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2 **COMPARISON OF THREE TECHNIQUES FOR**
3 **ANALYSIS OF DATA FROM AN AEROSOL TIME-**
4 **OF-FLIGHT MASS SPECTROMETER**

5
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38 **ABSTRACT**

39 The Aerosol Time-of-Flight Mass Spectrometer (ATOFMS) is one of few instruments able to
40 measure the size and mass spectra of individual airborne particles with high temporal resolution.
41 Data analysis is challenging and in the present study, we apply three different techniques (PMF,
42 ART-2a and K-means) to a regional ATOFMS dataset collected at Harwell, UK. For the first time,
43 Positive Matrix Factorization (PMF) was directly applied to single particle mass spectra as opposed
44 to clusters already generated by the other methods. The analysis was performed on a total of 56898
45 single particle mass spectra allowing the extraction of 10 factors, their temporal trends and size
46 distributions, named CNO-COOH (cyanide, oxidised organic nitrogen and carboxylic acids), SUL
47 (sulphate), NH₄-OOA (ammonium and oxidized organic aerosol), NaCl, EC⁺ (elemental carbon
48 positive fragments), OC-Arom (aromatic organic carbon), EC⁻ (elemental carbon negative
49 fragments), K (potassium), NIT (nitrate) and OC-CHNO (organic nitrogen). The 10 factor solution
50 from single particle PMF analysis explained 45% of variance of the total dataset, but the factors are
51 well defined from a chemical point of view. Different EC and OC components were separated: fresh
52 EC (factor EC⁻) from aged EC (factor EC⁺) and different organic families (factors NH₄-OOA, OC-
53 Arom, OC-CHNO and CNO-COOH). A comparison was conducted between PMF, K-means cluster
54 analysis and the ART-2a artificial neural network. K-means and ART-2a give broadly overlapping
55 results (with 9 clusters, each describing the full composition of a particle type), while PMF, by
56 effecting spectral deconvolution, was able to extract and separate the different chemical species
57 contributing to particles, but loses some information on internal mixing. Relationships were also
58 examined between the estimated volumes of ATOFMS PMF factors and species concentrations
59 measured independently by GRAEGOR and AMS instruments, showing generally moderate to
60 strong correlations.

61

62 **KEYWORDS:** Aerosol, ATOFMS, PMF, single particle analysis, K-means, ART-2a

63

64 **1. Introduction**

65 In the last decade numerous epidemiological studies have revealed a significant correlation between
66 environmental particulate matter concentrations and adverse health effects. However, since most
67 studies have used PM₁₀ or PM_{2.5} mass concentrations to investigate correlations with human health
68 outcomes it is likely that the health impacts of PM have been in most cases underestimated
69 (Harrison et al., 2010). Atmospheric aerosol is especially problematic to characterize because of its
70 complex and variable composition, wide size range and a broad spectrum of both natural and
71 anthropogenic sources. In this connection, on-line measurements deploying Mass Spectrometric
72 techniques are very promising in order to characterize both aerosol size and chemical composition
73 for a wide range of substances (Pratt and Prather, 2011). Aerosol Time-of-Flight Mass
74 Spectrometry (ATOFMS) is particularly attractive as it allows size and chemical characterisation by
75 measuring the aerodynamic diameter and positive and negative ion mass spectra of individual
76 particles in real time within the diameter range of 0.1 to 3 µm (Rebotier and Prather, 2007; Gard et
77 al., 1997; Dall'Osto et al., 2004; Drewnick et al., 2008). The ATOFMS can measure in a single
78 campaign hundreds of thousands of single particle mass spectra which present a considerable data
79 analysis challenge.

80 Successful analysis of ATOFMS data requires fast and reliable processing and interpretation of the
81 huge amount of data generated. In order to reduce the time of analysis and the pre-deterministic
82 nature of the manual classification, statistical methods can be used. The general aim of
83 classification is to find a structure, i.e, groups of similar or related objects in the available data set
84 (Hinds, 1999). The main difference between a clustering method and manual classification is that
85 the clustering method has the ability to perform analysis over the whole spectrum, rather than as
86 individual peaks. By applying a statistical algorithm to the ATOFMS dataset, the user bias of
87 determining which chemical information is more important in the spectra is minimised. Therefore
88 single particle data are usually treated with a clustering algorithm, such as K-means or ART-2a, in

89 order to group particles of similar size range and chemical composition (Rebotier and Prather, 2007;
90 Gross et al., 2010; Healy et al., 2009; Pekney et al., 2006).

91 In environmental studies, factor analysis techniques (PCA, PCFA, PMF) are widely used to perform
92 source apportionment from data taken at receptor sites. PMF analysis has been successfully applied
93 to 24h averaged data from analysis of particles collected on filters (Stortini et al., 2009; Jia et al.,
94 2010; Dogan et al., 2008; Bari et al., 2009; Alleman et al., 2010) whose principal limitation is the
95 possibility of losing the point source contributions as the characteristic time of plumes from local
96 sources is short. Thus the results obtained are usually limited to the extraction of the 3 or 4 main
97 sources like crustal, marine, combustion sources and secondary particulate matter, while other
98 sources can be extracted only with a wide range of chemical analyses, size segregation and more
99 frequent measurements (Pekney et al., 2006; Wexler et al., 2008). On the other hand, PMF applied
100 to high-resolution data (only obtainable for long periods with an on-line technique) can be a useful
101 tool for this purpose. For example PMF analysis was successfully applied to 1h semi-continuous
102 characterization data of both particulate and gas phase composition leading to the extraction of 6
103 main sources, while by combining ATOFMS and AMS (aerosol mass spectrometry) data to the
104 original dataset the PMF was able to identify 16 factors during a field campaign in Riverside, CA
105 (Eatough et al., 2008).

106 PMF has previously been applied to ATOFMS data *after* clustering by another technique (e.g.
107 McGuire et al., 2011), but not to data *before* clustering. In the present study, for the first time, PMF
108 analysis is directly applied to single particle mass spectra in order to deconvolve the different
109 chemical species which contribute to ambient particulate matter in a rural background location in
110 Harwell (UK). A comparison among three different data treatment techniques (PMF, K-means,
111 ART-2a) is also conducted. Hourly temporal trends of the factors extracted from single particle
112 analysis are compared to each other in order to highlight possible correlations and to study the
113 mixing state of ambient particles. Moreover, temporal trends of factors and clusters are compared

114 with independent ion (and non refractory organic carbon) measurements to evaluate the
115 performance of the data analysis.

116

117 **2. Methodology**

118 *2.1. Measurement Site and Instrumentation*

119 The sampling campaign was conducted in Harwell (51°34'32"N, 1°18'49"W), a rural background
120 site in Oxfordshire (UK) from the 4th October to the 17th October 2008 deploying two on-line mass
121 spectrometric instruments, an Aerosol Time-of-Flight Mass Spectrometer (ATOFMS TSI Model
122 3800-100) and an Aerosol Mass Spectrometer (Aerodyne high-resolution-ToF-AMS) (Drewnick et
123 al., 2005; DeCarlo et al., 2006; Canagaratna et al., 2007; Jimenez et al., 2003), and a GRAEGOR
124 (Thomas et al., 2009), which performs semi-continuous measurements of water-soluble trace gas
125 species (NH₃, HNO₃, HONO, HCl and SO₂) collected by two wet-annular rotating denuders and
126 their related particulate compounds (NH₄⁺, NO₃⁻, Cl⁻, SO₄²⁻) collected in series by two steam-jet
127 aerosol collectors (SJAC). Sample solutions are analyzed on-line by ion chromatography for anions
128 and flow injection analysis for ammonia and ammonium (Thomas et al., 2009). During the
129 campaign, the two inlets of GRAEGOR were placed at the same height (roughly XXX above
130 ground) collecting PTS and PM_{2.5} simultaneously.

