



**Cite this article:** Trevail AM *et al.* 2019 Environmental heterogeneity decreases reproductive success via effects on foraging behaviour. *Proc. R. Soc. B* **286**: 20190795. <http://dx.doi.org/10.1098/rspb.2019.0795>

Received: 8 April 2019

Accepted: 16 May 2019

**Subject Category:**

Ecology

**Subject Areas:**

behaviour, ecology, environmental science

**Keywords:**

competition, seabird, hidden Markov model, heterogeneity gradient, optimal foraging theory, resource availability

**Author for correspondence:**

Alice M. Trevail

e-mail: [alice.trevail@liv.ac.uk](mailto:alice.trevail@liv.ac.uk)

Electronic supplementary material is available online at <https://doi.org/10.6084/m9.figshare.c.4514741>.

# Environmental heterogeneity decreases reproductive success via effects on foraging behaviour

Alice M. Trevail<sup>1</sup>, Jonathan A. Green<sup>1</sup>, Jonathan Sharples<sup>1</sup>, Jeff A. Polton<sup>2</sup>, Peter I. Miller<sup>3</sup>, Francis Daunt<sup>4</sup>, Ellie Owen<sup>5</sup>, Mark Bolton<sup>6</sup>, Kendrew Colhoun<sup>7,8</sup>, Stephen Newton<sup>9</sup>, Gail Robertson<sup>10</sup> and Samantha C. Patrick<sup>1</sup>

<sup>1</sup>School of Environmental Sciences, University of Liverpool, Liverpool, UK

<sup>2</sup>National Oceanography Centre, Liverpool, UK

<sup>3</sup>Remote Sensing Group, Plymouth Marine Laboratory, Plymouth, UK

<sup>4</sup>Centre for Ecology and Hydrology Edinburgh, Bush Estate, Penicuik, Midlothian, UK

<sup>5</sup>RSPB Centre for Conservation Science, RSPB Scotland, Etive House, Beechwood Park, Inverness, UK

<sup>6</sup>RSPB Centre for Conservation Science, The Lodge, Sandy, Bedfordshire, UK

<sup>7</sup>RSPB Centre for Conservation Science, Belfast, UK

<sup>8</sup>School of Agriculture and Food Science, University College Dublin, Belfield, Dublin 4, Ireland

<sup>9</sup>BirdWatch Ireland, Kilcoole, Wicklow, Ireland

<sup>10</sup>School of Mathematics, University of Edinburgh, Edinburgh, UK

AMT, 0000-0002-6459-5213

Environmental heterogeneity shapes the uneven distribution of resources available to foragers, and is ubiquitous in nature. Optimal foraging theory predicts that an animal's ability to exploit resource patches is key to foraging success. However, the potential fitness costs and benefits of foraging in a heterogeneous environment are difficult to measure empirically. Heterogeneity may provide higher-quality foraging opportunities, or alternatively could increase the cost of resource acquisition because of reduced patch density or increased competition. Here, we study the influence of physical environmental heterogeneity on behaviour and reproductive success of black-legged kittiwakes, *Rissa tridactyla*. From GPS tracking data at 15 colonies throughout their British and Irish range, we found that environments that were physically more heterogeneous were associated with longer trip duration, more time spent foraging while away from the colony, increased overlap of foraging areas between individuals and lower breeding success. These results suggest that there is greater competition between individuals for finite resources in more heterogeneous environments, which comes at a cost to reproduction. Resource hotspots are often considered beneficial, as individuals can learn to exploit them if sufficiently predictable. However, we demonstrate here that such fitness gains can be countered by greater competition in more heterogeneous environments.

## 1. Introduction

The spatial and temporal distribution of resources places a major constraint on foraging success [1–3]. Therefore, heterogeneity in resource distribution, which is considered a universal feature of natural environments [4,5], has played a defining role in the evolution of animal foraging behaviour [1,3]. Theory predicts that key to an individual's success is the ability to maximize gains from areas with high resource density and minimize energy expenditure locating resources, and therefore optimize energy allocation to fitness [2,3]. This theory is supported by numerous empirical studies (e.g. [6–8]). In response to resource heterogeneity, selection will therefore favour efficient foraging




behaviour, whereby individuals minimize the energetic costs of searching and transiting between high-resource locations and maximize resource intake [1,3].

However, not all heterogeneous environments are equal [9–11], as high-prey locations vary in distribution, predictability and numbers of competing individuals. Studies often present these ‘prey hotspots’ as beneficial resource patches [12–14]; however, the optimality of foraging strategies in response to resource heterogeneity may be constrained by both the nature of resource heterogeneity [15,16] and the behaviour of other foragers [17,18]. First, the travel distance to reach foraging patches in heterogeneous environments will determine the trade-off between resource intake and the additional energetic costs to the animal’s own fitness [16,19,20]. Second, higher levels of intraspecific competition at resource patches in heterogeneous environments may also limit resource acquisition from a patch [17,18,21,22] through competitive exclusion [18,22] and prey disturbance [17] and depletion [23]. The key knowledge gap is whether greater environmental heterogeneity has positive or negative consequences for fitness.

Underlying variability in the physical environment is a strong driver of heterogeneous resource distributions, and therefore can be used as a proxy for resource heterogeneity, particularly where resource availability to foragers is difficult to measure directly. Indeed, because of effects on resources, physical environmental heterogeneity, hereafter ‘environmental heterogeneity’, is known to be an important driver of community dynamics [11,24] and life-history strategies [25,26]. Marine environments provide a model study system of environmental heterogeneity, with numerous physical features (such as fronts, eddies and currents) that together define resource availability to foragers [12,27]. Furthermore, the degree to which any given marine environment is heterogeneous can vary [9], and therefore offers the opportunity to study the influence of heterogeneity on behaviour and fitness.

In this study, we test the influence of environmental heterogeneity on behaviour and reproductive success using data from black-legged kittiwakes (*Rissa tridactyla*, hereafter ‘kittiwakes’) at 15 colonies across their UK and Irish breeding range. Studying such a comprehensive dataset is ideal to understand how environmental heterogeneity affects behaviour and fitness. As with many seabirds, kittiwakes are central place foragers during the breeding season, and are therefore constrained to forage within their local environment. As such, greater travel distances away from the breeding location are considered indicative of poorer resource availability nearby [28,29]. Furthermore, as surface feeders, kittiwakes are thought to suffer from direct competition with conspecifics for prey as fish schools are forced lower down in the water column to inaccessible depths [30,31]. We first calculate a measure of local environmental heterogeneity at each colony based on six environmental metrics that can all influence kittiwake prey distributions. Second, we consider kittiwake foraging behaviour along the gradient of environmental heterogeneity between study colonies, and then test the link between the degree of environmental heterogeneity and reproductive success. Our analyses tested the following alternative hypotheses (see table 1) based on the literature reviewed above. (H1) Foraging opportunity hypothesis: greater environmental heterogeneity is associated with higher fitness because it features greater amounts of

**Table 1.** Hypotheses of the potential effects of greater environmental heterogeneity on resources, foraging dynamics and fitness at the population level. Upwards and downwards arrows indicate an expected increase and decrease, respectively, and crosses indicate no expected change.

hypothesis	potential effect of greater environmental heterogeneity on resources	expected change in population foraging dynamics	expected change in foraging metrics			competition (overlap between individuals)	expected change in fitness
			travel distance (maximum, total and proportion of time transiting)	time foraging (trip duration and proportion of time foraging)			
(H1) foraging opportunity	resource patches present higher quality foraging opportunities	 smaller foraging range	↓	↓	X	↑	
(H2) reduced patch density	resource patches offering sufficient foraging opportunities are further apart	 longer foraging distance	↑	↑ or X	X or ↓	↓	
(H3) competition	resources concentrated into smaller patches	 increased competition	X	↑	↑	↓	

profitable habitat within the foraging range of the colony that animals can learn to exploit, which enables individuals to remain closer to the colony [16], provision offspring more frequently [32] and relieve partners of nest-attendance duties [33]. (H2) Reduced patch density hypothesis: greater environmental heterogeneity is detrimental to fitness because habitat patches with sufficient resources to support foraging are located further apart within the foraging range. This, therefore, would prompt individuals to more readily switch between patches [2], requiring an increase in travel distance away from the colony, time spent commuting and foraging area size [16], and resulting in greater expenditure to transit between patches. (H3) Competition hypothesis: greater environmental heterogeneity is detrimental to fitness because it increases competition between individuals at relatively profitable habitats, which results in greater overlap between individuals, greater time investment in foraging behaviour and increased duration of foraging trips [17].

