

# Biogeochemical responses to nutrient, moisture and temperature manipulations of soil from Signy Island, South Orkney Islands in the Maritime Antarctic

SUN BENHUA<sup>1</sup>, P.G. DENNIS<sup>2</sup>, V.A. LAUDICINA<sup>3</sup>, V.J. ORD<sup>4</sup>, S.P. RUSHTON<sup>4</sup>, A.G. O'DONNELL<sup>5</sup>, K.K. NEWSHAM<sup>6</sup> and D.W. HOPKINS<sup>7\*</sup>

<sup>1</sup>College of Natural Resources and Environment, Northwest A&F University, Yangling, Shaanxi Province, 712100, People's Republic of China

<sup>2</sup>Australian Centre for Ecogenomics and the Advanced Water Management Centre, University of Queensland, Brisbane, QLD 4072, Australia

<sup>3</sup>Dipartimento di Scienze Agrarie e Forestali, Università degli Studi di Palermo, Viale delle Scienze, Edificio 4 – 90128 Palermo, Italy

<sup>4</sup>School of Biology, University of Newcastle, Newcastle upon Tyne NE1 7RU, UK

<sup>5</sup>Institute of Agriculture, Faculty of Science, University of Western Australia, 35 Stirling Highway, Crawley, WA 6009, Australia

<sup>6</sup>British Antarctic Survey, NERC, High Cross, Madingley Road, Cambridge CB3 0ET, UK

<sup>7</sup>School of Life Sciences, Heriot-Watt University, Edinburgh EH14 4AS, UK

\*david.hopkins@hw.ac.uk

**Abstract:** We have investigated how the microbially-driven processes of carbon (C) mineralization (respiration) and nitrogen (N) mineralization/immobilization in a soil from the northern Maritime Antarctic respond to differences in water availability (20% and 80% water-holding capacity) and temperature (5°C and 15°C) in the presence and absence of different organic substrates (2 mg C as either glucose, glycine or tryptone soy broth (TSB) powder (a complex microbial growth medium)) in a controlled laboratory experiment over 175 days. Soil respiration and N mineralization/immobilization in the presence of a C-rich substrate (glucose) increased with increases in water and temperature. These factors were influential individually and had an additive effect when applied together. For the N-rich substrates (glycine and TSB), microbial responses to increased water or temperature alone were weak or not significant, but these factors interacted to give significantly positive increases when applied together. These data indicate that under the expected changes in environmental conditions in the Maritime Antarctic, where temperature and the availability of water and organic substrates will probably increase, soil microbial activity will lead to more rapid C and N cycling and have a positive feedback on these biogeochemical processes, particularly where or when these factors increase concurrently.

Received 18 June 2013, accepted 13 December 2013, first published online 10 March 2014

**Key words:** carbon mineralization, nitrogen mineralization, organic substrates, soil respiration, warming, water addition

## Introduction

Mean annual air temperatures in the Maritime Antarctic, particularly along the western Antarctic Peninsula, have risen at approximately double the rate of the global mean surface temperature over the past 50–100 years (up to 0.4°C decade<sup>-1</sup> compared to 0.2°C decade<sup>-1</sup>; Hansen *et al.* 2006, Adams *et al.* 2009). The warming of terrestrial Maritime Antarctic ecosystems has been accompanied by changes to precipitation patterns, with increases in snow accumulation in recent decades (Adams *et al.* 2009, Thomas *et al.* 2009), and alterations to the population size and range of plant species (Fowbert & Smith 1994, Smith 1994, Convey & Smith 2006, Convey 2011). However, Convey *et al.* (2011) concluded that the recently established southern limit for flowering plants in Antarctica was unlikely to be linked to the recent

warming trend. In the coming decades, the expansion of Maritime Antarctic plant populations will probably enhance and alter nutrient inputs to soils (Hill *et al.* 2011), with inputs of carbon (C) and nitrogen (N) leading to significant increases in soil microbial biomass (Davey & Rothery 1992, Malosso *et al.* 2004, 2005, Dennis *et al.* 2012).

Warming influences the size and composition of soil microbial communities in Maritime Antarctic soils, with substantial increases in cyanobacteria (autotrophs) and heterotrophic microorganisms being associated with experimental warming in the northern Maritime Antarctic (Wynn-Williams 1996, Yergeau *et al.* 2012). Dennis *et al.* (2013a) investigated the effects of increased soil temperature, water and substrate availability on soil bacterial communities on Signy Island (60°S) over 12 months. The study showed that bacterial communities

responded positively to organic substrates, but that responses to warming alone were limited. Little is known of the combined effects of warming and changes in water and substrate availability on the microbial processes in the Maritime Antarctic; however, this is the more probable scenario as changes in environmental factors rarely occur singly. Allison *et al.* (2010) suggest that the soil microbial response to warming depends on the efficiency of the soil microbes in using C, with less C allocated to microbial growth under warming conditions, and that this response differs with organic substrate quality.

We report a controlled laboratory-based study on soils from Signy Island. This study was designed to extend the field experiments described by Dennis *et al.* (2013a), as logistical constraints prevented biogeochemical process measurements in the field. We tested the effects of warming, water addition and substrate supply on the key soil microbial process of C mineralization/soil respiration and inorganic N dynamics (mineralization/immobilization).