131 Hourly data for gaseous pollutant concentrations measured as part of the UK national air quality
132 network and local weather were obtained from the UK national air quality archive
133 (www.airquality.co.uk). Five day air mass back-trajectories arriving at Harwell at three different
134 altitudes (100, 500 and 1000 metres) were obtained using HYSPLIT (Hybrid Single Particle
135 Lagrangian Integrated Trajectory Model) (Draxler and Rolph, 2003). Details of Harwell aerosol
136 characterization and air mass trajectories have been provided in supplementary material.

137

138 *2.2. ATOFMS Technique*

139 The ATOFMS (TSI 3800-100) collects, in real-time, bipolar mass spectra of individual aerosol
140 particles. The instrument is constituted by an aerosol inlet, a sizing region and a mass spectrometer
141 detector. In the aerosol inlet, particles are introduced into a vacuum system region through a
142 converging nozzle, then focused through aerodynamic lenses into a narrow particle beam, which
143 travels through the sizing region. The aerodynamic diameter of individual particles is determined
144 from the time of flight between two continuous-wave laser beams ($\lambda = 532$ nm). After that, particles
145 enter into the mass spectrometer region where a pulsed high power desorption/ionization laser ($\lambda =$
146 266 nm) is triggered on the basis of the transit time of the particle measured in the sizing region.
147 Mass analysis is then provided by a bipolar time of flight reflectron mass spectrometer (Gard et al.,
148 1997; Dall'Osto et al., 2004; Drewnick et al., 2008).

149 During the campaign, the ATOFMS sampled aerosol through a 3/4 inch diameter copper pipe
150 mounted vertically and in-line with the Aerodynamic Focussing Lens (AFL). The inlet of the
151 copper pipe (roughly 4m above the ground) was protected using a simple hockey stick rain cap. The
152 ATOFMS itself was fitted with a TSI 3800-100 AFL which admitted the aerosol at nominal
153 volumetric flow rate of 0.1 L/min operating at a pressure of 2 torr. The device has a quoted size
154 range of 100-3000 nm (Su et al., 2004) although in practice during the sampling campaign our
155 system was capable of hitting 56898 particles with a measured aerodynamic diameter up to 3019
156 nm.

157 Before data analysis, single particles mass spectra were exported using the TSI MS-Analyze
158 software. The peak-list were constructed using the following parameters: minimum peak height of
159 20 units above the baseline, minimum area of 20 units and representing at least the 0.005% of the
160 total area in the particle mass spectrum. The data obtained were analysed using positive matrix
161 factorization (PMF), K-means cluster analysis and artificial neural network (ART-2a) analysis.

162

163 *2.3. Positive Matrix Factorization (PMF) Analysis*

164 The PMF analysis was performed using the program PMF2 (Paatero and Tapper, 1994; Paatero,
165 1998). Briefly, the positive matrix factorisation model (whose principles are detailed elsewhere
166 (Paatero, 1994; Paatero, 1998)) solves the following equation $\mathbf{X} = \mathbf{GF} + \mathbf{E}$ where \mathbf{X} is the original
167 $n \times m$ data matrix, \mathbf{G} is the $n \times p$ scores matrix (factors weight) and \mathbf{F} is the $p \times m$ loadings matrix
168 (factors profile), \mathbf{E} represents the $n \times m$ residuals matrix. In the present case n is the number of
169 particles, m is the number of m/z signals of the spectra and p is the number of factors. The exact
170 number of factors to use was determined by monitoring the parameters suggested by Lee et al.
171 (1999) and the chemical interpretation of the factors profile.

172 *Data matrices.* Before the PMF analysis the dataset was reduced to 106 major m/z values (-146, -
173 144, -124, -121, -119, -104, -101, -99, -98, -97, -96, -95, -89, -88, -85, -84, -81, -80, -79, -76, -73, -
174 72, -71, -64, -63, -62, -61, -60, -59, -49, -48, -46, -45, -44, -43, -42, -37, -36, -35, -27, -26, -25, -24,
175 -17, -16, -15, -14, -13, -12, 7, 12, 15, 18, 23, 24, 27, 36, 37, 39, 41, 43, 46, 48, 49, 50, 51, 52, 53,
176 54, 55, 56, 57, 58, 59, 60, 61, 62, 63, 64, 69, 70, 71, 72, 73, 74, 75, 77, 81, 83, 84, 85, 86, 87, 88,
177 91, 94, 96, 108, 115, 118, 120, 128, 132, 138, 139, 207) and 55357 particles by eliminating the bad
178 variables (the ones that have more than 55000 zero point values on a total number of particles of
179 56898) and the particles with a diameter below the calibration range. Absolute area of peaks was
180 considered for the analysis, which was directly applied to single particle mass spectra.

181

182 2.3.1. Data uncertainties

183 Positive Matrix Factorization relies on the accuracy of error estimates to produce reliable non-
184 negative results and uses the estimates of the error in the data to provide both variable and sample
185 weighting. This is particularly important when less robust datasets have to be used because of the
186 presence of many missing or below detection limit values, as in the case of mass spectra, that could
187 have the ability to define real sources or even be source markers (Owega et al., 2004; Paatero and
188 Taper, 1994; Paatero, 1998; Zhang et al., 2008). The original noise of the data ($\bar{x}_b = 4$, $\sigma_b = 4$),
189 evaluated in zones of particle mass spectra without peaks, was added to the input matrix by

190 simulating it with random numbers between 0 and 8, to avoid multiple zero entries. In fact, circa
191 70% of data in the input matrix are null values. The detection limit was evaluated as the blank value
192 plus three times its standard deviation by integrating the mass spectra signals in several regions
193 without peaks. The uncertainty of the data was evaluated in a laboratory experiment in which
194 equimolar solutions of various salts were nebulised and analyzed with the ATOFMS. The data
195 reproducibility was about 50% and 80% on the average signals for positive and negative ions
196 respectively. Moreover, the particle diameter does not influence the signal intensity. These high
197 uncertainties reflect the principal limits of the ATOFMS analyzer which reside in the size-
198 dependent transmission losses (Allen et al., 2000, Wenzel et al., 2003), laser intensity shot-to-shot
199 variations (Bhave et al., 2002), ionization matrix effects (Reilly et al., 2000), different sensitivities
200 among chemical species that make a semi-quantitative analysis possible to achieve only beside
201 independent sampling measurements (Bhave et al., 2002; Gross et al., 2000; McGuire et al., 2011).
202 The data uncertainties used for the PMF analysis were then calculated as follow $s_{ij} = t + v \cdot x_{ij}$,
203 where $t=4DL=64$ and $v=0.4$ in order to give the same weight to both low and high intensity signals
204 and to avoid the effect of background noise upon the analysis. The data uncertainty of 40% was
205 chosen because there were no further improvements by using a higher uncertainty or different
206 uncertainties for positive and negative ions in terms of quality of the fit and explained variations.
207 Although $Q/Q_{exp} = 0.43$ could indicate a slight overestimation of real data uncertainty, the
208 optimized value seems to be a good compromise considering laboratory experimental data.

