotplate at 40 $^{\circ}$ C overnight due to the slow nature of the oxidation reaction. The Pu in the solution was then separated using a TEVA column separation.

2.3.2 Plutonium column separation. The separation used was modified using the method described in Ketterer et al. (2004).²⁶ Fig. 2 shows the outline of steps within the column separation. The first step involved the pre-conditioning of the column with the first mobile phase to be used, and this ensured all air bubbles had been removed from the column bed and equilibrated the column chemistry for the sample solution matrix. Following this, the sample was loaded onto the column, and subsequently, the column was washed using 2 M HNO3 and 8 M HCl. These two steps remove both major and trace elements from the TEVA resin, reducing the formation of molecular interferences within the plasma, primarily UH⁺ but also interferences originating from Hf, Ir, Pt, Hg, Tl, Pb, Bi, Th, and Pb. Due to the acid dependency (k') of tetravalent Pu on the TEVA resin being greatest in the region of 2-4 M HNO3, it is well retained during the washing step with 2 M HNO₃ while hexavalent U is eluted. Following the washing steps, the Pu was eluted from the column by using an appropriate complexing agent or a dilute acid concentration.21 The separation was

optimised through a step-by-step process and is detailed in the results section. These steps include determining the volume of 2 M HNO₃ needed to ensure maximum elution of U isotopes to limit UH⁺ interference on Pu measurements, the most suitable eluent for the elution of Pu isotopes (oxalate *vs.* dilute HCl), and the volume of this eluent to ensure optimal elution of Pu isotopes. The steps were optimised to allow a cost-effective, robust method and to minimise laboratory waste disposal – applicable to laboratories with minimal resources for analysis and waste disposal. The column was then flushed and stored with a 0.01 M HNO₃ matrix to be stored until the quantitative recovery was determined.

2.4 Instrumentation and setup

An ICP-MS/MS (Agilent 8900, Agilent Technologies, Japan) was used for the measurement of Pu isotopes in the soil samples. This instrument was equipped with a collision reaction cell (CRC) between two quadrupole mass filters. For the measurement of plutonium isotopes, O2 was used in the CRC to mass shift both Pu and U to PuO₂⁺ and UO₂⁺ ions for measurement, as demonstrated in Zhang et al. (2021).4 The operating gas modes and default internal standards for each isotope can be seen in Table 1. This mass shifting has the advantage of removing the major interference of UH⁺ on Pu isotope measurements. The ICP-MS/MS was equipped with an Agilent IaS micro-autosampler and a Cetac Aridus 2 desolvating nebuliser (Teledyne CETAC Technologies, Omaha, USA). This combination required only 1 ml to be used for analysis, ensuring the maximum sensitivity of Pu in the sample through large pre-concentration factors for a given sample size. The instrument was auto-tuned using the Agilent Masshunter software using a 1 µg kg⁻¹ tune solution (SPEX CertiPrep #CL-TUNE-1) for general performance. Agilent typically enables autotuning up to mass 260, and therefore, autotuning on U⁺ at mass 270 is not possible. The instrument was then manually tuned on ²⁷⁰U⁺ specifically for Pu and U in oxygen mode. The key tuning parameters were O_2 flow rate, Q2 bias, energy discrimination, OctP bias, wait time offset, cell focus and axial acceleration. Optimised parameters used are shown in ESI Table 1.† The instrument was calibrated for sensitivity using a ²³⁸U standard as no certified Pu standard was available. It was assumed that the sensitivity of ²³⁸U was the same as ²³⁹Pu allowing for isotopic abundance (ESI Table 2[†]).