## Materials and methods

### *Site and soil*

The soil samples were collected at Wynn Knolls (60°41'56"S, 45°38'10"W), Signy Island, South Orkney Islands (northern Maritime Antarctica) during the 2008–09 summer and transported frozen at -20°C to the UK. Wynn Knolls is situated *c.* 500 m from the western shoreline of Signy Island at an altitude of 199 metres above sea level. Signy Island has an oceanic climate, characterized by dense cloud cover during the summer and precipitation of between 350–500 mm water equivalent per annum, with much of the water falling as rain in the summer (Dennis *et al.* 2013a). Mean air temperatures at Signy Island were between -2–3 °C during the summer and -2–-17°C during the winter (Dennis *et al.* 2013a). At 1–5 cm depth, the mean annual soil temperature was -2.5°C, the highest monthly mean temperature was 3.6°C (January) and the lowest was -9.5°C (July) (Dennis *et al.* 2013a). The soil contained 1.4 mg organic C g<sup>-1</sup> soil and 0.4 mg total N g<sup>-1</sup> soil, and had a pH<sub>water</sub> of 7.5. Further site and soil details are reported in Dennis *et al.* (2012, 2013a).

### *Experimental design*

The soil samples were thawed and sieved (2 mm) in the field-moist state then kept at <5°C for no more than 24 hours before the experiment was established. Sub-samples of soil (40 g dry weight equivalent) were weighed into 64 glass incubation jars with 125 cm<sup>3</sup> volume. The experiment comprised two water addition treatments, two incubation temperatures and addition of three different organic substrates plus a no-substrate control. Each combination was replicated four times (i.e. two water additions x two temperatures x four substrate

treatments x four replicates = 64). The incubation jars were fitted with gas-tight septa and incubated for 175 days. These conditions were designed to support the parallel field experiment (Dennis *et al.* 2013a) and broadly simulate increases in soil nutrients associated with organic input from algae and cyanobacteria.

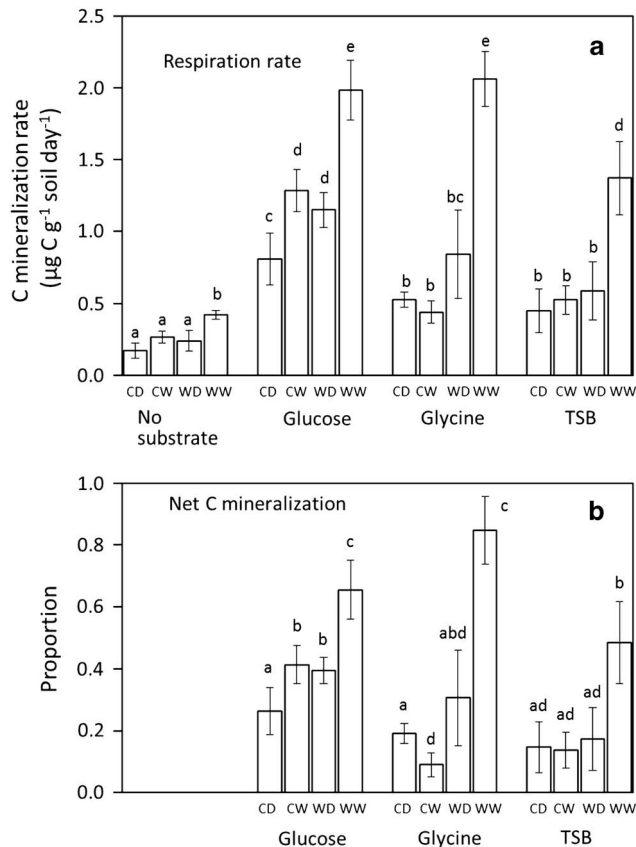
The water content was adjusted by the addition of distilled water to achieve either 20% or 80% of the soil water-holding capacity (WHC; 100% WHC was determined on separate sub-samples of the soil). Soil WHC of 20% and 80% were selected to represent the extremes of dryness and wetness expected on Signy Island. When the soil samples were collected the WHC was between 80–100%. Water content was maintained by re-weighing and addition of distilled water as necessary after each gas sampling occasion (see below).

The soils were incubated at either 5°C or 15°C in thermostatically controlled, fan-assisted incubators. These temperatures were selected to provide a comparison between a warm day in summer (*c.* 5°C) and extreme warming events (15°C).

The substrate treatments included glucose (a simple C-rich organic substrate; Aldrich, UK), glycine (a combined organic C and N substrate; Aldrich, UK) or tryptone soy broth (TSB) powder (a complex organic C and N substrate; Difco, USA), or no-substrate as a control. Each substrate provided an additional 2 mg C g<sup>-1</sup> dry weight soil. The glycine and TSB supplied organic N at the rate of 0.58 and *c.* 0.2 mg N g<sup>-1</sup> dry weight soil, respectively (the N content of TSB is slightly variable because it is not a defined medium).