209

210 2.3.2. PMF solution

211 The robustness of factor solutions was inspected by comparing the temporal trends of factors
212 through the different PMF solutions. The global minimum of the factor solution was achieved by
213 starting from 50 seeds (pseudorandom starting points). The rotational ambiguity was also tested by
214 modifying the F_{peak} parameter from -2.5 to 2.5. The effect of this variation was not significant with
215 values in the range -0.5 – 0.5 while PMF analysis did not converge with larger F_{peak} values. Thus

216 the PMF solution obtained could be considered unique and $F_{\text{peak}}=0$ was used for the final analysis.

217 After the PMF analysis, factor loadings (**F**) and scores (**G**) obtained were respectively normalized

218 and weighted as follows: each factor loading vector was normalized by dividing it by a scalar value

219 $b_h = \sum_{j=1}^m f_{hj}$ and the corresponding score vector was weighted by multiplying it by the same scalar

220 b_h .

221

222 2.4. Cluster Analysis

223 2.4.1. K-means

224 ATOFMS particle mass spectra were directly imported into ENCHILADA, an open source single

225 particle mass spectra software package (Gross et al., 2010), and 56898 single particle mass spectra

226 were clustered using the K-means/Euclidean square algorithm (McQueen, 1967). K-means, which is

227 a non hierarchical clustering technique, starts with the random subdivision of objects (in this case

228 single particles) into a number of clusters previously defined by the operator. The algorithm

229 computes the total heterogeneity of the system $E_T = \sum_{c=1}^C \sum_{i=1}^{I_c} \sum_{v=1}^{V_c} (x_{ivc} - \bar{x}_{vc})^2$, which is related

230 to the Euclidean distance of every object to the centroid of the cluster to which the object belongs

231 to, and moves objects from a cluster to another until it finds the minimum of system heterogeneity

232 (McQueen, 1967; Gross et al., 2010). In the current study, data analysis was repeated several times

233 with increasing numbers of clusters. The exact number of clusters to use was chosen by monitoring

234 E_T and the chemical interpretation of the cluster centroid mass spectra.

235

236 2.4.2. ART-2a

237 The ATOFMS dataset was imported into YAADA (Yet Another ATOFMS Data Analyzer) and

238 single particle mass spectra were grouped with Adaptive Resonance Theory neural network, ART-

239 2a (Song et al., 1999). The parameters used for ART-2a in this experiment were: learning rate 0.05,

240 vigilance factor 0.85 and iterations 20. These are standard setting used in the ART-2a procedure on

241 ATOFMS data and further details of the parameters can be found elsewhere (Song et al., 1999;
242 Dall'Osto and Harrison, 2006; Rebotier and Prather, 2007). An ART-2a area matrix (AM) of a
243 particle cluster represents the average intensity for each m/z for all particles within a group. An
244 ART-2a AM therefore reflects the typical mass spectrum of the particles within a group.

245

246 *2.5. Positive Matrix Factorization of AMS data*

247 Standard unit mass resolution PMF analysis was carried out on the organic matrix of the AMS
248 dataset (Ulbrich et al. 2009). Two general factors were found: LV-OOA (low-volatile oxidized
249 organic aerosol) and a SV-OOA (semi-volatile oxidized organic aerosol). Whilst the mass spectrum
250 of LV-OOA factor was found to be equivalent to previous standard factor (Ulbrich et al. 2009), the
251 factor SV-OOA contains the standard aliphatic series together with a high m/z 44 and m/z 60
252 signals, indicating a contribution from biomass burning (Lanz et al. 2007).

253

254

255 3. Results and discussion

256 3.1. PMF Analysis on Individual Particle Mass Spectra

257 Single particle mass spectra were subjected to Positive Matrix Factorization analysis with solutions
258 varying from 3 to 15 factors. According to both mathematical parameters and chemical
259 interpretation of factor profiles, the 10 factor solution was selected. The factors extracted are:

- 260 • F1 “CNO-COOH”, explaining 2% of variance, presents peaks of (CN⁻) (m/z -26) and oxidised
261 species (CNO⁻) (m/z -42), (CHOO⁻) (m/z -45) and (CH₃COO⁻) (m/z -59), i.e. carboxylic acids
262 and organic nitrogen species (Angelino et al. 2001, Dall’Osto and Harrison 2006, Moffett et al.,
263 2008);
- 264 • F2 “SUL” explaining 2 % of variance, is characterized by the main peak of sulphate (m/z -97);
- 265 • F3 “NH4-OOA” with an explained variation of 4%, is characterized by peaks of (NH₄⁺) (m/z 18)
266 and secondary organic species (C₂H₃⁺) (m/z 27) and (C₂H₃O⁺) (m/z 43);
- 267 • F4 “NaCl” explaining 6% of variance, is characterized by peaks of (Na⁺) (m/z 23), (Na₂⁺) (m/z
268 46), (Na₂O⁺) (m/z 62), (Na₂OH⁺) (m/z 63) and (Na₂Cl⁺) (m/z 81/83);
- 269 • F5 “EC+” explaining 7% of data variation, contains the elemental carbon positive ions (C⁺, C₂⁺,
270 C₃⁺ at m/z=+12,+24,+36);
- 271 • F6 “OC-Arom” explaining 5% of variance, contains signals related to organic carbon and the
272 benzene fragment (m/z 27, 41, 43, 51, 53, 55, 57, 63, 69, 77, 87, 91, 115) (McLafferty, 1983);
- 273 • F7 “EC-” explaining 3%, is characterized by elemental carbon signals in the negative mass
274 spectrum (C⁻, C₂⁻, C₃⁻ at m/z=-12,-24,-36);
- 275 • F8 “K” explaining 7%, contains the potassium signals (m/z 39/41);
- 276 • F9 “NIT” explaining 4%, is characterized by the nitrate peaks (m/z -46/-62);
- 277 • F10 “OC-CHNO” with an explained variation of 5%, is characterized by organic carbon and
278 organic carbon related to nitrogen signals (m/z -26, 27, 37, 49-52, 60-63, 84-87).

279 The 10 factors obtained can explain only 45% of the total data variance but they are characterized
280 by clear and well defined chemical patterns (Figure 1). Despite the low explained variance, the

281 main signals constituting the factors are well represented and they account for up to 89% of the
282 variance of potassium for example. Sulphate is explained at 84%, while the majority of the bad
283 variables (m/z values with low signal/noise ratio, i.e. $m/z=-146, -144, -124, -121, -119, -104, -101$)
284 are not explained at all.

285 From inspection of residuals (Figure S1) it appears that the PMF analysis failed to extract a few
286 components: this includes chloride signals ($m/z = -35, -37$), which are not present in the NaCl factor,
287 water signals and some other signals probably related to m/z miscalibration problems (Dall'Osto
288 and Harrison, 2006); however, these signals do not influence the interpretation of factors. It should
289 be noted that despite the limited explained variance, which could be a problem in relation to
290 quantification, the factors' chemical profiles obtained are clear and well-defined and thus of
291 qualitative value with the only exceptions of chloride and water signals.

292 The results obtained demonstrate that Positive Matrix Factorization analysis applied to individual
293 particle mass spectra allows the deconvolution of the mass spectra into the contributing specific
294 chemical species (factors K, NIT, SUL, NaCl) or their related classes (factors EC+, EC-, OC-Arom,
295 OC-CHNO, CNO-COOH, NH4-OOA) as well as the extraction of their temporal trends and size
296 distributions (Figure S2). Positive and negative m/z signals are split into different factors (EC+ and
297 EC-, K, NIT, SUL for example) due to different temporal trends either representing changing
298 source contributions or varying relative ionization efficiencies (Bhave et al., 2002; Dall'Osto et al.,
299 2006; Gross et al., 2000). Unlike K-means or ART 2a, PMF does not cluster whole spectra, but
300 disaggregates them into chemical constituents, or groups of constituents. The factors are used to
301 reconstitute actual particle mass spectra as shown in Figure S3. From Figure S3, it may be seen that
302 more than one factor is necessary to reconstruct each particle mass spectrum, demonstrating the
303 mass spectral deconvolution made by PMF analysis on single particles. For example, in Figure S3a,
304 the particle mass spectrum is reconstructed by 10% CNO-COOH, 3% SUL, 6% NH4-OOA, 26%
305 OC-Arom, 10% EC-, 8% K, 8% NIT, and 29% remains unexplained.

