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1 **Quantifying the impact of ozone on crops in Sub-Saharan Africa**
2 **demonstrates regional and local hotspots of production loss.**

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23 African crops.

24 **Abstract**

25 Tropospheric ozone can have a detrimental effect on vegetation, including reducing the quantity of
26 crop yield. This study uses modelled ozone flux values (POD₃IAM; phytotoxic ozone dose above 3
27 nmol m⁻² s⁻¹, parameterised for integrated assessment modelling) for 2015, together with species-
28 specific flux-effect relationships, spatial data on production and growing season dates to quantify
29 the impact of ozone on the production of common wheat (*Triticum aestivum*) and common beans
30 (*Phaseolus vulgaris*) across Sub-Saharan Africa (SSA). A case study for South Africa was also done
31 using detailed data per province. Results suggest that ozone pollution could decrease wheat yield by
32 between 2 and 13%, with a total annual loss of 453,000 tonnes across SSA. The impact on bean
33 production depended on the season, however estimated yield losses were up to 21% in some areas
34 of SSA, with an annual loss of ~300,000 tonnes for each of the two main growing seasons.
35 Production losses tended to be greater in countries with the highest production, for example
36 Ethiopia (wheat) and Tanzania (beans). This study provides an indication of the location of areas at
37 high risk of crop losses due to ozone. Results emphasise that efforts to reduce ozone precursors
38 could contribute to reducing the yield gap in SSA. More stringent air pollution abatement policies are
39 required to reduce crop losses to ozone in the future.

40 **Keywords:** Wheat, beans, Africa, crop production, ozone, yield loss.

41 **1. Introduction**

42 Tropospheric (ground-level) ozone is a secondary air pollutant, formed from sunlight-driven
43 reactions between precursor molecules including carbon monoxide (CO), methane (CH₄), non-
44 methane hydrocarbons and nitrogen oxides (NO_x) (Monks *et al.*, 2015). Surface ozone
45 concentrations vary spatially at the global scale. Data from ozone measurement stations around the
46 globe for 1995 – 2014 showed that while ozone levels had become more stable in Europe and
47 declined in North America, levels in rapidly developing regions such as East Asia continued to rise

48 (Mills *et al.*, 2018a). As a powerful oxidant, ozone can have a detrimental effect on vegetation,
49 entering the leaves via the stomata and disrupting metabolic pathways (Emberson *et al.*, 2018),
50 which can lead to physiological changes, including decreased photosynthesis and early leaf
51 senescence (Ainsworth *et al.*, 2012).

52 For the continent of Africa, there is little measured data on ozone concentrations, however model
53 simulations suggest that ozone increased over the last four decades in central Africa (Ziemke *et al.*,
54 2019). In South Africa, ozone monitoring is carried out across a network of air quality monitoring
55 stations (<https://saaqis.environment.gov.za/>). Data for 2019/2020 (accessed 1st June 2020) shows
56 high maximum daily ozone concentrations (above 100ppb) during the crop growing season (for
57 common wheat and beans) in states with high crop production (FAOSTAT, 2017). During biomass
58 burning season, surface O₃ concentrations can reach peaks of up to 70 ppb in Rwanda (DeWitt *et al.*,
59 2019).

60 There are a number of sources of ozone precursors in Africa, including emissions from biomass
61 burning, biogenic VOCs, lightning (NO_x), and anthropogenic activities (Bouarar *et al.*, 2011). Large
62 amounts of CO, NO_x and VOCs are emitted from burning biomass associated with savanna and forest
63 fires, which takes place during dry and monsoon periods over West and Central Africa, respectively
64 (Sauvage *et al.*, 2005). Ozone is also a transboundary pollutant as precursors can travel far from their
65 point of production before ozone is formed (Cooper & Derwent 2013). For example, dry season
66 ozone concentrations in Rwanda are increased by transport of precursor gases from biomass
67 burning in northern and southern Africa (DeWitt *et al.*, 2019). Increasing application of fertiliser to
68 crops is also predicted to lead to an increase in NO_x emissions in Sub-Saharan Africa (SSA), and in
69 turn lead to an increase in ozone concentrations (Huang *et al.*, 2018). Ozone concentrations are
70 predicted to continue to increase in developing regions around the world in the future (Turnock *et*
71 *al.*, 2018) unless precursor emissions are further controlled.

72 Tropospheric ozone can have a negative effect on the quality and quantity of crop yield in sensitive
73 species. A number of experimental studies on African crop species show a decrease in yield with
74 elevated ozone. For example, Hayes *et al.* (2019) reported reductions in total yield and 1,000-grain
75 weight for African cultivars of common wheat (*Triticum aestivum*) and widely grown cultivars of
76 common bean (*Phaseolus vulgaris*) in high treatment levels of ozone (mean daily max concentration
77 of 84.0 ± 4.5 ppb and 93.0 ± 4.3 ppb respectively). Exposure to elevated ozone led to a reduction in
78 100-seed weight and seeds per pod for the two most widely cultivated varieties of cowpea (*Vigna*
79 *unguiculata* L.) grown in Ghana (Tetteh *et al.*, 2015).

80 Modelling studies are particularly useful for providing estimates of potential ozone impacts for
81 regions with few physical ozone measurements and/or where it may be difficult or impractical to
82 carry out field experiments. Previous modelling studies, using ozone concentration based metrics,
83 estimated that 4–15% of wheat yields, 3–4% of rice yields, 2–5% of maize yields and 5–15% of
84 soybean yields are lost globally due to ozone pollution (Van Dingenen *et al.*, 2009; Avnery *et al.*,
85 2011). An alternative method is the flux-based approach (Emberson *et al.*, 2000, Pleijel *et al.*, 2007),
86 which calculates ozone uptake via the stomata and takes into account environmental conditions that
87 may influence uptake, for example, temperature, humidity and light. Flux-based models have been
88 found to perform better than concentration-based models (Pleijel *et al.*, 2007) and to be better
89 predictors of the distribution of ozone damage (Mills *et al.*, 2011). A recent global flux-modelling
90 study estimated that ozone had reduced the annual global yield of soybean (12.4%), wheat (7.1%),
91 maize (6.1%), and rice (4.4%), leading to a total of 227 million tonnes of lost yield (Mills *et al.*,
92 2018c).

93 The aim of this study is to quantify the impact of ozone on common wheat and common bean across
94 SSA using spatial data on crop production, modelled ozone flux values and flux-effect relationships
95 from experimental data. In 2017, wheat was among the top ten crops produced in Africa (FAOSTAT,
96 2017). Common beans were the most common legume produced in East Africa in 2017 (FAOSTAT,

97 2017), providing an important source of dietary protein. While global flux modelling studies have
98 been carried out for soybean, wheat, maize and rice (e.g. Mills *et al.*, 2018b,c), this new study
99 focuses on SSA, using finer resolution data (~ 0.33 by 0.33°), and a new crop (common bean), using a
100 species-specific flux relationship for widely grown bean cultivars from experimental data. A more
101 detailed case study for South Africa was also carried out, using spatial data per province.

102 **2. Materials and Methods**

103 *2.1 The EMEP-WRF model*

104 The EMEP MSC-W (European Monitoring and Evaluation Programme, Meteorological Synthesising
105 Centre-West) chemical transport model (Simpson *et al.*, 2012) has been used for air quality
106 assessments for more than 30 years. For this study, EMEP-WRF Africa version rv4.33 (Vieno *et al.*,
107 2016), which is based on the official EMEP MSC-W model (Simpson *et al.*, 2012) was used. While the
108 EMEP MSC-W model uses data from the European Centre for Medium Range Weather Forecasting
109 integrated Forecasting System (ECMWF-IFS) model, EMEP-WRF Africa uses the Weather Research
110 and Forecast (WRF) model (Skamarock *et al.*, 2019). The EMEP-WRF Africa model uses a latitude-
111 longitude grid and 21 vertical layers with thickness varying from ~ 40 m at the surface to ~ 2 km at the
112 top of the vertical boundary (~ 16 km). The WRF version 3.9.1.1 was used to calculate hourly 3D
113 meteorological data used to drive the EMEP-WRF Africa model for the year 2015. The WRF model is
114 initialised and nudges every six hours using the Global Forecast system final reanalysis (GFS-FNL)
115 data (National Centres for Environmental Prediction, 2015). Nudging is a method for ensuring that
116 model simulations resemble observed values. For further details on the WRF model methodology,
117 including different types of nudging, see Skamarock *et al.*, (2019). The model domain covers the
118 globe with a horizontal resolution of $1.0 \times 1.0^\circ$ and a nested domain covering the African continent
119 with a horizontal resolution of $\sim 0.33 \times 0.33^\circ$ ($\sim 37 \times 37$ km at the equator). The emissions used were
120 based on the IIASA ECLIPSE v6a (Evaluating the Climate and Air Quality Impacts of Short-Lived
121 Pollutants) GAINS (Greenhouse gas – Air pollution Interactions and Synergies) model for the year

122 2015 (Stohl *et al.*, 2015). The WRF model output is not compatible with the convection
123 parametrisation used in the EMEP MSC-W model, so this feature has not been included in this work.
124 Model output data went through a process of quality assurance/control before use in subsequent
125 analyses.

126 The EMEP-WRF Africa model was used to provide daily surface ozone concentrations (ppb) and
127 ozone flux (POD₃IAM; phytotoxic ozone dose above 3 nmol m⁻² s⁻¹, parameterised for integrated
128 assessment modelling) values for Africa with a 0.33 x 0.33° grid cell resolution for 2015. The DO₃SE
129 (Deposition of O₃ for Stomatal Exchange) model, originally developed in 2000 (Emberson *et al.* 2000,
130 2001) for estimating ozone deposition and stomatal flux, has been embedded within the EMEP
131 model. The DO₃SE modelling approach is based on the multiplicative stomatal conductance (gs)
132 model originally established by Jarvis (1976). The DO₃SE model simulates the effect of phenology,
133 light, temperature, vapour, pressure deficit (VPD) and soil moisture deficit on stomatal conductance.
134 These factors are considered most important in determining the opening and closing of stomatal
135 pores. The model has been developed to allow the evaluation of species-specific stomatal
136 conductance (gs) if appropriate experimental data are available to parameterise the model. The
137 DO₃SE model has been continuously improved and updated since the original modelling approach
138 was first developed by Emberson *et al.*, (2000). For further details on the equations used by the
139 DO₃SE model within the EMEP model, see Simpson *et al.*, (2012).

