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1 **Detecting macroecological patterns in bacterial communities across independent studies of**
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3
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42

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45

46 **The emergence of high-throughput DNA sequencing methods provides unprecedented**

47 **opportunities to further unravel bacterial biodiversity and its worldwide role from human**

48 **health to ecosystem functioning. However, in spite of the abundance of sequencing studies,**

49 **combining data from multiple individual studies to address macroecological questions of**

50 **bacterial diversity remains methodically challenging and plagued with biases. Here, using a**

51 **machine learning approach that accounts for differences among studies and complex**

52 **interactions among taxa, we merge 30 independent bacterial datasets consisting of 1,998**

53 **soil samples from across 21 countries. While previous meta-analysis efforts have focused on**

54 **bacterial diversity measures or abundances of major taxa, we show that disparate**

55 **amplicon sequence data can be combined at the taxonomy-based level to assess bacterial**

56 **community structure. We find that rarer taxa are more important for structuring soil**

57 **communities than abundant taxa, and that these rarer taxa are better predictors of**
58 **community structure than environmental factors, which are often confounded across**
59 **studies. We conclude that combining data from independent studies can be used to explore**
60 **novel patterns in bacterial communities, identify potential ‘indicator’ taxa with an**
61 **important role in structuring communities, and propose new hypotheses on the factors that**
62 **shape bacterial biogeography previously overlooked.**

63

64 Soil microbial communities are more diverse and contain more individuals than any species
65 groups on the planet^{1,2}. Over the last decade, the use of high-throughput sequencing (HTS)
66 methods has substantially advanced our understanding of the worldwide biogeography and
67 ecology of soil bacterial and fungal communities³⁻⁶. Recent work has further demonstrated that
68 inclusion of microbial composition and functional attributes improves earth system models^{7,8},
69 which is of paramount importance for predicting effects of global change on ecosystem services
70 such as climate regulation or soil fertility^{9,10}. Yet, opposite to the long-standing view that every
71 organism may occur everywhere¹¹, even at small scales bacterial communities turn out to be
72 more patchy than previously expected^{12,13}, raising questions regarding dispersal constraints,
73 temporal dynamics, and niche breadth at the global scale¹⁴⁻¹⁷. Due to these knowledge gaps,
74 combined with practical challenges of exhaustive sample collection and the massive diversity of
75 communities, global assessment of soil microbial diversity remains an ongoing research
76 challenge^{18,19}.

77

78 For plants and animals, the integration of data from independent studies has been a valuable
79 option for generating an understanding of global biogeography patterns, answering ecological

80 questions (e.g. biodiversity-functioning relationships), and identifying threats to biodiversity
81 from global changes²⁰⁻²³. Similarly, our understanding of soil microbial diversity would greatly
82 improve from such worldwide assessments. However, the integration of microbial community
83 HTS data from different studies is not so unlike the merging of museum species records where
84 information and data is constrained by variations in nomenclature over space and time, among
85 many other challenges^{24,25}. Like plant and animal records, molecular microbial community
86 records and information can be incomplete, processing and naming varies greatly between
87 studies and over time²⁶, data storage is inconsistent, and there are few curated databases with
88 high quality data (especially for short read sequences)^{27,28}. Further, most microbial community
89 data and metadata are still available only in independently published studies that have been
90 carried out according to their own standards and procedures, and the extent of these confounding
91 factors has never been quantified across studies.

92
93 Regardless of the challenges, as indicated by the many open access data initiatives²⁹⁻³¹, merging
94 microbial sequence data is a potential option to address global scale questions, whether relating
95 to the human microbiome³², marine systems³³, or predicting the response of soil organisms to
96 global environmental change³⁴. For soil systems, the need to merge sequence data is supported
97 by the emerging role of bacterial phyla and classes as indicators of particular soil conditions such
98 as soil pH and nutrient concentrations^{35,36}. Until now, attempts to meta-analyze sequence data
99 have been limited to assessing diversity measures or abundances of major taxa, because the
100 merging of community data is constrained by methodological differences between sequencing
101 studies^{13,30,37-39}. However, a recent systematic review found that measures of microbial
102 community structure were more often linked to microbial process rates than diversity or

103 presence/absence data⁴⁰, and abundance ratios among phyla may be less important than previous
104 believed⁴¹. Together indicating that information on variation in microbial community structure is
105 potentially more ecologically relevant than measures of diversity and abundances of major taxa.