2.5 Quality control

A variety of laboratory control samples were used to track instrumental performance within and between instrument

Table 1	Operating gas	s modes and defa	ult internal standards fo	or each isotope
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Isotope	Gas mode	Reaction product	Internal standard
$238 \rightarrow 270$	O_2	²³⁸ UO ₂ ⁺	¹⁹³ Ir
$239 \rightarrow 271$	$\overline{O_2}$	239 PuO ₂ ⁺	¹⁹³ Ir
$240 \rightarrow 272$	$\overline{O_2}$	$^{240}PuO_{2}^{+}$	¹⁹³ Ir
$242 \rightarrow 274$	O_2	242 PuO ₂ ⁺	¹⁹³ Ir
	Isotope $238 \rightarrow 270$ $239 \rightarrow 271$ $240 \rightarrow 272$ $242 \rightarrow 274$	IsotopeGas mode $238 \rightarrow 270$ O_2 $239 \rightarrow 271$ O_2 $240 \rightarrow 272$ O_2 $242 \rightarrow 274$ O_2	Isotope Gas mode Reaction product $238 \rightarrow 270$ O_2 $^{238}_{239}UO_2^+$ $239 \rightarrow 271$ O_2 $^{239}_{239}PuO_2^+$ $240 \rightarrow 272$ O_2 $^{240}PuO_2^+$ $242 \rightarrow 274$ O_2 $^{242}PuO_2^+$

analyses. These included a certified reference soil obtained from the IAEA, specifically IAEA-384 (Radionuclides in Fangataufa Lagoon sediment), with a certified value of 108 \pm 13 Bq kg⁻¹ ²³⁹⁺²⁴⁰Pu, which was digested in duplicate within each dissolution batch and then analysed. Background signals were identified by analysis of laboratory control samples (water with 5% HNO₃ and 2.5% HCl) and method blanks (20 g silicate sand, which was subjected to the same dissolution procedure as the CRM and samples). Detector performance was monitored using a 1 µg kg⁻¹ tune solution (SPEX CertiPrep #CL-TUNE-1). Internal standard (¹⁹³Ir) was analysed to correct for instrumental drift and any matrix suppression effects. Soils were analysed in triplicate within each dissolution batch to verify intra-batch analytical measurements for precision.

3. Results and discussion

3.1 Soil dissolution

Fig. 3 shows the relative standard deviation (RSD) of ²³⁹⁺²⁴⁰Pu measurements according to different sample masses (g). To ensure the environmental impact (primarily the volume of acid required) of the method was accounted for, the minimum mass of sample with the lowest variability in measurement was selected, and the optimum mass for a UK sample was determined to be 20 g. The ²³⁹⁺²⁴⁰Pu concentration within the UK reference soil was found to be 38.86 \pm 1.93 pg kg⁻¹. Due to the much lower fallout in the tropics compared to the mid-latitudes of the northern hemisphere, Pu activity in Europe is on average two-to-three times greater than that in Africa.² As a result, when working with African soils, it was decided that 50 g would be the preferred mass whilst retaining optimum sensitivity of Pu measurement. Previous studies utilising ICP-MS for the determination of Pu in soils in the northern hemisphere have used masses between 1 and 10 g,^{6,27-31} however, the results suggest that the error in measurements at these lower masses justifies the use of a greater mass for separation. The use of 50 g as the soil mass for dissolution is supported elsewhere in the literature, e.g. Wilken et al. (2021)³ used 50 g of soil for the determination of Pu in African soils.



Fig. 3 Relative standard deviation (n = 3) in measurement of $^{239+240}$ Pu isotope concentrations by ICP-MS/MS according to different sample mass of UK soil used for dissolution.



Fig. 4 ²⁴²Pu spike recovery dependant on the volume of nitric acid used for the dissolution of soil samples.

Fig. 4 shows the recovery of the ²⁴²Pu spike as a function of the volume of HNO3 per gram of soil. Each sample was digested in triplicate, with the error bars representing two times the standard deviation of the measurements. The spike recovery of ²³⁹⁺²⁴⁰Pu, representative of the whole separation procedure, peaked at 2 ml with the addition of more digest acid, resulting in no significant increase in Pu extraction as the graph plateaus. This is likely as a result of both the decrease in viscosity of the digest solution passing through the column and the increased efficiency of dissolution. The maximum spike recoveries exceed 100% as a result of the overall uncertainty. This compliments with the method suggested by Ketterer et al. (2004),²⁶ which was widely reported in the literature. The benefits of the complete extraction of Pu isotopes from the soil using 2 ml of acid outweigh the environmental and cost implications of using a larger volume.