140 Within the EMEP model (via DO₃SE), the estimation of stomatal ozone flux follows the assumption
141 that the concentration of ozone at the canopy top represents a reasonable estimate of the
142 concentration near the sunlit upper canopy leaves (or flag leaf for wheat), at the upper surface of
143 the laminar layer. The model uses (hourly mean modelled) ozone concentration at canopy height,
144 deposition rate of ozone (allowing for boundary layer and leaf surface resistance), and the fraction
145 of ozone taken up by the stomata to calculate ozone uptake i.e. flux. For further details and full

146 equations, see the Modelling and Mapping Manual (CLRTAP, 2017). Daily flux values are calculated
147 from modelled hourly stomatal uptake and accumulated during daylight hours.

148 For this study, POD_3IAM was calculated with the model parameterised for Integrated Assessment
149 Modelling (IAM). This involves using simplified flux models for a vegetation type (e.g. crops) rather
150 than one species, which allow estimates to be provided at a larger scale than POD_YSPEC (values for a
151 specific species). There are difficulties in estimating ozone flux at higher values of Y (e.g. $Y = 6 \text{ nmol}$
152 $\text{m}^{-2} \text{ s}^{-1}$) due to the increase in uncertainty with increasing Y (CLRTAP, 2017). The simplified models
153 used for IAM do not include the modifying effect of phenology or soil moisture as inputs to calculate
154 ozone flux. When used within the EMEP model (Simpson *et al.*, 2012) however, a simplified soil
155 moisture index is included in the final calculations of ozone flux. The EMEP model outputs both
156 irrigated (calculated without soil moisture limitation) and non-irrigated (calculated using a modelled
157 soil moisture index) POD_3IAM values per grid cell. For crops, POD_3IAM is calculated over 90 days,
158 centred on the timing of mid-anthesis (flowering) and example model parameterisation values are
159 already available (CLRTAP, 2017), based on European wheat.

160 The EMEP-WRF Africa model was parameterised using the POD_3IAM crop inputs, but these were
161 adjusted for African wheat and common beans using published experimental data (for example, g_{max} ,
162 optimal temperature) from work carried out on African crops in the UK CEH Bangor solardomes 2017
163 – 2019 (Hayes *et al.*, 2019, 2020). See Supplementary Material, Tables S1 and S2 for the model input
164 values. Hayes *et al.* (2020) provide further details on the experimental set up.

165 *2.2 Crop production data*

166 Crop production data were obtained from the Spatial Production Allocation Model (SPAM; You *et al.*,
167 2014), resolution 0.0833 by 0.0833° for the year 2010. Both irrigated and non-irrigated production
168 (tonnes) data for SSA were downloaded for wheat and beans.

169 Using ArcGIS (version 10.6), a 0.33 x 0.33° fishnet grid (aligned with the EMEP data grid) was created
170 and for each crop, production per cell was summed. Cells were designated a SSA country, based on
171 the area of the country within the cell. For each country, a conversion factor was calculated using
172 national production data from the Food and Agriculture Organisation of the United Nations (FAO)
173 (2015 production/2010 production), then the production per grid cell was multiplied by this value to
174 provide an estimate of production in 2015.

175 Similarly, for the South African case study, each grid cell was designated a province, based on the
176 area of the province within the cell. For each province, a conversion factor was calculated using crop
177 production data for 2010 and 2015 published by the South African government (Directorate of
178 Statistics and Economic Analysis 2018) (2015 production/2010 production), then the production per
179 grid cell was multiplied by this value to provide an estimate of production in 2015.

180 *2.3 Calculating 90-day POD₃IAM*

181 For each crop, a 90-day period before harvest was designated for the SSA countries listed in the
182 SPAM data as crop producers (Tables S3 & S4). This time-period includes the ozone-sensitive period
183 between anthesis and end of grain fill (Soja *et al.*, 2000). Growing season information was taken
184 from the FAO crop calendar (<http://www.fao.org/agriculture/seed/cropcalendar/welcome.do>) and
185 the FAO Global Information and Early Warning System country profiles
186 (<http://www.fao.org/gIEWS/country-analysis/en/>). If there was no information available for a
187 particular country, then the growing season for the nearest country with the crop growing in the
188 same climate zone was used, based on the global “Climatic Zone” GIS raster layer produced by the
189 European Soil Data Centre (ESDAC) at JRC (Joint Research Centre).

190 The POD₃IAM over the designated 90-day period was calculated for each 0.33 by 0.33° grid cell in SSA
191 and mapped for each crop. Due to the variation in climate zone and altitude across Africa, crops are
192 grown throughout the year and beans can have more than one growing season in a year in some
193 African countries. The POD₃IAM was therefore calculated for two growing seasons for beans, a)

194 growing seasons spanning March – July (spring - summer in Northern Hemisphere); b) growing seasons
195 spanning August to February (autumn - winter in Northern Hemisphere; Table S4). Irrigated and non-
196 irrigated accumulated POD_3IAM per grid cell were calculated. Grid cells were classed as irrigated (>75%
197 irrigated production per cell) and non-irrigated (<75% irrigated production per cell). For irrigated grid
198 cells, irrigated POD_3IAM values were used, for non-irrigated cells, non-irrigated POD_3IAM values were
199 used.

200 The start and end of the growing season for each crop may vary within a country, for example, due
201 to weather conditions or elevation. Hence, we used South Africa as a case study to allow detailed
202 information on growing seasons for crops in each province to be used. The growing season for wheat
203 production in South Africa varies depending on province and irrigation practices. Growing season
204 information for irrigated and rainfed wheat was taken from the USDA Major World Crop Areas and
205 Climate Profiles (<http://www.usda.gov/oce/weather/pubs/Other/MWCACP/index.htm>), the Abstract
206 of Agricultural Statistics (Directorate of Statistics and Economic Analysis, 2018) and Guidelines for
207 the Production of Small Grains in the Summer Rainfall Region (ARC-Small Grain, 2019). Using the
208 SPAM production data for 2010, wheat production totals (irrigated and non-irrigated) were
209 calculated for each South African province (Table S5) and subsequently converted to 2015
210 production data as described above.

211 For South African provinces with the majority of production being rainfed (Western Cape, Free State,
212 North West, KwaZulu-Natal, Mpumalanga and Gauteng), non-irrigated ozone data were used to
213 calculate the 90-day POD_3IAM . For provinces with >40% irrigated production overall (Northern Cape,
214 Limpopo and Eastern Cape), grid cells were split into irrigated (with >75% irrigated production per grid
215 cell) and non-irrigated (<75% irrigated production per grid cell). For irrigated grid cells in these
216 provinces, the irrigated growing period and ozone data were used to calculate POD_3IAM . For non-
217 irrigated cells, non-irrigated ozone data and growing period were used. See Table S6 for the wheat
218 growing seasons used for each South African province.

219 Bean planting times per province were taken from [https://www.horticulture.org.za/green-bean-](https://www.horticulture.org.za/green-bean-planting-times/)
220 [planting-times/](https://www.horticulture.org.za/green-bean-planting-times/). Production data per province (SPAM 2010 spatial data) indicated that while
221 irrigation of beans varied per state, the majority of beans produced were rainfed (Table S7). No
222 specific information on growth periods for irrigated and non-irrigated beans could be sourced,
223 therefore the primary growing periods for each province were used for all grid cells. When irrigated
224 production per grid cell was >75%, irrigated ozone data were used to calculate the POD₃IAM over 90
225 days, otherwise non-irrigated ozone data were used. Bean growing period information also indicated
226 that the growing season varied depending on the presence of frost in the winter. South Africa has
227 areas known as the Highveld, (1500m < elevation < 2100m) and the Lowveld (150m < elevation < 600
228 m). Beans growing in the Highveld (exposed to winter frost) grow at a different time of year from
229 beans growing in the Lowveld (hot summers and frost-free winters). Using SRTM (Shuttle Radar
230 Topography Mission, v4.1) elevation data (90m resolution) to calculate the average elevation per
231 grid cell, cells in the provinces of Mpumalanga, Limpopo and Kwazulu-Natal were split into 'Highveld'
232 and 'Lowveld' and assigned 90-day accumulation periods based on this classification (Table S8).

233 *2.4 Yield loss calculations*

234 Using the POD₃IAM values per grid cell, the percentage yield loss due to ozone for each crop was
235 calculated. For wheat, the African wheat yield data collected in the solardome experiments was not
236 a large enough sample to calculate a robust flux effect relationship (Hayes *et al.*, 2020). However,
237 Hayes *et al.* (2019) show that the African wheat cultivars tested have a similar ozone sensitivity as
238 European cultivars. Therefore, the POD₃IAM flux-effect relationship for European wheat from the
239 Modelling and Mapping Manual (CLRTAP, 2017) was used, using the most up to date methodology
240 adopted by the convention:

$$241 \quad \% \text{ yield loss} = (POD_3IAM - 0.1) * 0.64 \quad \text{Equation 1}$$

242 where 0.1 is a reference value used to represent ozone uptake at pre-industrial or natural ozone levels
243 and subtracted before yield loss is calculated and 0.64 is the slope of the relationship between

244 POD₃IAM and percentage yield reduction (Fig. S1), and represents the percentage reduction per mmol
245 m⁻² POD₃IAM.

246 For beans, experimental data from the UK CEH Bangor solardomes (2017-2019) was used to
247 calculate a flux-effect relationship (Fig. S2). POD₃IAM was estimated using the DO₃SE model stomatal
248 algorithm (Emberson *et al.*, 2000, 2001, CLRTAP, 2017). Following Zuur *et al.* (2009), generalised
249 least squares (gls) regression (using the R 'nlme' package (Pinheiro *et al.*, 2018)) was used to model
250 the relationship between relative yield and POD₃IAM, to allow for the spread in the data as POD₃IAM
251 increased. The reference value was calculated for the experimental conditions under which the flux-
252 effect relationship data were collected, assuming constant 10 ppb ozone across the 90-day period.
253 For the bean experiment, this reference value was zero.

254 Therefore the equation for the bean flux-effect relationship was:

$$255 \quad \% \text{ yield loss} = \text{POD}_3\text{IAM} * 1.175 \quad \text{Equation 2}$$

256 *2.5 Production loss calculations*

257 Following Mills *et al.*, (2018b), production loss due to ozone was calculated using the following
258 equation:

$$259 \quad \text{Production loss} = (\text{Total production}/(\text{Relative yield})) - \text{Total production} \quad \text{Equation 3}$$

$$260 \quad \text{Where 'relative yield'} = 1 - (\% \text{ yield loss}/100) \quad \text{Equation 4}$$

261 For SSA countries where beans were grown in two seasons, annual production was divided by two
262 before calculating production loss for each season.