106

107 Here, we show that, despite the outlined challenges, published microbial community data from
108 independent studies can be analyzed together to address questions about the global structuring of
109 communities. Using a novel machine learning approach, we take methodological and technical
110 biases into account, factor in interactions among taxa, and produce an improved assessment of
111 the abiotic and biotic drivers of soil community structure. The objectives of this study were two-
112 fold: (1) to identify the biases and incompatibilities of microbial community HTS studies (and
113 confounding factors) so as to strengthen our ability to integrate data from disparate studies, and
114 (2) to reveal worldwide soil microbial community patterns by merging independent taxonomy-
115 based datasets.

116

117 **Results and Discussion**

118 **Taxonomy-based merging of disparate amplicon sequence data**

119 We identified 30 individual HTS bacterial studies from 21 countries for our analysis (Figure 1A
120 and Supplementary Table 1). While we aimed to merge HTS data of both soil bacterial and
121 fungal datasets, our approach was only successful for bacterial data (Figure 1B and 1C), and
122 highlights the well-known dilemma of fungal databases, where extremely high diversity
123 combined with high endemism and mismatched taxonomy across continents make merging data
124 by taxonomy difficult and unusable for downstream analyses^{4,5,42}. For the bacterial studies, we
125 were able to successfully merge 30 individual OTU tables; using a taxonomy-based approach,

126 datasets were merged using the taxonomic affiliations of individual OTUs. Once filtered, and
127 singletons removed, the final ‘taxonomy-based’ community contained 1,998 individual soil
128 samples, and 8,287 taxa. Here ‘taxon’ is defined as a unique name in the classification; where a
129 name could be a specific phylum, genus, or other taxonomic level. For example, ‘Acidovorax’
130 (genus) and Proteobacteria (the phylum containing Acidovorax) were both considered as taxa).
131 To account for variation in sequencing depth between different studies, OTU relative abundances
132 were used per sample, rather than absolute read abundance. To test known biogeographical
133 patterns, metadata (information on geographical location, soil pH and soil core measurements)
134 were compiled for all studies. Technical and methodical information was also collected; all of
135 these 30 studies had conducted amplicon sequencing on hypervariable regions of the 16S rRNA
136 gene in soil samples using either Illumina or (Roche) 454 pyrosequencing (with any primer pair)
137 (Supplementary Table 1). For a validation step we retrieved all usable raw sequence data
138 available, resulting in 417 samples from locations across the globe (approximately 1/5 of all our
139 samples) (Figure 1A). Data not included in this sequence-matched analysis either had an
140 incompatible raw sequence format or simply no longer existed. Available raw sequence data
141 were combined into a single ‘sequence-matched’ community comprising 44,106 OTUs
142 (Supplementary Figure 1).

143

144 **Machine learning assessment of bacterial community structure**

145 Ordination of the taxonomy-based community reveals large amounts of structure both within and
146 between studies (structure that is removed by permuting taxa among samples (Supplementary
147 Figure 2), without greatly affecting diversity (Supplementary Table 3)), and the observation of
148 the well-established negative relationship between relative abundance of Acidobacteria and soil

149 pH (Figure 1D)⁴³ confirms our merging method. This visualization also suggests that some of the
150 community variation (e.g. the near absence of Acidobacteria in some studies, even at low pH) is
151 due to technical factors such as the particular primer sets chosen, region sequenced, and
152 sequencing platform (Supplementary Methods and Supplementary Table 2). However, we expect
153 that some taxa are not correlated with technical factors, and are non-randomly distributed with
154 respect to biotic and abiotic factors. Therefore, using a machine learning approach capable of
155 accounting for complex interactions among taxa (Random ForestsTM, see methods), we
156 determined the extent to which individual taxa could influence the community structure of
157 merged independent studies. Here community structure is defined by the presence and relative
158 abundances of individual taxa, along with co-occurrence relationships between those taxa. This
159 was done in two ways: first, we constructed a model that classified the study from which a
160 sample came based on the proportions of the 8,287 taxa it contained (1.5% [\pm 0.02% CI]
161 classification error, by internal cross-validation). Second, we determined the contribution of each
162 taxon to bacterial community structure by quantifying its importance in a model that separated
163 the observed data from synthetic data randomly drawn from the observed distributions of relative
164 abundances for each taxon^{44,45} (*see Methods*).

