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1 **EDITORIAL: Recent developments in surface complexation modeling**

2 This special section on surface complexation modeling contains papers presented during the
3 special session “Recent developments in surface complexation modelling: Trace element
4 speciation and reactive transport modelling” at the 12th International Conference on the
5 Biogeochemistry of Trace Elements (ICOBTE), which took place 16–20 June 2013 in Athens,
6 Georgia, USA.

7 Ion-binding and surface complexation models have been developed during the past 2 decades to
8 describe equilibrium trace element binding to complex reactive components of the surface and
9 near-surface environments, including natural organic matter, hydrous metal oxides, and clay
10 minerals. Because of their semi-mechanistic and process-based nature, these models are
11 applicable over a wide range of conditions and have potentially powerful uses in hazard and risk
12 assessment of metals and metalloids.

13 In the past, such models largely have been used to increase understanding of trace element
14 complexation and adsorption processes, although there also have been important applications in
15 hazard and risk assessment, particularly the biotic ligand model (BLM), and in multimedia fate and
16 transport modeling. With this special section, we aimed to stimulate broader application of these
17 models by establishing the state of the art and highlighting recent developments in the field.

18 The state of the art is established in two reviews, each of which deals with one of the two main
19 approaches for applying ion-binding models to natural systems comprising multiple reactive
20 components [1, 2]. The review by Goldberg [1] describes the generalized composite approach, in
21 which binding to the reactive components of a natural material such as a soil, sediment, or aquifer
22 is simulated using a single composite binding phase whose binding sites represent the overall
23 binding properties of the natural material. This requires separate fitting of conditional model
24 parameters for each natural material. The approach may be adapted for predictive purposes by
25 regressing model parameters against natural material properties such as cation exchange
26 capacity, surface area, organic carbon content, inorganic carbon content, aluminum oxide content,
27 and iron oxide content.

28 The review paper by Groenenberg and Lofts [2] describes the component additivity approach,
29 whereby the total binding of elements to the assemblage of reactive components is computed by
30 simulating and summing binding to the individual components. Model parameters, including
31 binding constants and site densities for each component, are determined a priori from binding
32 experiments with the specific components. The approach has the advantage of wide applicability in
33 principle, provided that ion binding to the in situ reactive components may be sufficiently well
34 characterized. Groenenberg and Lofts [2] describe the development of models for specific
35 components and review the literature on the application of such models to both artificial and natural
36 composites. “The characterization of natural assemblages in terms of their components (active
37 organic matter, reactive oxide surface) is key to successful model applications. Improved methods
38 for characterization of reactive components in situ will enhance the applicability of assemblage

39 models. Collection of compositional data for soil and water archetypes, or the development of
40 relationships to estimate compositions from geospatially available data, will further facilitate
41 assemblage model use for predictive purposes” [2].

42 Of the four research studies included in the section, three focus on modeling the adsorption of
43 oxyanions to soils. The study by Gabos et al. [3] is an example of the application of the generalized
44 composite approach to a series of tropical soils with a broad range in soil properties. The study
45 demonstrates how application of the generalized composite approach results in a range of
46 conditional binding parameters that could be related to soil properties to provide a framework for
47 predictive modeling of new soils. The two studies by Perez et al. [4, 5] describe oxyanion sorption
48 to a ferrallic soil using the charge distribution surface complexation model with goethite as a proxy
49 for the reactive surface; thus, the modeling is a limiting case of the component assemblage
50 approach. Modeling takes into account the competitive adsorption of organic matter onto the soil
51 and its effect on reducing the available site density for oxyanion sorption. The first
52 study [4] describes the modeling of phosphate adsorption to soil using standard binding constants
53 but with optimization of the soil surface characteristics, which yielded fitted values of the amount of
54 phosphate reversibly bound to the surface, the reactive surface area, and the site density of
55 adsorbed organic matter on the soil surface. The parameters obtained in the first study were used
56 to model binding of chromate, selenite, molybdate, and arsenate adsorption to the soil, described
57 in the second study [5].

58 In recent years, ion-binding modeling—specifically, binding to natural organic matter in surface
59 waters—has found a significant application in improved understanding and prediction of how metal
60 bioavailability and toxicity vary with exposure medium chemistry. Development and application of
61 bioavailability models, particularly the BLM, has reached a stage where the knowledge gained is in
62 the process of being applied to set site-specific environmental quality standards for certain metals
63 in some legislatures. This represents a significant success for speciation modeling in practical
64 application to risk assessment and chemical management. Research into the relationships
65 between speciation, bioavailability, and toxicity continues to develop. The final study in this section,
66 by Antunes and Kreager [6], shows how chemical speciation modeling can aid the description of
67 dose-response relations for biota under laboratory conditions.

68 Overall, advancements in the mechanistic modeling of ion-binding to natural surfaces mean that, in
69 many cases, models are sufficiently advanced to be “ground-truthed” against real field data, which
70 is necessary for their robust application in hazard and risk assessment. The next stage is to
71 establish how well models can perform in application to the field, particularly in relation to data
72 availability at different scales, and to assess how well current methods for field characterization
73 provide suitable data for modeling. Depending on the desired timescales of application, the need to
74 incorporate models into transport frameworks that can consider slower, non-equilibrium processes
75 alongside ion-binding needs to be considered.

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