### *Carbon mineralization/soil respiration*

Periodically (approximately once every 10–15 days) the headspace from each incubation jar was sampled (1 ml volume) with a syringe and CO<sub>2</sub> concentration was determined by gas chromatography (Hopkins & Shiel 1996; Varian 90 GC, fitted with a thermal conductivity detector, Poropak Q column and using helium as the carrier gas). After each gas sampling, the septa were removed from the incubation jars for *c.* 30 minutes to allow headspace gas to be refreshed, then the jars were resealed (as described above, gravimetric water content was corrected after each sampling occasion). The CO<sub>2</sub> concentration was used to calculate respiration rate and cumulative CO<sub>2</sub> production with time. The total CO<sub>2</sub> produced over 175 days from the substrate-amended treatments, minus that from the corresponding no-substrate controls, were used to estimate the equivalent proportion of added C lost as CO<sub>2</sub>. Incubation for 175 days was selected as this is sufficient time for the burst of CO<sub>2</sub> following substrate addition to have subsided (Hopkins *et al.* 2011). Total CO<sub>2</sub> could not be used to accurately estimate the actual substrate C

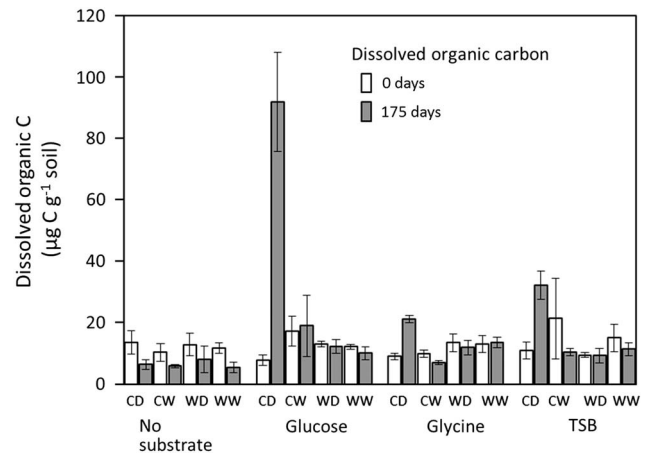


**Fig. 1.** Carbon (C) mineralization/soil respiration for soil from Wynn Knolls, Signy Island after treatment with no substrate, glucose, glycine or tryptone soy broth (TSB) incubated under cold and dry (CD) conditions (5°C and 20% water-holding capacity, WHC), cold and wet (CW) conditions (5°C and 80% WHC), warm and dry (WD) conditions (15°C and 20% WHC), or warm and wet (WW) conditions (15°C and 80% WHC) for 175 days. Each value is the mean of four replicates and the bars are  $\pm$  standard error. **a.** The respiration rate. **b.** The net C mineralization expressed as the proportion of substrate C mineralized (i.e. the additional CO<sub>2</sub> produced over 175 days above that produced in the no-substrate control) with no allowance for priming. Columns with the same letter are not significantly different ( $P < 0.05$ ).

mineralization because a correction could not be made for priming, i.e. the amount of C mineralized from the indigenous soil organic matter as a result of the added substrate. The error on the estimate for priming was probably small due to the low C content of the indigenous soil. However, it cannot be discounted because significant priming has been indicated in another soil sample from Signy Island (Malosso *et al.* 2004).

#### Extractable organic carbon

Extractable organic C in the soil was determined prior to substrate addition and after 175 days of incubation. Organic C was extracted from 10 g (wet weight) soil



**Fig. 2.** Extractable organic carbon (C) concentrations for soil from Wynn Knolls, Signy Island after treatment with no substrate, glucose, glycine or tryptone soy broth (TSB) incubated under cold and dry (CD) conditions (5°C and 20% water-holding capacity, WHC), cold and wet (CW) conditions (5°C and 80% WHC), warm and dry (WD) conditions (15°C and 20% WHC), or warm and wet (WW) conditions (15°C and 80% WHC) at day 0 and day 175. Each value is the mean of four replicates and the bars are  $\pm$  standard error.

with 40 cm<sup>3</sup> of 1.0 M potassium chloride shaken for 30 minutes, centrifuged and frozen prior to flow injection analysis (Skalar Analytical BV, The Netherlands).

#### Nitrogen mineralization/immobilization

The same samples prepared for the extractable organic C analyses were used to determine extractable NH<sub>4</sub><sup>+</sup>-N (ammonium) and NO<sub>3</sub><sup>-</sup>-N (nitrate) plus NO<sub>2</sub><sup>-</sup>-N (nitrite) by flow injection analysis (Skalar Analytical BV). These data were used to estimate net mineralization/immobilization of inorganic N in the soil, but could not be used to reliably estimate total N mineralized from the added substrates due to errors from primed N mineralization and N loss to the gaseous phase by denitrification which was not determined.