165
166 Merging of disparate microbial sequence data is known to be plagued with potential biases
167 including: lack of standardization of sample collection, methodological issues regarding DNA
168 extraction and primer choice, incomplete metadata, the technical biases of different sequencing
169 platforms, sequencing depth, PCR Bias, different clustering methods, and the use of different
170 taxonomic classification pipelines⁴⁶⁻⁵². We therefore took the novel step to quantify the
171 importance of both technical and environmental factors alongside taxa in the Random Forests

172 models (Figure 2). Of note, ‘owner’, which encompasses the technical biases and uniqueness of a
173 given dataset, is very effective for differentiating between studies (i.e. the owner is far to the
174 right in Figure 2) yet is entirely uninformative about community structure (i.e. owner is at the far
175 bottom in Figure 2). In fact, *all* technical factors included are better than 98.5% of all taxa to
176 differentiate between studies, indicating that the observed differences among studies in taxon
177 relative abundances are strongly confounded with technical factors. Independent of taxonomy,
178 certain environmental factors, such as country of origin, latitude and longitude, and soil pH, were
179 highly important in differentiating studies but not in determining community structure. By
180 contrast, minimum soil sampling depth was not very important in separating studies, and was
181 more associated with community structure. It is well known that bacterial diversity decreases
182 with soil depth⁵³ and our results show that in a global assessment, soil depth remains a strong
183 predictor of bacterial community composition. Perhaps most useful for future research, this result
184 highlights that not all environmental factors are equally confounded by technical factors, and
185 shows that by combining data from across many independent studies we may identify previously
186 overlooked taxa and factors relevant for structuring communities.

187

188 **Importance for structuring soil bacterial communities**

189 Although all studies were confounded by technical and environmental covariates, there remained
190 many taxa that were non-randomly distributed and were not confounded with technical
191 differences among studies (upper left in Figure 2). When assessing the role of these different taxa
192 in structuring the community, we found a trade-off between taxon abundance and importance in
193 community structure, such that low abundance taxa are disproportionately important in the non-
194 random structure of communities, where the most important taxa are rarer than expected

195 compared to the randomly permuted data (Figure 3). Thus, the importance of taxa for
196 determining community structure is negatively correlated with the average abundance of those
197 taxa, whereas taxon abundance is positively correlated with importance for separating studies (ρ
198 = -0.79 and $\rho = +0.51$ respectively, rank correlation, cf. null expectations of $\rho = -0.62$ and -0.12
199 respectively in permuted data). The taxa most closely associated with differences between
200 studies tend to be those present at or greater than 0.1% relative abundance, but those most
201 important in determining community structure tend to be present at 0.0001% abundance or less
202 (with a null expectation of around 0.01-0.001% in each case, Figure 3). This result is only found
203 by considering the full set of studies and is neither apparent within single studies (Supplementary
204 Fig. 4A-B) nor a subset of studies (whether matched by name or sequence Supplementary Fig.
205 5). It corresponds to the long tail in frequency-abundance distributions of soil microbial
206 communities⁵⁴, where many taxa in the soil are known to occur at low abundance. Thus, if rarer
207 taxa tend to be more important for distinguishing between communities, it is within this long tail
208 that we might identify taxa that could indicate ecological or functional differences among soil
209 communities^{33,55,56}.

210

211 To be ecological indicators^{57,58}, taxa need to vary in abundance in response to environmental
212 factors and have high occurrence across studies, as is the case for the phylum Acidobacteria⁴³.
213 Acidobacteria, however, are typically abundant and our analysis suggests that the most abundant
214 taxa are *not* the most important in determining community structure. While dominant taxa like
215 Acidobacteria do change with environmental factors such as pH (Figure 1D), those changes are
216 of lesser importance for the ‘non-randomness’ of community structure, and more confounded
217 with technical effects, than changes in less dominant, pH responsive taxa (Supplementary Figure

218 3A). Therefore, we assessed which taxonomic ranks are more or less distinguished from the
219 randomly permuted data. Although differences among domains and phyla are strongly
220 associated with differences among studies (Figure 4B) only taxa at a rank lower than phyla are
221 consistently better than random at identifying community structure (Figure 4A).