## 2. Methods

### (a) Quantifying environmental heterogeneity

To quantify environmental heterogeneity, we used a multivariate dispersion analysis [34] to identify the dissimilarity of spatial environmental characteristics. Multivariate dispersion analyses have primarily been used for species diversity studies [34]; however, they have also been used to quantify environmental heterogeneity using multiple continuous variables in studies of freshwater ecosystems [35,36], marine ecosystems [34] and grasslands [37]. Multivariate dispersion analysis is suitable for this study because it incorporates variance in multiple environmental parameters that can all influence resource distribution into a single metric, in contrast to measures such as standard deviation or range of a single continuous variable [10,38], or the diversity of categorical habitat variables [39].

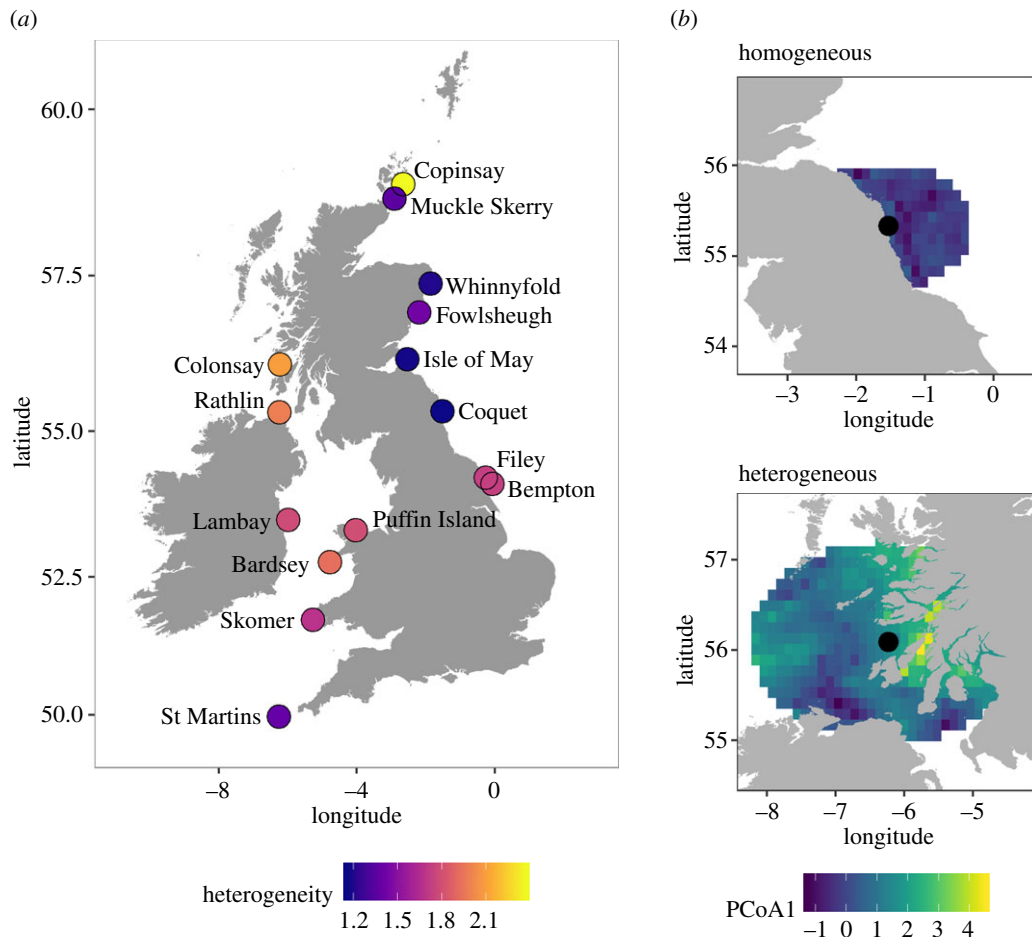
We calculated environmental heterogeneity using six environmental metrics: (1) bathymetry, (2) potential tidal stratification, (3) sea surface temperature, and ocean front (4) strength, (5) distance and (6) persistence, all of which have been shown to influence resource location for foraging seabirds. (1) Bathymetry, or sea floor depth, can shape the flow of horizontal water currents and control vertical water column structure [40,41], both of which are physical processes that can influence the availability and accessibility of prey fish to surface foragers such as kittiwakes [27,40,41]. (2) Potential tidal stratification incorporates both depth and tidal currents [12], to quantify the vertical water column structure—a key physical driver of marine ecosystem dynamics [42], prey fish distribution [43] and seabird distribution [27]. (3) Sea surface temperature can be a proxy for oceanographic processes that influence nutrient availability, such as upwelling of cold nutrient-rich water [44], and has been linked to the at-sea distribution and breeding success of kittiwakes [27,42,45,46]. (4–6) Ocean fronts are horizontal boundaries between different water masses where physical processes cause upwelling of deeper, nutrient-rich water and entrain plankton at the surface [47,48]. Fronts are known to be an important feature of marine environments, shaping resource distribution and thus marine vertebrate behaviour [43,49]. Full details of data sources are described in electronic supplementary material, appendix A.

We used a principal coordinate analysis (a type of multivariate dispersion analysis) [50] to determine the heterogeneity of environmental conditions at each colony and year (hereafter 'colony-year') from within the maximum foraging range of kittiwakes. We used the overall maximum foraging range across all

years as a measure of the environment available to each colony (electronic supplementary material, appendix B). Principal coordinate analyses place values from all colonies along all axes (or principal coordinates) in unconstrained ordination space based on a Euclidean distance matrix of standardized environmental data, using the functions *vegdist* and *betadisper* in the R package *vegan* [51]. Herein, we use the average distance of observations from the colony-year centroid (or spatial median) in the principal coordinate analysis ordination space (using all axes) as a continuous measure of environmental heterogeneity, with higher values indicating greater heterogeneity. As such, environmental heterogeneity can vary independently of the absolute values of the six environmental variables. Permutation tests of dispersion (PERMDISP [34]) calculate an *F*-statistic to compare the average distances of observations from the colony-year centroid between each colony-year in the analysis to test for differences in heterogeneity. We used a two-way ANOVA to test whether environmental heterogeneity differed between colonies and between years (as factors), and Tukey HSD *post hoc* tests for pairwise differences. To understand whether environmental heterogeneity was simply associated with availability of a particular habitat type or was a proxy of overall prey abundance within the foraging range of kittiwakes (maximum foraging distance across years at each colony from tracking data; H1 and H2, table 1), we used linear regression to test whether environmental heterogeneity was linked to the mean value of any of the individual environmental metrics. To determine whether environmental heterogeneity was influenced by the size of the foraging radius used to extract environmental data (maximum foraging distance across years at each colony), we compared environmental heterogeneity values with the maximum foraging range of kittiwakes at each colony across all years using linear regression.