263 **3. Results**

264 *3.1 Surface ozone concentrations in Africa*

265 Daily mean ozone concentrations varied with season (Fig. 1). During the first three months of the
266 year, concentrations in SSA were highest in western and central areas, including the Democratic
267 Republic of the Congo (DR Congo) and the Central African Republic (Fig. 1a). During April to June,
268 ozone concentrations in SSA were slightly lower, with the highest values still in central/western
269 areas, including Chad, Nigeria, Cameroon, Burkina Faso and the Central African Republic (Fig. 1b).
270 Between July and September 2015, the highest ozone concentrations were found slightly further
271 south, particularly in DR Congo and Angola (Fig. 1c). The final period, October to December 2015,
272 showed the lowest ozone concentrations, with pockets of slightly higher values across SSA, for
273 example in Nigeria, Botswana and Ethiopia (Fig. 1d).

274 *3.2 Ozone impact on wheat production in Sub-Saharan Africa*

275 Wheat is produced across SSA, with production levels highest in Ethiopia, South Africa and Sudan
276 (Fig. 2a, Table 1). FAO national data suggested that wheat production increased for many countries
277 between 2010 and 2015, particularly Rwanda and Mali (Table 1). In 2015, average ozone flux
278 (POD_3IAM) values were highest in Cameroon ($20.2 \text{ mmol m}^{-2} \pm 7.9\text{sd}$) and Zambia ($14.3 \text{ mmol m}^{-2} \pm$
279 3.3sd) (Fig. 2b, Table S9). In the countries with the highest wheat production, average POD_3IAM
280 values were lower, for example, in Ethiopia ($9.7 \text{ mmol m}^{-2} \pm 3.8\text{sd}$) and South Africa ($4.3 \text{ mmol m}^{-2} \pm$
281 2.5sd) (Table 1). Average percentage yield loss estimates for wheat varied between 2 and 13%, with
282 high losses in Cameroon ($13.0\% \pm 5.0\text{sd}$) and Zambia ($9.1\% \pm 2.1\text{sd}$) (Fig. 2c, Table S9). Figure 2b&c
283 also show that there were pockets of high ozone flux and yield loss estimated within some countries,
284 for example in northern and western areas of Ethiopia and southern areas of South Sudan.
285 Production losses in SSA due to ozone, which account for the combination of the total amount of
286 wheat grown and the % yield loss, were highest in the countries with the highest wheat production,
287 including Ethiopia and Sudan (Fig. 2d, Table 1). Rwanda also had a relatively high production loss of
288 46,000 tonnes. The total estimate for wheat production loss due to ozone for SSA in 2015 was
289 453,000 tonnes (Table S9).

290 **Table 1.** Wheat production and ozone-induced losses per country in Sub-Saharan Africa for the 10
 291 countries with the highest estimated production loss in 2015. SPAM production data are used for
 292 2010, values for production in 2015 have been estimated using conversion factors from FAO national
 293 totals.

Country	2010 prod. (Th. t)	2015 prod. (Th. t)	Av. POD ₃ IAM (mmol m ⁻²)	±sd	Av. % yield loss	±sd	Total prod. loss (Th. t)
Ethiopia	2950	4805	9.7	3.8	6.1	2.4	233
Sudan	442	853	7.5	4.5	4.7	2.9	69
Rwanda	72	758	10.8	2.4	6.9	1.5	46
South Africa	1792	1826	4.3	2.5	2.7	1.6	46
Zambia	201	250	14.3	3.3	9.1	2.1	27
Mali	24	87	9.2	3.6	5.8	2.3	7
Tanzania	88	102	10.3	3.0	6.5	1.9	5
Kenya	331	154	6.9	3.6	4.4	2.3	4
Nigeria	115	63	4.1	3.8	2.5	2.4	3
Zimbabwe	19	49	5.8	3.7	3.7	2.4	3

294

295 3.3 Ozone impact on bean production in Sub-Saharan Africa

296 3.3.1 Season 1

297 Common beans are widely grown across SSA. Countries with the highest production include
 298 Tanzania, Kenya and Uganda (Fig. 3a, Table 2). FAO national data suggested that bean production
 299 increased for the majority of countries between 2010 and 2015, particularly Togo, DR Congo and
 300 Kenya (Table 2). In season 1, ozone flux values in 2015 were highest in the central countries of SSA
 301 (Fig. 3b), with the highest average values in Congo (17.9mmol m⁻² ± 3.1sd) and DR Congo (15.8mmol
 302 m⁻² ± 4sd) (Table S10). In countries with the highest bean production, average flux values were
 303 lower, for example, Tanzania (10.1 mmol m⁻² ± 2.9sd) and Kenya (11.3 mmol m⁻² ± 4.6sd) (Table 2).
 304 Estimated (average) percentage yield losses due to ozone were again highest in the central area of
 305 SSA, with values ranging from 0.9% ± 0.7sd (Mauritania) to 21.0% ± 3.7sd (Congo) (Fig. 3c, Table
 306 S10). Other countries with high estimates for average yield loss included DR Congo (18.6% ± 4.7sd),
 307 Togo (17.5% ± 3sd) and Uganda (16.5% ± 5.1sd) (Table 2). Countries with the highest bean

308 production showed slightly lower estimates for average percentage loss, for example Tanzania
 309 (11.9% ± 3.4sd) and Kenya (13.3% ± 5.4sd). Production losses due to ozone were generally highest in
 310 the countries with the highest bean production, for example Tanzania (55,000 tonnes), and Kenya
 311 (47,000 tonnes) (Fig. 3d, Table 2). As Uganda had a high estimate for percentage yield loss (16.5%),
 312 estimated production loss was also high (43,500 tonnes), despite the country not being one of the
 313 top producers. Total estimated production loss for beans due to ozone in season 1 was 370,000
 314 tonnes (Table S10).

315 **Table 2.** Bean production and ozone-induced losses (season 1) per country in Sub-Saharan Africa for
 316 the 10 countries with the highest estimated production loss in 2015. SPAM production data are used
 317 for 2010, values for 2015 have been estimated using conversion factors from FAO national totals. For
 318 countries where beans were grown in two seasons, annual production and estimated production
 319 losses were divided by two.

Country	2010 prod. (Th. t)	2015 prod. (Th. t)	Av. POD ₃ IAM (mmol m ⁻²)	± sd	Av. % yield loss	± sd	Total prod. loss (Th. t)
Tanzania	382	529	10.1	2.9	11.9	3.4	54.86
Kenya	240	469	11.3	4.6	13.3	5.4	47.05
Uganda	249	266	14.0	4.3	16.5	5.1	43.49
Ethiopia	182	318	11.5	4.2	13.5	4.9	37.62
Cameroon	170	175	13.7	7.3	16.1	8.6	34.81
Ghana	210	212	12.9	2.5	15.1	2.9	31.64
Togo	76	177	14.9	2.5	17.5	3.0	28.33
Rwanda	156	207	8.9	2.5	10.4	3.0	20.05
DR Congo	75	156	15.8	4.0	18.6	4.7	18.13
Benin	110	101	12.1	5.0	14.2	5.9	16.77

320

321 3.3.2 Season 2

322 Bean production in the second growing season was more focused on the central area of SSA
 323 (whereas in the first growing season, there was more production in Western Africa) (Fig. S3).

324 Countries with the highest production include Tanzania, Kenya and Angola (Fig. S3a, Table S11). FAO
 325 national data suggested that bean production increased for the majority of countries between 2010

326 and 2015 (Tables S11, S12). Ozone flux values were generally highest in the central part of SSA,
327 however high values were also seen in South Africa (Fig. S3b). Countries with the highest average
328 flux values in season 2 included Cameroon ($13.4 \text{ mmol m}^{-2} \pm 6.4\text{sd}$), DR Congo ($12.6 \text{ mmol m}^{-2} \pm$
329 3.0sd) and Angola ($12.1 \text{ mmol m}^{-2} \pm 3.1\text{sd}$) (Table S11). In this second growing season, some
330 countries with high bean production had relatively high average flux values, for example Angola,
331 however values were lower in Tanzania ($5.3 \text{ mmol m}^{-2} \pm 3.0\text{sd}$) and Kenya ($6.7 \text{ mmol m}^{-2} \pm 3.6\text{sd}$)
332 (Table S11). Similarly, estimated percent yield losses due to ozone were highest in central areas of
333 SSA, with high values also seen in South Africa (Fig S3c). The highest estimated average percentage
334 yield losses were seen in Cameroon ($15.7\% \pm 7.5\text{sd}$), DR Congo ($14.8\% \pm 3.5\text{sd}$) and Mozambique
335 ($14.4\% \pm 1.8\text{sd}$) (Table S12). Some countries with high bean production, had relatively high average
336 yield loss values, for example Angola, however values were lower in Tanzania ($6.3\% \pm 3.5\text{sd}$) and
337 Kenya ($7.8\% \pm 4.3\text{sd}$) (Table S11). The highest production losses due to ozone were estimated for
338 Angola (60,000 tonnes), Uganda (42,500 tonnes) and Tanzania (41,500 tonnes) (Fig. S3d, Table S11).
339 While Uganda was not one of the highest producing countries, the relatively high yield loss led to a
340 high estimate of production loss. Total estimated production loss for beans due to ozone in season 2
341 was 337,000 tonnes (Table S12).

342 *3.4 South Africa case study*

343 *3.4.1 Ozone impact on wheat production*

344 Key wheat production areas in South Africa were the Western Cape, Free State and Northern Cape
345 provinces (Fig. 4a), with 37%, 28% and 15% production respectively in 2010 (Table 3). Government
346 data per province show that in 2015 wheat production had increased in some states (e.g. Western
347 Cape and Northern Cape) but decreased in others (e.g. Free State) (Table 3). In 2015, ozone flux
348 (POD_3IAM) values for wheat were highest in the north of the country, in Limpopo province, with values
349 $>12 \text{ mmol m}^{-2}$ (Fig. 4b). The highest average flux value was for the Free State ($8.8 \text{ mmol m}^{-2} \pm 2.0\text{sd}$;
350 Table 3), where 28% of South African wheat was produced in 2010. In the Western Cape, the province

351 with the highest wheat production in 2010 (and 2015), ozone flux values were low. The average flux
 352 for South Africa was 6.6 mmol m⁻² (± 3.3sd). Figure S4 shows where irrigated wheat cells were located
 353 in the Northern Cape, Limpopo and the Eastern Cape. It can be seen that ozone flux values for irrigated
 354 cells were higher than other cells within these provinces, particularly in Limpopo. Irrigated wheat is
 355 grown later in the season when ozone concentrations are slightly higher and irrigation increases the
 356 ozone uptake by the plants.