The final step of the dissolution was to centrifuge the mixture to separate the digest from the residual solid. In the initial tests, it was found that spike recoveries were low (30-50%), and so the washing of the solid pellet after centrifugation was investigated. After the initial centrifugation step, the supernatant was transferred into a PTFE beaker, leaving behind a solid pellet. This was then re-distributed using an equivalent volume of DI water to the volume of acid used in the digestion. The centrifuge tube was then shaken vigorously until the solid pellet was fully redistributed into the water. The mixture was then centrifuged again at 3000 rpm min⁻¹ for 15 minutes and the new supernatant was subsequently added to the PTFE beaker containing the digest to adjust the concentration of HNO₃ to 8 M. The results showed that the spike recovery was increased by 30% with the addition of a washing step, suggesting that some of the Pu remained within residual acid retained between pores of the centrifugated soil.

3.2 Optimisation of column separation

The Pu in the samples was separated from the sample matrix using a TEVA column. The separation involved preconditioning of the column, loading of the sample, elution of matrix elements, including U isotopes, and finally the elution of Pu isotopes. As part of the method development, the elution of U



Fig. 5 Elution profile of ²³⁸U isotope from TEVA resin using eluent 2 M nitric acid.

isotopes and the elution of Pu isotopes were investigated. Due to the major interference caused by UH⁺ on Pu isotopes during ICP-MS analysis, it was vital to elute and remove the greatest amount of the U isotopes present in the dissolution prior to analysis. The elution of U isotopes was achieved by loading the column with 2 M HNO₃ and discarding the eluent. The tetravalent Pu provided maximum uptake in the region of 2–4 M HNO₃, while the hexavalent U was eluted, Pu was retained on the resin during this step.³² After loading the sample onto the column, 2 M HNO₃ was added to the column 5 ml at a time up to 80 ml. Each of the eluents for every addition of 5 ml was collected and subsequently analysed by ICP-MS/MS for ²³⁸U to determine at what volume the maximum U had been eluted from the resin.

Fig. 5 shows the percentage U eluted between 0 and 80 ml 2 M HNO₃. The greatest proportion of U was eluted within the first 10 ml of acid (98.5%); however, an additional 1.0% was removed up to 60 ml of acid, thereafter minimal U was eluted. For this reason, a volume of 60 ml acid was used to elute the U from the column to optimise the analysis due to the reduction in UH⁺ formation in the plasma. Portes *et al.* (2018)²⁷ and Wilken *et al.* (2021)³ used 5 ml per 30 mg of TEVA resin as the rinse volume of 2 M HNO₃ which would be equivalent to approximately 35 ml for the TEVA column used. Although 99% of the U had been eluted when using a volume of 30 ml, the benefit of adding an additional 30 ml of rinse acid had the benefit of reducing the overall U in the final sample for analysis, which is particularly important when working with ultra-trace Pu isotopes in African soils.