### (b) Quantifying kittiwake foraging behaviour

To determine the foraging behaviour of kittiwakes around the UK, adults from multiple colonies were tracked using GPS loggers (Mobile Action i-GotU GT-120), while raising small chicks. Tracked individuals were selected randomly with respect to brood size and were assumed to be representative of each study population. Loggers were attached to the back feathers between the wings (or infrequently to the tail) using waterproof tape, and total instrument mass was less than or equal to 5% of body mass (or less than or equal to 3% where tail attachments were used; mean  $\pm$  s.e. body mass at Skomer, Rathlin and Puffin Island:  $327.9 \pm 5.1$  from Trevail *et al.* [9]). Full details of tracking procedures can be found in the first publications of the data: Wakefield *et al.* [27] and Trevail *et al.* [9]. Here, we use data from a total of 1567 trips from 415 chick-rearing kittiwakes at 15 colonies in Britain and Ireland between 2010 and 2017 (figure 1): Bardsey (NW Wales; 2011,  $n = 8$ ), Bempton Cliffs (E England; 2010–2013 and 2015,  $n = 59$ ), Copinsay (Orkney Islands; 2010–2012,  $n = 26$ ), Coquet (NE England; 2011–2012,  $n = 26$ ), Colonsay (W Scotland; 2010–2014,  $n = 69$ ), Filey (E England; 2013 and 2015,  $n = 26$ ), Fowlshough (E Scotland; 2012,  $n = 13$ ), Isle of May (E Scotland; 2013,  $n = 16$ ), Lambay (E Ireland; 2010,  $n = 10$ ), Muckle Skerry (Orkney Islands; 2012–2014,  $n = 26$ ), Puffin Island (NW Wales; 2010–2016,  $n = 63$ ), Rathlin (Northern Ireland; 2017,  $n = 17$ ), Skomer (SW Wales; 2016–2017,  $n = 14$ ), St Martins (Isles of Scilly; 2010–2011,  $n = 28$ ) and Whinnyfold (E Scotland; 2012,  $n = 14$ ). Full sample sizes, including colony coordinates, tracking dates and number of individuals per year are given in electronic supplementary material, table B1. For further analyses, we excluded points closer than 500 m to the colony, and attributed sequential points to a foraging trip if the total trip duration was over 14 min [9] to eliminate departures from the colony due to disturbance [52]. At all colonies, we included trips



**Figure 1.** (a) Map of study kittiwake colonies, coloured by mean environmental heterogeneity and (b) environment within the foraging range of two example colonies according to the position along the first axis (PCoA1) from the principal coordinate analysis used to calculate environmental heterogeneity, here for 2015 as an example. Colony environmental heterogeneity is a single measure of variance calculated as the mean distance in Euclidian space (using all PCoA axes) of all locations from the colony centroid. At the homogeneous colony (Coquet, top), values are concentrated together along the first PCoA axis. At the heterogeneous colony (Colonsay, bottom), values range along the first PCoA axis. (Online version in colour.)

where individuals were away from the colony overnight. At Rathlin, Skomer and Puffin Island, loggers did not record data between 23.00 and 03.00, to save battery power overnight while kittiwakes exhibit minimal foraging activity [9,53]. At all other colonies, we excluded locations during this period.

To understand the influence of environmental heterogeneity on foraging behaviour, we calculated three different measures of behaviour, all predicted to vary with each hypothesis (table 1). First, for each year and at each colony, we calculated the following trip metrics: mean trip duration, mean total distance travelled during a foraging trip, and mean maximum distance from the colony, all important indicators of resource accessibility for central place foragers as they seek to remain close to the colony and minimize travel times [16,17,28]. Second, we examined movement behaviours while away from the colony using a hidden Markov model to classify behaviour into rest, forage (including searching) or transit [54]. Time spent in each behaviour can signal the energetic trade-off between travel costs and resource gains from exploiting prey patches [55]. We used the R package *moveHMM* [56] for behavioural classification based on distributions of step lengths and turning angles, after interpolating GPS data to regular time steps to fulfil HMM assumptions, using the R package *adehabitatLT* [57]. We used a gamma distribution to describe step lengths and a von Mises distribution to describe turning angles, and the Viterbi algorithm to estimate the most likely sequence of movement states based on the fitted hidden Markov model (electronic supplementary material, appendix C). We used values from the previous classification of

kittiwake behaviour to inform model starting parameters [9], and found that model outputs were robust to different values of starting parameters when tested on a subset of tracking data. For each bird, we quantified the proportion of time away from the colony while on a foraging trip spent in each behaviour classified by the HMM (forage, transit and rest). Third, we determined at-sea area use of kittiwakes by calculating the size of 50% core foraging areas of individuals from utilization kernels on a 1 km grid using the *kernelUD* function in the R package *adehabitatHR* [57]. The appropriate smoothing parameter ( $h$ ) was determined by the default ad hoc method, which assumes a bivariate normal distribution [57]. As a proxy for intra-specific competition, we calculated the overlap of 50% core foraging areas between all individuals tracked in the same year at each colony using Bhattacharya's affinity (BA). Values of BA range from 0 when there is no overlap between foraging areas to 1 when utilization distributions are identical [58].

### (c) Quantifying kittiwake reproductive success

To test the effect of environmental heterogeneity on kittiwake reproductive success, we used colony-average reproductive success data from the Seabird Monitoring Programme, collated by the UK Joint Nature Conservation Committee (JNCC; <http://jncc.defra.gov.uk/smp>) and the Centre for Ecology & Hydrology for the Isle of May [59]. Reproductive success data were available for 11 colonies, for 1–8 years between 2010 and 2017 (electronic supplementary material, table B3). Reproductive success was

calculated as the total number of chicks fledged divided by the number of nests/pairs monitored at each colony in each year (electronic supplementary material, table B4 and figure B1).

#### (d) Effect of environmental heterogeneity on kittiwake foraging behaviour and reproductive success

In all analyses described below, explanatory variables were standardized to a mean of 0 and standard deviation of 1. Model structure and effect significance were determined using ANOVA comparisons ( $\chi^2$  for linear regressions and GLMMs, and F tests for quasi-binomial), for which *p*-values are presented.

To understand the effects of environmental heterogeneity on foraging metrics and reproductive success, we used the mean environmental heterogeneity for each colony across all years because colony and year, by definition, explained a large proportion of the variation in environmental heterogeneity (electronic supplementary material, figure A3), and did not include colony or year as variables in regression analyses. We refer to this mean value as 'colony-mean environmental heterogeneity'. To understand the effect of environmental heterogeneity on foraging behaviour in relation to the hypotheses (table 1), we undertook the following statistical tests. First, we compared the colony-mean environmental heterogeneity to the annual mean of trip metrics for each colony (trip duration, total distance and maximum distance; log-transformed to meet the assumptions of Gaussian models) using linear regression. Second, we compared the colony-mean environmental heterogeneity to the proportion of time spent away from the colony in each behavioural state (forage, transit and rest) by each individual using linear regression with a quasi-binomial logit-link to account for overdispersion. Lastly, we compared the colony-mean environmental heterogeneity with the size of 50% core foraging area of each bird using linear regression, and overlap between trips of all pairs of individuals using a GLMM with the focal BirdID as a random effect and a Gaussian distribution. To understand the effect of environmental heterogeneity on reproductive success, we compared the colony-mean environmental heterogeneity with the annual reproductive success for each colony using linear regression.

To verify that observed patterns in foraging dynamics and resource success could be attributed to environmental heterogeneity, we tested for potentially confounding effects of colony size and individual environmental variables on reproductive success (electronic supplementary material, appendix D). We used data from the most recent census of UK breeding populations, Seabird 2000 [60], to compare breeding success with colony size and the number of breeding kittiwakes within the foraging radius of each colony using linear regression. Seabird 2000 data may no longer provide currently accurate estimates of breeding numbers; however, they offer the most useful indicator of relative colony size for the purpose of this study. In support of results presented below, we found no link between reproductive success and any environmental metric in isolation (bathymetry, stratification, sea surface temperature and ocean front metrics; electronic supplementary material, table D1), suggesting that heterogeneity in resource distribution is key in this system.