306 Factor time-series were calculated as hourly sum of factor scores (not shown) and in equivalent
307 numbers of particle (Figure 2a). Factor time series in equivalent number of particles are calculated
308 as sum over each hour of particle fractions attributable to each factor by first calculating the fraction
309 of particle i attributable to factor h as

$$310 \quad fF_{ih} = \frac{\sum_{j=1}^m g_{ih} f_{hj}}{\sum_{j=1}^m (g_{ih} f_{hj} + e_{ij})} \quad (1)$$

311 and then summing over each hour of particle fractions attributable to factor h :

$$312 \quad NfF_{h,hour} = \sum_{hour} fF_{ih} \quad (2)$$

313 The number size distributions were calculated by summing the factor scores of particles within the
314 same size bin (size bin width of 0.01 μm). The factor size distributions are very similar to each
315 other and all are dominated by the accumulation mode. The only exceptions are F4-NaCl, which
316 presents a coarse distribution because of its origin from sea spray, and F9-NIT which presents both
317 an accumulation and a coarse mode (Figure S2). Moreover, EC- and OC-CHNO factors clearly
318 show a distribution that is shifted towards smaller particles with a tail in the direction of the Aitken
319 mode particles. Despite not being corrected for size-dependent inlet efficiencies, these distributions
320 show predictable differences.

321 The analysis of the correlations between temporal trends of the factors, obtained by summing the
322 score values of each factor within an hour, may give deeper insight into particle components and
323 their sources. Correlations between factors were studied through the correlation coefficients (Figure
324 3 and Table S1) in the Pearson correlation test. Almost every correlation is statistically significant
325 ($p\text{-value} < 0.05$) but to different degrees. The NaCl seems to be an independent factor because it has
326 no strong correlations with the other factors, according to the Cohen classification (Cohen, 1988)
327 and it is not correlated to sulphate, EC-, potassium, nitrate and OC-CHNO ($p\text{-value} > 0.05$). Sulphate
328 is strongly correlated with potassium ($r=0.64$), nitrate ($r=0.76$) and the organic carbon factors, OC-
329 Arom (0.82) in particular. Potassium and nitrate are strongly correlated with almost every factor and

330 are the dominant species, present in the majority of the particles collected. This reflects, at least in
331 part, the very high sensitivity of the ATOFMS to these species (Gross, 2000).
332 Hourly temporal trends of EC⁻ and EC⁺ present only a correlation of medium strength ($r=0.41$).
333 EC⁻ is strongly correlated with OC-CHNO ($r=0.64$) while EC⁺ is correlated more with secondary
334 species (r coefficients for NH₄-OOA, 0.87 and NIT, 0.61). This result suggests that the splitting of
335 elemental carbon signals into two factors may not only reflect different ionization and detection
336 efficiencies between positive and negative ions. It seems that the ionization pattern is influenced by
337 the matrix composition (Reilly, 2000) distinguishing two different elemental carbon components:
338 one probably freshly emitted (EC⁻) and one more aged (EC⁺), modified by oxidation reactions, and
339 internally mixed with secondary species. In fact, as proposed by Reinard and Johnston (2008)
340 secondary species like nitrate and sulphate could limit the electron availability, leading to a
341 suppression of elemental carbon fragments negatively charged, while potassium and sodium, on the
342 contrary, could lead to an enhancement of them. Moreover, the temporal trend of the EC⁻ is
343 characterized by a peak event on 16/10/2008 probably due to a combustion event near the sampling
344 site.

345

346 *3.2. Cluster Analyses*

347 *3.2.1. K-means*

348 The K-means analysis separated 13 clusters. Clusters obtained from miscalibrated mass spectra
349 were eliminated and clusters with similar profiles and temporal trends were recombined to generate
350 a total of 9 clusters (mass spectra are reported in Figure S4a and their temporal trends, expressed as
351 the number of particles are reported in Figure 2b). The clusters are:

- 352 • K (14140 particles, 25%), which presents high potassium signals and some signals of low
353 intensity due to Na⁺, cyanide, nitrate and sulphate;

- 354 • K-EC (3252 particles, 6%), which presents negative ions signals related to elemental carbon, and
355 to a lesser extent nitrate and sulphate signals, while in the positive mass spectrum it presents
356 signals of a low intensity, related to oxidized organic carbon, potassium and sodium;
- 357 • NaCl (10872, 19%), which mainly presents signals of sodium, chloride, potassium and nitrate;
- 358 • EC (9436 particles, 17%), which presents both positive and negative signals related to elemental
359 carbon and signals of nitrate and sulphate;
- 360 • K-SUL-OC-NIT (1832 particles, 3%) presents CN^- , NO^- , NO_2^- , SO_3^- , HSO_3^- , HSO_4^- signals in
361 the negative mass spectrum and potassium and OC aromatic signals in the positive mass
362 spectrum;
- 363 • OC (4625 particles, 8%) presents both aromatic, amine and oxygenated carbon signals and traces
364 of ammonium, nitrate, sulphate and cyanide;
- 365 • K-NIT (6829 particles, 12%) is mainly characterized by potassium and nitrate signals along with
366 the presence of cyanide, sulphate, ammonium and oxidized organic aerosol fragments ($m/z =$
367 $+27/+43$);
- 368 • OOA (2006 particles, 4%), composed of signals corresponding to C_2H_3^+ , $\text{C}_2\text{H}_3\text{O}^+$ and carboxylic
369 acids along with ammonium, potassium, nitrate and sulphate;
- 370 • Fe-V (840 particles, 1%), characterized by signals at $m/z = +51/+56/+67$ that could be assigned
371 respectively to V^+ , Fe^+ and VO^+ and by signals at $m/z = +58/+60$ that could be attributed to
372 nickel and, to a lesser extent, by sodium, potassium, elemental carbon and nitrate.

373

374 3.2.2. ART-2a

375 The ART-2a algorithm generated 389 clusters used to describe the dataset (total particles 56898).

376 The 50 most populated clusters represent more than 63% of the mass spectra from the study and

377 thus were used for the results presented in this paper. The remaining clusters were mostly made up

378 of a majority of miscalibrated mass spectra. By manually merging similar clusters according to their

379 chemical and temporal profiles with the standard procedure elsewhere described (Dall'Osto and

380 Harrison, 2006), the total number of clusters describing the whole database was reduced to 9,
381 representing about 63% of the total number of particles sampled (Figure S4b). The rest of the
382 particles presented low signal to noise ratios and therefore were not classified. The 9 clusters are:

- 383 • K-NIT (9613 particles, 17%) composed by potassium, nitrate, cyanide and sulphate;
- 384 • NaCl (7852 particles, 14%) characterized by a mass spectrum identical to K-means NaCl;
- 385 • OC (3172 particles, 6%) composed mainly by oxidized organic aerosol and aromatic
386 compounds, and potassium, cyanide, nitrate and sulphate signals;
- 387 • K-SUL (2355 particles, 4%) with high potassium and sulphate signals, along with ammonium,
388 nitrate and aromatic organic compounds;
- 389 • EC (5416 particles, 10%) which present a mass spectrum identical to the K-means EC;
- 390 • K (1656 particles, 3%) with an high potassium signal and Na^+ , C_3^+ , nitrate and sulphate signals;
- 391 • EC-Fe-V (1337 particles, 2%) composed by high signals of elemental carbon, and V^+ , Fe^+ and
392 VO^+ in the positive mass spectrum while it does not present significant signals in the negative
393 mass spectrum;
- 394 • SOA (1066 particles, 2%) composed by ammonium, C_2H_3^+ , $\text{C}_2\text{H}_3\text{O}^+$, potassium, low elemental
395 and organic carbon signals, nitrate and sulphate;
- 396 • K-EC (3390 particles, 6%) elemental carbon signals, potassium, sodium, cyanide, nitrate and
397 sulphate.