357 Estimates of (average) wheat yield loss due to ozone in 2015 varied between <2 and >8%, with grid
 358 cell values highest in the north of the country, in Limpopo province (Fig. 4c). The highest average yield
 359 loss was seen in the Free State (5.6% ± 1.3sd; Table 3). In the provinces with the greatest levels of
 360 wheat production, yield loss varied between primarily <2% for the Western Cape, areas of 6-8% loss
 361 for the Free State and primarily 4-6% yield loss for the Northern Cape. Across South Africa, the average
 362 yield loss for wheat was 4.2% (± 2.1sd). Wheat production losses due to ozone in 2015 were highest
 363 in the Northern and Western Cape provinces of South Africa (Fig. 4d, Table 3), with 15, 700 and 14,
 364 000 tonnes lost respectively. The Free State (13, 700 tonnes) and Limpopo (10, 700 tonnes) also
 365 showed relatively high losses. The total estimated wheat production loss due to ozone for 2015 in
 366 South Africa was 61,700 tonnes (Table 3).

367 **Table 3.** Wheat production and ozone-induced losses per province in South Africa in 2015. Production
 368 data are from the SPAM 2010 model. Production was estimated for 2015 using a conversion factor
 369 value, calculated per province from South African government production data for the years 2010 and
 370 2015.

Province	2010 prod. (Th. T)	2015 prod. (Th. T)	Av. POD ₃ IAM (mmol m ⁻²)	± sd	Av. % yield loss	± sd	Total prod. loss (Th. T)
Western Cape	670	881	2.5	0.8	1.6	0.5	14.04
Northern Cape	272	279	8.2	1.2	5.2	0.7	15.68
Free State	520	242	8.8	2.0	5.6	1.3	13.65
Limpopo	109	240	7.0	4.9	4.4	3.2	10.69
North West	123	85.97	4.0	2.1	2.5	1.4	2.39
KwaZulu-Natal	39.88	54.50	7.2	3.5	4.6	2.2	2.94
Mpumalanga	38.65	32.21	4.7	3.0	2.9	1.9	1.21

Eastern Cape	22.99	19.16	7.8	1.6	5.0	1.0	0.97
Gauteng	11.27	2.25	7.9	1.4	5.0	0.9	0.12
South Africa	1,806	1,837	6.6	3.3	4.2	2.1	61.69

371

372 3.4.2 Ozone impact on bean production

373 The majority of bean production was in the Free State, Mpumalanga and the Eastern Cape (Fig. 5a),
374 with small amounts of production in the other provinces (Table 4). Bean production in 2015 increased
375 in the Free State and Eastern Cape compared to in 2010, however decreased in other provinces,
376 particularly in Mpumalanga and KwaZulu-Natal (Table 4). In 2015, ozone flux (POD₃IAM) values for
377 bean were highest in the north of the country, particularly in Limpopo and also parts of the North
378 West, Gauteng and Mpumalanga provinces (Fig. 5b). The highest average flux value was for Gauteng
379 province (14.3 mmol m⁻² ± 1.9sd; Table 4). Flux values were also relatively high in the Free State (up to
380 15-18 mmol m⁻²), which is where the majority of bean production in South Africa occurs. The average
381 flux for South Africa for beans was 11.3 mmol m⁻² (± 3.9sd). Figure S5 shows where Lowveld cells are
382 located in Limpopo, Mpumalanga and KwaZulu-Natal. It can be seen that ozone flux values for Lowveld
383 cells were generally low compared to other cells within these provinces. Beans in the Lowveld region
384 are grown at a different time in the season (planted in March-April compared to Dec-Jan) when ozone
385 concentrations are slightly lower, ozone concentrations also tend to increase with altitude.

386 Estimated yield loss for beans in 2015 was considerably higher than for wheat, with maximum values
387 >20% (Fig. 5c). Losses were highest for grid cells in the northern part of South Africa, particularly in
388 Limpopo and Gauteng. The highest average yield loss was seen in Gauteng (16.8% ± 2.2sd, Table 4).
389 Areas of losses >20% were also seen in the North West, Mpumalanga and KwaZulu-Natal provinces.
390 In the Free State, the province with the greatest bean production in South Africa, estimated yield
391 losses were relatively high, particularly in the eastern half of the province, with estimates of 17.5-20%
392 yield losses. The Lowveld grid cells showed the lowest estimated yield loss, <5% (See Fig. S5 for map
393 of Lowveld cells). The average yield loss across South Africa for beans was 13.3% (± 4.6sd). In 2015,

394 bean production losses due to ozone were lower than estimates for wheat. The highest production
 395 losses were seen in the Free State (6,591 tonnes) and Mpumalanga (1,542 tonnes) (Fig. 5d, Table 4).
 396 The total estimated bean production loss due to ozone for South Africa was 11,800 tonnes.

397 **Table 4.** Bean production and ozone-induced losses per province in South Africa in 2015. Production
 398 data are from the SPAM 2010 model. Production was estimated for 2015 using a conversion factor
 399 value, calculated per province from South African government production data for the years 2010
 400 and 2015.

Province	2010 prod. (T)	2015 prod. (T)	Av. POD_3IAM (mmol m^{-2})	\pm sd	Av. % yield loss	\pm sd	Total prod. loss (T)
Free State	20568	33300	13.7	2.3	16.2	2.7	6591
Mpumalanga	11789	8322	11.6	5.3	13.7	6.2	1542
Eastern Cape	1800	7201	10.9	1.7	12.8	2.0	1070
Gauteng	3288	4326	14.3	1.9	16.8	2.2	889
KwaZulu-Natal	6837	3703	10.2	5.1	12.0	6.1	522
Limpopo	7894	3107	13.5	5.1	15.9	6.0	650
North West	4957	2163	13.4	2.9	15.8	3.4	459
Northern Cape	1023	614	8.8	1.9	10.4	2.2	69.4
Western Cape	630	420	8.6	2.4	10.1	2.8	41.5
South Africa	58,786	63,156	11.3	3.9	13.3	4.6	11,833

401

402 **4. Discussion**

403 *4.1 Spatial and seasonal variation in ozone impacts*

404 This study indicates that ground-level ozone can have a substantial negative impact on the production
 405 of both wheat and beans in SSA. However, the spatial patterns and the location of the highest losses
 406 varied for the two crops. Results indicated that the areas with the highest estimated percentage losses
 407 do not necessarily have the highest production losses, because the highest ozone flux values might
 408 occur where production is low. For wheat, the greatest production losses were estimated for countries
 409 with the most production, including Ethiopia and Sudan. For beans in season 1, production losses due
 410 to ozone were generally highest in the countries with the highest production, while in season 2, not

411 all of the countries with the greatest production loss were those with the greatest production (for
412 example, Uganda).

413 For South Africa, average yield loss due to ozone for wheat and beans was 4.2% and 13.3%
414 respectively, however as more wheat is produced in South Africa than beans, the overall production
415 losses for wheat were higher than for beans. The region with high ozone flux and percentage yield loss
416 values for beans (north-east) is the largest industrial area of South Africa with emissions of
417 atmospheric pollutants from anthropogenic activities (Lourens *et al.* 2011, 2012). Spring maximum
418 ozone concentrations (the key growing season for beans), can also be attributed to regional biomass
419 burning (Laban *et al.*, 2018). Due to their high protein content, it is predicted that efforts may be made
420 to increase bean yield in the future. Estimates of percentage yield loss provide important information
421 on where production loss due to ozone could become a significant problem, if crop production were
422 to increase in the future.

423 Results for South Africa in the broad-scale and more detailed study differed slightly. For example, for
424 wheat, although estimated total production losses were similar, the case study using finer scale data
425 identified additional areas 'at risk,' which were not detected at a coarser resolution. A small difference
426 in growing season could be important if this causes a greater overlap with the time-period of highest
427 ozone flux. Detailed studies for individual countries can be more informative, particularly if data are
428 also available on ozone flux-effect relationships with local crop cultivars. However if detailed data are
429 difficult to source, broad-scale studies are a useful first step to provide an overview of the potential
430 risk of ozone impacts. These studies highlight high risk areas, which could then be followed up with
431 further investigation, including the placement of relatively low cost methods of measuring local ozone
432 concentrations, for example, passive (diffusion) samplers (Saitanis *et al.*, 2020).

433 The South African study shows that crops sown in slightly different growing seasons, due to timing of
434 irrigation or elevation, can be exposed to differing levels of ozone flux. For example, for wheat, ozone
435 flux values for irrigated grid cells tended to be higher than other neighbouring cells. Irrigated wheat is

436 grown later in the season when ozone concentrations are slightly higher and irrigation can increase
437 ozone uptake as leaf stomata open to allow water intake. Increased ozone flux and yield loss were
438 demonstrated for irrigated wheat in a recent global modelling study (Mills *et al.*, 2018b). There may
439 be potential for crop growers to mitigate the impact of ozone on yield by practicing irrigation and crop
440 management (e.g. sowing date) strategies. Harmens *et al.* (2019) investigated the effect of irrigation
441 regimes on Kenyan wheat growing in low (30 ppb) and high (80 ppb) ozone treatments. Reduced
442 irrigation stimulated grain weight and harvest index, which compensated for ozone induced
443 reductions in well-watered plants.

444 4. 2 Sources of variation in modelling results

445 4.2.1 Comparison with other global modelling studies

446 Mills *et al.* (2018b) evaluated the POD₃IAM modelling approach (at the global scale) using the EMEP-
447 MSC West model, finding that the model captured spatial and temporal variations when modelled
448 and measured ozone data were compared. In the current study, the EMEP-WRF model (Vieno *et al.*,
449 2016) was used, which works with a different source of meteorological data. Also more up to date
450 crop production data (SPAM 2010, You *et al.*, 2014) were used in the current study. A comparison of
451 wheat yield loss estimates for SSA between Mills *et al.* (2018b, c) and the current study does show
452 higher estimates in some countries for the latter. Further differences between the studies include
453 the resolution and the years of ozone flux data used. Due to computational constraints, the EMEP-
454 WRF model was run for only one year (2015), while Mills *et al.* (2018b, c) ran the EMEP-MSW West
455 model for the years 2010-2012. Due to the lack of measured ozone data for SSA, it was difficult to
456 test if differences in model output were due to the time-periods used. Ozone concentrations (and
457 flux values) can be expected to vary between years (e.g. Balashov *et al.*, 2014), primarily due to
458 variations in weather and emissions of precursors. To build on the current results, future modelling
459 should be carried out with further years of data (or an average across years) to allow for annual
460 variation in ozone levels.