### 3. Results

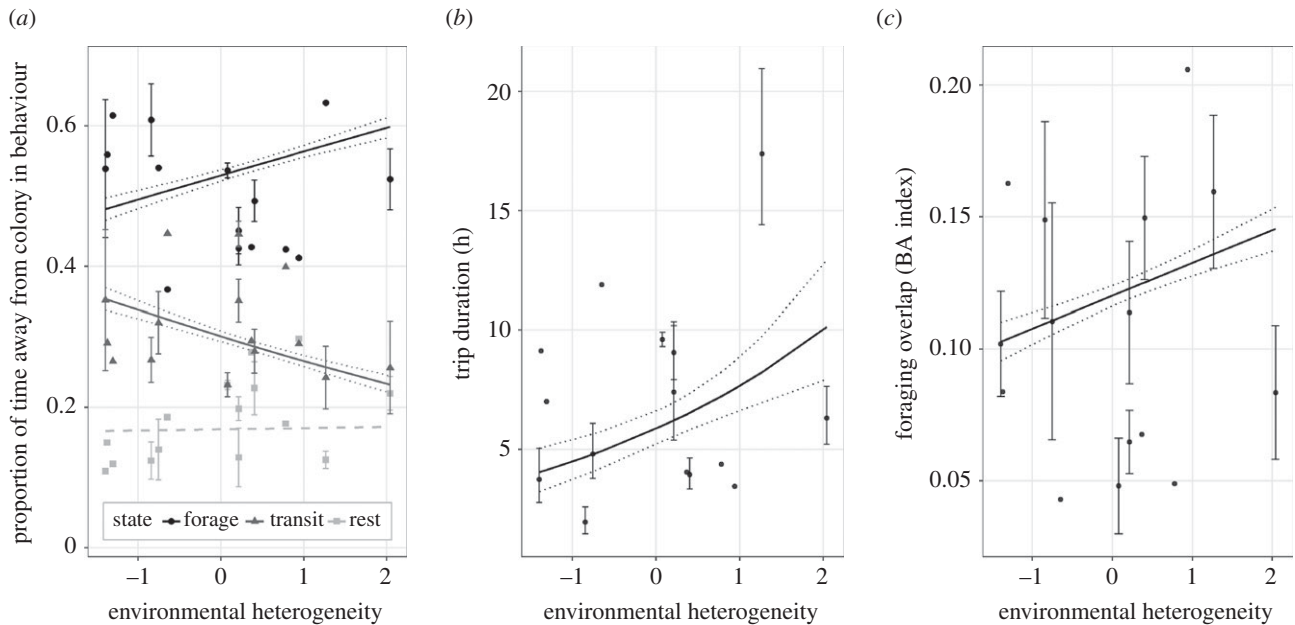
#### (a) Environmental heterogeneity at colonies

Environmental heterogeneity varied significantly between colony and year combinations in the principal coordinate analysis (figure 1;  $F_{119,17880} = 16.6$ ,  $p < 0.001$ ). The first two coordinate axes from the principal coordinate analysis together explained 63% of the total variation between

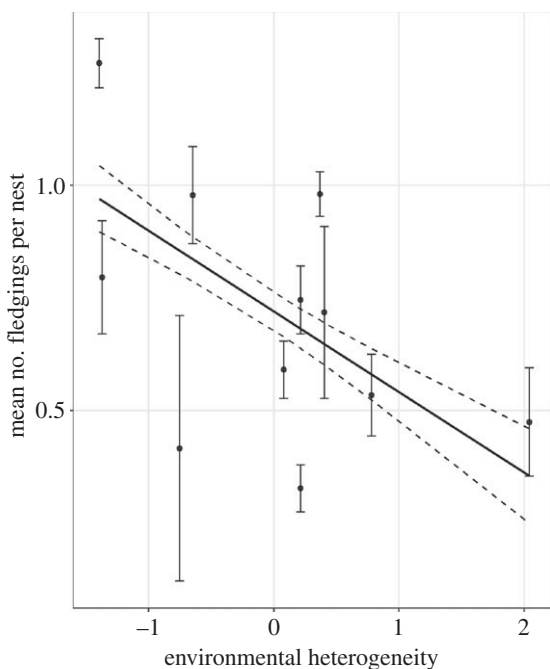
colonies (first axis: 43.1%, all others presented in electronic supplementary material, table A1). Environmental heterogeneity differed significantly between colonies (figure 1; ANOVA:  $F_{14,98} = 42.8$ ,  $p < 0.001$ ), and between years (ANOVA:  $F_{7,98} = 3.0$ ,  $p = 0.007$ ), although the effect of year was driven by a significant difference between 2011 and 2014 (electronic supplementary material, appendix A). Environmental heterogeneity was highest at Copinsay (mean  $\pm$  s.e. between years =  $2.37 \pm 0.05$ ), and was lowest at Coquet ( $1.16 \pm 0.06$ ), the Isle of May ( $1.17 \pm 0.04$ ) and Whinnyfold ( $1.19 \pm 0.07$ ). Values of environmental heterogeneity at each colony, and pairwise comparisons between colonies and years are given in electronic supplementary material, appendix A. Comparisons of environmental heterogeneity with individual environmental metrics showed no strong relationships (electronic supplementary material, appendix E). There was no link between environmental heterogeneity from the principal coordinate analysis and the size of the radius (maximum foraging distance from the colony across all years) used to select environmental data ( $F_{(1,118)} = 0.76$ ,  $p = 0.386$ ).

#### (b) Hypothesis testing: effect of environmental heterogeneity on kittiwake foraging behaviour and reproductive success

We found most support for the competition hypothesis (H3) that environmental heterogeneity was associated with greater competition between individuals, and consequently lower fitness. We found that trip duration (time spent away from the colony) was positively correlated with environmental heterogeneity (figure 2*b*; parameter estimate  $\pm$  s.e.:  $0.27 \pm 0.12$ ,  $F_{1,33} = 5.11$ ,  $p = 0.03$ ). Furthermore, the proportion of individuals' time spent foraging was significantly higher in more heterogeneous environments (figure 2*a*; parameter estimate  $\pm$  s.e.:  $0.14 \pm 0.03$ ,  $F_{1,415} = 18.8$ ,  $p < 0.01$ ), and environmental heterogeneity was positively correlated with overlap of the 50% core foraging area between individuals (figure 2*c*; parameter estimate  $\pm$  s.e.:  $0.012 \pm 0.004$ ,  $\chi^2_1 = 9.85$ ,  $p < 0.01$ ). Reproductive success was significantly lower in colonies with greater environmental heterogeneity (figure 3; parameter estimate  $\pm$  s.e. =  $-0.18 \pm 0.05$ ;  $F_{1,59} = 15.44$ ,  $p < 0.01$ ), equivalent to a 63% decrease in reproductive success across the observed range of environmental heterogeneity. This relationship is robust to removal of the apparent outlier of Copinsay. We did not find support for the foraging opportunity hypothesis (H1) that environmental heterogeneity was associated with greater amounts of profitable habitat: there was no link between environmental heterogeneity and the mean maximum distance kittiwakes travelled from the colony (table 2;  $F_{1,33} = 1.11$ ,  $p = 0.30$ ). Lastly, we did not find support for the reduced patch density hypothesis (H2) that environmental heterogeneity is associated with greater distances between relatively profitable foraging areas, since there was no link between environmental heterogeneity and the mean maximum distance travelled (detailed above) or the total distance travelled (table 2;  $F_{1,33} = 2.59$ ,  $p = 0.12$ ). The proportion of individuals' time spent transiting was significantly lower in more heterogeneous environments (figure 2*a*; parameter estimate  $\pm$  s.e.:  $-0.17 \pm 0.04$ ,  $F_{1,415} = 23.5$ ,  $p < 0.01$ ), and there was no change in the time spent resting (figure 2*a*;  $F_{1,415} = 0.08$ ,  $p = 0.78$ ). There was no link between



**Figure 2.** Comparison between environmental heterogeneity and foraging behaviour of kittiwakes. (a) The proportion of individuals' time while away from the colony in different behavioural states varied over the observed range of heterogeneity: time spent foraging significantly increased ( $F_{1,415} = 18.8$ ,  $p < 0.01$ ), time spent transiting significantly decreased ( $F_{1,415} = 23.5$ ,  $p < 0.01$ ), and there was no change in time spent resting (dashed line,  $F_{1,415} = 0.08$ ,  $p = 0.78$ ). (b) Trip duration significantly increased over observed range of heterogeneity ( $F_{1,33} = 5.11$ ,  $p = 0.031$ ). (c) Overlap between pairs of individuals' 50% core foraging areas significantly increased over observed range of heterogeneity ( $\chi^2_1 = 9.85$ ,  $p = 0.002$ ). Colony environmental heterogeneity is a measure of variance using a principal coordinate analysis. In all cases, error bars show standard error around the mean where GPS data were collected in multiple years, and solid lines show significant regressions with standard error (dotted lines).