398 The mass spectra of the 9 clusters are shown in Figure S4b and their time-series, expressed as
399 number of particles are reported in Figure 2c. Results obtained from ART-2a analysis are very
400 similar to the K-means results. The two NaCl and EC clusters present the same mass spectrum. The
401 two K-EC clusters are similar but the ART-2a cluster is characterized by higher cyanide, nitrate and
402 sulphate signals. The OC ART-2a cluster presents higher aromatic signals than OC K-means
403 cluster. The K-NIT K-means cluster presents aromatic organic carbon signals which are indeed not
404 present in the K-NIT ART-2a cluster. The OOA K-means cluster has a mass spectrum similar to the
405 SOA ART-2a cluster, with a high contribution of NH_4^+ , C_2H_3^+ and $\text{C}_2\text{H}_3\text{O}^+$, but the former presents

406 carboxylic acids signals while the latter presents more aromatic organic compounds signals. The
407 main differences reside in the abundance of the K cluster, which is probably overestimated by K-
408 means (25% of particles in the K-means clustering and 3% of particles in the ART-2a clustering),
409 the EC-Fe-V ART-2a cluster which exhibits strong elemental carbon signals that are not present in
410 the Fe-V K-means cluster, and K-SUL ART-2a cluster which has a different positive mass
411 spectrum, dominated by the potassium signal, while the K-SUL-OC-NIT K-means cluster has OC
412 aromatic signals in the positive mass spectrum.

413 The differences between the two techniques could reside in the different approach to clustering the
414 data. In K-means cluster analysis, all particles are assigned to the clusters by dividing them into
415 groups of similarity. The number of clusters is chosen by the operator who proceeds with a trial-
416 and-error approach by incrementing the number of clusters until the division into more clusters is
417 chemically meaningless (13 clusters in this case). On the contrary, ART-2a (running with standard
418 parameters) usually produces a huge number of clusters (389 in this case). After that, clusters made
419 by only few particles are eliminated and only the main contributing clusters are considered, and
420 clusters of similar composition and size distribution are merged manually. Thus, ART-2a may give
421 more clear and well defined clusters than K-means which considers more particles than the former
422 in the final solution.

423

424 *3.3. Comparison between Results of PMF Analysis on Single Particles, K-means Cluster Analysis* 425 *and ART-2a Artificial Neural Network Analysis*

426 PMF and cluster analysis can be viewed as complementary techniques. While K-means and Art-2
427 give a rapid classification of whole particles by dividing them into classes of similarity the PMF
428 analysis on single particle mass spectra permits the extraction of the chemical species constituting
429 the particles. Much of the information on internal mixing is lost.

430 The results of the correlation analysis (Pearson correlation test) among cluster and factor temporal
431 trends (in equivalent number of particles) which have a similar chemical profile show a good

432 agreement (Table 1). For instance, taking into account K-means results, PMF F8-K is correlated to
433 the K cluster with $r^2 = 0.99$ and $p\text{-value} = <0.001$ (Pearson correlation test); F4-NaCl is correlated to
434 the NaCl cluster ($r^2 = 0.91$, $p\text{-value} = <0.001$) and F3 NH₄-OOA is correlated to the OOA cluster (r^2
435 $= 0.83$, $p\text{-value} = <0.001$). The cross correlations between factors and cluster temporal trends
436 confirm the conclusions obtained from the cross correlations between PMF factors (Figure 3 and
437 Table S1). The NaCl cluster presents a strong correlation only with the NaCl factor. In fact, NaCl is
438 an independent particle type which is directly associated with the sea spray source. The EC⁺ factor
439 is strongly correlated with clusters characterized by secondary aerosol (OOA $r^2=0.53$, K $r^2=0.59$, K-
440 NIT $r^2=0.36$) while EC⁻ is not strongly correlated with any cluster, confirming the two different
441 elemental carbon contributions to aged (EC⁺) and fresh (EC⁻) particles. The cluster Fe-V is strongly
442 correlated with the EC⁺ factor ($r^2=0.49$) probably because of a common origin from oil based fuel
443 combustion (Korn et al., 2007) or transported from coal-fired power plants in Central Europe. In
444 fact, EC⁺ abundance increased during long-range transport of air masses from Central Europe (see
445 SI). Moreover, the K-EC cluster which is moderately correlated to organic factors as well as EC⁻,
446 could represent a biomass burning signature (Bi et al., 2011; Healy et al., 2012). CNO-COOH, SUL
447 and NH₄-OOA PMF factors, as expected, are present in multiple clusters as they are highly
448 oxidized aerosol components produced during aging processes.

449

450 *3.4. Comparison of PMF Analysis Results with Independent Measurements*

451 Alongside the ATOFMS, inorganic water soluble components in the TSP and PM_{2.5} size fractions
452 were measured by GRAEGOR, and in non-refractory PM₁ (NR-PM₁) by the AMS, defined as those
453 components within PM₁ that volatilise rapidly at the vaporiser temperature of 600°C. In order to
454 validate the PMF factor temporal trends, a correlation analysis (r-Pearson test) was made between
455 them and these independent measurements. In Figure 5 the sulphate, nitrate, chloride, ammonium
456 and organic concentrations are reported compared to the corresponding PMF ATOFMS factors.

457 For this purpose, factor temporal trends were calculated under the simplifying assumptions that all
458 particles are homogenous, spherical and a constant mass of material is ionized from each particle,
459 irrespective of their size (Dall'Osto et al., 2006). Particle volume was multiplied by the percentage
460 contribution of each factor to it. The hourly time-series (in volume) of the factors were then
461 calculated by summing the partial volume of each particle attributable to each factor (Figure 4). For
462 comparison with AMS PM₁ concentrations, PMF factor partial volumes were integrated for
463 particles of < 1 µm diameter. It is important to note that ATOFMS time-series were not corrected
464 for size-dependent inlet efficiencies (Dall'Osto et al., 2006).

465 The SUL factor (expressed in volume of particles) is significantly correlated with sulphate
466 concentrations in PM_{2.5} ($r^2 = 0.34$, p-value = <0.001) and in AMS PM₁ ($r^2 = 0.41$, p-value =
467 <0.001). In the case of nitrate, the NIT PMF factor temporal trend is weakly correlated with nitrate
468 concentration in PM_{2.5} ($r^2 = 0.07$, p-value = <0.001), but is strongly correlated with nitrate in NR-
469 PM₁ ($r^2 = 0.54$, p-value = <0.001). The difference in the correlations may reflect different
470 instrumental inlet characteristics leading to different large particle contributions to the temporal
471 patterns. In fact, while NR-PM₁ is fairly specific to NH₄NO₃, PM_{2.5} can contain also significant
472 amount of NaNO₃, produced by sea salt processing through HNO₃. NIT PMF factor presents both
473 an accumulation and a coarse mode, and the latter could be measured with higher efficiency than
474 the former, and would also contain contributions that are not included in the NR-PM₁. The high
475 correlations seen for the NR-PM₁ fraction are however reassuring.