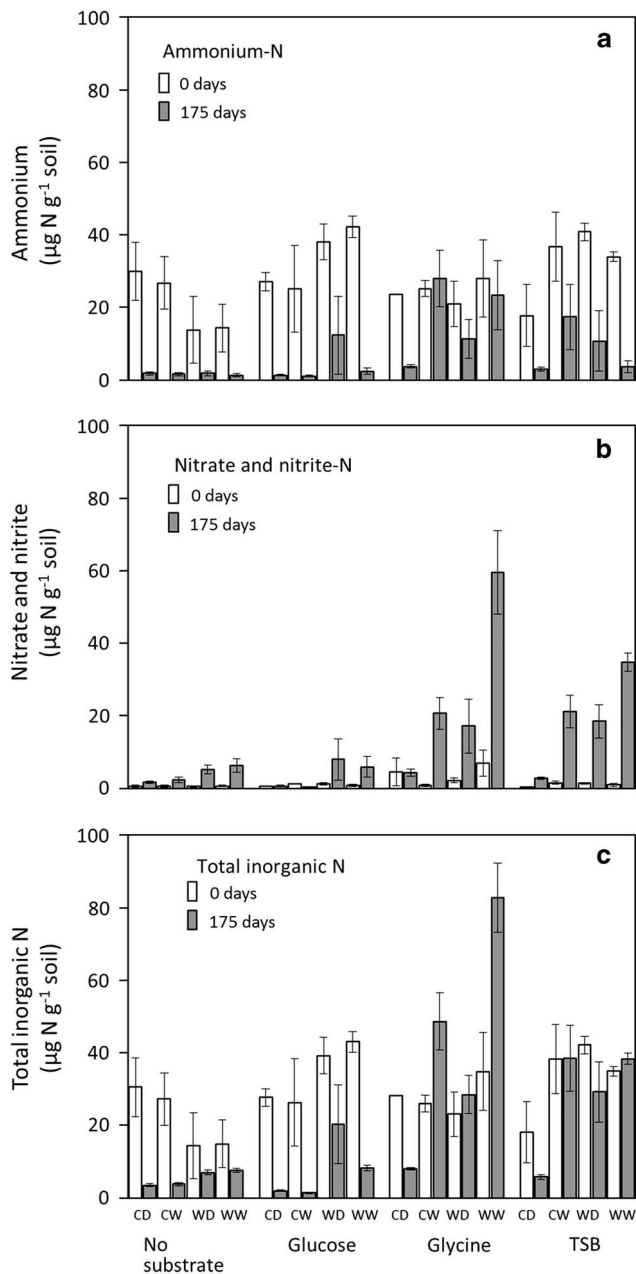
#### Statistical analyses

The data for C mineralization/soil respiration rate and the concentrations of extractable organic C and inorganic N ions were each subject to three-way ANOVA with two timepoints using SPSS-21. The significance differences between means were determined using Tukey's Honestly Significant Difference calculated at  $P < 0.05$ .

## Results

#### Carbon mineralization/soil respiration

The addition of water to 20% and 80% WHC led to a significantly increased basal soil respiration rate. The effect of warming on the basal soil respiration rate was



**Fig. 3.** Inorganic nitrogen (N) concentrations for soil from Wynn Knolls, Signy Island after treatment with no substrate, glucose, glycine or tryptone soy broth (TSB) incubated under cold and dry (CD) conditions ( $5^{\circ}\text{C}$  and 20% water-holding capacity, WHC), cold and wet (CW) conditions ( $5^{\circ}\text{C}$  and 80% WHC), warm and dry (WD) conditions ( $15^{\circ}\text{C}$  and 20% WHC), or warm and wet (WW) conditions ( $15^{\circ}\text{C}$  and 80% WHC) at day 0 and day 175. Each value is the mean of four replicates and the bars are  $\pm$  standard error. **a.** Ammonium ( $\text{NH}_4^+$ ) concentration. **b.** Sum of the nitrate ( $\text{NO}_3^-$ ) and nitrite ( $\text{NO}_2^-$ ) concentrations ( $\text{NO}_2^-$  was present in trace concentrations or undetectable). **c.** Combined inorganic N concentrations (sum of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{NO}_2^-$  concentrations).

not significant. However, there was a significant positive interaction between water addition and warming, i.e. water and warming operated additively to increase the basal respiration rate (Fig. 1a).

Addition of glucose, glycine or TSB led to increased respiration rate for all warming and moisture combinations relative to the no-substrate control (Fig. 1a). In the presence of glucose, both water addition and warming increased respiration rate when applied singly, and there was a positive interaction between water addition and warming (Fig. 1a). Contrasting with the glucose treatment, addition of water or warming alone had no significant effect on respiration in the presence of either glycine or TSB (Fig. 1a). However, both treatments demonstrated significant positive interactions between water addition and warming.

The total C mineralized to  $\text{CO}_2$ , expressed as the proportion of the substrate C, followed the same pattern as the respiration rate (Fig. 1b). Under optimal conditions of water and warming, the total C mineralized was equivalent to 66% for glucose, 85% for glycine and 48% for TSB (Fig. 1b).

#### Extractable organic carbon

In the no-substrate controls, the extractable organic C concentrations declined with time (Fig. 2). Addition of glucose, glycine or TSB led to significantly increased extractable organic C concentrations after 175 days compared with the no-substrate control (Fig. 2). For the cold and dry conditions, the concentration of extractable organic C was greater after 175 days compared with day 0, but for all other water addition and warming combinations the extractable organic C concentration after 175 days was the same as at day 0 (Fig. 2).