**Figure 3.** Kittiwake reproductive success compared with standardized environmental heterogeneity. Solid line shows a significant regression  $\pm$  s.e. (dashed lines) between environmental heterogeneity and reproductive success ( $F_{1,59} = 15.44$ ,  $p < 0.001$ ,  $R^2 = 0.21$ ). Colony environmental heterogeneity is a measure of variance using a principal coordinate analysis. Error bars show standard error around the mean reproductive success from multiple years.

environmental heterogeneity and the size of an individual's 50% core foraging area (table 2;  $F_{1,414} = 0.34$ ,  $p = 0.56$ ). In support of the above results that environmental heterogeneity is an important mechanism driving fitness, reproductive success

was not linked to colony size ( $F_{1,51} = 0.96$ ,  $p = 0.33$ ) or the number of kittiwakes breeding within the foraging radius of the colony ( $F_{1,59} = 1.64$ ,  $p = 0.21$ ).

## 4. Discussion

Heterogeneous resources are inherent within nature [4,5,61], and are typically assumed to be beneficial to foragers [12,13,62]. However, our study demonstrates that in areas of higher environmental heterogeneity (or greater patchiness), kittiwakes undertook longer foraging trips, spent proportionally more time foraging while away from the colony, overlapped more with other individuals and had reduced breeding success. Together, these results are consistent with our hypothesis that environmental heterogeneity may have concentrated resources into relatively more profitable patches; however, this resulted in greater intraspecific competition, with negative consequences for fitness.

Heterogeneous environments can concentrate resources into patches that animals can adapt their behaviour to, in theory to optimize foraging efficiency [1,63]. Indeed, here we show differences in foraging behaviour with environmental heterogeneity; specifically, in more heterogeneous environments kittiwakes undertook longer foraging trips, and while away from the colony spent more time foraging. If overall resource availability was higher in heterogeneous environments, such changes in foraging behaviour could be an adaptation to increase resource acquisition. However, by contrast, we found that reproductive success was lower in heterogeneous environments, suggesting that greater time investment in foraging behaviour was not compensated for by higher energetic returns [64]. Furthermore, we show that in colonies with more heterogeneous local environments,

**Table 2.** Changes in kittiwake behaviour and reproductive success over the range of environmental heterogeneity observed in this study. Rows in italic type showed a significant relationship ( $p < 0.05$ ).

response variable	relationship with increasing heterogeneity	parameter estimate	units	test statistic	p-value
<i>proportion of time foraging</i>	<i>increase</i>	$0.09 \pm 0.04$	<i>proportion</i>	$F_{(1,415)} = 18.8$	$p = 0.029$
<i>proportion of time transiting</i>	<i>decrease</i>	$-0.17 \pm 0.04$	<i>proportion</i>	$F_{(1,415)} = 23.5$	$p < 0.001$
proportion of time resting	no difference	$0.04 \pm 0.05$	proportion	$F_{(1,415)} = 0.08$	$p = 0.479$
<i>mean trip duration</i>	<i>increase</i>	$0.27 \pm 0.12$	<i>hours (log-scale)</i>	$F_{(1,33)} = 5.11$	$p = 0.031$
mean total distance	no difference	$0.19 \pm 0.12$	km (log-scale)	$F_{(1,33)} = 2.59$	$p = 0.117$
mean maximum distance	no difference	$0.12 \pm 0.11$	km (log-scale)	$F_{(1,33)} = 1.11$	$p = 0.299$
<i>foraging area: overlap</i>	<i>increase</i>	$0.01 \pm 0.00$	<i>BA index</i>	$\chi^2_1 = 9.85$	$p = 0.002$
foraging area: size	no difference	$42.2 \pm 71.7$	km <sup>2</sup>	$F_{(1,414)} = 0.34$	$p = 0.561$
<i>breeding success</i>	<i>decrease</i>	$-0.18 \pm 0.05$	<i>fledglings per nest</i>	$F_{(1,59)} = 15.4$	$p < 0.001$

pairs of individuals overlapped more in their core foraging areas, despite no difference in individual foraging area size. These results suggest that in more heterogeneous environments there is more competition between individuals for finite resources, with costs for reproductive success. While this may be balanced by lower competition elsewhere, lower resource availability away from resource patches will limit resource gains, and where resources are concentrated, resource density may still not be sufficient to benefit all competing individuals. Increased competition between individuals also explains extended foraging trip duration, as acquiring sufficient resources takes more time [17,31], which could incur additional energetic costs on adults, reduce offspring provisioning rates and increase the risk of offspring predation during brood neglect [17,18]. Bio-logging devices can cause a slight increase in trip duration [65]; however, we would expect such effects to be equal across colonies. As such, fitness gains from resource patches may in fact be limited by the degree of environmental heterogeneity, because of the potential cost of competition.

Environmental heterogeneity may also decrease reproductive success if a greater variability of habitat types reduces the amount of productive habitat and/or is associated with generally lower primary productivity. If that were the case, we would expect foragers in heterogeneous environments to have to travel further from the colony in order to access high-quality habitat [16,29,66,67]. However, we found no difference in how far kittiwakes travelled away from the colony in heterogeneous environments, even accounting for the size of breeding populations. Maximum foraging distances recorded here (mean maximum distance:  $23.3 \pm 0.8$  km) were within both theoretical and observed ranges of the species (e.g. theoretical based on Isle of May data and kittiwake flight speeds:  $73 \pm 9$  km [27,53], observed at Pribilof Islands, Bering Sea, Alaska:  $206.7 \pm 6.7$  km [68] and observed at Sør-Gjæslingan, Norway:  $303.7 \pm 6.1$  km [69]). We can therefore assume that individuals were not foraging at, or near, their maximum physiological capability, but rather that sufficient resource availability facilitated individuals to remain within relative proximity of the colony. Alternatively, heterogeneity may decrease reproductive success if profitable resource patches are more dispersed in space [18], requiring greater travel distances to reach sufficient resource patches [16]. However, we found no

difference with environmental heterogeneity in the total distance travelled during a foraging trip, and no increase in the proportion of a trip spent transiting or the size of an individual's 50% core foraging area, suggesting no increase in space use to acquire resources. Heterogeneous environments may, however, require behavioural adaptations that, if not compensated for by energetic gains, could contribute to the reduced breeding success observed in this study [70]. For example, environmental heterogeneity can drive the magnitude of temporal variability in resources, which in turn prompts a greater behavioural response to temporal cycles in heterogeneous environments [9].

Foraging behaviour (in particular, foraging range) is typically linked to colony size in central place breeders. Density-dependent prey depletion can increase the colony foraging radius [29,67], up to the physiological constraints of a species, which can then limit the carrying capacity [66]. As such, when considering foraging adaptations and reproductive consequences of environmental heterogeneity here, it is important to recognize the potential effect of colony size. However, we found no link between reproductive success and colony size, nor the number of kittiwakes breeding within the foraging range of the colony, in contrast to previous studies of seabird population dynamics [31]. Our results therefore suggest that the spatial distribution of resources, as shaped by environmental heterogeneity, could be the predominant mechanism driving differences in levels of intraspecific competition, and therefore reproductive success, between kittiwake colonies in the UK and Ireland. Colony size data were from the most recent full census of the UK and Ireland seabird breeding colonies in 2000 [60]. Many sites have documented population declines both before and after the Seabird 2000 survey, and kittiwakes have been reclassified as vulnerable on the IUCN Red List of threatened species [71]. It is therefore likely that during the years of this study, population numbers were well below historic carrying capacity; however, the Seabird 2000 data provide a useful indication of population numbers for this study.