222

223 A very similar pattern was found for the sequence-matched community, emphasizing the
224 importance of taxa at the level of Class and below (Supplementary Figure 7A and 7B). However,
225 this was not apparent in individual studies (Supplementary Figure 4C-D), where phyla were
226 relatively important. A subset of the taxonomy-matched studies showed a pattern intermediate
227 between the single studies and the full dataset (phyla with some importance, but less than Class,
228 Order or Family, Supplementary Figure 7C). This, along with abundance analyses (Figure 3 and
229 Supplementary Figure 5), suggests that our name matching approach is consistent with, but less
230 powerful than a full sequence-matched analysis. At the same time, the taxonomy-matching is
231 worthwhile because, as with the findings on abundance (Figure 3), macroecological patterns (the
232 importance of taxa below phyla and of relatively low abundance in community structure) are
233 evident when we consider thousands of samples from tens of studies, that are not apparent from
234 hundreds of samples from one or a handful of studies.

235

236 To be a good ecological indicator a taxon should occur in most studies; we therefore looked
237 explicitly at the relationship between a taxon's importance in community structure and its
238 occurrence across studies. Low abundance taxa and taxa of lower taxonomic rank are
239 consistently important in determining community structure, but tend to be detected in fewer
240 studies ($\rho = 0.59$ and 0.31 respectively Supplementary Figure 3B and 3C). We discovered a

241 novel relationship between taxon occurrence across studies and importance for structuring
242 communities for all taxa (Figure 5, Supplementary Table 4). Comparison with the null
243 expectation reveals a range of taxa, occurring in multiple samples from most studies, which are
244 much more important in determining community structure than expected by chance. A similar
245 pattern is apparent in the sequence-matched dataset (Supplementary Figure 8A) and the same
246 subset of studies when taxonomy-matched (Supplementary Figure 8B). Altogether, the analyses
247 clearly illustrate the significance of taxonomic rank, for example *class* Gemmatimonadetes is
248 relatively unimportant for community structure but *genus* Gemmatimonadetes is relatively
249 important. The result also shows rarer taxa being more important in structuring communities and
250 suggests rarer bacterial taxa play overlooked ecologically important roles for bacterial
251 community dynamics⁵⁶. This result is robust to artifacts caused by the rarest taxa (e.g.
252 differences between 0 and 1 reads in a sample could be significant for a model, without being
253 biologically significant) – a very similar pattern is seen when only taxa present at above 0.003%
254 in any given sample were included in this analysis (typically removing the rarest 10% of taxa
255 from any given sample, Supplementary Figure 9). Conversely, many taxa of high taxonomic rank
256 with high occurrence across samples, such as the phyla Actinobacteria, Acidobacteria,
257 Proteobacteria, and Bacteroidetes, were much less important for community structure than the
258 null expectation. These taxa have been reported elsewhere as ‘core’ members of the soil
259 community^{43,59,60}, and even been included in source-tracking of microbial communities due to
260 their ubiquitous presence in soil⁶¹. Yet, it is the consistent presence of the core taxa across
261 samples and studies that makes them inadequate for assessing community structure.

262

263 **Conclusions**

264 Our results demonstrate the power of combining global bacterial HTS data from multiple
265 independent sources for the detection of biogeographical patterns and for identifying community
266 patterns that can be used to generate hypotheses on the roles of certain taxa. Though our
267 assessment was on soil communities, our methods can be applied to broadly to other microbial
268 datasets and disciplines. Taxonomy-based merging gives results that are consistent with raw
269 sequence data, and expands opportunities for extracting information about microbial
270 communities from the wealth of existing and future studies. Moreover, we find that rarer
271 bacterial taxa are more important in differentiating communities than previously assumed, and
272 hold potential as overlooked soil indicators or keystone species. Still, there are considerable
273 challenges associated with merging large sequence datasets beyond the well-known biases that
274 accompany any molecular HTS study. Perhaps the most concerning was that so few raw
275 sequence datasets for publically deposited analyses could be retrieved. This highlights the need
276 for wider community adoption of open and accessible short read sequence databases⁶², open
277 reference clustering⁶³, standardized databases⁶⁴ and—as always—that metadata should be
278 consistent and accessible. Regardless of these challenges, as HTS methods rapidly advance we
279 must find ways to simultaneously curate and carry our research knowledge forward³². Only then,
280 in combination with the many novel and classical approaches, can we uncover the full breadth of
281 soil diversity and the roles soil microbes play for ecosystem processes.