476 The NaCl factor is weakly but significantly correlated to the chloride measurements in PM_{2.5} ($r^2 =$
477 0.11, p-value = <0.001). On the contrary AMS chloride is not significantly correlated with the
478 GRAEGOR chloride measurements (p-value = 0.32 for TSP and 0.46 for PM_{2.5}), which shows
479 much larger concentrations, because the AMS only detects the non-refractory fraction which is
480 thought to be dominated by NH₄Cl. The NH₄-OOA factor, which contains both OOA and
481 ammonium signals, is correlated with the ammonium concentration in PM_{2.5} ($r^2 = 0.61$, p-value =
482 <0.001) and in NR-PM₁ ($r^2 = 0.59$, p-value = <0.001) and to the organic component measured by

483 the AMS ($r^2 = 0.60$, $p\text{-value} = <0.001$). The non-refractory organic concentration measured by AMS
484 is strongly correlated with ammonium concentration and presents the highest correlation with the
485 $\text{NH}_4\text{-OOA}$ PMF factor rather than with other organic factors. ATOFMS factors were also compared
486 to TSP ion measurements, but because of different inlet characteristics, the correlations are in
487 general weak or not significant and the results are not reported.

488 The analysis shows clearly that PMF factors are highly significantly correlated with the
489 corresponding chemical species mass concentrations, with a better agreement with NR-PM_1 (if
490 ATOFMS PMF factors are integrated for particles $<1 \mu\text{m}$). On the contrary, clustering analytical
491 techniques such as K-means and ART-2a cannot disaggregate the contribution of the different
492 chemical species present in the particles. For this reason, a direct comparison between the time-
493 series of a cluster and the mass concentration of one of its components is not appropriate. In fact,
494 such correlation would be highly dependent on particle mixing-state. Thus, the disaggregation of
495 species made by the PMF analysis (on single particles) proves very useful for quantification
496 purposes of the principal substances or classes of substances constituting the particles. The
497 determination coefficient, slope and intercept of the linear regressions between ATOFMS factors
498 and the species concentrations measured by AMS in NR-PM_1 are reported in Supplementary
499 Material. (Table S2). Moreover, the correlation between PMF factors and the corresponding species
500 concentrations may be even stronger if ATOFMS data are corrected for size-dependent transmission
501 losses (Jeong et al., 2011).

502

503 *3.5. Comparison between ATOFMS-PMF factors and AMS-PMF factors for secondary organic* 504 *aerosol*

505 In order to further validate the PMF analysis on single particle ATOFMS spectra, the factors
506 obtained were compared with standard factors (Ulbrich et al., 2009) extracted by PMF analysis on
507 the organic matrix of the AMS measurements (Table S3, Supplementary Materials). The
508 comparison was conducted considering ATOFMS-PMF factor time-series in volume (integrated

509 over particles of $< 1 \mu\text{m}$ diameter) because AMS-PMF factors are expressed in mass concentration
510 ($\mu\text{g}/\text{m}^3$). The results show that ATOFMS-PMF factors associated with aged aerosol (NH₄-OOA and
511 EC+) are better correlated with the most aged LV-OOA AMS-PMF factor ($r^2=0.66$ and 0.67 for
512 NH₄-OOA and EC+ respectively) rather than with SV-OOA ($r^2=0.55$ and 0.43 for NH₄-OOA and
513 EC+ respectively). On the contrary, fresh or less aged components (ATOOFMS-PMF factors OC-
514 Arom and OC-CHNO) are better correlated with the less aged SV-OOA AMS-PMF factor ($r^2=0.54$
515 and 0.37 for OC-Arom and OC-CHNO respectively) rather than with LV-OOA ($r^2=0.36$ and 0.07
516 for OC-Arom and OC-CHNO respectively).
517 Unexpectedly, EC- presented a correlation of medium intensity with both AMS-PMF factors ($r^2 =$
518 0.45 for LV-OOA and $r^2 = 0.43$ for SV-OOA). However, the correlations are stronger, especially
519 with respect to the less aged SV-OOA if the EC- time-series is expressed as the equivalent number
520 of particles ($r^2 = 0.62$). This may be due to the fact that using time-series calculated in volume we
521 may further underestimate the contribution of small particles because of size-dependent
522 transmission losses (Gross et al., 2000; Dall'Osto et al., 2006).

523

524 *3.6. Harwell Aerosol Characterization*

525 From the study of the back-trajectories of air masses arriving in Harwell during the sampling
526 campaign (detailed in S.M.), it was clear that the NaCl factor was dominant during the sampling of
527 marine-polar air masses, while during periods of sampling continental air masses (from Central
528 Europe) elemental carbon, potassium, nitrate and sulphate concentrations increased. This is as
529 expected as Harwell is a rural background site and it should not be influenced substantially by local
530 primary sources. More interesting were two marine-continental periods. The first was characterized
531 by air masses coming from the ocean, crossing Scotland and England before arriving at the Harwell
532 site. It was characterized by high concentrations of NO_x and other primary gaseous pollutants, and a
533 high abundance of the OC-CHNO and EC- factors. The second was characterized by air masses
534 coming from the west coast of France, low concentrations of primary gaseous pollutants and a high

535 amount of CNO-COOH and EC+ PMF factors. Thus, the first period was characterized by freshly
536 emitted aerosol while the second period is characterized by aged and chemically oxidized particles.
537

538 **4. Conclusions**

539 PMF analysis has been applied to single particle ATOFMS mass spectra and allows the extraction
540 and separation of significant contributing chemical components. In general, PMF factor profiles
541 identify well defined chemical species or classes of substances from inorganic (NaCl, K, NIT, SUL)
542 to organic families (EC+, EC-, OC-Arom, OC-CHNO, CNO-COOH, NH4-OOA). There is a partial
543 loss of information on internal mixing of particles.

544 From the cross correlation analysis among temporal trends of PMF factors it was possible to
545 identify two elemental carbon components: the EC- factor, correlated to OC-CHNO, probably
546 related to anthropogenic primary emissions and the EC+ factor present in aged particles internally
547 mixed with secondary species. Furthermore, this is the first time in which different families of
548 organic carbon have been extracted from ATOFMS data, including aromatic, oxidized organic
549 compound and two different organic nitrogen components: primary (OC-CHNO) and oxidized
550 (CNO-COOH). Oxidized carbon in the form of oxidised organic nitrogen and carboxylic acids is
551 found only in aged aerosol while nitrate and sulphate are found in different proportions: the former
552 in less aged aerosol such as in urban plumes while sulphate arose predominantly from long-range
553 transport from continental sources.

554 From the comparison of different data treatment techniques it emerges that K-means cluster
555 analysis and ART-2a artificial neural network analysis give similar results, with particles grouped in
556 clusters of similar composition, reflective of aerosol sources, chemical processes and a combination
557 of both, while PMF analysis of single particle mass spectra allows the deconvolution of the mass
558 spectra and the extraction of some constituent components. Moreover, when expressed in volume,
559 the temporal trends of PMF factors are highly significantly correlated to the corresponding chemical
560 species concentration measured by independent instruments, even in the case of highly internally

561 mixed particles, while the correlation between cluster temporal trends and corresponding chemical
562 species concentration is highly dependent upon particle mixing state. Thus PMF analysis may prove
563 useful for the quantification of the main components of PM data collected with the ATOFMS
564 instrument. However, better repeatability of the ionization process and higher efficiency of particle
565 detection would improve its quantification capability.