The degree of environmental heterogeneity at each colony remained relatively consistent over time, which may favour an individual to switch breeding colony in favour of homogeneous sites where reproductive success was higher [72,73]. Reproductive success was, however, generally low;

at all but one colony in this study (Coquet) kittiwakes reared less than one fledgling per nest on average. This may mean that the potential increase in reproductive success in more homogeneous environments is not worth the risk of switching breeding site, but instead is outweighed by other factors driving strong site fidelity common among seabirds such as pair bonds [74], familiarity with conspecifics [75] and natal philopatry [76]. Future study could, however, shed light on the effect of environmental heterogeneity on recruitment of prospecting breeders, as well as long-term population trends [77].

## 5. Conclusion

In this study, we demonstrate that, in contrast to common assumptions, environmental heterogeneity is detrimental to breeding success in this species. Environmental heterogeneity can concentrate resources into hotspots, which could offer foraging opportunities; however, it may also increase competition between individuals. Reproductive success is an important driver of population dynamics across taxa [78,79], including adult recruitment in kittiwakes [77], and therefore the results of this study highlight the potential importance of environmental heterogeneity for driving population success and species distributions. Furthermore, environmental heterogeneity may be a key consideration in future studies of species resilience to environmental stressors, particularly given that many species, including kittiwakes, are undergoing population declines.

**Ethics.** Permission was granted for all fieldwork to collect seabird tracking data, as detailed in Wakefield *et al.* [27] and Trevail *et al.* [9].

## References

- Bell WJ. 1990 *Searching behaviour: the behavioural ecology of finding resources*. London, UK: Chapman and Hall.
- Stephens DW, Krebs JR. 1986 *Foraging theory*. Princeton, NJ: Princeton University Press.
- MacArthur RH, Pianka ER. 1966 On optimal use of a patchy environment. *Am. Nat.* **100**, 603–609. (doi:10.1086/282454)
- Sparrow AD. 1999 A heterogeneity of heterogeneities. *Trends Ecol. Evol.* **14**, 422–423. (doi:10.1016/S0169-5347(99)01735-8)
- Tilman D, Kareiva PM. 1998 *Spatial ecology? The role of space in population dynamics and interspecific interactions*. Princeton, NJ: Princeton University Press.
- Thygesen UH, Sommer L, Evans K, Patterson TA. 2016 Dynamic optimal foraging theory explains vertical migrations of bigeye tuna. *Ecology* **97**, 1852–1861. (doi:10.1890/15-1130.1)
- Tyson RB, Friedlaender AS, Nowacek DP. 2016 Does optimal foraging theory predict the foraging performance of a large air-breathing marine predator? *Anim. Behav.* **116**, 223–235. (doi:10.1016/J.ANBEHAV.2016.03.034)
- Watanabe YY, Ito M, Takahashi A. 2014 Testing optimal foraging theory in a penguin–krill system. *Proc. R. Soc. B* **281**, 20132376. (doi:10.1098/rspb.2013.2376)
- Trevail AM, Green JA, Sharples J, Polton JA, Arnould JPY, Patrick SC. 2019 Environmental heterogeneity amplifies behavioural response to a temporal cycle. *Oikos* **128**, 517–528. (doi:10.1111/oik.05579)
- Yang Z, Liu X, Zhou M, Ai D, Wang G, Wang Y, Chu C, Lundholm JT. 2015 The effect of environmental heterogeneity on species richness depends on community position along the environmental gradient. *Sci. Rep.* **5**, 15723. (doi:10.1038/srep15723)
- Stein A, Gerstner K, Kreft H. 2014 Environmental heterogeneity as a universal driver of species richness across taxa, biomes and spatial scales. *Ecol. Lett.* **17**, 866–880. (doi:10.1111/ele.12277)
- Scott BE, Sharples J, Ross ON, Wang J, Pierce GJ, Camphuysen CJ. 2010 Sub-surface hotspots in shallow seas: fine-scale limited locations of top predator foraging habitat indicated by tidal mixing and sub-surface chlorophyll. *Mar. Ecol. Prog. Ser.* **408**, 207–226. (doi:10.3354/meps08552)
- Worm B, Sandow M, Oschlies A, Lotze HK, Myers RA. 2005 Global patterns of predator diversity in the open oceans. *Science* **309**, 1365–1369. (doi:10.1126/science.1113399)
- Abrahms B, Scales KL, Hazen EL, Bograd SJ, Schick RS, Robinson PW, Costa DP. 2018 Mesoscale activity facilitates energy gain in a top predator. *Proc. R. Soc. B* **285**, 20181101. (doi:10.1098/rspb.2018.1101)
- Bernstein C, Kacelnik A, Krebs JR. 1991 Individual decisions and the distribution of predators in a patchy environment. II. The influence of travel costs and structure of the environment. *J. Anim. Ecol.* **60**, 205–225. (doi:10.2307/5455)
- Pinaud D, Chérel Y, Weimerskirch H. 2005 Effect of environmental variability on habitat selection, diet, provisioning behaviour and chick growth in yellow nosed albatrosses. *Mar. Ecol. Ser.* **298**, 295–304. (doi:10.3354/meps298295)
- Lewis S, Sherratt TN, Hamer KC, Wanless S. 2001 Evidence of intra-specific competition for food in a pelagic seabird. *Nature* **412**, 816–819. (doi:10.1038/35090566)
- López-Bao JV, Palomares F, Rodríguez A, Ferreras P. 2011 Intraspecific interference influences the use of prey hotspots. *Oikos* **120**, 1489–1496. (doi:10.1111/j.1600-0706.2011.19194.x)
- Barrette M, Boivin G, Brodeur J, Giraldeau LA. 2010 Travel time affects optimal diets in depleting patches. *Behav. Ecol. Sociobiol.* **64**, 593–598. (doi:10.1007/s00265-009-0876-5)

**Data accessibility.** Data and code used for analyses are available from the Dryad Digital Repository: <https://doi.org/10.5061/dryad.0898ss6> [80] and were made available to editors and reviewers upon submission. Kittiwake tracking data can be accessed via the BirdLife International Seabird Tracking Database.

**Authors' contributions.** A.M.T., J.A.G., J.S. and S.C.P. designed the study. A.M.T. processed the data, conducted the statistical analyses and wrote the manuscript. A.M.T., J.A.G., S.C.P., F.D., E.O., M.B., K.C., S.N. and G.R. collected kittiwake tracking data. J.A.P. and P.I.M. contributed oceanographic data. All authors contributed critically to the drafts, and gave final approval for publication.

**Competing interests.** We declare we have no competing interests.

**Funding.** Tracking work during 2016 and 2017 was funded by the Natural Environment Research Council (grant no. NE/L0020450/1). At the Isle of May, funding was from NERC's National Capability. Funding for the FAME and STAR projects was provided by the European Regional Development Fund through its Atlantic Area Programme, Marine Scotland, Scottish Natural Heritage, Natural England, the Natural Environment Research Council, the Joint Nature Conservation Committee, Natural Resources Wales, Environment Wales, Argyll Bird Club and the RSPB.

**Acknowledgements.** We are grateful to many people for their help with data collection, including David Aitken, Guy Anderson, Julia Baer, Chris Bell, Anthony Bellamy, Ruth Brown, Birgitta Büche, Richard Bufton, Tessa Coledale, Steve Dodd, Derren Fox, Carrie Gunn, Maggie Hall, Robert Hughes, Juliet Lamb, Elizabeth Mackley, Liam McFaul, Mark Newell, Tegan Newman, Mara Nydegger, Shanna Rice, Emily Scragg, Katherine Snell, Yvan Stage, Ed Stubbings, Chris Taylor, Jenny Taylor and Matt Wood. Many thanks to Daisy Burnell at the JNCC for her assistance with kittiwake reproductive success data, which were extracted from the Seabird Monitoring Programme Database at [www.jncc.gov.uk/smp](http://www.jncc.gov.uk/smp). Data were provided to the SMP by the generous contributions of its partners, other organizations and volunteers throughout Britain and Ireland. Thanks to Judith Thomas and Matthew Stoodley at UKHO for their assistance with bathymetry data. We are grateful for comments from two anonymous reviewers that improved the manuscript.