282

283 **Methods:**

284 *Description of datasets:*

285 Metadata from the 30 studies and 1998 samples were collected and compiled into a summary
286 data file. To do so, we standardized the metadata of each study using the dplyr package
287 (Wickham & Francois, 2016) of the R statistical platform (R Core Team, 2016). Samples were
288 collected from 21 countries representing all continents except Antarctica. In addition to location
289 and pH data (median = 6.1, quartile range=5.3-7.0), which were available from all studies,
290 information on altitude (10 m, 10-860 m), soil moisture (19.5%, 14.1-27.4%), and total soil
291 nitrogen (0.36 mg kg⁻¹, 0.23-0.51 mg kg⁻¹), carbon (4.7%, 1.9-7.5%) and phosphorus (20.7 mg
292 kg⁻¹, 7.0-223.0 mg kg⁻¹) was noted where available. Depth of sample collection was also noted
293 and ranged from surface collections to a maximum depth of 70 cm, with 83% of samples
294 originating from 0-10 cm below the soil surface. Samples represented anthropogenically
295 managed (59%) and natural (40%; remaining samples undefined) systems, and were taken from
296 arable, grassland, peatland, forest, scrub (including tundra) and urban habitats. The majority of
297 samples (71%) were described as non-experimental, meaning no treatments were applied, with
298 the remainder described as experimental. Sequencing data were either produced using Roche 454
299 technology (22%) or one of the Illumina platforms (78%). Primer pairs were defined for 92% of
300 the samples and nine different pairs were identified from the study meta data (27F:338R;
301 341F:518R; 341F:806R; 341F:907R; 357F:926R; 515F:806R; 577F:926R; 799F:1193R and
302 341F:805R) with the majority of samples (66%) using 515F and 806R to produce amplicons.
303 Post sequencing processing varied, but 81% of samples were run through the QIIME workflow
304 at some point. An OTU table for 1 study comprising 43 samples was programmatically retrieved
305 from the MG-RAST public metagenome repository⁶⁵. Taxonomy for the different studies was

306 mainly assigned using the Greengenes database (84 %), but RDP (6 %;⁴⁶ and the Silva database
307 (9 %)⁶⁶ were also used.

308

309 *Merging OTU tables:*

310 For the OTU tables from the 30 individual studies to be merged, extensive data cleaning was
311 carried out on the OTU and taxonomy files to maximize the possibility of matching taxa across
312 datasets. This comprised several steps: (1) Most datasets contained a seven-level taxonomy,
313 recorded in a variety of ways, which was converted to a standardized format. (2) Individual
314 taxon names were cleaned, to give a single name at each taxonomic level (e.g. removing special
315 characters and extra annotations, such as ‘candidate division’ or details of containing taxa). (3)
316 For the many cases where a taxon was not assigned at a particular taxonomic level, a unified
317 ‘unassigned’ label was created. Repeating analyses with all these taxa removed made no
318 qualitative difference to the results (Supplementary Figure 10). Merging at the taxonomy-based
319 level has the added benefit of lessening the impacts of hypervariable regions. For example, the
320 identification of an organism at a specific level in one sample also contributes to the
321 identification of the containing genus for that sample, allowing direct comparison with a sample
322 where, because a different region was sequenced, that same organism is only resolved to the
323 genus level. Next, relative abundance data were, where necessary, re-scaled to sum to 1 for a
324 sample, using original OTU count files where possible. These values were then manipulated to
325 give data tables usable for modeling using custom R scripts. For some analyses (Figures 3-5), a
326 dataset without community structure was created by randomly permuting the relative abundance
327 of each taxon across all samples. Unless otherwise stated, the analyses performed on the
328 permuted dataset was identical to that performed on the observed data.