461 Modelling studies using concentration-based metrics to estimate yield loss (e.g. Van Dingenen *et al.*,
462 2009, Avnery *et al.*, 2011) show varying results for SSA, depending on the model and metric used.
463 Van Dingenen *et al.* (2009) estimate wheat yield loss (in year 2000) for some areas of SSA, e.g.
464 Zambia, at ~30%, however the model used was found to overestimate ozone levels for some African
465 areas and model resolution was reduced over Africa. Avnery *et al.* (2011) present national relative
466 yield loss for wheat (year 2000), for example 2-4% (M12) and 6-8% (AOT40) in South Africa, which
467 are more similar to the results for the current study.

468 4.2.2 Influence of growing season and crop cultivar

469 The growing seasons used for wheat and bean in this study were chosen to cover the majority of the
470 production of each crop. However suggested planting times for beans can span several months. The
471 optimum planting time for each country/province was chosen where possible, however the 90-day
472 ozone flux for beans planted in October would differ from those planted in December. Other factors,
473 including temperature, may also influence growing season length. An additional factor to include if
474 possible is how planting times for irrigated crops vary from rainfed crops. There are many different
475 wheat cultivars grown across SSA. Within South Africa, wheat cultivars have slight variations in
476 optimum planting time (ARC-Small Grain, 2019). This could become increasingly important into the
477 future if new cultivars are developed that might be selected to grow in a slightly different growing
478 season to account for changes in climate.

479 Different cultivars of a crop can show differing sensitivity to ozone, for example, different bean
480 cultivars showed varying effects on yield after ozone exposure (Hayes *et al.*, 2019). Flux-effect
481 relationships are calculated using as many cultivars as possible to allow for this variation. However,
482 for particularly sensitive cultivars, losses could be higher than estimated in this study. The flux-effect
483 relationship used to calculate yield loss for African wheat (CLRTAP, 2017) uses data compiled from 13
484 experiments on five wheat cultivars, carried out in four European countries (Pleijel *et al.*, 2007). Data
485 were collected on the response of African wheat yield to ozone (Hayes *et al.*, 2020) however there

486 were not enough data points for a robust relationship. Studies have indicated that the ozone
487 sensitivity of crop cultivars could vary between continents (e.g. Emberson *et al.*, 2009, Osborne *et al.*,
488 2016). However, Hayes *et al.* (2019) report that the African wheat cultivars tested had a similar ozone
489 sensitivity as European cultivars. Further experimental work is required to develop a flux-effect
490 relationship specifically for African wheat cultivars to improve the predictions of potential production
491 losses. This should include as many cultivars of varying ozone sensitivity as possible to provide a robust
492 estimate of ozone impact on African crop production. For crop cultivars that are used widely and/or
493 of particular economic importance in a region or country, it would also be beneficial to gather enough
494 data to develop individual flux-effect relationships.

495 4.2.3 Crop production estimates

496 The SPAM crop production data are estimated using a combination of regional/national production
497 data and spatial modelling (You *et al.*, 2014). When compared with South African government data
498 (totals per province), values were in the same order of magnitude, however totals were not identical.
499 Conversion factors (national or per province for South Africa) were used to convert between 2010 and
500 2015 production data. This method does not allow for any changes in distribution of production
501 between 2010 and 2015, for example, new areas of production. The SPAM data may also not include
502 all countries producing the crops of interest, for example, there are no data for bean production in
503 Mali however, FAOSTAT (2017) provides a national production total for this country. Studies such as
504 this could be used to highlight regions where it may be possible to expand production of a crop into a
505 new area, but could also show where expansion of production areas should be treated with caution,
506 as the potential yield losses due to ozone might outweigh the benefits of changing to a different crop.
507 Additional factors must be taken into account however, for example, Rippke *et al.* (2016) report that
508 60% of the present suitable areas for common beans in SSA are likely to require a “transformational
509 change,” for example switching crop types or moving away from agriculture, by 2100 due to predicted
510 changes in climate.

511 4.3 Measured vs. modelled values

512 Due to the potential reduction in crop production reported in this study, it is important to validate the
513 predictions. There is a focus on air pollution monitoring in cities, but this study shows that it is
514 important to carry out more standardised ozone monitoring measure in agricultural areas, for
515 example, in central countries of SSA. The estimated values of surface concentration from the EMEP-
516 WRF model can be compared with measured ozone data from monitoring stations in South Africa
517 (Laban *et al.*, 2018). Both the observed and modelled data show higher values towards the end of the
518 year (spring-summer). However, the modelled EMEP data also show high values at the start of the
519 year, which are not seen so clearly in the measured data. The mapped EMEP data are daily mean
520 values (between 6am and 6am) presented for every 3 month period in 2015, while the observed data
521 are hourly median values for each month, spanning a period of 2006 – 2015, therefore the outputs
522 can't be directly compared. Further comparison between modelled and measured data (across SSA)
523 would be useful. Laban *et al.* (2020) highlight how regional-scale ozone precursors and meteorological
524 conditions can both influence daily ozone levels in South Africa. The EMEP model uses emissions of
525 precursors as input data. While emission inventories are well developed for Europe, there are less up
526 to date data available for Africa (including on how annual emissions are distributed across the year),
527 which shows that it is very important to have improved information on the emission of precursors,
528 such as from biomass burning, which has been shown to be an important local/regional source.

529 4.4 Future work

530 This study highlights that tropospheric ozone has the potential to reduce the production of staple
531 food crops in Africa. In 2015, the total requirement for wheat in South Africa was higher than the
532 total production (Department of Agriculture, Forestry and Fisheries, 2016), leading to importation of
533 wheat from other countries. Van Ittersum *et al.* (2016) demonstrate the threat to food security in
534 SSA due to the contrast between rapidly increasing population and demand and slow growing cereal
535 yields, highlighting the need to focus on methods of closing the yield gap. To build on current

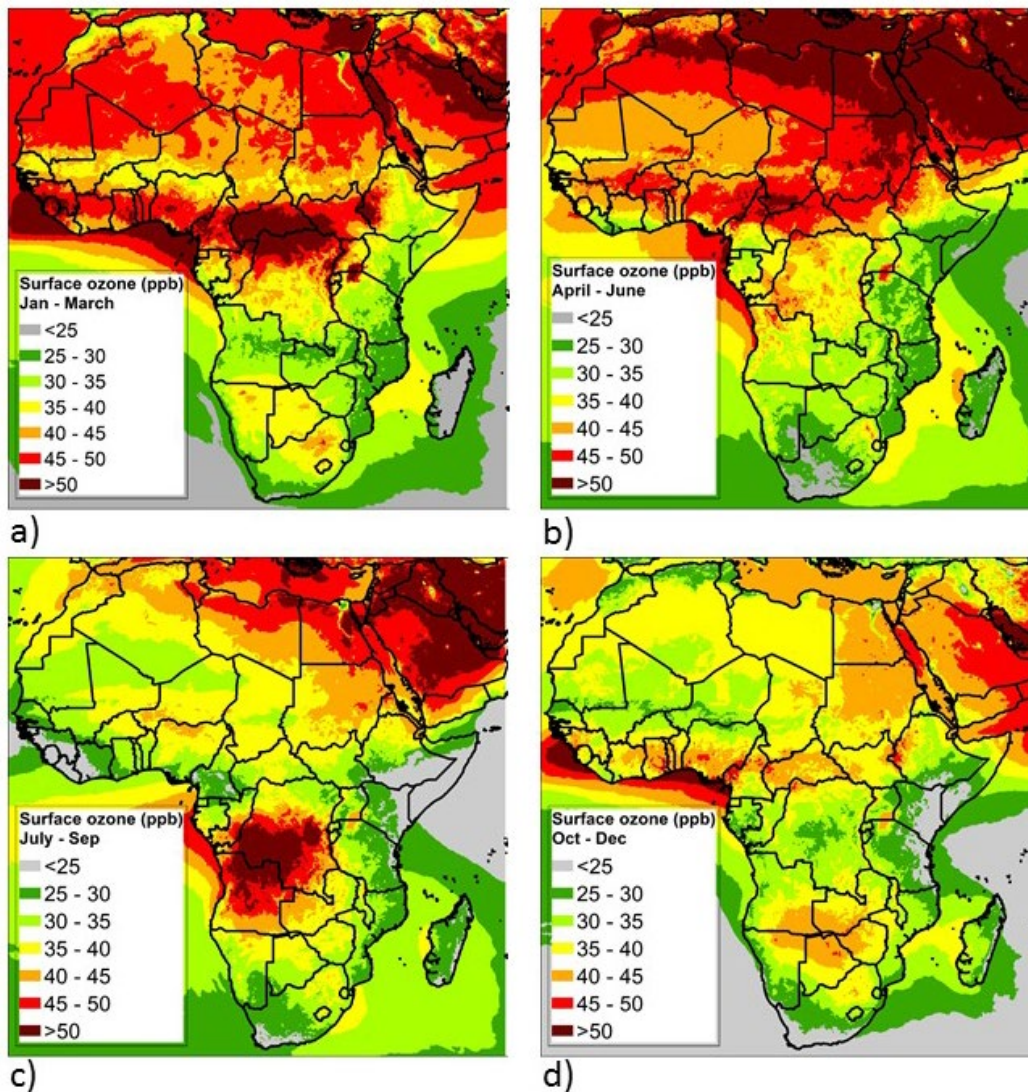
536 knowledge and understanding, future studies should test the impact of ozone on a wider variety of
537 tropical and temperate crop species, under realistic growing conditions if possible. As different
538 cultivars of the same crop can show varying levels of ozone sensitivity, (demonstrated for African
539 crops by Hayes *et al.* 2019, 2020), it would be beneficial to carry out screening of crop cultivars,
540 including further commonly used cultivars of African wheat and common bean and other important
541 SSA crop types for relative ozone sensitivity. This variation in response to ozone also highlights the
542 potential scope for breeding ozone tolerant cultivars. Mills *et al.* (2018c) present an ideotype for an
543 ozone tolerant crop but highlight the need to balance ozone tolerance with other favourable crop
544 characteristics, such as high yield and disease resistance. It would therefore be useful to aim to
545 include ozone as a factor in future regional SSA crop breeding programs. Lastly, measurements of
546 ozone (for example, using passive samplers) from areas of SSA where models predict a high risk of
547 crop production losses would be beneficial to validate model outputs and provide supporting
548 evidence for policy makers.

549 **5. Conclusions**

550 Overall, while the methodology does contain some sources of variation, this study is the first to
551 quantify the impact of ozone on two important SSA crops using finer resolution data than previous
552 global studies and species-specific flux-effect relationships. The results demonstrate that efforts to
553 reduce ozone precursors could contribute to reducing the yield gap in SSA, which would be
554 particularly useful as ozone stress often occurs in areas where other stresses (such as pests &
555 diseases or heat stress) are also a problem (Mills *et al.*, 2018c). More stringent air pollution
556 abatement policies are required to achieve this in the future. In the shorter term, approaches based
557 on crop management (e.g. irrigation, sowing date) and breeding ozone tolerant cultivars could
558 provide potential solutions for reducing ozone impact on yield.