The elution of Pu isotopes from the TEVA resin has commonly used 0.05 M ammonium oxalate in the literature, as described in Ketterer *et al.* (2004).²⁶ However, more recently, Metzger *et al.* (2019)¹⁶ found that Pu can alternatively be removed from the TEVA resin using a low acid concentration. The Idaho National Engineering and Environmental Laboratory (INEEL) method uses 0.5 M HCl as the eluent of Pu, and so for this study, the use of both 0.05 M ammonium oxalate and 0.5 M HCl was investigated to optimise the elution of Pu from TEVA columns. The UK reference sample was prepared using the dissolution method detailed and loaded onto the column. After acid rinsing of the column, each sample was eluted using the different eluent matrices in triplicate. The oxalate samples were evaporated to dryness on a hot plate, followed by an addition of concentrated nitric acid and heating to decompose hydroxylamine in the samples, which would otherwise detriment the ICP-MS plasma efficiency. They were then reconstituted into 1 ml of the analysis matrix (5% HNO3 and 2.5% HCl) and subsequently diluted to $\times 5$ with the addition of the internal standard for analysis. The HCl samples did not require removal of the matrix and were simply diluted using a solution of 5% HNO3 and the internal standard. The samples were then analysed by ICP-MS/MS to determine the ²⁴²Pu spike recovery. The spike recovery for the ammonium oxalate was $61 \pm 11\%$, compared to the HCl spike recovery of 87 \pm 17%. These findings agree with those of Metzger et al. (2019)¹⁶ and show that dilute HCl is a more suitable eluent for the determination of Pu than the commonly used ammonium oxalate. The presented method shows comparable spike recoveries with other methods utilising ICP-MS/MS analysis where reported values range between 70 and 90%.15,23,33 Using HCl not only improved spike recovery, but it also eliminated the need for extra steps in the process of removing oxalate from the sample. This increased throughput by 20% and reduced the environment burden through the removal of the complexing agent.

The product sheet for TEVA resin indicated that the retention of U isotopes in dilute HCl is also very low and as a result any remaining U on the column after the rinsing steps is likely to elute alongside the Pu isotopes. In total 0.7 μ g kg^{-1 238}U was eluted in the collection step, contributing to 0.5% of the total ²³⁸U in the sample. In comparison 0.2 μ g kg^{-1 238}U was eluted when using ammonium oxalate as the eluent. With the capabilities of ICP-MS/MS to remove the ²³⁸UH⁺ interference through mass shifting of the Pu isotopes the increased throughput of the method when using dilute HCl outweighs the additional U isotopes being eluted into the sample.

To ensure that the maximum Pu was eluted from the column within the smallest volume to avoid dilution, the volume of eluent acid was investigated. The dilute HCl was added stepwise in 0.5 ml increments to the column and collected to determine the elution profile (Fig. 6). To avoid dilution of the Pu isotopes



prior to analysis, the eluent was collected between 1 ml and 4 ml where the maximum Pu was eluted (99%).

3.3 Method performance

Blank samples were prepared in conjunction with the analysed soil samples, these consisted of finely milled silicate sand and enabled the determination of U and Pu originating from the whole dissolution and separation processes. While Pu does not occur naturally in the environment, naturally occurring U can be found throughout the environment, including in labware as well as within reagents. The use of clean acid-leached vials and ultra-pure reagents was employed to minimise the U background detected. In addition, analysis blanks containing the analysis matrix of 5% HNO3 and 2.5% HCl were analysed as part of the analytical procedure. The amounts of U and Pu in the column blanks and ICP-MS measurement only blanks (analytical blanks) were estimated based on count rates and the sensitivity of the day. The analytical blanks yielded an average of $0.009\pm 0.014~\mu g~kg^{-1}~^{238}\text{U}$ and $0.108\pm 0.105~pg~kg^{-1}~^{239+240}\text{Pu}$ (n = 32) with limits of detection of 0.021 µg kg⁻¹ and 0.15 pg kg^{-1} respectively. The method blanks yielded an average $(\pm 2\sigma)$ concentration of 0.012 \pm 0.008 $\mu g~kg^{-1}$ ^{238}U and 0.066 \pm 0.121 pg kg^{-1 239+240}Pu (n = 32) with limits of detection of 0.012 µg kg^{-1} and 0.18 pg kg^{-1} respectively. These detection limits were comparable with previous studies that used O_2 as the reaction gas, such as Zhang et al. (2021)4 who reported detection limits of 0.06 pg kg^{-1} . The detection limits using this method are improved compared to methods using NH₃ as the reaction gas $(0.16-0.55 \text{ pg kg}^{-1})$ owing to the formation of PuO_2^+ in the reaction cell, effectively mass shifting the Pu isotopes to m/z 271 away from interfering 238 UH⁺ at m/z 239.^{8,10,23}