20. Lihoreau M, Chittka L, Raine NE, Kudo G. 2011 Trade-off between travel distance and prioritization of high-reward sites in traplining bumblebees. *Funct. Ecol.* **25**, 1284–1292. (doi:10.1111/j.1365-2435.2011.01881.x)
21. Goldberg JL, Grant JWA, Lefebvre L. 2001 Effects of the temporal predictability and spatial clumping of food on the intensity of competitive aggression in the Zenaida dove. *Behav. Ecol.* **12**, 490–495. (doi:10.1093/beheco/12.4.490)
22. Bonin MC, Bostrom-Einarsson L, Munday PL, Jones GP. 2015 The prevalence and importance of competition among coral reef fishes. *Annu. Rev. Ecol. Evol. Syst.* **46**, 169–190. (doi:10.1146/annurev-ecolsys-112414-054413)
23. Birt VL, Birt TP, Goulet D, Cairns DK, Montevecchi WA. 1987 Ashmole's halo: direct evidence for prey depletion by a seabird. *Mar. Ecol. Prog. Ser.* **40**, 205–208. (doi:10.2307/24825734)
24. Tews J, Brose U, Grimm V, Tielbörger K, Wichmann MC, Schwager M, Jeltsch F. 2004 Animal species diversity driven by habitat heterogeneity/diversity: the importance of keystone structures. *J. Biogeogr.* **31**, 79–92. (doi:10.1046/j.0305-0270.2003.00994.x)
25. Giesel JT. 1976 Reproductive strategies as adaptations to life in temporally heterogeneous environments. *Annu. Rev. Ecol. Syst.* **7**, 57–79. (doi:10.1146/annurev.es.07.110176.000421)
26. Berdahl A, Torney CJ, Schertzer E, Levin SA. 2015 On the evolutionary interplay between dispersal and local adaptation in heterogeneous environments. *Evolution* **69**, 1390–1405. (doi:10.1111/evo.12664)
27. Wakefield ED *et al.* 2017 Breeding density, fine-scale tracking, and large-scale modeling reveal the regional distribution of four seabird species. *Ecol. Appl.* **27**, 2074–2091. (doi:10.1002/eap.1591)
28. Hamer KC, Monaghan P, Uttley JD, Walton P, Burns MD. 1993 The influence of food supply on the breeding ecology of kittiwakes *Rissa tridactyla* in Shetland. *Ibis (Lond. 1859)* **135**, 255–263. (doi:10.1111/j.1474-919X.1993.tb02842.x)
29. Elliott KH, Woo KJ, Gaston AJ, Benvenuti S, Dall'Antonia L, Davoren GK. 2009 Central-place foraging in an arctic seabird provides evidence for Storer-Ashmole's Halo. *Auk* **126**, 613–625. (doi:10.1525/auk.2009.08245)
30. Fauchald P. 2009 Spatial interaction between seabirds and prey: review and synthesis. *Mar. Ecol. Prog. Ser.* **391**, 139–151. (doi:10.3354/meps07818)
31. Ainley DG, Ford RG, Brown ED, Suryan RM, David B. 2003 Prey resources, competition, and geographic structure of kittiwake colonies in Prince William sound. *Ecology* **84**, 709–723. (doi:10.1890/0012-9658(2003)084[0709:PRCAGS]2.0.CO;2)
32. Davoren GK, Montevecchi WA. 2003 Consequences of foraging trip duration on provisioning behaviour and fledging condition of common murre *Uria aalga*. *J. Avian Biol.* **34**, 44–53. (doi:10.1034/j.1600-048X.2003.03008.x)
33. Boersma PD, Rebstock GA. 2009 Foraging distance affects reproductive success in Magellanic penguins. *Mar. Ecol. Prog. Ser.* **375**, 263–275. (doi:10.3354/meps07753)
34. Anderson MJ, Ellingsen KE, McArdle BH. 2006 Multivariate dispersion as a measure of beta diversity. *Ecol. Lett.* **9**, 683–693. (doi:10.1111/j.1461-0248.2006.00926.x)
35. Heino J. 2013 Environmental heterogeneity, dispersal mode, and co-occurrence in stream macroinvertebrates. *Ecol. Evol.* **3**, 344–355. (doi:10.1002/ece3.470)
36. Bini LM, Landeiro VL, Padial AA, Siqueira T, Heino J. 2014 Nutrient enrichment is related to two facets of beta diversity for stream invertebrates across the United States. *Ecology* **95**, 1569–1578. (doi:10.1890/13-0656.1)
37. Conradi T, Kollmann J. 2016 Species pools and environmental sorting control different aspects of plant diversity and functional trait composition in recovering grasslands. *J. Ecol.* **104**, 1314–1325. (doi:10.1111/1365-2745.12617)
38. Oliver T, Roy DB, Hill JK, Brereton T, Thomas CD. 2010 Heterogeneous landscapes promote population stability. *Ecol. Lett.* **13**, 473–484. (doi:10.1111/j.1461-0248.2010.01441.x)
39. Hendrickx F *et al.* 2007 How landscape structure, land-use intensity and habitat diversity affect components of total arthropod diversity in agricultural landscapes. *J. Appl. Ecol.* **44**, 340–351. (doi:10.1111/j.1365-2664.2006.01270.x)
40. Genin A. 2004 Bio-physical coupling in the formation of zooplankton and fish aggregations over abrupt topographies. *J. Mar. Syst.* **50**, 3–20. (doi:10.1016/j.jmarsys.2003.10.008)
41. Stevick PTP, Incze LLS, Kraus SDS, Rosen S, Wolff N, Baukus A. 2008 Trophic relationships and oceanography on and around a small offshore bank. *Mar. Ecol. Prog. Ser.* **363**, 15–28. (doi:10.3354/meps07475)
42. Carroll MJ *et al.* 2015 Effects of sea temperature and stratification changes on seabird breeding success. *Clim. Res.* **66**, 75–89. (doi:10.3354/cr01332)
43. Waggitt JJ, Cazenave PW, Howarth LM, Evans PGH, van der Kooij J, Hiddink JG. 2018 Combined measurements of prey availability explain habitat selection in foraging seabirds. *Biol. Lett.* **14**, 20180348. (doi:10.1098/rsbl.2018.0348)
44. Benazzouz A, Mordane S, Orbi A, Chagdali M, Hilmi K, Atillah A, Lluís Pelegrí J, Hervé D. 2014 An improved coastal upwelling index from sea surface temperature using satellite-based approach: the case of the canary current upwelling system. *Cont. Shelf Res.* **81**, 38–54. (doi:10.1016/J.CSR.2014.03.012)
45. Frederiksen M, Edwards M, Mavor R, Wanless S. 2007 Regional and annual variation in black-legged kittiwake breeding productivity is related to sea surface temperature. *Mar. Ecol. Prog. Ser.* **350**, 137–143. (doi:10.3354/meps07126)
46. Johnston A *et al.* 2015 Modelling the abundance and distribution of marine birds accounting for uncertain species identification. *J. Appl. Ecol.* **52**, 150–160. (doi:10.1111/1365-2664.12364)
47. Franks PJS. 1992 Phytoplankton blooms at fronts: patterns, scales, and physical forcing mechanisms. *Rev. Aquat. Sci.* **6**, 121–137.
48. Franks PJS. 1992 Sink or swim: accumulation of biomass at fronts. *Mar. Ecol. Prog. Ser.* **82**, 1–12. (doi:10.3354/meps082001)
49. Scales KL, Miller PI, Embling CB, Ingram SN, Pirotta E, Votier SC. 2014 Mesoscale fronts as foraging habitats: composite front mapping reveals oceanographic drivers of habitat use for a pelagic seabird. *J. R. Soc. Interface* **11**, 20140679. (doi:10.1098/rsif.2014.0679)
50. Gower JC. 1966 Some distance properties of latent root and vector methods used in multivariate analysis. *Biometrika* **53**, 325–338. (doi:10.2307/2333639)
51. Oksanen J *et al.* 2018 vegan: community ecology package. See <https://cran.r-project.org/web/packages/vegan/index.html>.
52. Warwick-Evans V, Atkinson PW, Arnould JPY, Gauvain R, Soanes L, Robinson LA, Green JA. 2016 Changes in behaviour drive inter-annual variability in the at-sea distribution of northern gannets. *Mar. Biol.* **163**, 156. (doi:10.1007/s00227-016-2922-y)
53. Daunt F, Benvenuti S, Harris MP, Dall'Antonia L, Elston DA, Wanless S. 2002 Foraging strategies of the black-legged kittiwake *Rissa tridactyla* at a North Sea colony: evidence for a maximum foraging range. *Mar. Ecol. Prog. Ser.* **245**, 239–247. (doi:10.3354/meps245239)
54. Patterson TA, Basson M, Bravington MV, Gunn JS. 2009 Classifying movement behaviour in relation to environmental conditions using hidden Markov models. *J. Anim. Ecol.* **78**, 1113–1123. (doi:10.1111/j.1365-2656.2009.01583.x)
55. Ropert-Coudert Y, Grémillet D, Kato A, Ryan PG, Naito Y, Le Maho Y. 2004 A fine-scale time budget of Cape gannets provides insights into the foraging strategies of coastal seabirds. *Anim. Behav.* **67**, 985–992. (doi:10.1016/J.ANBEHAV.2003.09.010)
56. Michelot T, Langrock R, Patterson TA. 2016 moveHMM: an R package for the statistical modelling of animal movement data using hidden Markov models. *Methods Ecol. Evol.* **7**, 1308–1315. (doi:10.1111/2041-210X.12578)
57. Calenge C. 2006 The package 'adehabitat' for the R software: a tool for the analysis of space and habitat use by animals. *Ecol. Modell.* **197**, 516–519. (doi:10.1016/J.ECOLMODEL.2006.03.017)
58. Fieberg J, Kochanny CO. 2005 Quantifying home-range overlap: the importance of the utilization distribution. *J. Wildl. Manage.* **69**, 1346–1359. (doi:10.2193/0022-541X(2005)69[1346:QHOTI]2.0.CO;2)
59. Newell MA, Harris MP, Wanless S, Burthe SJ, Bogdanova MI, Gunn CM, Daunt F. 2016 The Isle of May long-term study (IMLOTS) seabird annual breeding success 1982–2016. (doi:10.5285/02c98a4f-8e20-4c48-8167-1cd5044c4afe)
60. Mitchell PI, Newton SF, Ratcliffe N, Dunn TE. 2004 *Seabird populations of Britain and Ireland: results of the Seabird 2000 census (1998–2002)*. London, UK: T & AD Poyser.