#### Nitrogen mineralization/immobilization

For both the no-substrate control and the glucose treatment, the  $\text{NH}_4^+$  concentration declined significantly with time for all water addition and warming combinations (Fig. 3a). For the glycine treatment, the  $\text{NH}_4^+$  concentration after 175 days was significantly less than at day 0 for the cold and dry treatment, but there were no significant differences for the other combinations (Fig. 3a). For the TSB treatments, the  $\text{NH}_4^+$  concentrations all declined significantly with time (Fig. 3a). For both the glycine and TSB treatments, the  $\text{NH}_4^+$  concentration was significantly greater than the no-substrate control after 175 days, except under cold and dry conditions (Fig. 3a).

In all cases, the  $\text{NO}_2^-$  concentrations were very small (often undetectable), thus they are described alongside the  $\text{NO}_3^-$  data. In the no-substrate control,  $\text{NO}_2^-$  and  $\text{NO}_3^-$ -N concentrations increased with time, there were positive effects of water addition and warming, and a significant

interaction between these factors (Fig. 3b). In the presence of glucose, warming led to a significant increase in  $\text{NO}_2^-$  and  $\text{NO}_3^-$ -N concentrations (Fig. 3b). For the glycine and TSB treatments, there were significant increases in  $\text{NO}_2^-$  and  $\text{NO}_3^-$ -N concentrations for both water addition and warming treatments, with a significant interaction between these two factors with glycine (Fig. 3b).

Significant declines in total inorganic N concentration (sum of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{NO}_2^-$ ) were observed over time for all of the water addition and warming combinations for the no-substrate control and glucose treatment (Fig. 3c). For the glycine and TSB treatments, the total inorganic N concentration increased with time in all cases, except under the cold and dry conditions (Fig. 3c).

## Discussion

### *Carbon mineralization/soil respiration*

The addition of an organic substrate increased C mineralization, but the respiratory response to water addition and warming differed according to the substrate. The proportion of C mineralized was consistent with previous studies, where the total amount of C mineralized was equivalent to *c.* 80% of the C added as glucose or amino acids, these studies were comparatively short-term and had warmer incubation temperatures (Hopkins *et al.* 1997, Meli *et al.* 2003). Mineralization of the indigenous soil organic C increased significantly with water addition alone but not warming alone. By contrast, in the presence of glucose mineralization of C responded positively to both water addition and warming. This observation is supported by the higher concentration of extractable organic C in cold and dry conditions in the presence of glucose. The organic substrates that supplied both organic C and N (glycine and TSB) only responded to water addition and increased temperature when these factors were combined. These results may be the consequence of differential responses within the soil microbial community, with some organisms able to utilize glucose at low temperatures provided water is not limiting, and other organisms able to utilize glucose when water availability is low provided the temperature is high enough (Newsham *et al.* 2010).

The positive effect of water addition could be due to more rapid diffusion of the substrate, thus under wetter conditions substrates are more readily available to a larger proportion of the soil microbial community. Such effects have been observed for other organic substrates, the mineralization of which is influenced by the antecedent water content (Zak *et al.* 1999, Newsham *et al.* 2010).

Mineralization of C from glucose and the basal respiration rate were more sensitive to temperature than mineralization from the N-rich substrates. Mineralization

of C from the N-rich substrates only responded positively to temperature in the presence of additional water. The variable temperature sensitivity of C mineralization in soils is associated with substrate complexity; mineralization of more complex substrates shows a stronger temperature response than is seen with simple substrates based on thermodynamic theory (Hartley & Ineson 2008, Hartley *et al.* 2008, Allison *et al.* 2010). This probably applies to TSB, which contains an undefined mixture of peptides, possibly requiring several different catabolic enzymes for complete mineralization. However, glycine is a relatively simple molecule which deaminates to ethanoate (acetate), a precursor in the tricarboxylic acid pathway. It would appear that there is either a thermodynamic constraint on glycine catabolism or that this amino acid is not being mineralized but being used in anabolic metabolism incorporated into microbial biosynthetic processes. Hill *et al.* (2011) identified plants in the Maritime Antarctic with a high affinity for amino acids as an N source, thus it is possible that at least some of the soil microorganisms in this environment do the same.

Irrespective of the mechanism, it is clear that C mineralization is limited by both water and temperature to a greater extent in the presence of the N-rich substrates than C-rich substrates.

### *Nitrogen mineralization/immobilization*

The decline in  $\text{NH}_4^+$  and total inorganic N over time in the no-substrate control and with glucose treatment is evidence for net N immobilization, the result of the soil microbial community being more limited by N supply than C supply, thus N is immobilized from the external environment (Harmsen & van Schreven 1955). Such a response in soils is not uncommon, especially following the addition of a C-rich substrate such as glucose (Dungait *et al.* 2013). Total inorganic N also declined over time in response to water addition and warming in the no-substrate control, a finding consistent with an N limitation to microbial activity.

With the glycine and TSB treatments, the effect of the organic N in the added substrate on net N mineralization is evident. Total inorganic N concentration increased over time in nearly all cases for the glycine treatments and did not decline over time in nearly all cases for TSB, with the exception of cold and dry conditions for both treatments. This is broadly consistent with the C mineralization responses. Constraints on microbial activity probably explain the decline in total inorganic N concentration under cold and dry conditions. Fraser *et al.* (2013) demonstrated a clear temperature sensitivity for enzymes that contribute to N mineralization in soils.

The increases in  $\text{NO}_3^-$  concentrations in nearly all cases coincide with the C mineralization responses to water addition and warming for the different treatments.