566

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573

574 **Appendix. Supplementary material**

575 Supplementary material related to this article can be found at ...

576

577

578 **References**

579

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721 Table 1. Coefficient of determination (r^2) values of the linear regressions between hourly temporal
 722 trends of PMF factors (equivalent number of particles) and K-means clusters or ART-2a clusters*.

PMF factors	r^2 (PMF factors vs K-means clusters)									r^2 (PMF factors vs ART-2a clusters)								
	K	K-EC	NaCl	EC	K-SUL-OC-NIT	OC	K-NIT	Fe-V	OOA	K-NIT	NaCl	OC	K-SUL	EC	K	EC-Fe-V	SOA	K-EC
CNO-COOH	0.43	0.34	0.00	0.09	0.28	0.09	0.25	0.02	0.12	0.29	0.00	0.19	0.27	0.10	0.09	0.01	0.04	0.16
SUL	0.53	0.20	0.00	0.28	0.51	0.12	0.45	0.10	0.37	0.47	0.00	0.33	0.42	0.31	0.30	0.08	0.22	0.07
NH4-OOA	0.76	0.01	0.11	0.68	0.01	0.17	0.64	0.27	0.83	0.78	0.12	0.19	0.46	0.76	0.73	0.23	0.48	0.03
NaCl	0.06	0.06	0.91	0.12	0.06	0.00	0.03	0.06	0.08	0.07	0.82	0.00	0.05	0.12	0.12	0.05	0.07	0.01
EC+	0.59	0.00	0.14	0.99	0.00	0.02	0.36	0.49	0.53	0.53	0.16	0.04	0.38	0.87	0.60	0.66	0.24	0.01
OC-Arom	0.54	0.31	0.00	0.25	0.36	0.29	0.59	0.08	0.42	0.53	0.00	0.45	0.34	0.30	0.29	0.05	0.26	0.11
EC-	0.53	0.35	0.02	0.42	0.06	0.17	0.42	0.16	0.39	0.43	0.02	0.17	0.26	0.42	0.32	0.18	0.23	0.08
K	0.99	0.03	0.04	0.56	0.03	0.04	0.47	0.17	0.61	0.78	0.05	0.11	0.64	0.58	0.60	0.21	0.25	0.07
NIT	0.60	0.12	0.00	0.37	0.12	0.19	0.92	0.12	0.64	0.77	0.00	0.50	0.21	0.50	0.54	0.08	0.42	0.06
OC-CHNO	0.14	0.40	0.00	0.08	0.11	0.83	0.49	0.03	0.30	0.24	0.00	0.51	0.03	0.14	0.12	0.01	0.36	0.11

723 *strong correlated results ($r^2 > 0.5$) are presented in bold

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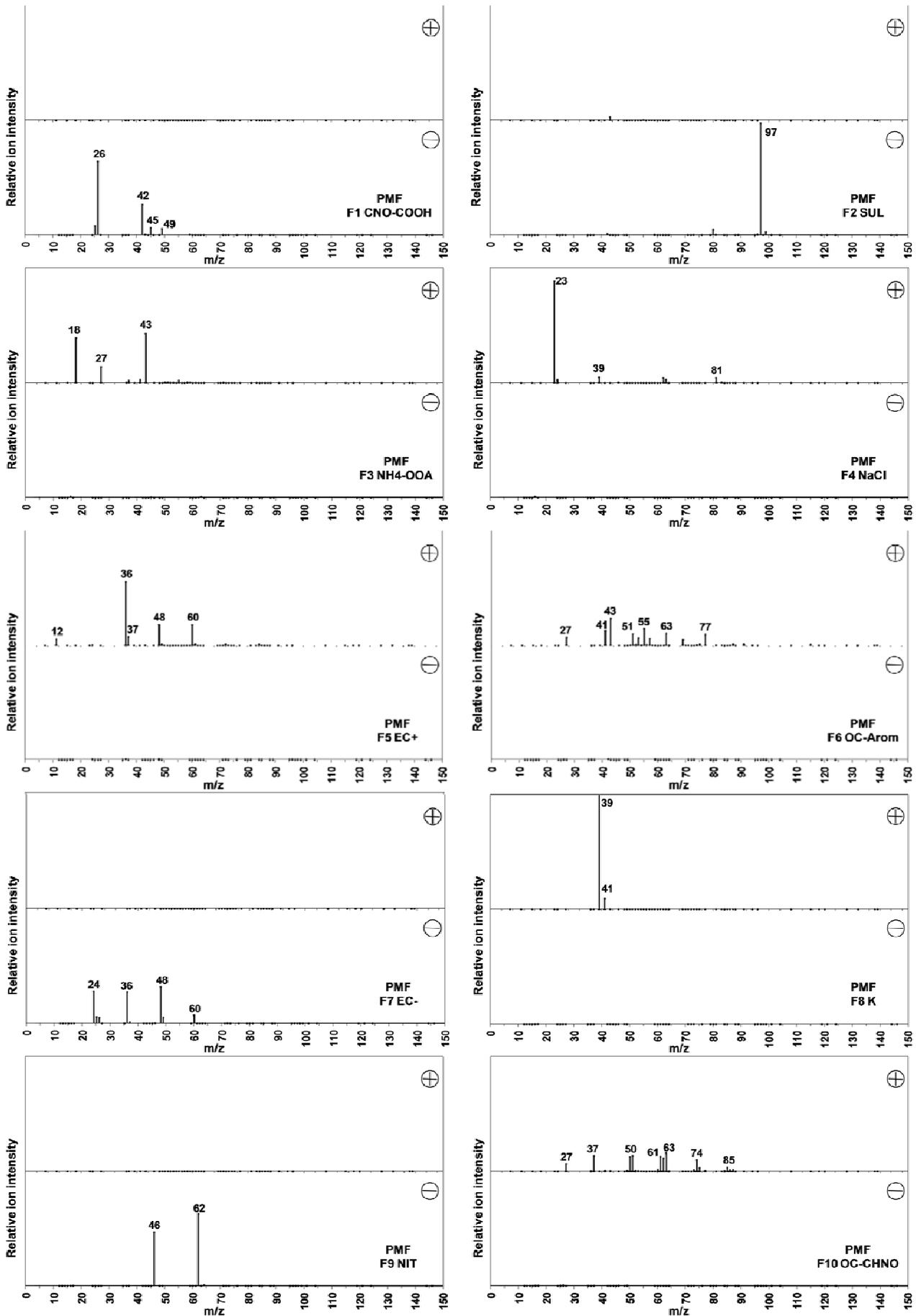


Figure 1. Mass Spectra of the 10 PMF factors.