61. Chesson P, Case TJ. 1986 Overview: nonequilibrium community theories: chance, variability, history. In *Community ecology* (eds J Diamond, TJ Case), pp. 229–239. New York: Harper and Row.
62. Lascelles BG, Langham GM, Ronconi RA, Reid JB. 2012 From hotspots to site protection: identifying marine protected areas for seabirds around the globe. *Biol. Conserv.* **156**, 5–14. (doi:10.1016/j.biocon.2011.12.008)
63. Barton RA, Whiten A, Strum SC, Byrne RW, Simpson AJ. 1992 Habitat use and resource availability in baboons. *Anim. Behav.* **43**, 831–844. (doi:10.1016/S0003-3472(05)80206-4)
64. Ballance LT, Ainley DG, Ballard G, Barton K. 2009 An energetic correlate between colony size and foraging effort in seabirds, an example of the Adélie penguin *Pygoscelis adeliae*. *J. Avian Biol.* **40**, 279–288. (doi:10.1111/j.1600-048X.2008.04538.x)
65. Bodey TW, Cleasby IR, Bell F, Parr N, Schultz A, Votier SC, Bearhop S. 2018 A phylogenetically controlled meta-analysis of biologging device effects on birds: deleterious effects and a call for more standardized reporting of study data. *Methods Ecol. Evol.* **9**, 946–955. (doi:10.1111/2041-210x.12934)
66. Jovani R, Lascelles B, Garamszegi LZ, Mavor R, Thaxter CB, Oro D. 2016 Colony size and foraging range in seabirds. *Oikos* **125**, 968–974. (doi:10.1111/oik.02781)
67. Burke CM, Montevecchi WA. 2009 The foraging decisions of a central place foraging seabird in response to fluctuations in local prey conditions. *J. Zool.* **278**, 354–361. (doi:10.1111/j.1469-7998.2009.00584.x)
68. Paredes R *et al.* 2014 Foraging responses of black-legged kittiwakes to prolonged food-shortages around colonies on the Bering Sea shelf. *PLoS ONE* **9**, e92520. (doi:10.1371/journal.pone.0092520)
69. Christensen-Dalsgaard S, May R, Lorentsen S-H. 2017 Taking a trip to the shelf: behavioral decisions are mediated by the proximity to foraging habitats in the black-legged kittiwake. *Ecol. Evol.* **8**, 866–878. (doi:10.1002/ece3.3700)
70. Gils JA, Spaans B, Dekinga A, Piersma T. 2006 Foraging in a tidally structured environment by red knots (*Calidris canutus*): ideal, but not free. *Ecology* **87**, 1189–1202. (doi:10.1890/0012-9658(2006)87[1189:FIATSE]2.0.CO;2)
71. BirdLife International. 2017 *Rissa tridactyla*. The IUCN Red List of Threatened Species 2017: e.T22694497A118366481.
72. Doligez B, Cadet C, Danchin E, Boulinier T. 2003 When to use public information for breeding habitat selection? The role of environmental predictability and density dependence. *Anim. Behav.* **66**, 973–988. (doi:10.1006/ANBE.2002.2270)
73. Danchin E, Boulinier T, Massot M, Danchin E, Boulinier T, Massot M. 2013 Conspecific reproductive success and breeding habitat selection? Implications for the study of coloniality. *Ecology* **79**, 2415–2428. (doi:10.1890/0012-9658(1998)079[2415:CRSABH]2.0.CO;2)
74. Rebke M, Becker PH, Colchero F. 2017 Better the devil you know: common terns stay with a previous partner although pair bond duration does not affect breeding output. *Proc. R. Soc. B* **284**, 20161424. (doi:10.1098/rspb.2016.1424)
75. Grabowska-Zhang AM, Wilkin TA, Sheldon BC. 2012 Effects of neighbor familiarity on reproductive success in the great tit (*Parus major*). *Behav. Ecol.* **23**, 322–333. (doi:10.1093/beheco/arr189)
76. Coulson JC, Coulson BA. 2008 Measuring immigration and philopatry in seabirds; recruitment to black-legged kittiwake colonies. *Ibis (Lond. 1859)* **150**, 288–299. (doi:10.1111/j.1474-919X.2007.00777.x)
77. Suryan RM, Irons DB. 1991 Colony and population dynamics of black-legged kittiwakes in a heterogenous environment. *Auk* **118**, 636–649. (doi:10.1642/0004-8038(2001)118)
78. Jenouvrier S, Barbraud C, Cazelles B, Weimerskirch H. 2005 Modelling population dynamics of seabirds: importance of the effects of climate fluctuations on breeding proportions. *Oikos* **108**, 511–522. (doi:10.1111/j.0030-1299.2005.13351.x)
79. Sandvik H, Erikstad KE, Sæther BE. 2012 Climate affects seabird population dynamics both via reproduction and adult survival. *Mar. Ecol. Prog. Ser.* **454**, 273–284. (doi:10.3354/meps09558)
80. Trevaill AM *et al.* 2019 Data from: Environmental heterogeneity decreases reproductive success via effects on foraging behaviour. Dryad Digital Repository. (doi:10.5061/dryad.0898ss6)