This indicates that nitrification was constrained by the same factors as both C mineralization and net N mineralization/immobilization. Net nitrification (i.e. accumulation of  $\text{NO}_3^-$ ) in closed soil incubation is frequently observed (e.g. Hopkins *et al.* 1988), particularly in soils with near neutral pH, because leaching losses are absent and oxygen diffusion in soil is rarely limited over the small distances that occur in laboratory microcosms. Nitrate accumulation suggests that N losses from the soil to the gaseous phase were minor, consistent with the lack of anoxic conditions, and that chemoautotrophic nitrifying bacteria are active in these soils, an observation consistent with an earlier report (Wilson *et al.* 1997). The amount of  $\text{CO}_2$  fixed by nitrifying bacteria is usually small by comparison with the amount of  $\text{CO}_2$  released by heterotrophic microorganisms and, therefore, probably does not influence the interpretation of the C mineralization data, not least because nitrification rates can be temporarily suppressed by high  $\text{CO}_2$  concentrations that could occur in closed incubation vessels (Keeney *et al.* 1985, Kinsbursky & Saltzman 1990).

#### *Relation to field conditions*

Water fluctuations in the field are highly variable depending upon melting and precipitation, both of which can be sudden and lead to rapid changes. During establishment of the field work, the liquid water content of the soil increased to approximately 100% WHC as a result of a rapid rise in temperature leading to snow melt followed by rain (Dennis *et al.* 2013a). For these reasons, the contrast between 20% and 80% WHC can be regarded as representative of field conditions.

The incubation temperatures selected for the laboratory study reflect the difference between mean soil temperatures on a warm day in summer (3.6°C; Dennis *et al.* 2013a) and extreme warming events during which the surface temperature may rise to more than 20°C (Convey 2013).

The additional C and N supplied by the substrates (2 mg C g<sup>-1</sup> soil and 0.58 mg N g<sup>-1</sup> soil or *c.* 0.2 mg N g<sup>-1</sup> soil for glycine and TSB, respectively) represent an approximate doubling in the organic C and total N contents (the soil organic C content was 1.4 mg C g<sup>-1</sup> soil and the total N content was 0.4 mg N g<sup>-1</sup> soil). Although this is a large increment, it is not unrealistic for discrete sites of nutrient deposition in an otherwise resource-poor environment, such as at the sites of guano deposition or directly beneath and surrounding plants (Hopkins *et al.* 2006). For these reasons, the laboratory conditions are considered to be representative of the field conditions.

#### *Comparison with other Antarctic studies*

Positive responses to warming have been reported for microorganisms (Yergeau *et al.* 2012) and nematodes

(Simmons *et al.* 2009) in Antarctic soils. The field-based research with analogous treatments reported by Dennis *et al.* (2013a) demonstrated microbial responses in terms of concentration of lipid biomarkers, i.e. growth, rather than physiological responses, over approximately one year. They showed a positive response to organic substrate additions, but no responses were observed with either water addition or warming and no interactions between factors were identified (Dennis *et al.* 2013a). Soil saturation early in the experiment may account for the lack of any measured effect of water addition in the field. The passive warming method (open-top chambers) employed by Dennis *et al.* (2013a) had a significant effect on temperature by increasing the monthly minimum, mean and maximum temperatures by 0.3, 0.6 and 0.7°C, respectively. However, by comparison with the incubation temperatures in the laboratory, warming in the field was relatively modest. Furthermore, passive warming did not have a significant effect on the number of freeze-thaw cycles in the soil (i.e. the periods when the soil was above 0°C were no more frequent with warming than for the control). This may explain the absence of a microbial response to warming in the field.

An analogous substrate-addition study in the more environmentally extreme Dry Valleys region of Antarctica showed increases in microbial activity (respiration and enzyme activities) over several years (Hopkins *et al.* 2008, Sparrow *et al.* 2011, Dennis *et al.* 2013b). The study included addition of inorganic N ( $\text{NH}_4\text{Cl}$ ) and reported microbial responses to inorganic N when glucose was added, this is consistent with our observations of net N immobilization when a C-rich substrate was added, but net N mineralization when N-rich substrates were added (Hopkins *et al.* 2008, Sparrow *et al.* 2011). Thus, the availability of N in these soils influences microbial activity alongside other factors, including organic C supply. These results are consistent with our findings and those of Dennis *et al.* (2013a). This study did not include warming or effective water addition (temperatures in the Dry Valleys are often too low for meaningful water manipulations except over very short periods), thus there are no results from other water addition or warming experiments with which to compare our results.

#### **Conclusions**

In a soil from the northern Maritime Antarctic, soil respiration and N mineralization/immobilization in the presence of a C-rich substrate (glucose) are positively affected by increases in water and temperature. These factors were influential individually and had an additive effect when applied together. For N-rich substrates (glycine and TSB), microbial responses to increased water or temperature alone were weak or not significant, but these factors interacted to give significantly positive

increases when applied together. These data indicate that under the expected changes in environmental conditions in the Maritime Antarctic, where temperature and the availability of water and organic substrates will probably increase, soil microbial activity will lead to more rapid C and N cycling and have a positive feedback on these biogeochemical processes, particularly where or when these factors increase concurrently.