559

560 **Figures**



562 **Fig. 1** EMEP-WRF Africa model output for surface ozone (ppb) daily mean (6am – 6am), averaged
563 across 3 month periods for 2015; a) Jan – March, max = 68.7ppb; b) April – June, max = 66.9ppb; c)
564 July – September, max = 84.2 ppb; d) October – December, max = 62.4ppb.

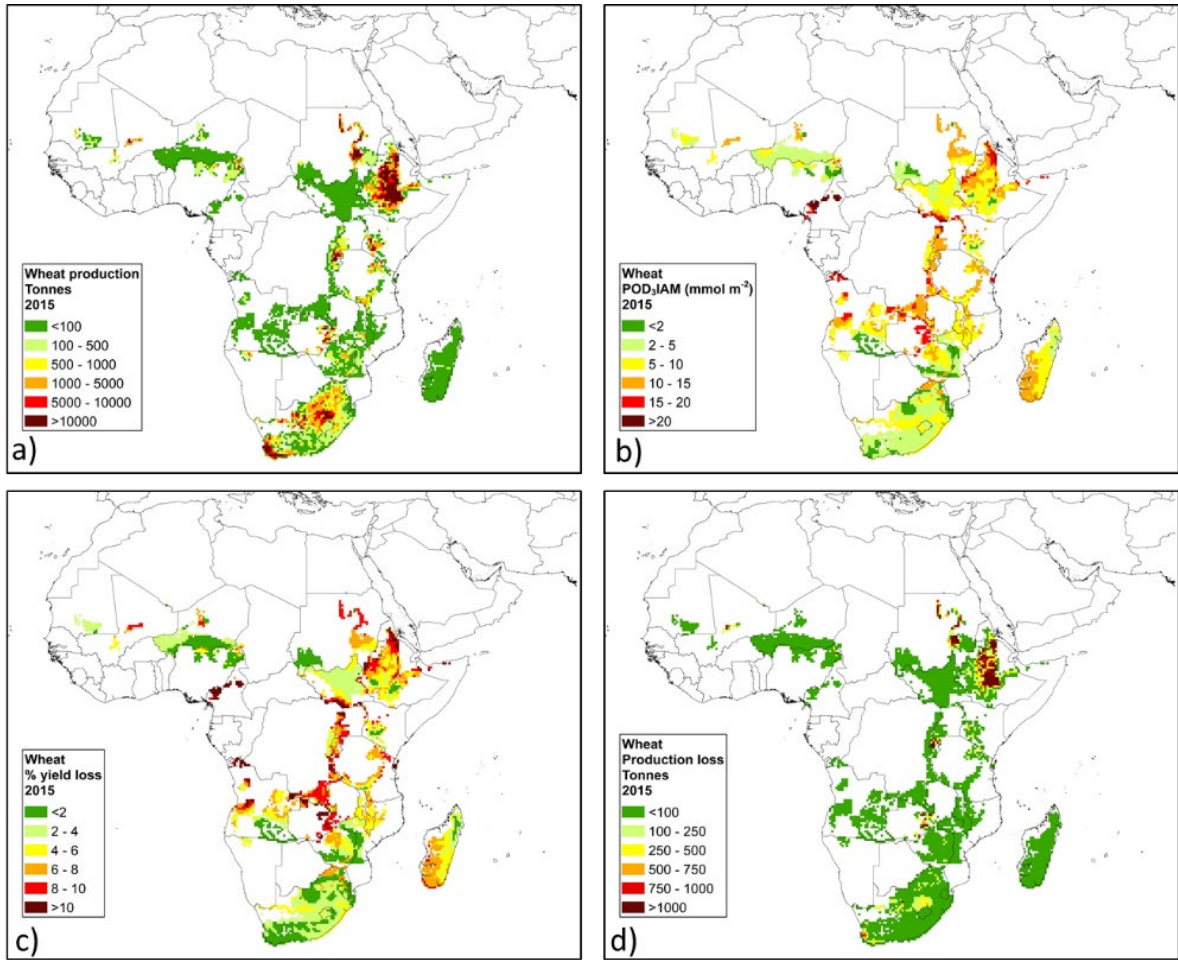
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571 **Fig. 2** Ozone impact on wheat production in Sub-Saharan Africa for the year 2015. a) Wheat
 572 production for 2015, using spatial data from the SPAM dataset for 2010 and FAO national conversion
 573 factors for 2015; b) POD₃IAM (mmol m⁻²) for wheat, accumulated over a 90-day period during the
 574 growing season; c) Percentage yield loss for wheat due to ozone; d) Production loss (tonnes) in 2015
 575 for wheat due to ozone.

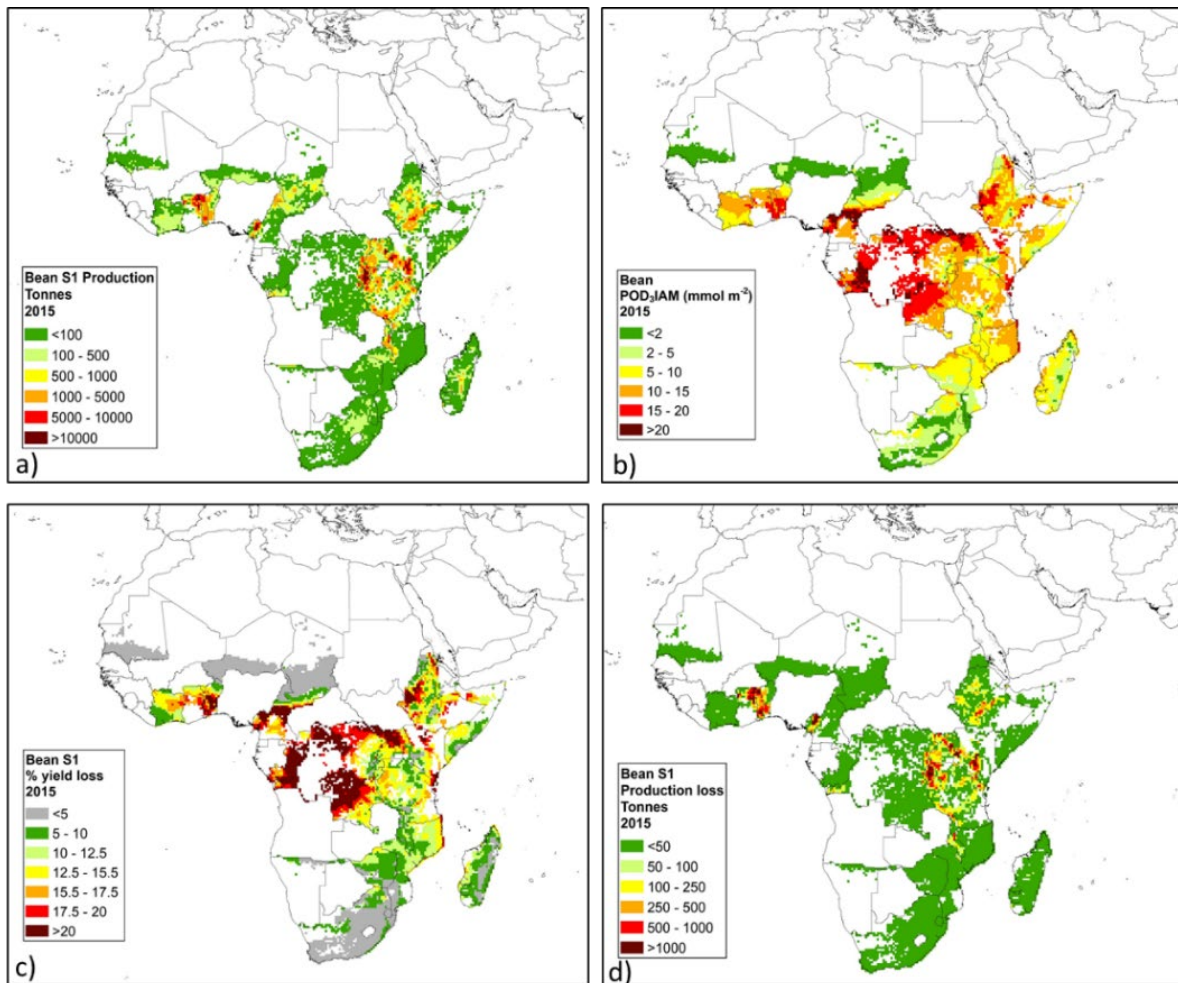
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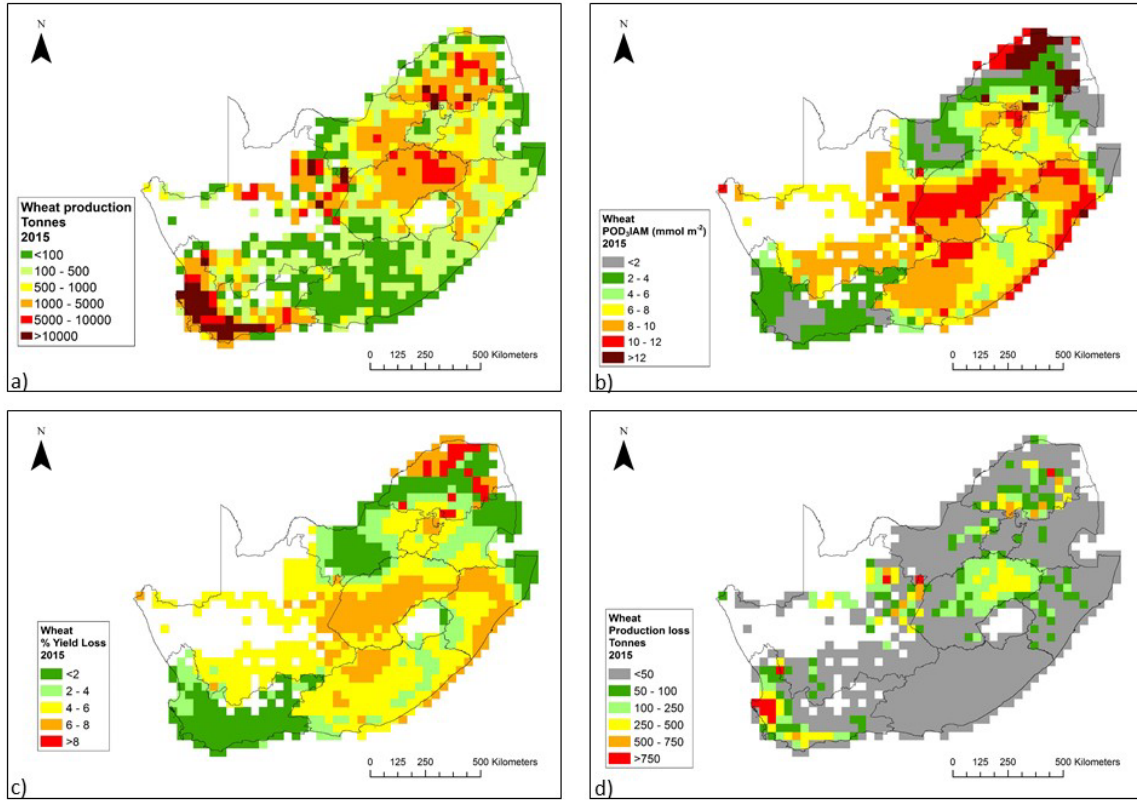
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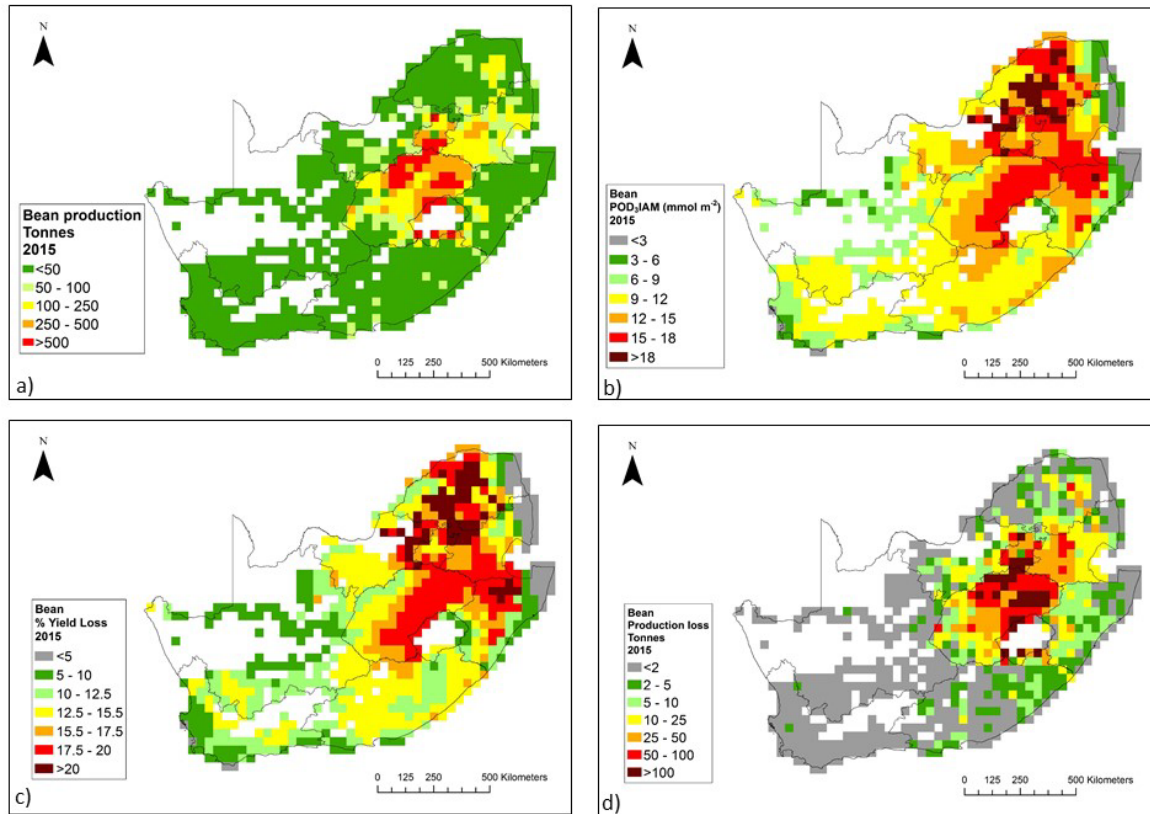
582 **Fig. 3** Ozone impact on bean production in Sub-Saharan Africa for the year 2015 (season 1).
 583 a) Bean production for 2015, using spatial data from the SPAM dataset for 2010 and FAO national
 584 conversion factors for 2015; b) POD_3IAM ($mmol\ m^{-2}$) for bean, accumulated over a 90-day period
 585 during the growing season; c) Percentage yield loss for bean due to ozone; d) Production loss
 586 (tonnes) in 2015 for beans due to ozone. For countries where beans were grown in two seasons,
 587 annual production and estimated production losses were divided by two.