329

330 *Merging raw sequence data and other validation datasets:*

331 While no dataset can currently provide a “ground truth” against which to judge our approach, we
332 can at least validate it. The primary validation of our taxonomy-matching approach was to merge
333 raw sequence data (‘sequence-matched’) from five studies (Supplementary Table 1). Per sample
334 fastq files were obtained for each individual dataset. Read files were quality filtered with sickle
335 ⁶⁷ for single end reads trimming bases below phred score 36 and shorter than 100bp. These
336 stringent filtering criteria were applied to keep only high quality reads and to make sure it is
337 possible to map reads to full length 16S rRNA gene sequences. Full length 16S rRNA gene
338 sequences from the Silva 119 release ⁶⁶ were obtained in Qiime compatible format from the [Silva](#)
339 [Download Archive](#) For each dataset, all reads were mapped to the full length 16S rRNA gene
340 sequences using the usearch global algorithm implemented in VSEARCH version 1.9.6 ⁶⁸. The
341 alignment results in usearch table format (uc) were directly converted to BIOM format using
342 biom version 2.1.5 ⁶⁹. Consensus/majority taxonomy was added as metadata to the biom file.
343 Finally, all BIOM files of each dataset were merged using Qiime version 1.9.1 ⁷⁰. All steps were
344 implemented in a workflow made with Snakemake version 3.5.4 ⁷¹ available: ([De Hollander](#)
345 [2016](#)). See Supplementary Fig 1 for workflow.

346

347 To use this sequence-matched dataset to validate our taxonomy-matching approach across
348 studies using different taxonomy databases (Supplementary Figures 5, 7 & 8) we created an
349 equivalent taxonomy-matched dataset from the same 5 studies. As with the full dataset, only taxa
350 occurring in at least two studies were included in either this or the sequence-matched dataset. To
351 test what is gained or lost by considering different numbers of studies simultaneously, we

352 considered, not only the full dataset (30 studies) and the subset of 5 studies used in the sequence-
353 matched dataset, but two of the largest individual studies: from Central Park, NYC
354 encompassing 594 samples (study #24) and a global dataset encompassing 103 samples (study
355 #30). In each case a simple subset of the full dataset was analyzed (Supplementary Figure 4). To
356 address PCR biases (Supplementary Table 2) and biases associated with rare taxa, we created a
357 filtered subset of the data where only taxa present at above 0.003% in any given sample were
358 considered, meaning that all taxa deemed present are represented by multiple sequence reads
359 (Supplementary Figure 9). To address the issue of differential 16S copy numbers skewing
360 abundance estimates, we created a binary dataset of the presence/absence of all taxa. The results
361 for a model separating studies using this dataset were very similar to the main dataset using
362 relative abundance, however, there was insufficient power to identify taxa important for
363 community structure (Supplementary Figure 6). Nonetheless, this analysis did agree with the
364 main analysis that phyla were the most stable taxonomic level, with lower importance than on
365 the permuted data (Supplementary Figure 6). Finally, to test the effect of ‘unknown’ or
366 unclassified bacterial taxa we created a reduced dataset where all taxa classified as ‘unassigned’
367 at any level were removed (Supplementary Figure 10).

368

369 *Random forest models.*

370 To test for the importance of different taxa in the structuring of the data we used Random Forest
371 models^{45,72} with the relative abundances of the taxa as explanatory variables. Random Forest
372 models have two principal advantages in this context: 1) they can deal easily with thousands of
373 explanatory variables and quantify their relative importance, and 2) they can run equivalently in
374 both supervised and un-supervised modes. In the latter, the importance of a variable describes

375 how effective it is at separating the observed data from randomized synthetic data⁴⁴. In both
376 cases, a proximity matrix may be generated, which can be used for ordination (Supplementary
377 Figure 2). The importance of individual taxa in a Random Forest relate to traditional ecological
378 measures. For instance, the importance in a supervised model, such as that used separating
379 studies (x-axis in Figure 2) is closely correlated with the sensitivity component of the indicator
380 value of each taxon ($\rho = 0.89$, Supplementary Figure 3D)⁵⁸. There are two key parameters that
381 may be adjusted in a Random Forest model, *mtry*, the number of variables randomly sampled as
382 candidates for a split in the constituent trees and *ntree*, the number of trees in the forest. *mtry* was
383 set at its default value (square root of the number of variables) *ntree* was set to 100,000 for each
384 forest. Such a large number of trees was found to be necessary to achieve stable importance
385 across taxa and was achieved by combining several forests run in parallel without normalizing
386 votes. Other parameters were left at default values, in particular, trees were grown to completion
387 (i.e. a minimum node size of 1). The un-scaled permutation importance of variables is used
388 throughout: Each variable importance is the difference between the classification error rate of a
389 tree on data not used to construct it (the ‘out of bag’ data) and the same error following random
390 permutation of the variable in question, averaged over all trees.