### Acknowledgements

Funding was supplied by the UK Natural Environment Research Council through the Antarctic Funding Initiative (AFI 7/05; NE/D00893X/1). Logistical support was provided by the British Antarctic Survey. Sun Benhua was supported by the Chinese 111 Project (No. B12007). We would like to thank the reviewers for their constructive comments.

### References

- ADAMS, B., ARTHURN, R. & ATKINSON, A. *et al.* 2009. The instrumental period. In TURNER, J., BINDSCHADLER, R., CONVEY, P., DI PRISCO, G., FAHRBACH, E., GUTT, J., HODGSON, D., MAYEWSKI, P. & SUMMERHAYES, C., eds. *Antarctic climate change and the environment*. Cambridge: Scientific Committee on Antarctic Research, Scott Polar Research Institute, 183–298.
- ALLISON, S.D., WALLENSTEIN, M.D. & BRADFORD, M.A. 2010. Soil-carbon response to warming dependent on microbial physiology. *Nature Geoscience*, **3**, 336–340.
- CONVEY, P. 2011. Antarctic terrestrial biodiversity in a changing world. *Polar Biology*, **34**, 1629–1641.
- CONVEY, P. 2013. Antarctic ecosystems. In LEVIN, S.A. *Encyclopedia of biodiversity*, Vol. 1. 2nd ed. Amsterdam: Academic Press, 179–188.
- CONVEY, P., HOPKINS, D.W., ROBERTS, S.J. & TYLER, A.N. 2011. Global southern limit of flowering plants and moss peat accumulation. *Polar Research*, **30**, 10.3402/polar.v30i08929.
- CONVEY, P. & SMITH, R.I.L. 2006. Responses of terrestrial Antarctic ecosystems to climate change. *Plant Ecology*, **182**, 1–10.
- DAVEY, M.C. & ROTHERY, P. 1992. Factors causing the limitation of growth of terrestrial algae in Maritime Antarctica during late summer. *Polar Biology*, **12**, 595–601.
- DENNIS, P.G., RUSHTON, S.P., NEWSHAM, K.K., LAUDUCINA, V.A., ORD, V.J., DANIELL, T.J., O'DONNELL, A.G. & HOPKINS, D.W. 2012. Soil fungal community composition does not alter along a latitudinal gradient through the maritime and sub-Antarctic. *Fungal Ecology*, **5**, 403–408. *Corrigendum: Fungal Ecology*, **5**, 759.
- DENNIS, P.G., NEWSHAM, K.K., RUSHTON, S.P., ORD, V.J., O'DONNELL, A.G. & HOPKINS, D.W. 2013a. Warming constrains bacterial community responses to nutrient inputs in a southern, but not northern, maritime Antarctic soil. *Soil Biology & Biochemistry*, **57**, 248–255.
- DENNIS, P.G., SPARROW, A.D., GREGORICH, E.G., NOVIS, P.M., ELBERLING, B., GREENFIELD, L.G. & HOPKINS, D.W. 2013b. Microbial responses to carbon and nitrogen supplementation in an Antarctic dry valley soil. *Antarctic Science*, **25**, 55–61.
- DUNGAIT, J.A.J., KEMMITT, S.J., MICHALLON, L., GUO, S.L., WEN, Q., BROOKES, P.C. & EVERSLED, R.P. 2013. The variable response of soil microorganisms to trace concentrations of low molecular weight organic substrates of increasing complexity. *Soil Biology & Biochemistry*, **64**, 57–64.
- FRASER, F.C., HALLETT, P.D., WOOKEY, P.A., HARTLEY, I.P. & HOPKINS, D.W. 2013. How do enzymes catalysing soil nitrogen mineralization respond to changing temperatures? *Biology & Fertility of Soils*, **49**, 99–103.
- FOWBERT, J.A. & SMITH, R.I.L. 1994. Rapid population increases in native vascular plants in the Argentine Islands, Antarctic Peninsula. *Arctic & Alpine Research*, **26**, 290–296.
- HANSEN, J., SATO, M., REUDY, R., LO, K., LEA, D.W. & MEDINA-ELIZADE, M. 2006. Global temperature change. *Proceedings of the National Academy of Sciences of the United States of America*, **103**, 14 288–14 293.
- HARMSEN, G.W. & VAN SCHREVEN, D.A. 1955. Mineralization of organic nitrogen in soil. *Advances in Agronomy*, **7**, 299–398.
- HARTLEY, I.P. & INESON, P. 2008. Substrate quality and the temperature sensitivity of soil organic matter decomposition. *Soil Biology & Biochemistry*, **40**, 1567–1574.
- HARTLEY, I.P., HOPKINS, D.W., GARNETT, M.H., SOMMERKORN, M. & WOOKEY, P.A. 2008. Soil microbial respiration in Arctic soil does not acclimate to temperature. *Ecology Letters*, **11**, 1092–1100.
- HILL, P.W., FARRAR, J., ROBERTS, P., FARRELL, M., GRANT, H., NEWSHAM, K.K., HOPKINS, D.W., BARDGETT, R.D. & JONES, D.L. 2011. Vascular plant success in a warming Antarctic may be due to efficient nitrogen acquisition. *Nature Climate Change*, **1**, 50–53.
- HOPKINS, D.W., O'DONNELL, A.G. & SHIEL, R.S. 1988. The effect of fertilization on soil nitrifier activity in experimental grassland plots. *Biology and Fertility of Soils*, **5**, 344–349.
- HOPKINS, D.W., O'DOWD, R.W. & SHIEL, R.S. 1997. Comparison of D- and L-amino acid metabolism in soils with differing microbial biomass and activity. *Soil Biology & Biochemistry*, **29**, 23–29.
- HOPKINS, D.W. & SHIEL, R.S. 1996. Size and activity of soil microbial communities in long-term experimental grassland plots treated with manure and inorganic fertilizers. *Biology and Fertility of Soils*, **22**, 66–70.
- HOPKINS, D.W., SPARROW, A.D., NOVIS, P.M., GREGORICH, E.G., ELBERLING, B. & GREENFIELD, L.G. 2006. Controls on the distribution of productivity and organic resources in Antarctic Dry Valley soils. *Proceedings of the Royal Society of London*, **B273**, 2687–2695.
- HOPKINS, D.W., SPARROW, A.D., SHILLAM, L.L., ENGLISH, L.C., DENNIS, P.G., NOVIS, P., ELBERLING, B., GREGORICH, E.G. & GREENFIELD, L.G. 2008. Enzymatic activities and microbial communities in an Antarctic dry valley soil: responses to C and N supplementation. *Soil Biology & Biochemistry*, **40**, 2130–2136.
- HOPKINS, D.W., WAITE, I.S. & O'DONNELL, A.G. 2011. Microbial biomass, organic matter mineralization and nitrogen in soils from long-term experimental grassland plots (Palace Leas meadow hay plots, UK). *European Journal of Soil Science*, **62**, 95–104.
- KEENEY, D.R., SAHRAWAT, K.L. & ADAMS, S.S. 1985. Carbon dioxide concentration in soil – effects on nitrification, denitrification and associated nitrous oxide production. *Soil Biology & Biochemistry*, **17**, 571–573.
- KINSBURSKY, R.S. & SALTZMAN, S. 1990. CO<sub>2</sub>-nitrification relationships in closed soil incubation vessels. *Soil Biology & Biochemistry*, **22**, 571–572.
- MALOSSO, E., ENGLISH, L., HOPKINS, D.W. & O'DONNELL, A.G. 2004. Use of <sup>13</sup>C-labelled plant materials and ergosterol, PLFA and NLFA analyses to investigate organic matter decomposition in Antarctic soil. *Soil Biology & Biochemistry*, **36**, 165–175.
- MALOSSO, E., ENGLISH, L., HOPKINS, D.W. & O'DONNELL, A.G. 2005. Community level physiological profile response to plant residue additions in Antarctic soils. *Biology and Fertility of Soils*, **42**, 60–65.