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593 **Fig. 4** Ozone impact on wheat in South Africa (a) 2015 wheat production (using SPAM 2010 spatial
 594 data, converted per province using Government data for 2015; b) POD_3IAM ($mmol\ m^{-2}$); c)
 595 Percentage yield loss due to ozone and d) Wheat production loss in 2015 due to ozone.

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605 **Fig. 5** Ozone impact on beans in South Africa (a) 2015 production (using SPAM 2010 spatial data,
 606 converted per province using Government data for 2015; b) POD_3IAM (mmol m^{-2}); c) Percentage
 607 yield loss due to ozone and d) Bean production loss in 2015 due to ozone.

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616 **References**

- 617 Ainsworth, E.A., Yendrek, C.R., Sitch, S., Collins, W.J. & Emberson, L.D. (2012) The effects of
618 tropospheric ozone on net primary productivity and implications for climate change. *Annual review*
619 *of plant biology*, 63, 637-661.
- 620 ARC-Small Grain (2019) *Guideline on the production of small grains in the summer rainfall area*. ARC-
621 Small Grain, University of the Free State, South Africa.
- 622 Avnery, S., Mauzerall, D.L., Liu, J. & Horowitz, L.W. (2011) Global crop yield reductions due to surface
623 ozone exposure: 1. Year 2000 crop production losses and economic damage. *Atmospheric*
624 *Environment*, 45, 2284– 2296.
- 625 Balashov, N.V., Thompson, A.M., Piketh, S.J. & Langerman, K.E. (2014) Surface ozone variability and
626 trends over the South African Highveld from 1990 to 2007. *Journal of Geophysical Research:*
627 *Atmospheres*, 119, 4323-4342.
- 628 Bouarar, I., Law, K.S., Pham, M., Liousse, C., Schlager, H., Hamburger, T., Reeves, C.E., Cammas, J.-P.,
629 Nédélec, P., Szopa, S., Ravegnani, F., Viciani, S., D'Amato, F., Ulanovsky, A. & Richter, A. (2011)
630 Emission sources contributing to tropospheric ozone over Equatorial Africa during the summer
631 monsoon. *Atmospheric Chemistry and Physics*, 11, 13395-13419.
- 632 CLRTAP (2017) Chapter 3 “Mapping critical levels for vegetation.” LRTAP Convention Modelling and
633 Mapping Manual. Retrieved from <http://icpvegetation.ceh.ac.uk/>
- 634 Cooper, O. & Derwent, D. (2013) Conceptual overview of hemispheric or intercontinental transport
635 of ozone and particulate matter. In *Hemispheric Transport of Air Pollution 2010* (Dentener, F.,
636 Keating, T. & Akimoto, H., eds). New York and Geneva: United Nations, pp. 1–24.

637 DeWitt, H.L., Gasore, J., Rupakheti, M., Potter, K.E., Prinn, R.G., Ndikubwimana, J.D., Nkusi, J. &
638 Safari, B. (2019) Seasonal and diurnal variability in O₃, black carbon, and CO measured at the Rwanda
639 Climate Observatory. *Atmospheric Chemistry and Physics*, 19, 2063-2078.

640 Department of Agriculture, Forestry and Fisheries, Republic of South Africa (2016) *Production*
641 *guideline for wheat*.

642 Directorate of Statistics and Economic Analysis (2018) *Abstract of Agricultural Statistics*. Department
643 of Agriculture, Forestry and Fisheries, Republic of South Africa.

644 Emberson, L.D., Ashmore, M.R., Cambridge, H.M., Simpson, D. & Tuovinen, J.P. (2000) Modelling
645 stomatal ozone flux across Europe. *Environmental Pollution*, 109, 403-413.

646 Emberson, L.D., Ashmore, M.R., Simpson, D., Tuovinen, J.P. & Cambridge, H.M. (2001) Modelling and
647 mapping ozone deposition in Europe. *Water, Air, and Soil Pollution*, 130, 577-582.

648 Emberson, L.D., Büker, P., Ashmore, M.R., Mills, G., Jackson, L.S., Agrawal, M., Atikuzzaman, M.D.,
649 Cinderby, S., Engardt, M., Jamir, C. & Kobayashi, K. (2009) A comparison of North American and
650 Asian exposure–response data for ozone effects on crop yields. *Atmospheric Environment*, 43, 1945-
651 1953.

652 Emberson, L.D., Pleijel, H., Ainsworth, E.A., Van den Berg, M., Ren, W., Osborne, S., Mills, G., Pandey,
653 D., Dentener, F., Büker, P. & Ewert, F. (2018) Ozone effects on crops and consideration in crop
654 models. *European Journal of Agronomy*, 100, 19-34.

655 FAOSTAT (2017) Food and agriculture Statistical Databases (United Nations). Retrieved from
656 <http://faostat3.fao.org/home/E>

657 Harmens, H., Hayes, F., Sharps, K., Radbourne, A. & Mills, G. (2019) Can reduced irrigation mitigate
658 ozone impacts on an ozone-sensitive African wheat variety? *Plants*, 8(7), 220.

659 Hayes, F., Sharps, K., Harmens, H., Roberts, I. & Mills, G. (2019) Tropospheric ozone pollution
660 reduces the yield of African crops. *Journal of Agronomy and Crop Science*, 206, 214 – 228.

661 Hayes, F., Harmens, H., Sharps, K. & Radbourne, A. (2020) Ozone dose-response relationships for
662 tropical crops reveal potential threat to legume and wheat production, but not to millets. *Scientific
663 African* (in press). <https://doi.org/10.1016/j.sciaf.2020.e00482>

664 Huang, Y., Hickman, J.E. & Wu, S. (2018) Impacts of enhanced fertilizer applications on tropospheric
665 ozone and crop damage over sub-Saharan Africa. *Atmospheric Environment*, 180, 117-125.

666 Jarvis, P.G. (1976) The interpretation of the variations in leaf water potential and stomatal
667 conductance found in canopies in the field. *Philosophical Transactions of the Royal Society, London
668 B*, 273, 593-610.