391

392 We used permuted data (see above) to create null distributions for taxon importance. For
393 unsupervised Random Forests analyses, such as the community structure model, this amounts to
394 calculating how important a taxon with a particular abundance distribution is for separating two
395 randomized distributions. This can then be compared to its importance for separating the
396 observed from a randomized distribution. This clarifies the fact that, even in null data without
397 community structure (Supplementary Figure 2), variable importance correlates with ecologically

398 important factors, such as abundance. This makes intuitive sense in as much as, even with
399 randomized samples, is easier to separate them on the basis of taxa that occur in only some of
400 them than on the basis of ubiquitous taxa. This, for instance, results in the negative slope of the
401 orange (permuted, null, data) line in Figure 5.

402

403 All analyses were completed with RandomForest package for R version 4.6.

404

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417

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577

578 **Figures:**

579

580 **Figure 1. Merging of data from 32 independent studies demonstrates wide geographic**
581 **breadth, community variation, and confirms the well-known importance of soil pH. A.** Map
582 of locations from which samples were collected, with zoom panels on the United States (left) and
583 western Europe (right). Points in blue were used in both the taxonomy-based and raw-unified
584 analyses and red points were only used in taxonomy-based analyses. **B.** Average proportion of
585 total prokaryotic abundance and **C.** eukaryotic abundance, represented by taxa shared among
586 different numbers of datasets at different taxonomic levels. Level 1 indicates the complete data,
587 levels 2-4 are subsets of the data containing only taxa present in a minimum of 2-4 separate
588 datasets. **D.** Correlation plot of Acidobacteria relative abundance to soil pH where each color
589 represents a different study ($r = -0.42$ $p = 8.6 \times 10^{-87}$).

590

591 **Figure 2: Regardless of technical differences between studies, many bacterial taxa are still**
592 **informative about bacterial community structure.** Machine learning models classify the study
593 from which samples came (x-axis) based on the relative abundance of taxa within samples and
594 distinguish the observed distribution of taxa among samples from random (y-axis). Plotted
595 alongside bacterial taxa (black) are technical factors (red) and ecological factors (purple),
596 including soil pH, minimum and maximum soil depth, longitude, latitude and degrees from the
597 equator. All values are variable importance from Random Forest models (see *Methods*) – points
598 further to the right on the x-axis have more importance in separating studies, while points higher
599 up on the y-axis, have more importance for community structure. Note the non-linear axes.

600

601 **Figure 3: Rarer taxa are more important for structuring communities than abundant taxa.**

602 Here we show the thousand most important bacterial taxa in community structure (A) and in
603 separating studies (B) with respect to their average relative abundance across samples. Plotted
604 are the ‘observed’ points (green) and ‘permuted’ points (orange) which are a null distribution
605 from performing the same analysis on a permuted dataset (see *Methods*). The y-axis reports the
606 rank variable importance in the Random Forests model of community structure (see *Methods*),
607 i.e. the taxon with the greatest importance in this model is ranked 1, the second greatest 2, etc.

608

609 **Figure 4: The importance of bacterial taxa classified at different taxonomic ranks.** Lower

610 taxonomic rank is more important for community structure (A), while high taxonomic rank is
611 more important for separating studies (B). For each taxon, the difference was calculated between
612 the variable importance (see *Methods*) of that taxon in a Random Forests model of either
613 community structure or separating studies and the equivalent value from an analysis performed
614 on the permuted dataset (see *Methods*). The lines and grey ribbons show the mean and standard
615 error respectively of these values across taxa at each taxonomic r considered.

616

617 **Figure 5: Importance of bacterial taxa in community structure related to their occurrence**

618 **in different studies.** The y-axis reports the variable importance in the Random Forests model of
619 community structure (see *Methods*). Green ‘observed’ points correspond to those taxa shown in
620 Figure 1. Orange ‘permuted’ points correspond to the same analysis on a null distribution (see
621 *Methods*). Lines are general additive model (gam) smoothers. Each line is shown with a
622 confidence interval (grey); where this is not visible it is narrower than the line it surrounds.

623