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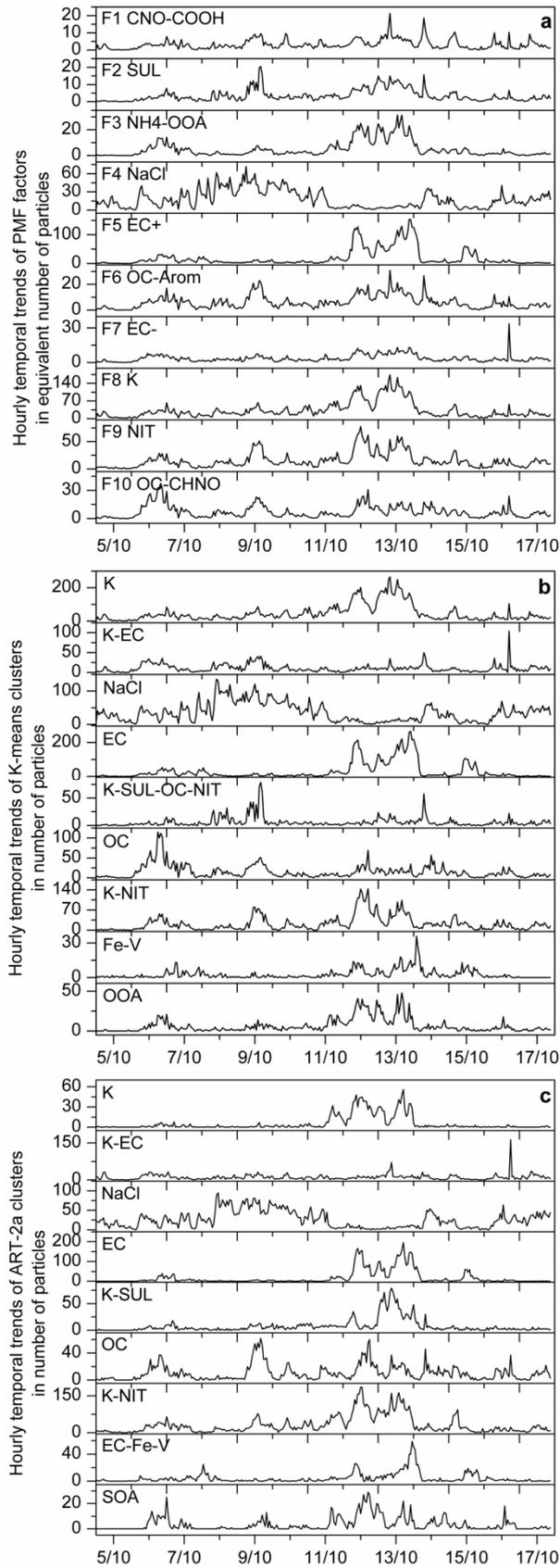


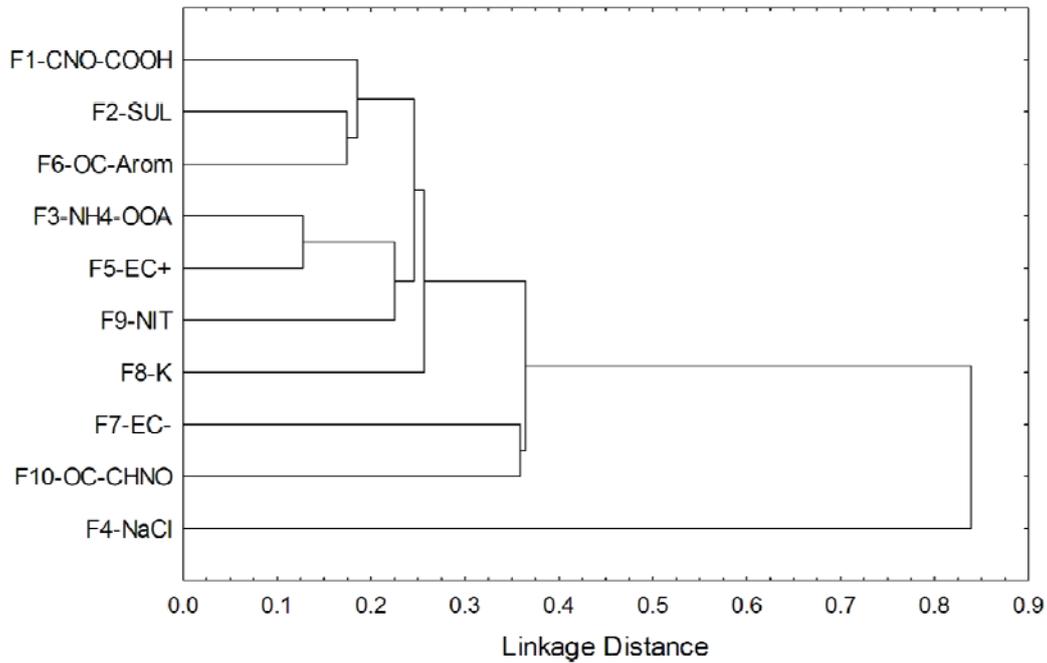
Figure 2. Temporal trends of (a) PMF factors expressed in equivalent number of particles, (b) K-means clusters and (c) ART-2a clusters in number of particles.

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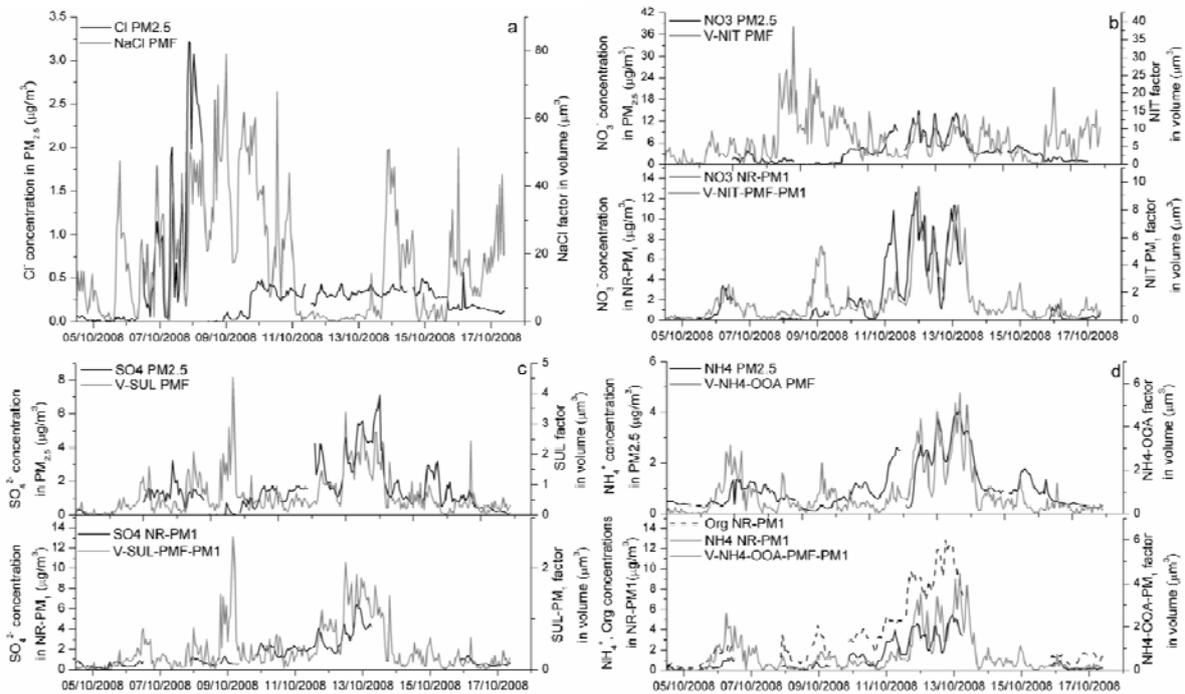
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Figure 3. Dendrogram obtained from the hierarchical cluster analysis of the temporal trends of PMF factors (single linkage method, r-Pearson correlation coefficient distance measure).



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Figure 4. Hourly time-series of (a) chloride concentrations and NaCl factor, (b) nitrate concentrations and NIT factor, (c) sulphate concentrations and SUL factor, (d) ammonium, organic concentrations and NH4-OOA factor.