- MELI, S.M., BADALUCCO, L., ENGLISH, L.C. & HOPKINS, D.W. 2003. Respiratory responses of soil micro-organisms to simple and complex organic substrates. *Biology and Fertility of Soils*, **37**, 96–101.
- NEWSHAM, K.K., PEARCE, D.A. & BRIDGE, P.D. 2010. Minimal influence of water and nutrient content on the bacterial community composition of a maritime Antarctic soil. *Microbiological Research*, **165**, 523–530.
- SIMMONS, B.L., WALL, D.H., ADAMS, B.J., AYRES, E., BARRETT, J.E. & VIRGINIA, R.A. 2009. Long-term experimental warming reduces soil nematode populations in the McMurdo Dry Valleys, Antarctica. *Soil Biology & Biochemistry*, **41**, 2052–2060.
- SMITH, R.I.L. 1994. Vascular plants as bioindicators of regional warming in Antarctica. *Oecologia*, **99**, 322–328.
- SPARROW, A.D., GREGORICH, E.G., HOPKINS, D.W., NOVIS, P., ELBERLING, B. & GREENFIELD, L.G. 2011. Resource limitations on soil microbial activity in an Antarctic dry valley. *Soil Science Society of America Journal*, **75**, 2188–2197.
- THOMAS, E.R., DENNIS, P.F., BRACEGIRDLE, T.J. & FRANZKE, C. 2009. Ice core evidence for significant 100-year regional warming on the Antarctic Peninsula. *Geophysical Research Letters*, **36**, 10.1029/2009GL040104.
- WILSON, K., SPRENT, J.I. & HOPKINS, D.W. 1997. Nitrification in Antarctic soils. *Nature*, **385**, 404.
- WYNN-WILLIAMS, D.D. 1996. Response of pioneer soil microalgal colonists to environmental change in Antarctica. *Microbial Ecology*, **31**, 177–188.
- YERGEAU, E., BOKHORST, S., KANG, S., ZHOU, J.Z., GREER, C.W., AERTS, R. & KOWALCHUK, G.A. 2012. Shifts in soil microorganisms in response to warming are consistent across a range of Antarctic environments. *ISME Journal*, **6**, 692–702.
- ZAK, D.R., HOLMES, W.E., MACDONALD, N.W. & PREGITZER, K.S. 1999. Soil temperature, matric potential, and the kinetics of microbial respiration and nitrogen mineralization. *Soil Science Society of America Journal*, **63**, 575–584.