669 Laban, T.L., Van Zyl, P.G., Beukes, J.P., Vakkari, V., Jaars, K., Borduas-Dedekind, N., Josipovic, M.,
670 Thompson, A.M., Kulmala, M. & Laakso, L. (2018) Seasonal influences on surface ozone variability in
671 continental South Africa and implications for air quality. *Atmospheric Chemistry and Physics*, 18,
672 15491–15514.

673 Laban, T.L., Van Zyl, P.G., Beukes, J.P., Mikkonen, S., Santana, L., Josipovic, M., Vakkari, V.,
674 Thompson, A.M., Kulmala, M. & Laakso, L. (2020) Statistical analysis of factors driving surface ozone
675 variability over continental South Africa. *Journal of Integrative Environmental Sciences*, 1-28.

676 Lourens, A.S., Beukes, J.P., Van Zyl, P.G., Fourie, G.D., Burger, J.W., Pienaar, J.J., Read, C.E. & Jordaan,
677 J.H. (2011) Spatial and temporal assessment of gaseous pollutants in the Highveld of South Africa.
678 *South African Journal of Science*, 107, 1-8.

679 Lourens, A.S., Butler, T.M., Beukes, J.P., Van Zyl, P.G., Beirle, S., Wagner, T.K., Heue, K.P., Pienaar, J.J.,
680 Fourie, G.D. & Lawrence, M.G. (2012) Re-evaluating the NO₂ hotspot over the South African
681 Highveld. *South African Journal of Science*, 108, 83-91.

682 Mills, G., Hayes, F., Simpson, D., Emberson, L., Norris, D., Harmens, H. & Büker, P. (2011) Evidence of
683 widespread effects of ozone on crops and (semi-)natural vegetation in Europe (1990–2006) in
684 relation to AOT40- and flux-based risk maps. *Global Change Biology*, 17, 592-613.

685 Mills, G., Pleijel, H., Malley, C. S., Sinha, B., Cooper, O., Schultz, M. G., ... Xu, X. (2018a) Tropospheric
686 Ozone Assessment Report: Present day tropospheric ozone distribution and trends relevant to
687 vegetation. *Elementa: Science of the Anthropocene*, 6: 47. <https://doi.org/10.1525/elementa.302>

688 Mills, G., Sharps, K., Simpson, D., Pleijel, H., Broberg, M., Uddling, J., Jaramillo, F., Davies, W.J.,
689 Dentener, F., Van den Berg, M. & Agrawal, M. (2018b) Ozone pollution will compromise efforts to
690 increase global wheat production. *Global change biology*, 24, 3560-3574.

691 Mills, G., Sharps, K., Simpson, D., Pleijel, H., Frei, M., Burkey, K., Emberson, L., Uddling, J., Broberg,
692 M., Feng, Z. & Kobayashi, K. (2018c) Closing the global ozone yield gap: Quantification and
693 cobenefits for multistress tolerance. *Global Change Biology*, 24, 4869-4893.

694 Monks, P.S.; Archibald, A.T.; Colette, A.; Cooper, O.; Coyle, M.; Derwent, R.; Fowler, D.; Granier, C.;
695 Law, K.S.; Mills, G.E.,... Williams, M.L. (2015) Tropospheric ozone and its precursors from the urban
696 to the global scale from air quality to short-lived climate forcer. *Atmospheric Chemistry and Physics*,
697 15, 8889–8973.

698 National Centers for Environmental Prediction/National Weather Service/NOAA/U.S. Department of
699 Commerce (2015). Data from year 2000, updated daily. NCEP FNL Operational Model Global
700 Tropospheric Analyses, continuing from July 1999. Research Data Archive at the National Center for
701 Atmospheric Research, Computational and Information Systems Laboratory.
702 <https://doi.org/10.5065/D6M043C6>. Accessed 01 Feb 2017.

703 Osborne, S.A., Mills, G., Hayes, F., Ainsworth, E.A., Büker, P. & Emberson, L. (2016) Has the sensitivity
704 of soybean cultivars to ozone pollution increased with time? An analysis of published dose–response
705 data. *Global change biology*, 22, 3097-3111.

706 Pinheiro, J., Bates, D., DebRoy, S. & Sarkar, D. (2018) R Core Team nlme: Linear and Nonlinear Mixed
707 Effects Models R package version 3.1-137. <https://CRAN.R-project.org/package=nlme>

708 Pleijel, H., Danielsson, H., Emberson, L., Ashmore, M.R., & Mills, G. (2007) Ozone risk assessment for
709 agricultural crops in Europe: further development of stomatal flux and flux–response relationships
710 for European wheat and potato. *Atmospheric Environment*, 41, 3022-3040.

711 Rippke, U., Ramirez-Villegas, J., Jarvis, A., Vermeulen, S.J., Parker, L., Mer, F., Diekkrüger, B.,
712 Challinor, A.J. & Howden, M. (2016) Timescales of transformational climate change adaptation in
713 sub-Saharan African agriculture. *Nature Climate Change*, 6, 605-609.

714 Saitanis, C.J., Sicard, P., De Marco, A., Feng, Z., Paoletti, E. & Agathokleous, E. (2020) On the
715 atmospheric ozone monitoring methodologies. *Current Opinion in Environmental Science & Health*,
716 (in press). <https://doi.org/10.1016/j.coesh.2020.07.004>

717 Sauvage, B., Thouret, V., Cammas, J.-P. , Gheusi, F., Athier, G. & Nédélec, P. (2005) Tropospheric
718 ozone over Equatorial Africa: regional aspects from the MOZAIC data. *Atmospheric Chemistry and*
719 *Physics*, 5, 311-335.

720 Simpson, D., Benedictow, A., Berge, H., Bergström, R., Emberson, L. D., Fagerli, H., Flechard, C. R.,
721 Hayman, G. D., Gauss, M., Jonson, J. E., Jenkin, M. E., Nyíri, A., Richter, C., Semeena, V. S., Tsyro, S.,
722 Tuovinen, J. P., Valdebenito, Á. & Wind, P. (2012) The EMEP MSC-W chemical transport model -
723 technical description. *Atmospheric Chemistry and Physics*, 12, 7825-7865.

724 Skamarock, W. C., Klemp, J.B., Dudhia, J., Gill, D.O., Liu, Z., Berner, J., Wang, W., Powers, J.G., Duda,
725 M.G., Barker, D.M. & Huang, X.-Y. (2019) A Description of the Advanced Research WRF Version
726 4. *NCAR Tech. Note NCAR/TN-556+STR*, 145 pp. [doi:10.5065/1dfh-6p97](https://doi.org/10.5065/1dfh-6p97)

727 Soja, G., Barnes, J. D., Posch, M., Vandermeiren, K., Pleijel, H., & Mills, G. (2000). Phenological
728 weighting of ozone exposures in the calculation of critical levels for wheat, bean and
729 plantain. *Environmental Pollution*, 109, 517– 524.

730 Stohl, A., Aamaas, B., Amann, M., Baker, L., Bellouin, N., Berntsen, T.K., Boucher, O., Cherian, R.,
731 Collins, W., Daskalakis, N. & Dusinska, M. (2015) Evaluating the climate and air quality impacts of
732 short-lived pollutants. *Atmospheric Chemistry and Physics*, 15, 10529-10566.

733 Tetteh, R., Yamaguchi, M., Wada, Y., Funada, R. & Izuta, T. (2015) Effects of ozone on growth, net
734 photosynthesis and yield of two African varieties of *Vigna unguiculata*. *Environmental Pollution*, 196,
735 230-238.

736 Turnock, S.T., Wild, O., Dentener, F.J., Davila, Y., Emmons, L.K., Flemming, J., Folberth, G.A., Henze,
737 D.K., Jonson, J.E., Keating, T.J., Kengo, S., Lin, M., Lund, M., Tilmes, S. & O'Connor, F. M. (2018) The
738 impact of future emission policies on tropospheric ozone using a parameterised approach.
739 *Atmospheric Chemistry and Physics*, 18, 8953–8978.

740 Van Dingenen, R., Raes, F., Krol, M.C., Emberson, L. & Cofala, J. (2009) The global impact of O₃ on
741 agricultural crop yields under current and future air quality legislation. *Atmospheric Environment*, 43,
742 604– 618.

743 Van Ittersum, M.K., Van Bussel, L.G., Wolf, J., Grassini, P., Van Wart, J., Guilpart, N., Claessens, L., de
744 Groot, H., Wiebe, K., Mason-D’Croz, D. & Yang, H. (2016) Can sub-Saharan Africa feed itself?
745 *Proceedings of the National Academy of Sciences*, 113, 14964-14969.

746 Vieno, M., Heal, M. R., Williams, M. L., Carnell, E. J., Nemitz, E., Stedman, J. R. & Reis, S. (2016) The
747 sensitivities of emissions reductions for the mitigation of UK PM_{2.5}. *Atmospheric Chemistry and*
748 *Physics*, 16, 265-276.

749 You, L., Wood, S., Wood-Sichra U. & Wu, W. (2014) Generating global crop distribution maps: From
750 census to grid. *Agricultural Systems*, 127, 53–60.

751 Ziemke, J.R., Oman, L.D., Strode, S.A., Douglass, A.R., Olsen, M.A., McPeters, R.D., Bhartia, P.K.,
752 Froidevaux, L., Labow, G.J., Witte, J.C. & Thompson, A.M. (2019) Trends in global tropospheric ozone

- 753 inferred from a composite record of TOMS/OMI/MLS/OMPS satellite measurements and the
- 754 MERRA-2 GMI simulation. *Atmospheric Chemistry and Physics*, 19, 3257-3269.
- 755 Zuur, A.F., Ieno, E.N., Walker, N.J., Saveliev, A.A. & Smith, G.M. (2009) *Mixed effects models and*
- 756 *extensions in ecology with R*. Springer, New York.

757 **Declarations**

758 Ethics approval and consent to participate:

759 Not applicable.

760 Consent for publication:

761 Not applicable.

762 Availability of data and materials:

763 The datasets used and/or analysed during the current study are available from the corresponding
764 author on reasonable request. Due to requirements of the NERC funding body, if requested, data
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767

768 Competing interests:

769 The authors declare that they have no competing interests.

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783 KS, FH and HH contributed to the study conception and design. Material preparation, data collection
784 and analysis were performed by KS and FH. MM and RB ran the EMEP-WRF model, performed
785 quality assurance/control and provided ozone flux data for further analysis/mapping by KS. The first
786 draft of the manuscript was written by KS and all authors commented on previous versions of the
787 manuscript. All authors read and approved the final manuscript.