Validation of the optimised separation method was conducted using the analysis of $^{239+240}$ Pu in CRM samples IAEA-384 over five separate analytical runs (n = 36). In addition, the UK reference soil was analysed as an indicator of precision within an analytical sequence as well as to monitor performance between different analytical batches. Table 2 reports the measured values for both the CRM sample and the UK reference site ($\pm 2\sigma$). The measured value for IAEA-384 indicates both good accuracy and similar precision to the certified value, which was determined by a combination of radiometric (alpha and gamma spectroscopy) and mass spectroscopy techniques. Using a similar method with detection by Thermo X Series II

	Certified value (Bq kg ⁻¹)	Measured value (Bq kg ⁻¹)
IAEA 384 (Fangataufa Lagoon sediment)	108 ± 13	114 ± 12
(n = 36) UK reference sample $(n = 27)$	_	0.14 ± 0.02

Table 3	Separation scheme outline for the optimized method with
volumes	and reagents for Pu separation using 2 ml volume TEVA resin

Step	Description	Optimum volume (ml)	Column volumes	Reagent
1	Precondition	4	2	2 M HNO ₂
2	Load sample	200	100	_
3	Wash TEVA	60	30	2 M HNO ₃
4	Wash TEVA	20	10	8 M HCl
5	Load column with	1	0.5	0.5 M HCl
	eluent			
6	Elute Pu isotopes	3	1.5	0.5 M HCl
7	Wash TEVA	2	1	0.5 M HCl
8	Wash TEVA	4	2	2 M HNO ₃
9	Wash TEVA	4	2	0.01 M
				HNO_3
10	Store TEVA column	3	1.5	0.01 M
				HNO_3

quadrupole ICP-MS (Bremen, Germany), Wilken *et al.* (2021)³ measured the IAEA-384 reference value to be 102 \pm 20 Bq kg^{-1} over 11 measurements. The method presented shows improved precision through the optimisation of column separation and the use of O₂ in a collision cell ICP-MS/MS to reduce the error in measurements.

Overall spike recoveries using the method in Table 3 ranged between 67 and 100% with an average of 86% (n = 32). This is comparable with other reported results of spike recovery using ICP-MS/MS in the literature, which range from 70 – 90%.^{15,23,33} This, along with the low detection limits, accuracy of measurement, and high precision supports the application of this separation and analytical method for the detection of Pu isotopes in soils.

3.4 Application for the determination of Pu isotopes in African soil samples

The method reported in Table 3 was then tested on the African soil, which was collected in Zambia by Hamilton *et al.* (2020).²⁵ The overall ²³⁹⁺²⁴⁰Pu in the topsoil samples was determined to be 29.59 \pm 0.97 pg kg⁻¹ (equivalent to 0.09 Bq kg⁻¹) with an average spike recovery of 81% (n = 3). This value is approximately 1.5× smaller than the ²³⁹⁺²⁴⁰Pu concentration found within the UK reference soil and supports the use of 50 g as the sample mass for African soils to achieve maximum sensitivity down the soil profile. The measurement of soil erosion rates

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using Pu has increases significantly in the literature, and with improvements in the separation and measurement of Pu in soils, the method has the potential to improve understanding of processes influencing erosion and inform mitigation strategies globally.^{5,14} Wilken et al. (2021)³ demonstrated the usefulness of Pu as an alternative soil erosion tracer in East Africa, informing soil degradation patterns and highlighting the need for additional studies into erosion rates in tropical soils. This study reported measured mean values at sloped cropland sites in the DR Congo and Uganda of a similar magnitude, with reported values between 0.012 and 0.046 Bq kg^{-1} . Through the detection of ultra-trace Pu in African soils as a tracer of soil erosion, the data collected can be used to reinforce sustainable soil conservation measures and aid in the validation of prediction models, allowing for a better understanding of the factors influencing accelerated erosion.

To validate the methods usability for studies into soil redistribution rates, Pu inventory was determined at a reference site in the Oroba valley, Nandi County, Kenya. The greatest concentration of ²³⁹⁺²⁴⁰Pu was found at the depth of 7.5 cm (32.30 pg kg⁻¹/0.11 Bq kg⁻¹), whereafter concentrations exponentially decreased (Fig. 7). The usability of the method for the detection of Pu isotopes in African soils is demonstrated by the ability to detect Pu at depths greater than 30 cm with concentrations being an order of magnitude greater than the methods detection limit. The total inventory for the reference site between 0 and 30 cm was determined as 16.86 Bq m⁻² and the depth profile was consistent with examples in the literature globally.5,26,30,34,35 The total inventory for the site is consistent with the estimated global fallout reported by Hardy et al. (1973)² and Kelley et al. $(1999)^{36}$ for Kenya of 19.2 Bq m⁻² and the average latitudinal distribution (0–10° S) of 11.1 \pm 7.4 Bq m⁻². In addition, the value is in line with the inventories reported by Wilken et al. (2021),3 where although the inventories in this study were lower, they agree with the differences in annual precipitation (960 and 1400 mm per year for Kenya and the White Nile-Congo rift respectively).37



Fig. 7 Depth profile of ²³⁹⁺²⁴⁰Pu inventory at reference site.

4. Conclusion

A modified, robust analytical sequence for the preconcentration and separation of ultra-trace Pu isotopes in African soils provided increased sensitivity using ICP-MS/MS with O₂ as a reaction gas to remove interferences. This improved the determination of fallout Pu activity concentrations in the southern hemisphere where Pu signals are relatively low compared to the northern hemisphere. Accuracy was improved through the elimination of the ammonium oxalate matrix in the eluent, with recoveries improved from 61 to 87%. Removal of the oxalate-sample matrix resulted in more stable plasma conditions, an increase in column separation throughput by 20%, and reduction in reagent consumption/disposal. This method reduced the acid requirement for separations by 80% compared to consensus literature reports, whilst maintaining maximum sensitivity and spike recovery. Additionally, O_2 use as a reaction gas provided low method blank measurements by ICP-MS/MS of 0.06 pg kg^{-1 239+240}Pu. The CRM sediment, IAEA-384, evidenced accuracy, with improved precision of measurement of: $(\pm 2\sigma)$ of 114 \pm 12 Bq kg $^{-1}$ $^{239+240}$ Pu. The total inventory (16.86 Bq m^{-2}) and depth profile at a reference site in Kenya is in strong agreement with the literature which reinforces this methods usefulness in the determination of soil redistribution rates in tropical soils.

This method presents a simple, cost-effective, robust sequence with reduced laboratory waste disposal, which is vital to ensure the separation method is applicable to low-resource laboratories. Analysis via ICP-MS/MS with O2 as a reaction gas offers a robust technique with high throughput compared to traditional techniques such as gamma spectroscopy, and therefore lends itself well to field and survey-scale soil erosion assessment. This outcome, along with the low detection limits that are comparable to alternative mass spectrometric methods, makes the method applicable to the detection of ultra-trace fallout Pu in African soils. Due to increasing concern regarding accelerated soil erosion and its impact on sustainable intensification of agriculture in developing countries, this work provides advancements in the detection of ²³⁹⁺²⁴⁰Pu which has proven to be a robust tracer for soil erosion. Furthermore, the optimised analytical method is a powerful tool to drive mitigation strategies through the analysis of ultra-trace Pu in African soils, ultimately improving the determination of soil erosion rates in tropical soils to better inform mitigation strategies.

Author contributions

SMD undertook fieldwork, method development, analyses and writing, TSB analyses, SRC analysis and validation, OSH fieldwork and conceptualisation, JB fieldwork and analyses, WHB conceptualisation and supervision, OO fieldwork, conceptualisation, validation, MJW conceptualisation, funding acquisition, administration, validation, writing and supervision. All authors contributed to the review and editing.

Conflicts of interest

There are no conflicts to declare.

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